

Impact of cover crop on soil carbon accrual in topographically diverse terrain

J. Beehler, J. Fry, W. Negassa, and A. Kravchenko

Abstract: Farmers must consider real-world problems and variability to maximize yields and minimize environmental impacts when using cover crops in corn (*Zea mays* L.)-based cropping systems. Much of the variability encountered by farmers of the Midwest Corn Belt is due to the topographical diversity of the undulating landscape. The objectives of this study are to explore the effects of cover crops on soil organic carbon (C), both total organic C and its labile form, particulate organic C (POC), and on water retention at contrasting topographies (summit, slope, and depression). A cereal rye (*Secale cereal* L.) cover crop was established in the fall of 2011 at two experimental sites. At each site, the experimental design was a split-split plot with whole plot factor, topographical position, in a randomized complete block design with two replications. Topographical position affected all studied plant and soil characteristics, including aboveground plant biomass, rye residue decomposition, POC, total soil C, and soil water retention. Across all topographical positions, rye residue decomposition was ~5% greater in the treatment with than without cover crop. In slopes and summits, the POC in the treatment with cover crop was significantly ($\sim 0.7 \text{ mg g}^{-1} \text{ soil}$ [$\sim 700 \text{ ppm}$]) greater than in the treatment without the cover crop; however, the difference was not statistically significant in depressions. The cover crop effect on both total organic C and soil water retention levels was not statistically significant. The study points to potential interactions between topography and C sequestration benefits of cover crops; however, longer experimental times are needed to detect significant differences in soil total C and water retention measurements.

Key words: cover crops—soil carbon—sustainable corn—topography

Cover crops can provide multiple benefits to row crop agricultural systems, including suppression of weeds and pests, improvement of soil and water quality, and stimulation of nutrient cycles (Snapp et al. 2005). Yet, when implementing cover cropping farmers must consider spatial variability within their fields to maximize yields and minimize environmental impacts. In the Midwest Corn Belt, much of the spatial variability that farmers encounter is due to the topographical diversity of the undulating landscape. The topographical diversity controls many soil properties, including the distribution of soil organic carbon (SOC).

Enhancing rotation through the inclusion of a cover crop has the potential to accumulate SOC (West and Post 2002; Follett 2001). Cereal rye (*Secale cereale* L.) is especially suited for this purpose because it is a high biomass grass known to increase SOC (Kuo and Jellum 2000; Kaspar et al. 2006;

Reicosky and Forcella 1998). However, the advantages of rye as a cover crop in regards to SOC accrual might not be homogeneously spread across topographically diverse landscapes. Cover crops, such as rye, can grow at variable rates across topographies, commonly with better biomass accumulation in flat areas and in depression positions (Muñoz et al. 2014). Thus, one can expect that greater SOC benefits from a rye cover crop can occur in topographical depressions. The recent observations from topographically diverse agricultural fields in Michigan demonstrated that the magnitude of the cover crop effects on SOC was higher in the summit and slope positions as compared to depressions (Ladoni et al. 2016). This indicates that other factors, besides the overall amount of the aboveground cover crop biomass production, influence the magnitude of the SOC benefits in topographically diverse agricultural landscapes.

Spatial distribution patterns of soil moisture and temperature are among the factors that likely contribute to SOC accrual. They influence the environment for crop growth and regulate the conditions for microbial growth and activity (Wickings et al. 2016), thus can affect the decomposition of freshly added plant inputs. Topographical differences are often the main controls of soil moisture and soil temperature in undulating terrain.

Topography plays a role in the spatial variability of soil physical properties, which are highly correlated to SOC content (Moore et al. 1992; Tromp-van Meerveld and McDonnell 2006; Romano and Palladino 2002). One such property of significant value to plant growth and strongly related to SOC is soil water retention (Hudson 1994; Adams 1973; Rawls 1983; Gupta and Larson 1979). An increase in organic matter in sandy soils, like the soils in this study, can lead to an increase of water retention (Rawls et al. 2003). Organic matter content in the soil also affects soil aggregation, a key component of soil structure and water retention capability (Franzuebbers 2002). By assessing the effects of topography and cover cropping on water retention we can better understand the long-term implications on soil C and soil health. However, relatively few studies so far have directly addressed the effect of cover crops on soil hydraulic properties, including soil water retention.

