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Reduced arctic tundra productivity linked with landform and climate change interactions

Mark J. Lara^{1,2}, Ingmar Nitze^{3,4}, Guido Grosse^{3,4}, Philip Martin⁵ & A. David McGuire⁶

Arctic tundra ecosystems have experienced unprecedented change associated with climate warming over recent decades. Across the Pan-Arctic, vegetation productivity and surface greenness have trended positively over the period of satellite observation. However, since 2011 these trends have slowed considerably, showing signs of browning in many regions. It is unclear what factors are driving this change and which regions/landforms will be most sensitive to future browning. Here we provide evidence linking decadal patterns in arctic greening and browning with regional climate change and local permafrost-driven landscape heterogeneity. We analyzed the spatial variability of decadal-scale trends in surface greenness across the Arctic Coastal Plain of northern Alaska (~60,000 km²) using the Landsat archive (1999–2014), in combination with novel 30 m classifications of polygonal tundra and regional watersheds, finding landscape heterogeneity and regional climate change to be the most important factors controlling historical greenness trends. Browning was linked to increased temperature and precipitation, with the exception of young landforms (developed following lake drainage), which will likely continue to green. Spatiotemporal model forecasting suggests carbon uptake potential to be reduced in response to warmer and/or wetter climatic conditions, potentially increasing the net loss of carbon to the atmosphere, at a greater degree than previously expected.

Over the past few decades, greening or increased vegetation productivity in Arctic tundra lowlands has been inferred from trends in satellite-derived Normalized Difference Vegetation Index (NDVI)^{1–3}. Researchers have speculated that these positive NDVI trends may be in response to reduced snow cover or warming, which may manifest on the landscape in the form of shrubification⁴, increased vegetation biomass (height, length, density)^{3,5}, changing phenoperiods⁶, or increased surface water associated with thermokarst⁷. However, others have hypothesized that greening trends may have been in response to summer sea ice retreat⁸ or increased rates of infrastructure development associated with oil drilling and exploration⁹. It is important to recognize that arctic tundra landscapes are highly heterogeneous and have historically been in a slow but continuous state of change associated with permafrost thaw and aggradation processes related to periglacial landscape dynamics¹⁰. Consequently, patches of browning or negative NDVI trends have also been commonly observed across arctic tundra regions, but until recently the greening signal prevailed¹¹. Since 2011, regional shifts toward surface browning have reversed the direction of the trend after nearly 33 years of arctic greening¹¹. If such change in greening indeed corresponds with a reduction in vegetation productivity or carbon uptake capacity via photosynthesis, then nearly all ecosystem and earth system models have not foreseen this shift. Therefore, it is urgent to understand what may be controlling this spatiotemporal shift in browning across the Arctic and if this change is anomalous or represents a new long-term trajectory towards reduced vegetation productivity and carbon uptake in the Arctic tundra¹².

To date, our knowledge of circumpolar patterns of greening are derived from coarse-resolution sensors on board satellites, such as the global 8 km and 1 km resolution Advanced Very High Resolution Radiometer (AVHRR) and to a lesser extent the 1 to 0.25 km resolution Moderate Resolution Imaging Spectroradiometer (MODIS)^{1,3}, all of which robustly produce spectral observations spanning gradients of time and space. However,

¹Department of Plant Biology, University of Illinois Urbana-Champaign, Urbana, Illinois, 61801, USA. ²Institute of Arctic Biology, University of Alaska Fairbanks, Fairbanks, Alaska, 99775, USA. ³Alfred Wegener Institute Helmholtz Centre for Polar and Marine Research, Periglacial Research Unit, 14473, Potsdam, Germany. ⁴Institute of Earth and Environmental Science, University of Potsdam, 14476, Potsdam, Germany. ⁵U.S. Fish and Wildlife Service, Fairbanks, Alaska, 99701, USA. ⁶U.S. Geological Survey, Alaska Cooperative Fish and Wildlife Research Unit, University of Alaska Fairbanks, Fairbanks, Alaska, 99775, USA. Correspondence and requests for materials should be addressed to M.J.L. (email: mjlara@illinois.edu)

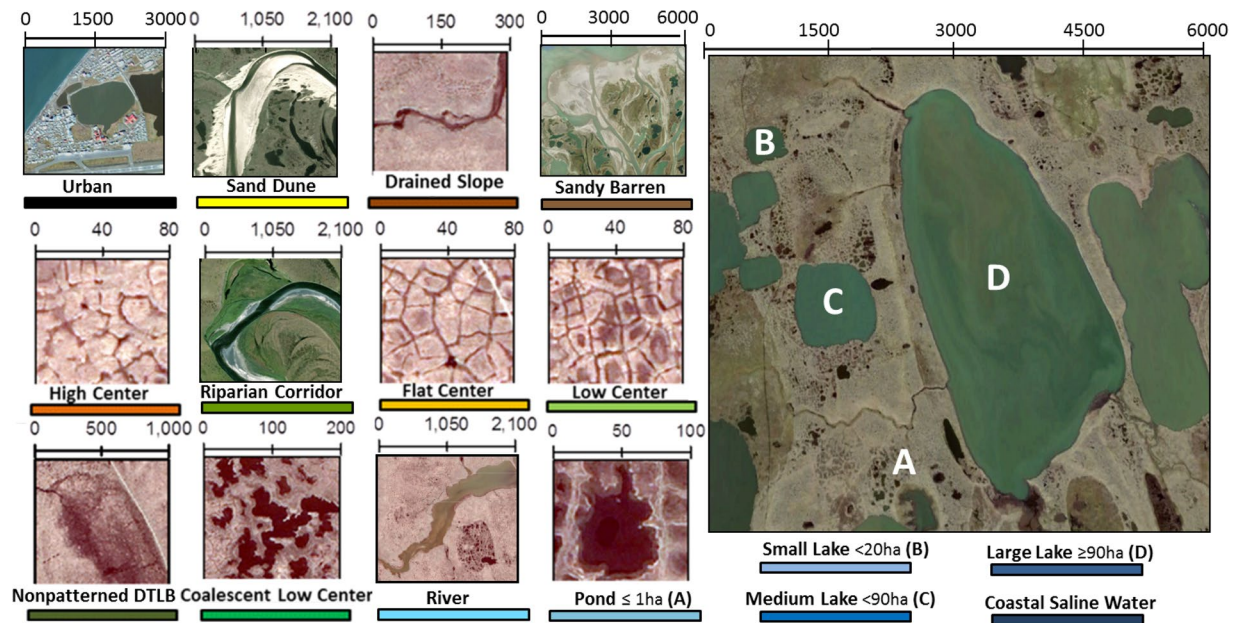


Figure 1. Dominant geomorphic types on the Arctic Coastal Plain of northern Alaska, as observed from high resolution satellite imagery⁴⁴ (copyright DigitalGlobe, Inc.). Figure created in Esri ArcMap 10.4.

data products generated by these observation platforms are limited in their ability to evaluate meso- to fine-scale patterns that may control greening and browning at coarser resolutions¹¹. Although, notable progress in the evaluation of landscape-level patterns of Arctic NDVI at fine-scales has been made^{9,13–17}, these studies typically feature a similar physiography and/or climate, limiting the evaluation of how similar vegetation types and/or landforms may respond to different climate regimes. It is difficult to assess such patterns across Arctic regions for a variety of reasons, but perhaps most importantly because of the limited availability of high quality land cover datasets, which do not exist, are only regionally specific, or do not adequately represent heterogeneity across the tundra landscape to decipher meaningful patterns¹⁸. Thus, without adequate spatial coverage and resolution of land cover data products used to link plot to landscape-level datasets, our ability to monitor and interpret patterns of change in the Arctic remains severely limited¹⁸.

