



Climate variability and cultural eutrophication at Walden Pond (Massachusetts, USA) during the last 1800 years

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Abstract

Recent shifts in the ecological condition of Walden Pond, MA, are of potentially wide interest due to its cultural, historical, and recreational resource in addition to its scientific value as an indicator of climate change. Algal microfossils in six sediment cores document changes in hydroclimate and trophic status over the last 1800 years and extend two previous sediment core records of shorter length. Low percentages of *Cyclotella* in cores (WAL-3, WAL-15) indicate shallowing and/or greater water clarity associated with a relative Climate Anomaly, ca. A.D. 1150–1300. Cultural eutrophication of the lake since the A.D. 1920s is indicated by increases in *Asterionella* and *Synedra* to increase in relative abundance at the expense of *Cyclotella*, *Discosphaera*, and *Mallomonas allorgei*. Percentages of *Asterionella* and *Synedra* have remained fairly stable since the sediment core study was conducted, but scaled chrysophytes have become more numerous. To date, no mitigation efforts have curtailed anthropogenic nutrient inputs to Walden Pond, the lake has not been described by Henry David Thoreau and may become increasingly vulnerable to further changes in a possibly wetter future.

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Data Availability: The data underlying this study have been uploaded to NOAA and are accessible at <https://www.ncdc.noaa.gov/paleo-search/study/23310>.

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Introduction

Best known as an inspiration for Henry David Thoreau's *Walden: or Life in the Woods* [1] during (Concord, MA) is also a heavily used recreational venue for hundreds of thousands of bathers, cabin site annually [2]. During the early 20th century, water clarity declined significantly due to a shoreline development and inputs of human wastes [2],[3]. Poisoning of the lake in 1968 with the repeated stocking with non-native sport fish, and regional climatic warming have also influenced the lake [2–4]. In recent decades, closure of a nearby town dump, shoreline stabilization, and improved management of Walden Pond State Reservation have mitigated some of the changes, but relatively few data of water clarity, lake levels, or other limnological parameters are available with which to assess the lake's health. Sediment cores, which can provide long, continuous records of phytoplankton community structure and environmental conditions, are therefore an important source of long-term perspectives on environmental conditions. Our analyses focus on the remains of diatoms and chrysophytes, which are microscopic plant-like organisms with resistant shells (known as "frustules" or "valves"), cysts, or scales that are often well preserved in sediments. They possess diagnostic features that allow the identification of species which, in turn, aids in the development of a lake's geochemistry, water clarity, and other features of the lake based upon the known habitat preferences of the organisms.

Two previous studies of sediment cores from the deepest basin in Walden Pond yielded 600-year-old pollen, microalgae, and geochemistry [7]. Both contained evidence of cultural eutrophication since the 1800s, indicated by shifts in the carbon and nitrogen isotope composition of the sediments and by the presence of the genera *Asterionella* and *Synedra* that are often associated with enhanced phytoplankton productivity.

In this paper, we present diatom, chrysophyte, and geochemical records from six additional cores to assess the status and phytoplankton community of Walden Pond since A.D. 2000, when the last coring was completed. We use diatom records of our two longest cores to derive for the first time a qualitative reconstruction of diatom community variability in the watershed, which provides new insights into the climate history of the northeast where relatively few such records are available.

Study site

Walden Pond (42°26.3'N, 71°20.4'W; Fig 1) occupies a glacial kettle depression that formed in the granite bedrock after the retreat of the Laurentide ice sheet [8]. It contains three basins of ca. 3–10 m depth depending upon the water level, which can vary over several meters between years, and it is the largest natural pond in Massachusetts (Fig 1) [3]. Walden is a flow-through seepage lake with no surface inlet or outlet; it receives about half of its volume from westward-flowing groundwater and the rest from direct precipitation. We previously showed that the hypolimnion (lower layer of the water column) was moderately oxygenated in August during a recent stratification period [3]. Secchi depth, a measure of water transparency, was reportedly within the 2–3 m range in the 19th century [1] and early 20th century [5], but more often within the lower end of the 2.5–9 m range in the 21st century. Values are low (65–93 $\mu\text{S}/\text{cm}$) [7] and pH varies from ca. 6.5 during the spring mixing period to ca. 8.5 in summer. The higher pH values have been attributed in part to productivity by benthic meadows of *Nitella*, that cover much of the lake bed between 6 and 13 m depths with a biomass 17 times that of the phytoplankton.