The principal hypothesis of this study is that topography does not only influence the spatial distribution patterns of SOC and water retention, but also moderates the cover cropping benefits, specifically the cover crop's contributions to SOC accrual. We hypothesize that topography's role is implemented, first, through its influence on the growth of the main crops and cover crops and on the subsequent biomass inputs into the soil, and second, through its influence on plant

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residue decomposition rates. We address the former by measuring aboveground biomass of the main crops and cover crops, and the latter by assessing decomposition of cover crop residue using the litterbag approach.

The objectives of this study are to explore the effect of a rye cover crop on soil C accrual and changes in soil water retention characteristics in topographically diverse agricultural landscapes. Total SOC responds relatively slowly to management practices (Plaza-Bonilla et al. 2014; Haynes 2000). However, labile SOC, such as particulate organic matter, can respond to changes in land use and management practices on the scale of a few days to a few years (Nascente et al. 2013). Thus, while reporting the observed total SOC results in this study, we used particulate organic C (POC) as a metric of the changes in SOC driven by the addition of a cover crop to the system.

Materials and Methods

Site Description. The two experimental sites of this study are located in central (Mason) and southwest (Kellogg) Michigan. Both sites average around 76 cm (29.9 in) of rainfall in addition to 76 cm of snowfall annually. Temperatures at both sites average -3°C (26.6°F) in winter and 20°C (68°F) in summer. At Mason, the soil is Marlette fine sandy loam (Oxyaquic Glossudalfs), 2% to 6% slope, consisting of mainly sandy loam with sandy clay loam present at the 20 cm (7.8 in) depth. At Kellogg, the soil is Kalamazoo loams (Typic Hapludalfs) with 2% to 6% slope and mainly sandy loam with fine to medium gravels.

At each of the sites, the experimental design was a split-split plot with the whole plot factor of topographical position in a randomized complete block design with two replications. The treatment design included the following three factors: topography with three levels (summit, slope, and depression), main crop with two levels (corn [*Zea mays* L.] and soybean [*Glycine max* L.]), and cover crop with two levels (presence and absence). Note that including both phases of the rotation (corn and soybean) in the experiment doubled the number of experimental plots, thus, at each topographical position each cover crop treatment had four experimental plots for a total of 12 plots of each cover crop treatment at each experimental site. The cover crop experimental plots were 9.1×9.1 m (29.8×29.8 ft) in size.

Conventional practices were used in tillage, planting, the application of fertilizer and herbicide, and harvesting consistent with practices in the region. The cereal rye cover crop was established each fall in early October, beginning in 2011. During 2011 to 2013, rye was sown at 112 kg ha^{-1} (99.9 lb ac^{-1}) with a John Deere 4.57 m (15 ft) no-till drill. In 2014 to 2015, rye population was increased to establish a better stand and was sown at 145 kg ha^{-1} (129.3 lb ac^{-1}). Rye was terminated by herbicide and then was chisel plowed, followed by a soil finisher to establish a seed bed for the subsequent main crop. Weeds were controlled using herbicides at least once per growing season with additional weed control by herbicide application if necessary.

Soil samples for total SOC and water retention were taken in the spring of 2015. Samples for SOC were taken from 0 to 10 cm, 10 to 20 cm, 20 to 40 cm, and 40 to 60 cm (0 to 4 in, 4 to 7.8 in, 7.8 to 15.7 in, and 15.7 to 23.6 in) depths. Samples for water retention were taken from 0 to 10 cm and 10 to 20 cm depths. For SOC analyses, three cores were taken per each plot and composited; for water retention analyses, three soil cores per each plot were taken and processed separately. Particulate organic matter samples were collected separately once a year in the springs from 2012 to 2014 at 0 to 10 cm depth.

Plant Biomass Sampling and Carbon Analysis. Biomass of corn and soybean at each plot was collected at harvest each year. Whole plants of corn and soybean were collected, then the grain was separated from the vegetative biomass and both were weighed to obtain measurements for yield and vegetative biomass.

Rye biomass was collected immediately before termination every year. Three random samples per plot were collected by using a 0.3×0.3 m (1×1 ft) sampling square. All the rye in the sampling square was cut at the soil surface to obtain the aboveground biomass. The rye was weighed, dried at 65°C (149°F), and then weighed again to obtain a dry biomass measurement.

Carbon content of plant biomass was measured by combustion analysis using a Costech ECS 4010 CHNSO Analyzer (Costech Analytical Technologies, Inc., Valencia, California). We choose to present the total C content in the aboveground biomass of the main and cover crops instead of biomass values as the metric most consistent with the SOC-processes focus of this study.