The Arctic Coastal Plain (ACP) of northern Alaska represents an expansive geographic region of tundra where decadal trends in greening have recently strengthened^{12,11}, yet the magnitude of change varied nearly two-fold between the western and eastern ACP (i.e. eastern Chukchi and Beaufort, respectively)¹¹. Concurrently, the climate of the ACP has changed and appears to be regionally specific, with some regions warming more extensively than others and increasing/decreasing in precipitation. Additionally, across the ACP, the spatial distribution of fine-scale tundra landforms (here after referred to as “geomorphic types”), varies significantly^{19–21}, likely with different sensitivities to climate variability and change^{12,22}. Here we focus on evaluating key factors controlling decadal scale surface NDVI (i.e. greening versus browning), and evaluate what regions and geomorphic types are most sensitive to future climate change. We calculated decadal NDVI trends using Landsat imagery from sensors, Thematic Mapper (TM), Enhanced Thematic Mapper Plus (ETM+), and the Observing Land Imager (OLI), and assessed the variability in greenness from local to regional scales by using novel mapping techniques¹⁹ to create a polygonal tundra map (30 × 30 m resolution), which represents fifteen of the most dominant tundra geomorphic types (e.g., high/low-center polygon; Fig. 1), nested within regional watersheds²³, and ecological landscapes²⁰ spanning the ACP.

Methods

We studied the effects of climate and tundra geomorphic types on decadal scale trends in greening across the ACP, which stretches from the western coast along the Chukchi sea to the Beaufort coastal plains at the Alaskan/Canadian border (latitude: 68–71°N; longitude: 140–167°W). This region is representative of ~1.9 million km² of Arctic coastal tundra²⁴, characterized by low topographic relief, with abundant ice wedge polygons²⁵, thick permafrost²⁰, and predominantly wet sedge or herbaceous vegetation²¹. Summer and winter temperatures range from 5 to 15 °C and –18 to –40 °C, respectively (www.ncdc.noaa.gov). Annual precipitation is variable but typically ranges from 120–200 mm. We defined the ACP spatial domain by the geographic land area within the Northern/Southern Chukchi Sea Coast, Beaufort Sea Coast, Beaufort Coastal Plain, and a section of the Brooks Range Foothills on the Krusenstern Coastal Plain, all of which are composed of a high density of polygonal tundra or patterned ground (Fig. 1).

We expand upon a novel automated object based image analysis (OBIA) geomorphic mapping approach¹⁹, for characterizing tundra geomorphology across the ACP (58,691 km²). The initial application of tundra mapping was developed for a polygonal coastal tundra ecosystem on the Barrow Peninsula (1800 km²)¹⁹. Twelve Landsat-8 OLI (summer) satellite images (Supplemental Table S1) were processed and mosaicked within ArcGISTM 10.4

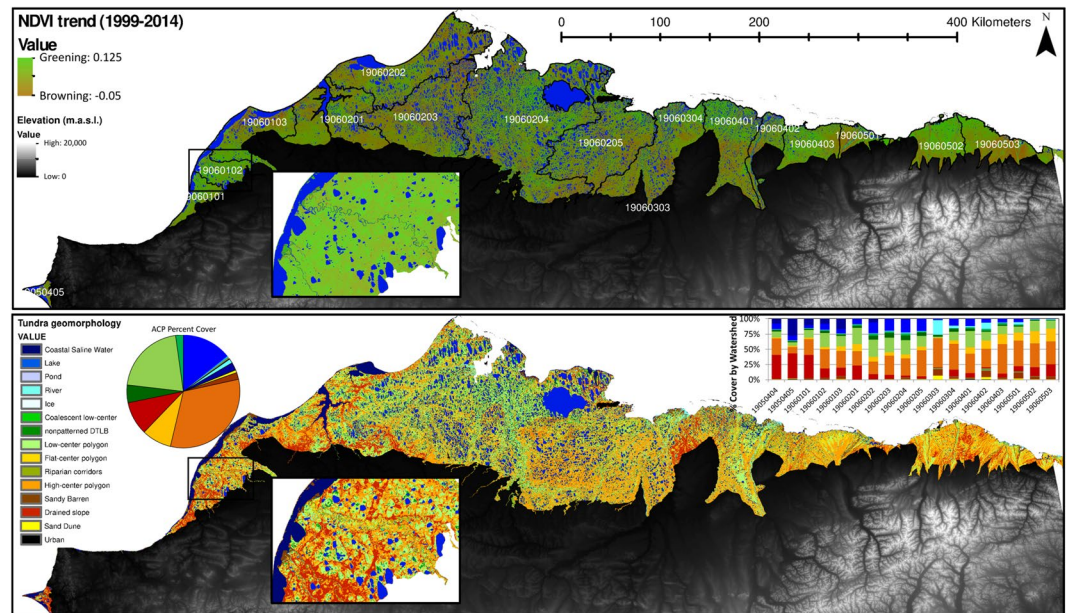


Figure 2. Decadal time scale Landsat derived greenness (NDVI) trend and regional watersheds (top panel), and tundra geomorphology map (bottom panel). Tundra geomorphology map was developed at a 30×30 m spatial resolution (see methods for more details). Note the variability in geomorphology distribution associated with regional watersheds (stacked bar chart). Greenness Trend map was created following established workflows¹, while tundra geomorphology map was created in Trimble eCognition v.9.1, both maps projected in Esri ArcMap 10.4.

(ESRI) for tundra geomorphology mapping. An OBIA land cover classifier (eCognition™ v.9.1, Trimble) was parameterized using various rules, thresholds, spectral indices, and proximity functions to differentiate between geomorphic types²⁶. Multiresolution segmentation and spectral difference algorithms were used to separate pixels into “image objects”, which were divided into open water, aquatic, wet, moist, dry classes using reference data (i.e. field/ground truth points and high resolution aerial/satellite imagery) and class thresholds based on Normalized Difference Water Index (NDWI).