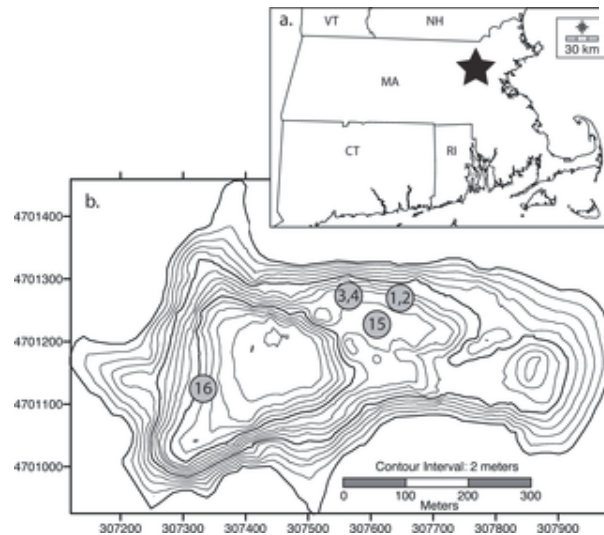


Fig 1. Location maps.

(a) Southern New England, with the location of Walden Pond indicated by a star. (b) Walden sites indicated. Numbers correspond to the numbers of the "WAL" cores discussed in the text. maps.com/carte.php?num_car=7175&lang=en under a CC BY license, with permission from 2007.

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Concord was settled by British colonists in A.D. 1635, and the forests surrounding Walden Pond extent over the next three centuries, during which time oak (*Quercus*) became less common in relative to birch (*Betula*) and pine (*Pinus*) [7]. A sequence of more recent human activities in the area is documented by Maynard [2] and summarized here, includes the building of a temporary shanty in 1843 for workers during construction of the Fitchburg Railroad, which lies close to the western shore since 1844. Sparks from steam engines frequently caused forest fires near the lake during the 1840s. Large fires in 1894 and 1896. Wood-cutting operations left hill-slopes along the eastern and northwestern shore in 1840–1841, 1851–1852, and 1918, and high water levels killed many lake-side trees during the 1860s. Logging operations on the western shore during the 1860s through the 1890s, but with increased usage of the lake increased further, underbrush and trees were cleared for walking along the western shore more frequent.

After Walden Pond State Reservation was established in 1922, beach and bath-house facilities were built along the shore. By the early 1930s, hundreds of thousands of swimmers used the facility in summer, and a site beside a cove on the northwestern shore caused large amounts of soil to wash into the lake and enlargement of the beach by county commissioners in 1957. More than half of the summer beach now be attributable to urine released by swimmers [3]. Stocking of the lake with game fish has been frequently stocked with rainbow trout (*Oncorhynchus mykiss*) and brown trout (*Salmo trutta*) treated with rotenone in 1968. The Reservation has been managed since 1975 by the Massachusetts Department of Conservation and Recreation, during which time shoreline stabilization and restoration efforts have reduced sediment input. Nearby Concord landfill during the early 1990s reduced the prevalence of waste-deposition by closing the site.

Materials and methods

Field permits for sampling in Walden Pond were provided by the Massachusetts Department of Conservation and Recreation.

June, 2015, two cores (WAL-1, WAL-4; 25 cm lengths) were collected with a UWITECH gravity basin, where two gravity-driven piston cores (WAL-2, WAL-3; 114 cm and 84 cm lengths, respectively) were collected with a modified Kullenberg corer (Fig 1). Previous studies had focused on the deep western basin, but were selected for this investigation in order to provide supporting information from that understudied location; for this choice was the likelihood that benthic diatoms associated with shallower microhabitats were present in that location, thereby providing more robust time series of the ratios of planktonic to benthic diatom levels and hydroclimate [9–12].

As photosynthetic algae, benthic diatoms require sunlight, and their distribution on a lake bed is controlled by factors including water depth, clarity, and the proximity of littoral habitats [13]. Higher (lower) percentages of benthic taxa in a core, for example, can reflect increased light penetration to the bottom due to the low levels of dissolved organic carbon, a lateral shift of littoral habitats closer to the coring site, or a decrease in planktonic productivity, all of which are likely to occur during prolonged droughts [9–14]. Variations in the percentages of planktonic diatoms (referred to here as %PLANK) within the longer WAL-3 and WAL-15 cores are indicative of qualitative changes in effective moisture, with decreases of %PLANK representing generally drier conditions that favor the growth of benthic taxa through changes in light penetration to the bottom. This interpretation is supported by significant relationships between modern-day water depth and %PLANK in the surface sediments of other North American lakes [9–12].