Rye Decomposition. Rye was collected from each field site prior to termination to use in the decomposition assessment by litterbag approach. Litterbags were 20×20 cm (7.8×7.8 in) and were constructed with 0.028 cm (0.01 in) plastic mesh. The rye was oven dried at 40°C (104°F), and a random subsample of 5 to 7 g (0.17 to 0.24 oz) of hand-cut dried rye, measuring approximately 5 to 7 cm (2 to 2.7 in) in length, was placed in each bag. In 2015, 60 litterbags were buried at 10 cm (4 in) depth in the plots with cover crop at the Mason site. In 2016, 340 litterbags were buried at 10 cm depth in the plots with and without cover crop at both Mason and Kellogg sites. Litterbags were removed at three time points of approximately one, three, and five weeks after the initial placement, which was at the beginning of the main crop growing season in the spring. After the litterbags were removed, the decomposed rye was cleaned of soil and other debris, oven dried at 65°C (149°F) and weighed (Alef 1995).

Total Soil Organic Carbon. Soil samples were collected using a Giddings hydraulic probe (Giddings Machine Company, Windsor, Colorado) (7.6 cm [3 in] in diameter). Approximately 0.7 kg (1.5 lb) of field moist soil from each depth increment was sieved to pass a 2 mm (0.07 in) sieve and air-dried. A 5 g (0.17 oz) subsample was then ground to a fine powder using an 8500 Shatterbox (Spex Sample Prep, Metuchen, New Jersey) in preparation for flash combustion analysis by the Carlo Erba EA 1108 (CE Elantec Inc., Lakewood, New Jersey). Values were corrected for small amounts of inorganic C at the Mason site through the acetic acid neutralization method.

Particulate Organic Carbon. Approximately 20 g (0.7 oz) of soil was subsampled from the sieved and air-dried soil to use for POC analysis. Particulate organic C was chemically dispersed using a 5% sodium hexametaphosphate ($[\text{NaPO}_3]_6$) solution. The soil solution was then passed through a 53-micron sieve. The contents of the sieve were oven dried at 60°C (140°F) and ground in the 8500 Shatterbox (Cambardella and Elliott 1992). Final C analysis of the samples were conducted using a Costech ECS 4010 CHNSO Analyzer.

Water Retention. Soil samples for water retention analysis were collected in the spring before corn or soybean planting in brass rings 5.5 cm (2.1 in) in diameter and 3 cm (1.2 in) in height. Pressure plates were

used to extract water at pressures of 0.05 bar, 0.1 bar, 0.33 bar, 1 bar, and 3 bar. The water content at 15 bar was measured by drying subsamples of the soil used from the pressure plate method and placing them in a desiccator above an oversaturated potassium chloride (KCl) solution (500 g KCl L⁻¹ [66.76 oz KCl gal⁻¹] water) for two months. For both methods, the difference in mass was used to calculate water content.

Statistical Analysis. Statistical analysis was performed using the MIXED procedure of SAS 9.4 (SAS Institute Inc., Cary, North Carolina). The statistical model for the analysis included presence/absence of cover crop, topographical position, and the interaction between them as fixed factors. Blocks nested within topography at each site were included in the model as the random effect and were used as an error term for testing the main effect of topography. When the interaction between topography and cover crops was statistically significant ($p < 0.1$), we used slicing to assess the effect of the cover crop presence at each topography level and the effect of topography at each level of the cover crop factor. When slicing effects were statistically significant ($p < 0.1$), comparisons between the means were conducted using *t*-tests. Because of high variability of the collected field data, we reported the results that were statistically significant at both 0.05 and 0.1 levels of significance.

Results and Discussion

We present the results in the order consistent with the hypothesized influences of cover crops and topography on the soil C processes that were considered in this study. Specifically, we start with topographical and cover crop effects on the main crop biomass, followed by topographical effect on the rye biomass, then topographical and cover crop effects on litter decomposition, and finally on POC, total soil organic C, and soil water retention.

Plant Biomass Carbon. The highest aboveground biomass of main crops, corn and soybean, and hence the highest amount of C in the aboveground biomass, were observed in depressions, followed by summits, and were the lowest in the slopes (figure 1a). Since the differences in the C contents of main crops in the presence or absence of the cover crop were not statistically significant, the results from both cover crop treatments were combined. An opposite trend was observed in the aboveground biomass of the

rye cover crop, which tended to be the lowest in depressions of both experimental sites (figure 1b); however, the differences were not statistically significant.