A series of proceeding functions were developed using individual and combined spectral bands, geometric object shapes/sizes (i.e. perimeter, area, roundness), and proximity functions to further differentiate respective geomorphology classes²⁶. Using this approach we mapped fifteen geomorphological and hydrologically distinct geomorphic types (Figs 1 and 2) at 30×30 m spatial resolution, including (qualitatively ranked from wet to dry), 1) coastal saline water, which commonly encroach into terrestrial lakes due to processes related to coastal erosion or lagoon formation, 2) lakes (large: >90 ha, medium: ≤ 90 and >20 ha, small: ≤ 20 ha), 3) rivers, 4) ponds, 5) coalescent low-center polygons, 6) nonpatterned drained thaw lake basins, 7) low-center polygons, 8) flat-center polygons, 9) riparian corridors, 10) high-center polygons, 11) sandy barrens, 12) drained slopes, 13) sand dunes, 14) ice/snow, and 15) urban. Refer to Supplemental Table S2 for surface characteristics related to moisture, relief, and vegetation communities specific to geomorphic type. For this analysis, all lake sizes were grouped into one “Lake” category.

The tundra geomorphic map (Fig. 2) was validated using an array of oblique aerial/ground based photography²¹ and 249 high resolution (2.5 m resolution) SPOT-5 orthorectified image tiles covering $>80\%$ of the ACP (Supplemental Table S3; gina.alaska.edu)²⁶. We used a stratified random sampling design, where 1000 reference sites were used to evaluate map accuracy within both the Arctic Peaty Lowlands and the Arctic Sandy Lowlands²⁰.

We computed and analyzed NDVI trends using Landsat imagery following workflows developed for the Siberian, Lena River Delta¹³, where each pixel has a temporal coverage of 40 to 110 observations collected between 1999–2014²⁶. Throughout this paper, we refer to greening and browning as increased and decreased NDVI, respectively. Prior to data extraction from NDVI maps (Fig. 2), all coastal saline water, lakes, rivers, and urban pixels were removed. Primarily due to limited data acquisition across the ACP prior to 1999, we restrict decadal greenness trends (i.e. absolute change and percent change relative to 1999) to 1999–2014 throughout this paper.

Due to our use of multiple Landsat sensors (i.e. TM, ETM+, and OLI) within the NDVI trend map product, we calculated the sensor bias of NDVI at three different locations across Alaska, each containing a sample of 40,000 pixels²⁶. We found minor calibration differences between sensors (i.e. one percent of the signal), while sensor specific NDVI distributions were consistent²⁶.

Two data subsets were extracted from decadal NDVI products (i.e. absolute and percent change), where local to regional absolute and percent change in greenness was extracted for each geomorphic type in each Hydrological unit code 8 (HUC 8) watershed (Supplemental Fig. S4, Supplemental Table S5), for use in boosted regression tree (BRT)²⁷ and multivariate regression model analysis (Supplemental Table S6), respectively. Hydrological unit code

	Climate Model	RCP 4.5		RCP 8.5	
		Mean	Stdev	Mean	Stdev
Temperature change (°C)	MRI-CGCM3	1.37	0.22	1.21	0.18
	GISS-E2-R	1.4	0.767	1.94	0.114
	IPSL-CM5A	1.62	0.129	3.45	0.2
	NCAR-CCSM4	2.93	0.19	2.56	0.2
	GFDL-CM3	5.46	0.58	4.63	0.44
Precipitation change (mm)	MRI-CGCM3	9.27	15.96	17.18	10.28
	GISS-E2-R	31.27	10.31	16.52	11.98
	IPSL-CM5A	35.15	11.15	9.48	7.93
	NCAR-CCSM4	14.85	8.22	28.4	10.19
	GFDL-CM3	46.6	19.99	45.47	9.45

Table 1. Projected change in precipitation and temperature (2020–2029) relative to climate normals for the Arctic Coastal Plain of northern Alaska by climate model and RCP emission scenario. Models are generally listed from lowest to highest projected change. Temp. Normal (1960–1999): -11.51°C , ± 0.93 . Precip. Normal (1960–1999): 236.63 mm, ± 21.29 .

8 watersheds were computed by the Alaska Watershed and Stream Hydrography Enhanced Dataset Project²³, where each watershed was divided into two subunits (i.e. lowland/upland) associated with the second elevation quantile. Prior to spatial analysis, we identified outlier pixels by determining if NDVI change was $>75.0\%$ or $<-75.0\%$ per decade, and used a nearest neighbor filtering algorithm for the recalculation of pixels within a 5×5 pixel window, which represented $<0.1\%$ of all pixels on the ACP (residual unfiltered open water pixels).

Datasets used in all analyses included the following predictor variables: elevation (60 m resolution), climate normals (1960–1999), *change* (difference between 2000–2010 and “normals”), and *anomalies* (“change”/“normal”) for annual temperature, precipitation, and potential evapotranspiration downscaled to 771 m resolution by the Scenarios Network for Alaska and Arctic Planning²⁸. In addition, we calculated the percent cover of soil moisture regime for each 771×771 m pixel (resolution standardized with input climate data), estimated by associations between geomorphic type and field observations^{19,22,29}. We combined geomorphic classes into their respective moisture categories as follows: *open water* (coastal saline water, lakes, rivers), *aquatic* (ponds, coalescent low-center polygons), *wet* (nonpatterned drained thaw lake basins, low-center polygons), *moist* (flat-center polygons, riparian corridors), *dry* (high-center polygons, drained slopes) and *other* (sandy barrens, sand dunes, ice/snow, urban).

The TreeNet Gradient Boosting machine, developed within the Salford Predictive ModelerTM v.8.0 (Salford Systems), was used to run all BRT analyses. A “shaving” procedure was used to iteratively remove and rerun the BRT analysis to minimize the mean squared error, where 80% ($n = 276$) of the dataset was used for model development (i.e. learning) and 20% ($n = 53$) was used for independent model evaluation (i.e. testing). BRT learning rate, tree complexity, and loss criterion, was set to 0.1, 6, and Huber-M, respectively. To ensure reproducibility, we used a seed state of 987654321 for model initialization. Partial dependency plots were used to show the response of individual predictor variables to the BRT analysis, using fitted functions²⁷. Fitted functions detail the effect of a variable on the response after accounting for average effects of all other variables in the model²⁷.