Care was taken to avoid sites where *Nitella* mats interfered with penetration of the core barrels. A preliminary core was rejected in the field due to partial clogging with *Nitella*. Twelve samples were collected from the central basin at 2 to 19 m depths, and a second field crew collected a gravity core to 19 m depth near the middle of the central basin (Fig 1). In 2016, a gravity core (WAL-16; 31 cm length) was collected in the deep western basin in order to better document recent changes in the phytoplankton community in different sectors of the lake (Fig 1). All but one of the cores were extruded vertically in the field, sectioned into 10 cm increments, and stored in plastic sample bags under refrigeration. For WAL-15, the mud-water interface was marked for transport, and the core was split longitudinally for subsampling.

Radiocarbon ages of oven-dried organic sediments and sieved pollen fractions were obtained for WAL-2, WAL-3, and WAL-15 (Table 1), and calibrations were determined with CALIB version 7.1. ^{137}Cs and ^{210}Pb analyses on dried, powdered sediments were also performed by staff at the SCSU Geochronology Station (MN, USA).

Depth (cm)	^{14}C age (yr BP)	Calibrated Year A.D. 2-sigma (prob.)	Sample #
WAL-2			
27.5	1990 ± 30	x	134219
36.5	485 ± 30	x	134218
56.5	380 ± 2	x	134217
75.5	485 ± 35	x	134216
100.5	1330 ± 15	x	134215
111.5	1530 ± 20	x	131316
WAL-3			
40.5	345 ± 15	1472–1527 (0.40) 1554–1633 (0.60)	134223
51.5	490 ± 15	1416–1441	134222
69.5	1080 ± 15	899–923 (0.23) 946–1013 (0.77)	134221
78.5	1360 ± 20	645–678	134220
WAL-15			
35.5	535 ± 25	1322–1347 (0.19) 1392–1435 (0.81)	142334
47.5	980 ± 15	1017–1046 (0.72) 1090–1122 (0.24) 1139–1148 (0.04)	142335
57.5	1290 ± 20	668–726 (0.64) 738–768 (0.36)	142336
64.5	1660 ± 25	266–271 (0.01) 332–426 (0.99)	142337

<https://doi.org/10.1371/journal.pone.0191755.t001>

Table 1. Radiocarbon ages and calibrated age ranges of sediment samples from Walden Pond cores WAL-

The radiocarbon ages for WAL-2 denoted by "X" were not calibrated due to evidence of sedi
<https://doi.org/10.1371/journal.pone.0191755.t001>

An age model for core WAL-3 was constructed from the ^{137}Cs profile and the ^{210}Pb activity pro
method [16], in combination with four accelerator mass spectrometry (AMS) dates on pollen fra
radiocarbon ages on bulk organic sediment were used to develop an age model for WAL-15 (T₂
were not obtained for the shorter cores due to cost constraints, which also prohibited additional
other than WAL-3. Major changes in their microfossil stratigraphies were instead used for comp
which radiometric dates were available.

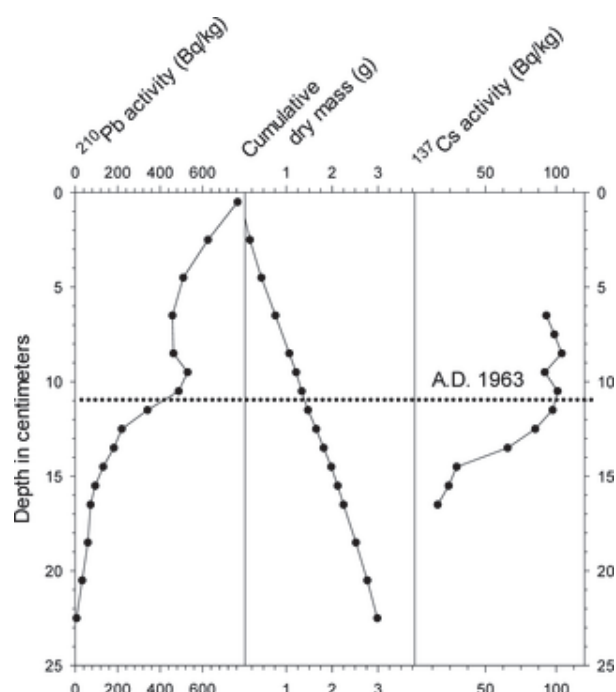


Fig 2. Lead-210 activity, cumulative dry mass, and ^{137}Cs activity profiles from core WAL-3.
Horizontal dotted line indicates the ca. A.D. 1963 interval as inferred from the ^{210}Pb and ^{137}Cs
<https://doi.org/10.1371/journal.pone.0191755.g002>

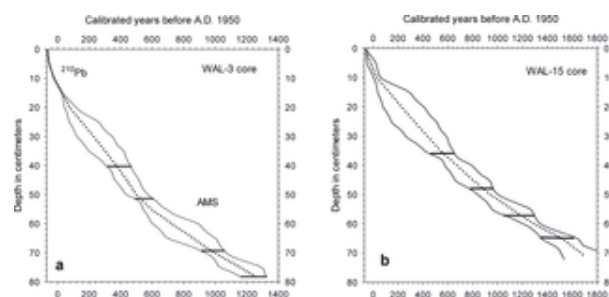


Fig 3. Age-depth relationships for Walden Pond cores WAL-3 and WAL-15.
Solid lines represent the upper and lower 95% confidence intervals of the BACON age mode

"best" model based on the weighted mean ages. Horizontal bars represent calibrated radiocarbon dates from different age scales in the two charts.