Positive effects of lower topography on the main crop yield observed in this study are consistent with published results where the yield of both corn and soybean were reported to be higher in depressions and/or concave sites (Kravchenko and Bullock 2000; Jiang and Thelen 2004; Da Silva and Alexandre 2005; Timlin et al. 1998). However, poor rye growth in depression positions of this study is contradictory to the results reported by Ladoni et al. (2016) from nearby large-scale agricultural fields. The discrepancy can be explained by greater extremity of depression positions used in this study, which represented poorly drained and often flooded areas. Depression positions used by Ladoni et al. (2016) were of a less extreme nature and typically had tile drainage. Negassa et al. (2015) in a study conducted at the same sites with a focus on rye growth and greenhouse gas emissions observed rye growth and impact consistent with our results.

Opposite trends in topographical effects on main and cover crops can be explained by their different growing seasons. Rye's growth primarily occurs during the cold and wet weather of falls and springs, when wet and cold depressions delay growth, in contrast to drier and warmer summits and slopes. The main crops, corn and soybean, grow during the summer when, if the weather is dry, it is advantageous to be in wetter depressions.

Rye Decomposition. Consistent with expectations, rye decomposition in depressions was higher than in slope and summit positions (figure 2). This trend was apparent in both 2015 and 2016, although the magnitude of decomposition was different between the two years. This was related to differences in weather patterns. In 2016, the presence of the cover crops tended to enhance decomposition at all topographical positions during first two sampling times (week 1 and week 2). By week 5, the amounts of decomposed residue were similar across all topographical positions and in both treatments with and without cover crops. Negassa et al.'s (2015) study at the same experimental sites reported greater carbon dioxide (CO₂) emissions at depression positions than on slopes and summits. They also observed greater CO₂ emissions due to cover crop presence, however only in topographical depressions.

Decomposition and CO₂ emissions are controlled in large part by soil moisture, temperature, and other variables that drive microbial activity (Raich and Potter 1995; Trumbore et al. 1996; Harrison-Kirk et al. 2013). Better soil moisture and soil temperature conditions leading to an increase of microbial activity and diversity explain the higher decomposition and CO₂ emissions in depressions. In nearby experimental fields, Wickings et al. (2016) observed microbial activity up to 55% higher in depression positions when compared to summits and slopes. Also, greater plant diversity is known to increase diversity and activity of soil microbial communities (Broughton and Gross 2000), which probably explains faster decomposition of rye residue in cover crop treatment's soil of this experiment.

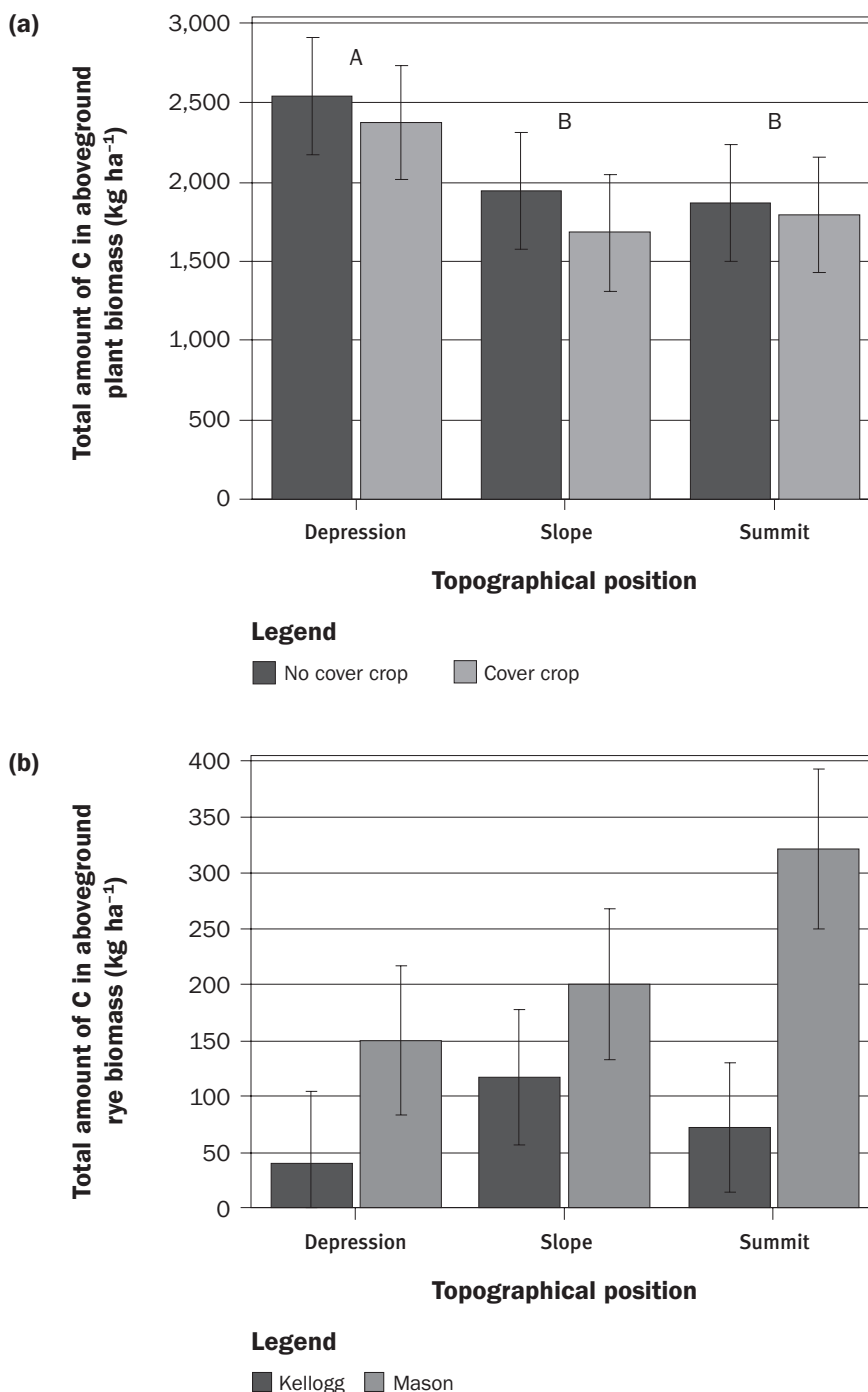
Particulate Organic Carbon. Overall, depressions contained more POC than summits and slopes (figure 3). Presence of cover crops resulted in significantly higher POC levels, but only in the slope and summit topographical positions ($\alpha = 0.1$).