Stepwise multivariate regression and Pearson’s correlation analyses were ran in Jmp ProTM v.10 (SAS). Input datasets used in the stepwise procedure (Supplemental Table S7) were all transformed to fit the assumptions of normality, and only important factors identified by the preceding BRT analysis were input into the stepwise procedure, used to predict regional greenness trends by fitting potentially important climate and/or environmental variables. A five-fold cross validation was concurrently performed, which divided the dataset into 5 subsets or 80:20 and iteratively used each 80% subset to predict the other 20% (e.g. k-1). The average R^2 of all models was then calculated. Multivariate regression models were used to forecast change in the NDVI over the next decade (i.e. 2020–29), forced by the top five climate models to accurately represent Arctic and Alaskan regions^{28,30} used in the IPCC Fifth Assessment Report (Table 1). These included the Community Earth System Model 4 (NCAR-CCSM4), Coupled Model 3.0 (GFDL-CM3), ModelE/Russell (GISS-E2-R), Institut Pierre-Simon Laplace Coupled Model v5A (IPSL-CM5A), and the Coupled General Circulation Model v3.0 (MRI-CGCM3). Selected model runs included the AR5 representative concentration pathways RCPs 8.5 (high) and 4.5 (low). We assumed no change in tundra geomorphic type distribution for model simulations.

Results

We found the regional distribution of tundra geomorphic types and greenness to vary markedly across the Arctic Coastal Plain of northern Alaska (Fig. 2, Supplemental Fig. S4). The newly developed tundra geomorphology map represented the spatial distribution of polygonal tundra well with an overall map accuracy and Cohen’s Kappa coefficient of 76% and 0.73, respectively (Supplemental Table S3)²⁶. Map statistics indicated that high-center polygons, low-center polygons, and lakes were the predominant features across the $\sim 60,000 \text{ km}^2$ ACP representing 69.3% of the total land cover area (Fig. 2). Watersheds ranged from 15 to $13,406 \text{ km}^2$ with a median of $2,128 \text{ km}^2$, which also varied in geomorphic type distribution (Fig. 2, Supplemental Fig. S4). Historical NDVI trends varied between regional watersheds (Supplemental Fig. S4), and across geomorphic types (Supplemental Tables S5 and S6), suggesting that indeed trends in greening and browning are locally and regionally specific. Generally, across the ACP historical trends in NDVI ($\pm$ standard deviation) differed by geomorphic type as observed in ponds (0.005 ± 0.22), coalescent low-center (0.035 ± 0.11), nonpatterned drained thaw lake basins (0.032 ± 0.14),

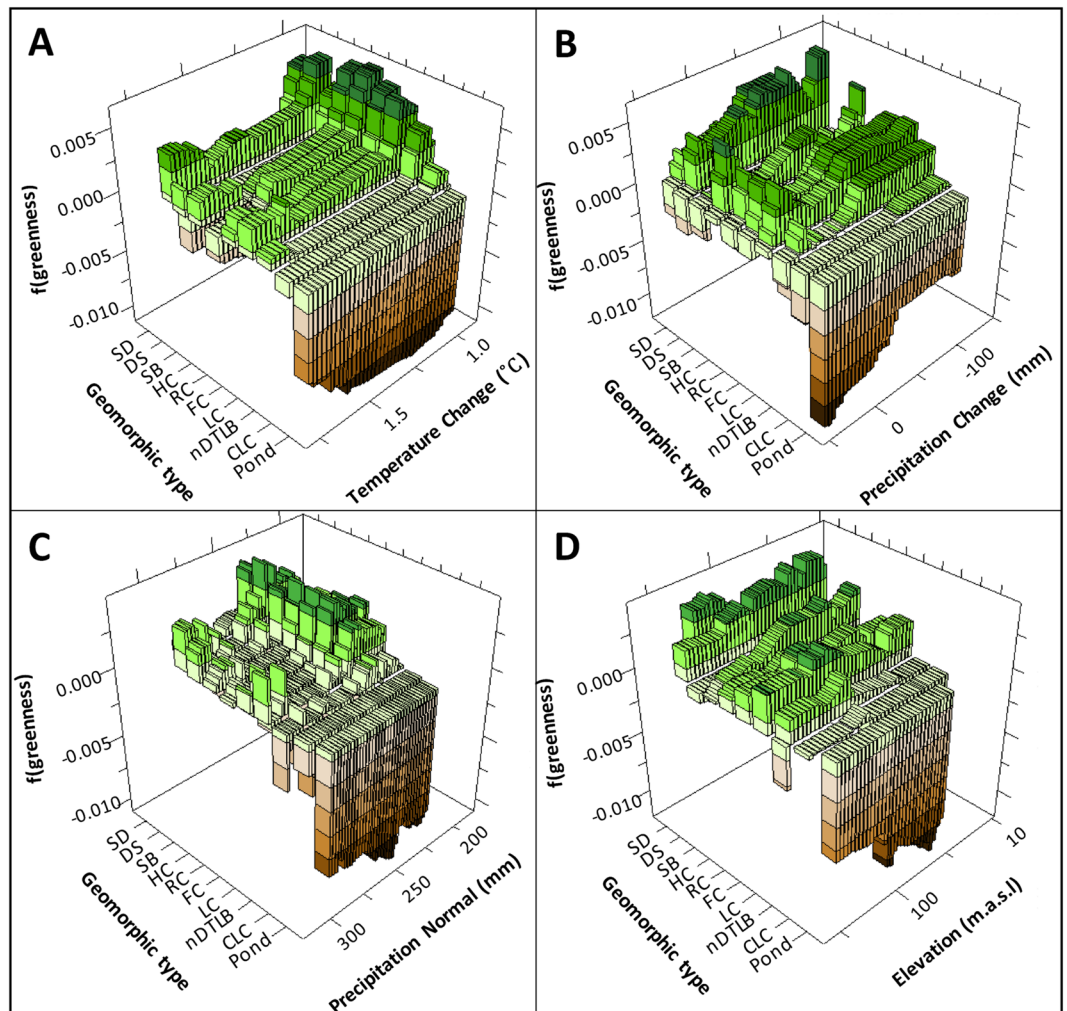


Figure 3. Three dimensional partial dependency plots from gradient boosting analysis, illustrating the strong interaction between geomorphic type and temperature change (A), precipitation change (B), precipitation normal (C), and elevation (D). Generally, green and brown colors indicate positive and negative NDVI trends, respectively. Geomorphic type acronyms correspond to sandy barren (SB), sand dune (SD), drainage slope (DS), high-center polygon (HC), flat-center polygon (FC), low-center polygon (LC), riparian corridors (RC), nonpatterned drained thaw lake basins (nDTLB), and coalescent low-center polygon (CLC). Figure created in Salford Systems TreeNet Salford Predictive Modeler v.8.0.

low-center (0.044 ± 0.06), flat-center (0.046 ± 0.05), riparian corridor (0.039 ± 0.13), high-center (0.041 ± 0.05), and drained slope (0.042 ± 0.04).