Greater amounts of POC in the depressions probably result from greater biomass inputs from the main crops there, as well as from historical deposition of plant residue transported from higher topographies. Particulate organic matter is prone to downhill redistribution through erosion, eventually becoming concentrated in depression areas (Dungait et al. 2013; Cambardella et al. 1994).

It should be noted that while historically both losses of POC from higher topographies and its gains at lower topographies were likely taking place, during this study only the losses of particulate organic matter from summit and slope positions could have occurred. No gains in POC in depressions were possible because the plots were separated by wide grass alleys (~10 m [~32.8 ft]). While erosion was not measured in this study, it is known that presence of cover crops can substantially reduce erosion (Jacinthe et al. 2004), thus it can be assumed that in summit and slope positions greater erosion was taking place in plots without cover crops than those with cover crops. Therefore, the observed increases in POC due to cover crop presence in summits and slopes can be attributed to two causes: (1) greater rye cover crop growth at summits and slopes as compared to depression positions (figure 1b) and (2) potential POC losses due to erosion from the experimental plots without cover crops.

Figure 1

Average total carbon (C) in (a) main crop (corn and soybean) biomass and (b) rye biomass across topographical positions for each site. In (a) the letters represent statistically significant differences among topographical positions across both sites ($p < 0.1$). In (b) the effect of topography was not significant at Mason or Kellogg. Bars represent standard errors.



Note that this study does not account for the portion of C inputs from underground biomass, such as roots and root exudates. Belowground biomass inputs might have contributed to the greater impact of the cover crop on the summit and slope positions (Gale et al. 2000). Plants in the summits and slopes could have been devoting more energy to finding soil water and expanding their root system than the plants in depressions (Unger and Kaspar 1994).

Total Soil Organic Carbon. Total organic soil C followed the trend depression > summit > slope, but there were no significant differences between treatments with and without cover crops (table 1). The changes in total SOC might not become detectable until 7 to 10 years or more after changing management practices (Duiker and Lal 1999; Al-Kaisi et al. 2005). Thus, lack of statistical significance in total SOC results of this study were expected and consistent with other studies that did not observe statistically significant increases in total soil C (Kaspar et al. 2006; Basche et al. 2016).

Water Retention. Topography affected soil water retention at 0 to 10 cm (0 to 4 in) depth and at a lesser extent at 10 to 20 cm (4 to 7.8 in) depth (table 2). Depression soil tended to have significantly higher water retention at saturation (table 2). More differences in topographical position were observed in both cover and no cover plots in lower pressures than in higher pressures at the 0 to 10 cm depth.