The gradient boosting analysis used 183 regression trees to construct a robust model that well represented the variability in historical greenness trends (learning/testing, $R^2 = 0.72, 0.62$). After “shaving” or recursive predictor elimination procedures were complete, the final BRT analysis determined geomorphic type (100.0), temperature change (64.5), precipitation change (61.5), elevation (49.1), and precipitation normal (50.3) to be the most important factors (i.e. “variable importance”, expressed as a percentage, scaled to the most important factor) controlling greenness trends (Fig. 3, Supplemental Fig. S8). Partial dependency plots illustrate the strong control of geomorphic type on NDVI (Fig. 3, Supplemental Fig. S8), as generally, the higher the soil moisture the lower the rate of decadal greening. This is in line with partial dependency plots for precipitation and temperature change, which indicate that NDVI increased (i.e. greening) with reduced precipitation and cooler temperatures, whereas the NDVI decreased (i.e. browning) with increased precipitation and warmer temperatures (Fig. 4). Specifically, precipitation change greater than +31 mm was associated with substantial decreases in NDVI, while precipitation change associated with drier conditions, less than -10 mm increased NDVI. Interestingly, partial dependency plots for temperature change reveal that a potential NDVI threshold exists, inferring that if warming is limited to below +1.06°C, the tundra continues to experience increased NDVI, but if warming exceeds that threshold NDVI is substantially reduced (Fig. 3, Supplemental Fig. S8). However, greening may slowly resume after +1.70°C of warming. Although elevation and precipitation normal did not account for most of the BRT model variability, they nevertheless notably impacted tundra geomorphology greenness (Fig. 3). Because temperature and precipitation change predictors did not meet the assumption of normality, we used log transformed temperature and precipitation anomalies in the ensuing multivariate analysis.

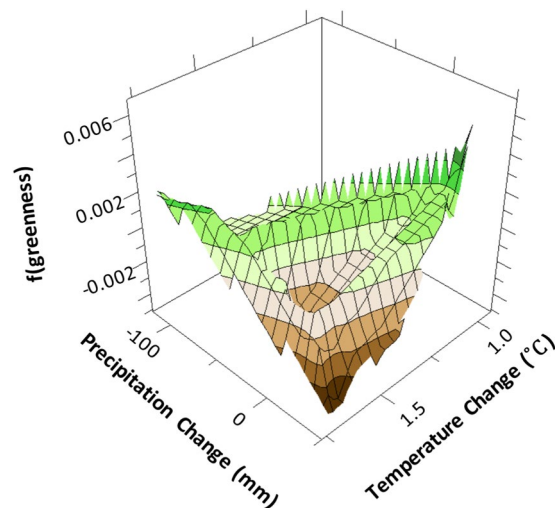


Figure 4. Three dimensional partial dependency plot from gradient boosting analysis, illustrating the impact both temperature change and precipitation change may have on NDVI trends. Figure created in Salford Systems TreeNet Salford Predictive Modeler v.8.0.

Geomorphic Type	Temperature Anomaly	Precipitation Anomaly	Elevation
SD	0.13	-0.43	0.14
DS	-0.15	-0.36	-0.07
SB	-0.58	0.13	0.12
HC	-0.42	-0.07	-0.27
RC	-0.48	0.23	0.1
FC	-0.27	-0.13	-0.02
LC	-0.49	0.18	-0.18
nDTLB	-0.31	-0.19	0.03
CLC	-0.21	-0.21	0.17
Pond	0.53	-0.60	0.27

Table 2. Pearson correlation coefficients for geomorphic type and potential drivers of NDVI change. Positive and negative correlations indicate greening and browning, respectively, with increasing climate or elevation parameters. See Fig. 3 for geomorphic type definitions. Bolded = $p \leq 0.05$; Italics = $p \leq 0.1$.

Regionally, the cross-validated multivariate regression model identified climate anomalies of precipitation and temperature, wet%, and elevation to be the most important predictors of historical NDVI trends ($R^2 = 0.49$, $p < 0.001$; five-fold cross validation $R^2 = 0.30$), which suggests that increasing the decadal precipitation and temperature anomalies will decrease NDVI trends (i.e. browning), whereas an increase in wet% (i.e. nonpatterned drained thaw lake basins and low-center polygons) area will increase NDVI trends, both consistent with the BRT analysis. Furthermore, NDVI correlations with temperature and precipitation anomalies specific to geomorphic type, provide insight into regional scale controls on decadal NDVI change, as nearly all geomorphic types were sensitive to either temperature or precipitation anomalies (Table 2). Significant decreases in NDVI with warmer temperatures were found for high-center polygons, low-center polygons, sandy barrens, nonpatterned drained thaw lake basins, and riparian corridors, whereas pond NDVI increased with warming. Decreases in NDVI were also identified for sand dunes, drained slopes, and ponds associated with increased precipitation (Table 2). The geomorphic type most sensitive to temperature-mediated NDVI change, inferred from the highest decadal rate of change, were ponds ($R^2 = 0.28$, $p = 0.002$, $n = 30$) and riparian corridors ($R^2 = 0.30$, $p = 0.001$, $n = 31$), which increased and decreased NDVI with warming, respectively. In contrast, the geomorphic types most sensitive to precipitation-mediated reduction in the NDVI trend, were ponds ($R^2 = 0.36$, $p < 0.001$, $n = 30$), followed closely by drained slopes ($R^2 = 0.22$, $p = 0.008$, $n = 31$). We did not identify any significant relationships linking increased greening with increased precipitation among geomorphic types (Table 2), consistent with BRT analysis and regression model. We associated climate sensitivities to patterns of greenness in sandy barrens and sand dunes to spectral differences associated with saturated versus unsaturated soils in response to warmer/wetter conditions and not explicitly a vegetation response in these sparsely vegetated types (Table 2). Cumulatively, the relative importance of regional climate change for predicting the trajectory of greening versus browning was made strikingly apparent as ~61% or 35,800 km² of the ACP were sensitive to change in temperature, whereas only ~10% or 5,900 km² were sensitive to change in precipitation.

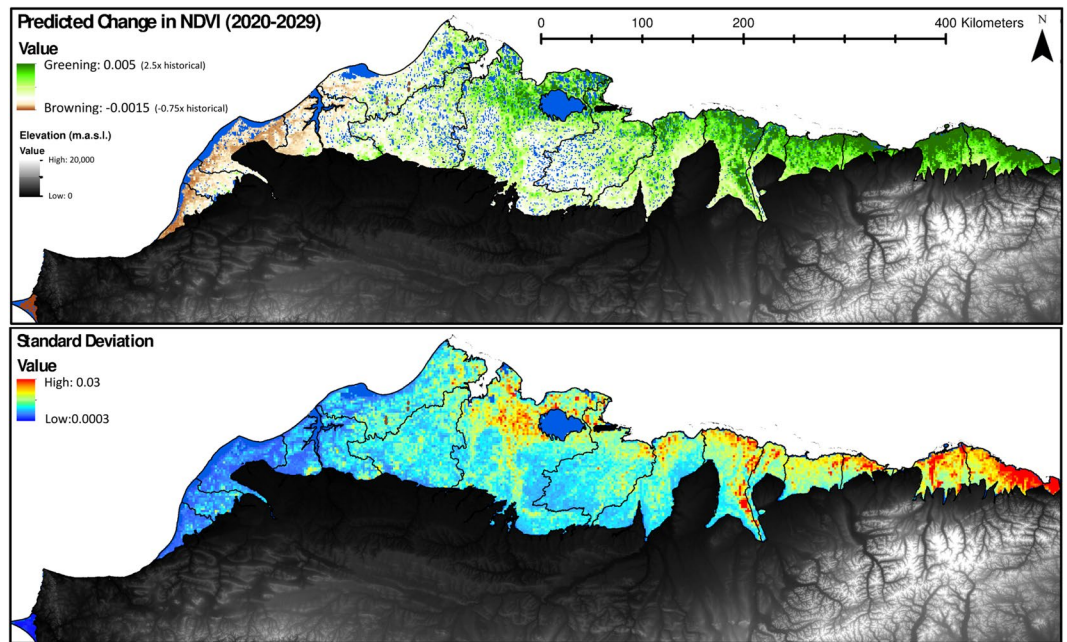


Figure 5. Forecasted change in greenness (2020–2029) relative to the long-term (1984–2014: +0.002) trend, using the five model mean and AR5 8.5 emission scenario (top panel). As a metric of future greenness uncertainty, the standard deviation is computed for all model outputs and emission scenarios (bottom panel). Maps created in Esri ArcMap 10.4.

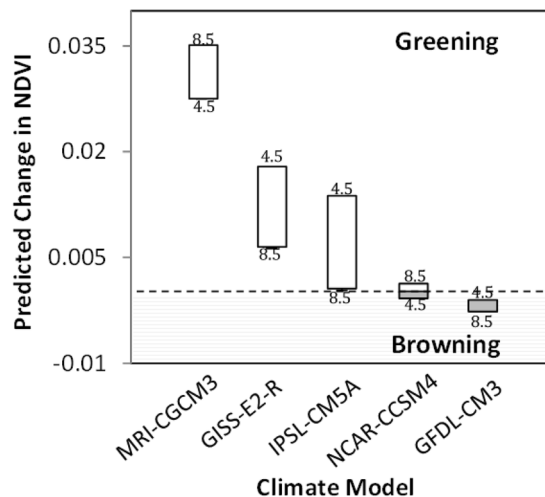


Figure 6. Projected variability in greenness (2020–2029) using five climate models and two emission scenarios. Above and below dotted line represent greening and browning relative to long-term trends, respectively, while RCP emission scenario highs and lows are indicated by 8.5 and 4.5, respectively.

To evaluate how NDVI trends may change over the next decade (i.e. 2020–2029), derived multivariate models were applied across the ACP using five IPCC climate models and two emission scenarios (i.e. RCP 8.5 and 4.5). Generally, we find the projected change in NDVI will likely vary in magnitude and spatial distribution (Fig. 5) associated with future climate change (Table 1). Simulations suggest the greatest magnitude of change in greening and browning relative to long term trends (1984–2014) were for the MRI-CGCM3 RCP 8.5 and GFDL-CM3 RCP 8.5 climates, respectively (Fig. 6). However, although, predicted change in NDVI notably varied between climate models, emission scenarios add another degree of uncertainty, as highlighted by the trajectory shift from greening to browning found with NCAR-CCSM4 and IPSL-CM5A (Fig. 6). Moreover, predicted change in NDVI may be expected to vary spatially (Fig. 5), as all models indicate the western Chukchi coast will experience the greatest browning (Fig. 5) and the northeastern Beaufort coastal plain will experience the most greening. However, regions anticipated to have the highest greening trends by 2020–2029, also have the highest uncertainty or disagreement among model outputs (Figs 5 and 6).

Discussion

As the climate continues to warm, we can expect an increased occurrence of browning across the Arctic Coastal Plain of northern Alaska (Figs 2 and 5), but the rate at which the landscape browns, depends on the magnitude of temperature and precipitation change (Fig. 4), with the exception of young geomorphic types from recently drained lakes (e.g. nonpatterned drained thaw lake basins and low-center polygons) that will likely continue to green. These patterns were highlighted in our analysis as historical NDVI trends varied by tundra geomorphic type (Fig. 3). Browning was particularly sensitive to elevated temperature and/or precipitation (Table 2), both of which are predicted to increase across arctic tundra regions³¹.

To evaluate how NDVI might change over the next decade (2020–2029) we developed a predictive multivariate model using a range of decadal scale climate and geospatial data inputs (Supplemental Table S7), selected as important predictors in the BRT analysis. However, because of the decadal scale input datasets³² we are unable to explicitly and directly evaluate year-specific observations of browning (i.e. 2011)¹¹ within this modeling framework. We still were able to provide insights into potential controls on greenness. Generally, it has been postulated that recent observations of browning, may have arisen from an array of annual/seasonal regional/site level changes that are largely correlated with decadal climate and land cover datasets, such as the delayed onset of the growing season and longer snow cover³², summer cloudiness³², winter warming^{33,34}, or thermokarst development⁹. Model NDVI forecasting (Figs 5 and 6) identified greening to likely continue on the ACP, but will be highly dependent on the amplitude of future climate change (Fig. 6 and Table 1). Predicted NDVI by 2020–2029 was found to be variable dependent on climate model and RCP (Fig. 6), but the magnitude in greening versus browning was generally explained by the inverse relationship between temperature and NDVI change, as the greater the temperature change the lower the NDVI. Generally, all simulations find regional specific browning to occur on the southwestern Chukchi coast and greening to occur on the northeastern Beaufort coastal plains but the magnitudes vary by climate model (Fig. 6).

Trends in MaxNDVI estimated from Global Inventory Modeling and Mapping Studies (GIMMS) datasets¹¹, were similar to that evaluated using Landsat imagery (Fig. 2), though, higher resolution imagery and newly developed maps enabled the evaluation of spatiotemporal heterogeneity in greenness, highlighting the strong interaction between landforms and climate influencing greenness trends (Table 2 and Fig. 3). We found trends in greenness were specific to geomorphic type and sensitive to either temperature or precipitation change (Table 2), with the exception of tundra ponds, which was found to be sensitive to both climatic drivers. The climatic sensitivity of tundra ponds is in line with hypothesized drivers of vegetation change^{22,35,36}, as warmer temperatures may thaw permafrost and increase available nutrients in the water column^{37,38}, potentially enabling the expansion of aquatic macrophytes^{35,36}. Increased precipitation is likely to increase the ratio of standing water to macrophyte distribution in ponds, manifesting as an increased browning signal⁹. Although, relationships between decreasing NDVI and increasing temperature were identified for high-center polygons, low-center polygons, and riparian corridors (Table 2), it is important to note that significant greening occurred across all geomorphic types until temperatures warmed beyond +1.06 °C, where NDVI decreases sharply (Fig. 3, Supplemental Fig. S8). However, the greening trend may begin to increase if warming is greater than ~+1.70 °C. Generally, at the landscape scale, we find greening to increase if precipitation decreases more than -10 mm relative to normal precipitation patterns, whereas browning increased with precipitation. This pattern was likely identified in response to either wetting/drying of water saturated tundra soils and/or increased cloudiness (associated with increased precipitation) that may decrease productivity in response to a reduction in photosynthetically active radiation³². However, although NDVI generally decreased with increased precipitation, this pattern was not able to explain the regionally specific variability in NDVI that occurred among landforms (Fig. 3).