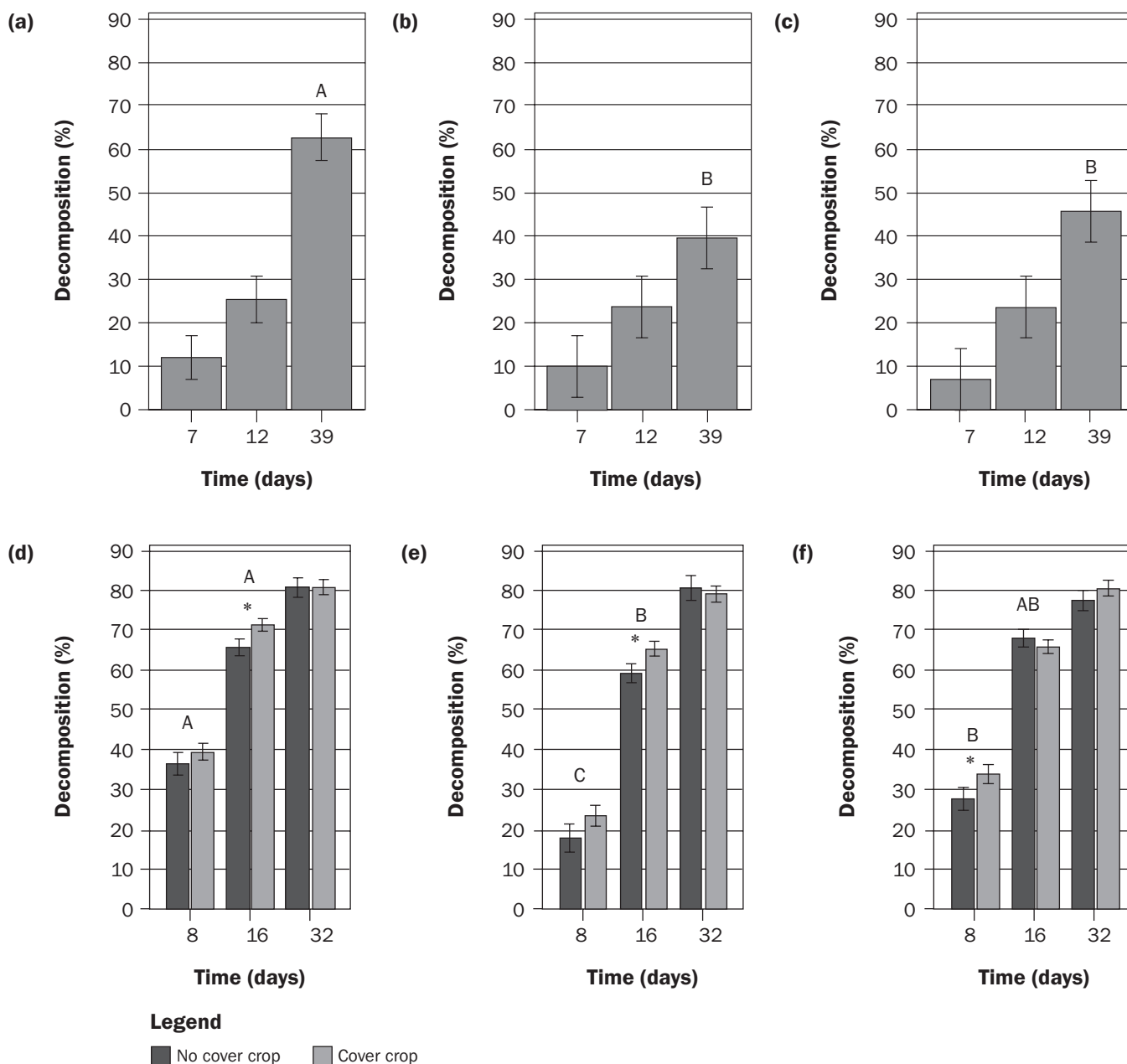
The differences between water retention data in the treatments with and without cover crops were not statistically significant. As with total SOC, a longer study duration appears to be necessary to observe significant changes in soil water retention due to cover crop use in the studied soil.

Summary and Conclusions

The study assessed the effects of adding rye cover crop to corn-soybean rotation in topographically diverse terrain at two experimental sites in Michigan. The aboveground biomass production by main (corn and soybean) crops was higher in topographical depressions than in summits and slopes, while the aboveground biomass of the rye cover crop followed an opposite trend. Thus, greater C inputs from the cover crop likely occurred in summit and slope positions. Decomposition of rye residue was greater in depressions than in summits and slopes.