Although this work increases our understanding of past and future greenness patterns in arctic coastal tundra ecosystems, it is unclear if differences in greening versus browning trends for each geomorphic type (Table 2) were associated with local disturbances. For example, the prevalence of thermokarst pits (upland tundra: drained slopes and high-center polygons) and associated increases in surface wetness^{7,39}, seasonally dependent patterns of herbivore activity (low-center polygons, nonpatterned drained thaw lake basins, and drained slopes)⁴⁰, and/or plant community change (ponds, nonpatterned drained thaw lake basins, low-center polygons)³⁶. The interpretation of greening and browning signals spanning small to large spatiotemporal scales (i.e. plot to pan-Arctic) is difficult, as multiple climatic and environmental factors influence NDVI trends, which are likely related but spatially dependent. For example, at the Pan-Arctic scale, greening has been linked to extra-Arctic processes such as CO₂ fertilization^{41,42}, atmospheric nitrogen deposition⁴¹, as well as marine-terrestrial interactions related to ocean circulation patterns and sea ice decline^{1,11}. At the landscape scale, other local processes such as the frequency and magnitude of wildfires², regional climate change^{1,2,11,41,42}, infrastructure development⁹, and shrubification⁴³ are important. We find the distribution of tundra geomorphology (Fig. 3 and Table 2, Supplemental Fig. S2) is another major factor modulating greening and browning signals in ways previously not recognized. At the fine scale, processes related to changing patterns of phenology⁴⁴, plant community change²², herbivory^{29,45} and surface hydrology driven by thermokarst^{7,46} can also notably impact vegetation productivity and NDVI. A holistic ecosystem perspective is required to unravel the spatiotemporal complexity involved with changing tundra greenness, which we are beginning to understand, but are limited by few observational data and comprehensive analyses across scales and platforms of observation.

Our findings indicate tundra geomorphic heterogeneity and regional climate change are dominant factors controlling decadal scale trends in greenness. Thus, a detailed understanding of how climate interacts with landforms is necessary for evaluating the spatiotemporal ecosystem interactions that impact regional-global patterns of plant productivity. Although, correlations between NDVI and vegetation productivity are robust across latitudinal gradients⁴⁷, our findings have several implications for local controls on vegetation productivity in the expansive (~1.9 million km²)²⁴ arctic coastal tundra ecosystem. Ecosystem and earth system models generally predict plant productivity to increase associated with projected climate change across northern latitudes.

However, assuming our observed greenness trends correspond with productivity trends, we predict a reduction in carbon uptake potential across much of the ACP of northern Alaska in response to projected warmer and/or wetter climatic conditions. In combination, with deeper active layer depths⁴⁸ exposing increased soil carbon to decomposition⁴⁸, this further increases the potential for a net loss of carbon to the atmosphere, at a greater degree than previously expected. It is important to better understand how regional-global trends in greening and browning correspond to both local and regional phenomena to enhance our predictive capacity and ability to detect change in plant productivity across the Pan-Arctic to constrain our predictive uncertainty related to the future state and fate of global climate change⁴⁹.