Figure 2

Rye decomposition (%) at Kellogg and Mason experimental sites in (a,b,c) 2015 and (d,e,f) 2016. Letters indicate significant differences among topographical positions within each year and each sampling time ($p < 0.1$), stars indicate the instances in 2016 when decomposition in cover crop treatment was significantly higher than in the no cover treatment ($p < 0.05$). (a) and (d) indicate depression. (b) and (e) indicate slope. (c) and (f) indicate summit.



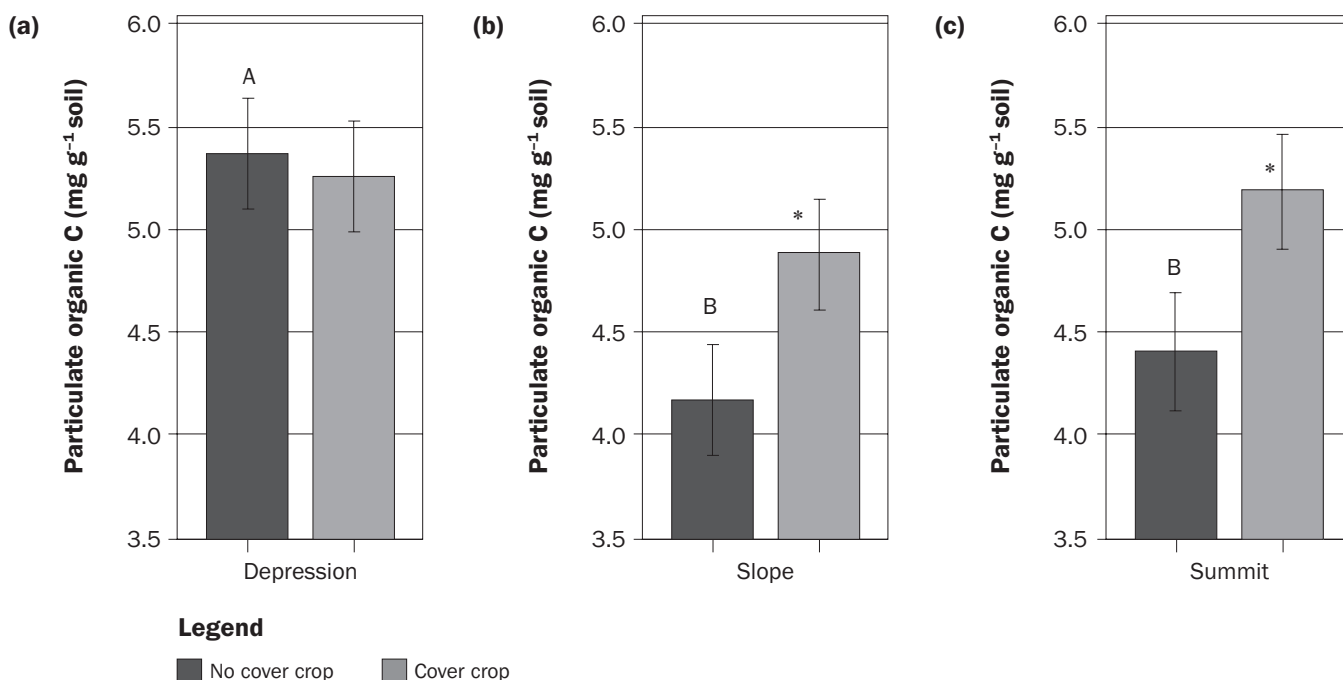
However, across all topographical positions, the presence of the cover crop in the rotation resulted in faster decomposition of the rye residue. Greater rye inputs in summit and slope positions along with potential reduction in erosion-driven soil losses from these positions led to greater POC levels in the treatment with the cover crop. No significant

differences between treatments with and without cover crop were observed in depression positions. Even though the observed increases in POC due to cover crop presence point to a tendency for soil C accrual, no significant differences between treatments with and without cover crop were observed in terms of the total soil C. In terms of soil water retention, there

were also no significant differences between treatments with and without cover crop. This result indicates the need for greater study duration to generate detectable effects of cover crops on these soil characteristics.

Figure 3

Particulate organic carbon (C) at the studied topographical positions and cover crop treatments at 0 to 10 cm depth across both experimental sites. Letters within no cover treatment represent statistically significant differences among the topographical positions ($p < 0.1$). There were no significant differences among the positions in the cover crop treatment. Asterisks mark the topographical positions where the difference between treatments with and without cover crop were statistically significant ($p < 0.1$).