References

- Bhatt, U. S. *et al.* Circumpolar Arctic Tundra Vegetation Change Is Linked to Sea Ice Decline. *Earth Interact* **14** (2010).
- Goetz, S. J., Bunn, A. G., Fiske, G. J. & Houghton, R. A. Satellite-observed photosynthetic trends across boreal North America associated with climate and fire disturbance. *Proc. Natl. Acad. Sci. USA* **102**, 13521–13525 (2005).
- Jia, G. S. J., Epstein, H. E. & Walker, D. A. Greening of arctic Alaska, 1981–2001. *Geophysical Research Letters* **30** (2003).
- Myers-Smith, I. H. *et al.* Shrub expansion in tundra ecosystems: dynamics, impacts and research priorities. *Environ Res Lett* **6** (2011).
- Elmendorf, S. C. *et al.* Plot-scale evidence of tundra vegetation change and links to recent summer warming. *Nat Clim Change* **2**, 453–457 (2012).
- Bokhorst, S., Bjerke, J. W., Street, L. E., Callaghan, T. V. & Phoenix, G. K. Impacts of multiple extreme winter warming events on sub-Arctic heathland: phenology, reproduction, growth, and CO₂ flux responses. *Global Change Biology* **17**, 2817–2830 (2011).
- Liljedahl, A. K. *et al.* Pan-Arctic ice-wedge degradation in warming permafrost and its influence on tundra hydrology. *Nat Geosci* **9**, 312 (2016).
- Bhatt, U. S. *et al.* Implications of Arctic Sea Ice Decline for the Earth System. *Annual Review of Environment and Resources*, **39**, 57 (2014).
- Raynolds, M. K. & Walker, D. A. Increased wetness confounds Landsat-derived NDVI trends in the central Alaska North Slope region, 1985–2011. *Environ Res Lett* **11** (2016).
- Jorgenson, M. T. & Grosse, G. Remote Sensing of Landscape Change in Permafrost Regions. *Permafrost Periglac* **27**, 324–338 (2016).
- Bhatt, U. S. *et al.* Recent Declines in Warming and Vegetation Greening Trends over Pan-Arctic Tundra. *Remote Sens-Basel* **5**, 4229–4254 (2013).
- Phoenix, G. K. & Bjerke, J. W. Arctic browning: extreme events and trends reversing arctic greening. *Global Change Biology* **22**, 2960–2962 (2016).
- Nitze, I. & Grosse, G. Detection of landscape dynamics in the Arctic Lena Delta with temporally dense Landsat time-series stacks. *Remote Sens Environ* **181**, 27–41 (2016).
- Pattison, R. R., Jorgenson, J. C., Raynolds, M. K. & Welker, J. M. Trends in NDVI and Tundra Community Composition in the Arctic of NE Alaska Between 1984 and 2009. *Ecosystems* **18**, 707–719 (2015).
- Frost, G. V., Epstein, H. E. & Walker, D. A. Regional and landscape-scale variability of Landsat-observed vegetation dynamics in northwest Siberian tundra. *Environ Res Lett* **9** (2014).
- Emmerton, C. A. *et al.* Net ecosystem exchange of CO₂ with rapidly changing high Arctic landscapes. *Global Change Biology* **22**, 1185–1200 (2016).
- McManus, K. M. *et al.* Satellite-based evidence for shrub and graminoid tundra expansion in northern Quebec from 1986 to 2010. *Global Change Biology* **18**, 2313–2323 (2012).
- Bartsch, A., Hofler, A., Kroisleitner, C. & Trofaier, A. M. Land Cover Mapping in Northern High Latitude Permafrost Regions with SatelliteData: Achievements and Remaining Challenges. *Remote Sens-Basel* **8** (2016).
- Lara, M. J. *et al.* Polygonal tundra geomorphological change in response to warming alters future CO₂ and CH₄ flux on the Barrow Peninsula. *Global Change Biology* **21**, 1634–1651 (2015).
- Jorgenson, T. M. & Grunblatt, J. Landscape-Level Ecological Mapping of Northern Alaska and Field Site Photography (2013).
- Jorgenson, M. T. & Heiner, M. Ecosystems of northern Alaska. 1:2.5 million-scale map produced by ABR, Inc., Fairbanks, AK and the Nature Conservancy, Anchorage, AK. (2003).
- Lara, M. J. *et al.* Estimated change in tundra ecosystem function near Barrow, Alaska between 1972 and 2010. *Environ Res Lett* **7** (2012).
- USGS. A Watershed and Stream Hydrography Enhanced Dataset Project, Alaska Watersheds -5th Level, compiled by the Conservation Biology Institute. (2006).
- Walker, D. A. *et al.* The Circumpolar Arctic vegetation map. *J Veg Sci* **16**, 267–282 (2005).
- Kanevskiy, M. *et al.* Ground ice in the upper permafrost of the Beaufort Sea coast of Alaska. *Cold Reg Sci Technol* **85**, 56–70 (2013).
- Lara, M. J., Nitze, I., Grosse, G. & McGuire, A. D. Tundra landform and vegetation productivity trend maps for the Arctic Coastal Plain of northern Alaska. *Scientific Data*, (in press).
- Elith, J., Leathwick, J. R. & Hastie, T. A working guide to boosted regression trees. *J Anim Ecol* **77**, 802–813 (2008).
- SNAP. *Scenarios Network for Alaska and Arctic Planning*, University of Alaska. (2017).
- Lara, M. J., Johnson, D. R., Andresen, C., Hollister, R. D. & Tweedie, C. E. Peak season carbon exchange shifts from a sink to a source following 50+ years of herbivore exclusion in an Arctic tundra ecosystem. *J Ecol* **105**, 122–131 (2017).
- Walsh, J. E., Chapman, W. L., Romanovsky, V., Christensen, J. H. & Stendel, M. Global Climate Model Performance over Alaska and Greenland. *J Climate* **21**, 6156–6174 (2008).
- ACIA. Arctic climate impact assessment scientific report. *Cambridge University Press, Cambridge, UK*. (2005).
- Bieniek, P. A. *et al.* Climate Drivers Linked to Changing Seasonality of Alaska Coastal Tundra Vegetation Productivity. *Earth Interact* **19** (2015).
- Bokhorst, S. F., Bjerke, J. W., Tommervik, H., Callaghan, T. V. & Phoenix, G. K. Winter warming events damage sub-Arctic vegetation: consistent evidence from an experimental manipulation and a natural event. *J Ecol* **97**, 1408–1415 (2009).
- Bokhorst, S., Tommervik, H., Callaghan, T. V., Phoenix, G. K. & Bjerke, J. W. Vegetation recovery following extreme winter warming events in the sub-Arctic estimated using NDVI from remote sensing and handheld passive proximal sensors. *Environ Exp Bot* **81**, 18–25 (2012).
- Andresen, C. G. & Loughheed, V. L. Disappearing Arctic tundra ponds: Fine-scale analysis of surface hydrology in drained thaw lake basins over a 65-year period (1948–2013). *Journal of Geophysical Research-Biogeosciences* **120**, 466–479 (2015).
- Villareal, S. *et al.* Tundra vegetation change near Barrow, Alaska (1972–2010). *Environ Res Lett* **7** (2012).
- Loughheed, V. L., Butler, M. G., McEwen, D. C. & Hobbie, J. E. Changes in Tundra Pond Limnology: Re-sampling Alaskan Ponds After 40 Years. *Ambio* **40**, 589–599 (2011).
- Reyes, F. R. & Loughheed, V. L. Rapid nutrient release from permafrost thaw in arctic aquatic ecosystems. *Arct Antarct Alp Res* **47**, 35–48 (2015).
- Jorgenson, M. T., Shur, Y. L. & Pullman, E. R. Abrupt increase in permafrost degradation in Arctic Alaska. *Geophysical Research Letters* **33** (2006).

40. Batzli, G. O., Pitelka, F. A. & Cameron, G. N. Habitat Use by Lemmings near Barrow, Alaska. *Holarctic Ecol* **6**, 255–262 (1983).
41. Zhu, Z. C. *et al.* Greening of the Earth and its drivers. *Nat Clim Change* **6**, 791 (2016).
42. Los, S. O. Analysis of trends in fused AVHRR and MODIS NDVI data for 1982–2006: Indication for a CO₂ fertilization effect in global vegetation. *Glob. Biogeochem. Cycle* **27**, 318–330 (2013).
43. Forbes, B. C., Fauria, M. M. & Zetterberg, P. Russian Arctic warming and ‘greening’ are closely tracked by tundra shrub willows. *Global Change Biology* **16**, 1542–1554 (2010).
44. de Jong, R., de Bruin, S., de Wit, A., Schaepman, M. E. & Dent, D. L. Analysis of monotonic greening and browning trends from global NDVI time-series. *Remote Sens Environ* **115**, 692–702 (2011).
45. Olofsson, J., Tommervik, H. & Callaghan, T. V. Vole and lemming activity observed from space. *Nat Clim Change* **2**, 880–883 (2012).
46. Reynolds, M. K. *et al.* Cumulative geoeological effects of 62 years of infrastructure and climate change in ice-rich permafrost landscapes, Prudhoe Bay Oilfield, Alaska. *Global Change Biology* **20**, 1211–1224 (2014).
47. Epstein, H. E. *et al.* Dynamics of aboveground phytomass of the circumpolar Arctic tundra during the past three decades. *Environ Res Lett* **7** (2012).
48. Koven, C. D. *et al.* A simplified, data-constrained approach to estimate the permafrost carbon-climate feedback. *Philosophical Transactions of the Royal Society a-Mathematical Physical and Engineering Sciences* **373** (2015).
49. Abbott, B. W. *et al.* Biomass offsets little or none of permafrost carbon release from soils, streams, and wildfire: an expert assessment. *Environ Res Lett* **11** (2016).

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Author Contributions

M.J.L. designed the study, analyzed the data, developed the polygonal tundra map, and wrote the manuscript. P.M. was pivotal in the conceptualization of the polygonal tundra map. I.N. and G.G. developed the Landsat time series dataset. A.D.M. assisted in model forecasting. All authors reviewed the manuscript and made significant contributions to the writing.

Additional Information

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