**Table 1**

Soil organic carbon (C) concentration (%) at the four studied depths of the two experimental sites in spring of 2015.

		Soil organic C concentration (%)							
		Depth (cm)							
Site	Topographic position	0 to 10		10 to 20		20 to 40		40 to 60	
		Yes*	No	Yes	No	Yes	No	Yes	No
Kellogg	Depression	0.81b†	0.83b	0.71b	0.71	0.44	0.45	0.24	0.30
	Slope	0.54a	0.52a	0.43a	0.51	0.33	0.39	0.17	0.25
	Summit	0.69ab	0.71ab	0.55ab	0.61	0.37	0.39	0.20	0.21
Mason	Depression	1.09b	0.99b	0.93b	0.80b	0.42	0.50	0.30	0.17
	Slope	0.79a	0.74a	0.56a	0.54a	0.42	0.34	0.39	0.23
	Summit	0.76a	0.76a	0.59ab	0.58a	0.35	0.27	0.28	0.25

*"Yes" corresponds to the treatment with cover crops. "No" corresponds to the treatment without cover crops.

†Within each site, each depth, and each cover treatment the letters mark the statistically significant differences among the topographical positions ($p < 0.1$). No letters are shown when the differences were not statistically significant.

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Table 2

Volumetric water content at all pressures at the two studied depths of the two experimental sites in the spring of 2015.

Site	Topographic position	Depth (cm)	Volumetric water content (cm ³ cm ⁻³)															
			Negative pressure (bar)															
			0		0.003		0.05		0.01		0.33		1		3		15	
			Yes*	No	Yes	No	Yes	No	Yes	No	Yes	No	Yes	No	Yes	No	Yes	No
Mason	Depression	0 to 10	0.46b†	0.41	0.40	0.38	0.34b	0.29	0.33b	0.28	0.28	0.24	0.26	0.22	0.26	0.21	0.026	0.022
	Slope	0 to 10	0.40a	0.40	0.36	0.35	0.31a	0.30	0.31ab	0.30	0.27	0.27	0.25	0.24	0.24	0.24	0.028	0.028
	Summit	0 to 10	0.42a	0.39	0.35	0.36	0.30a	0.30	0.29a	0.30	0.26	0.26	0.23	0.24	0.22	0.24	0.024	0.023
Mason	Depression	10 to 20	0.44b	0.44	0.39	0.37	0.30	0.29	0.33	0.28	0.25	0.24	0.24	0.22	0.23	0.22	0.025	0.026
	Slope	10 to 20	0.40a	0.39	0.34	0.33	0.30	0.29	0.29	0.28	0.27	0.26	0.25	0.24	0.24	0.24	0.027	0.030
	Summit	10 to 20	0.40ab	0.42	0.35	0.35	0.28	0.30	0.28	0.29	0.26	0.27	0.24	0.25	0.24	0.25	0.030	0.026
Kellogg	Depression	0 to 10	0.47	0.47	0.44b	0.44b	0.35b	0.36b	0.32b	0.33b	0.29	0.29	0.27b	0.27	0.25	0.27b	0.026a	0.025
	Slope	0 to 10	0.42	0.43	0.37a	0.39a	0.25a	0.26a	0.23a	0.24a	—	0.26	0.17a	0.18	0.18	0.18a	0.034b	0.031
	Summit	0 to 10	0.47	0.48	0.43b	0.45b	0.35b	0.35b	0.32b	0.32b	0.29	0.28	0.27b	0.26	0.25	0.24ab	0.023a	0.022
Kellogg	Depression	10 to 20	0.44	0.44	0.42	0.41	0.34b	0.34	0.32	0.32	0.30	0.30	0.28	0.28	0.26	0.27	0.027a	0.028a
	Slope	10 to 20	0.39	0.39	0.37	0.36	0.26a	0.27	0.25	0.25	0.19	0.27	0.20	0.22	0.19	0.22	0.037b	0.037b
	Summit	10 to 20	0.47	0.45	0.43	0.41	0.32a	0.32	0.30	0.30	0.27	0.28	0.26	0.27	0.24	0.25	0.023a	0.025a

*“Yes” corresponds to the treatment with cover crops. “No” corresponds to the treatment without cover crops.

†Within each site, each depth, and each cover treatment the letters mark the statistically significant differences among topographical positions ($p < 0.1$). No letters are shown when the differences were not statistically significant.

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