<https://doi.org/10.1371/journal.pone.0191755.g003>

Age-depth relationships were constructed using Bayesian age modeling on ^{210}Pb and AMS dat radiocarbon data alone for WAL-15 with the BACON version 2.2 modeling package in R (Fig 3) dates were input into the model as uncalibrated dates with their appropriate error ranges as est supply model. Pronounced inconsistencies in the radiocarbon ages of samples from core WAL- construction and detailed microfossil analyses for that core unfeasible.

Because of the delicate nature of the microfossils and the disaggregated nature of the sedimen extensive clumping or organic coatings was found in wet-mounted diatom samples, no chemica preparation of microscope slides for diatom and chrysophyte analyses. Subsamples were dried on glass slides with PermoutTM mounting medium. Diatom valves and chrysophyte scales anc under oil immersion, using standard references for identification [18–21]. At least 500 diatom va species level in all cores at varying increments of 1–10 cm in order to characterize the general i the last 18 centuries. The common planktonic diatom *Cyclotella bodanica* has recently been cla *bodanica*, *Handmannia bodanica*, and *Puncticulata bodanica* [22], but we refer to it here as *C. l* of consistency with previously published records [7].

Sample intervals of 1 cm were employed throughout the long cores WAL-3 and WAL-15 in orde conditions from the relative abundances of planktonic and benthic diatoms as determined from per sample. The most abundant diatom taxa were also enumerated to species level at 1 cm inc of cores WAL-1, WAL-3, WAL-4, WAL-15, and WAL-16, in order to examine the consistency of i multiple locations and thereby to better evaluate the degree to which the more rigorously dated history of Walden Pond.

Changes in sediment organic content, as estimated by percent weight loss on ignition (%LOI) a 1 cm increments in all of the cores except for WAL-15, which was analyzed for total carbon usir elemental analyzer. Carbon values for WAL-15 were converted to organic matter (%LOI) estimat assuming that elemental carbon comprises 44% of total organic matter [24].

Results

Radiometric analyses

Activity of ^{210}Pb in WAL-3 declined exponentially with depth with the exception of the 9–11 cm i disturbance of the sediments and/or a shift in deposition rates during the A.D. 1960s to A.D. 19 reached in the ca. 22–23 cm depth interval (Fig 2). Activity of ^{137}Cs in WAL-3 displayed a broad that was taken to represent the ca. A.D. 1963 global peak in atmospheric thermonuclear bomb the peak might represent post-depositional migration of ^{137}Cs upward in the unconsolidated org makes the ^{137}Cs data less reliable as a tool to test the accuracy of the ^{210}Pb age model.

The age-depth profiles derived for WAL3 and WAL-15 displayed relatively little variability other 1 increments per centimeter with depth (Fig 3). The radiocarbon ages of six subsamples from WA reversals that made the core unsuitable for detailed analysis (Table 1). The lack of such reverse depth profiles for WAL-3 and WAL-15 indicated that their stratigraphies were not significantly di: cores were roughly 1500 and 1800 calibrated radiocarbon years old, respectively (Table 1, Fig 1; stable in cores WAL-3 and WAL-15, and the prior distributions established for both accumulatio

the posterior distributions [17].

Microfossil analyses

Below the 7–12 cm level in all of the cores, the most abundant diatom taxa were *Discostella stelo*, and *Tabellaria flocculosa* var. IIIP, which together comprised roughly half to two thirds of the younger sediments of all of the cores, *Asterionella formosa* and *Synedra nana* were abundant, for example, a rise in *A. formosa* to ca. 10–20% occurred above the 14 cm level, followed by a similar rise in *Synedra nana*. *Asterionella ralfsii* was rare in all cores, but represented up to 2% of the assemblage within the 10–15 cm interval of WAL-3, the percentages of *C. bodanica* reached minimal values within the 5–15 cm interval of WAL-3, then increased again above the 15 cm of the core (Fig 4). A similar pattern of recent decline and recovery in *C. bodanica* was found in WAL-4 except for WAL-4 which displayed no recovery at the core top (Fig 6).

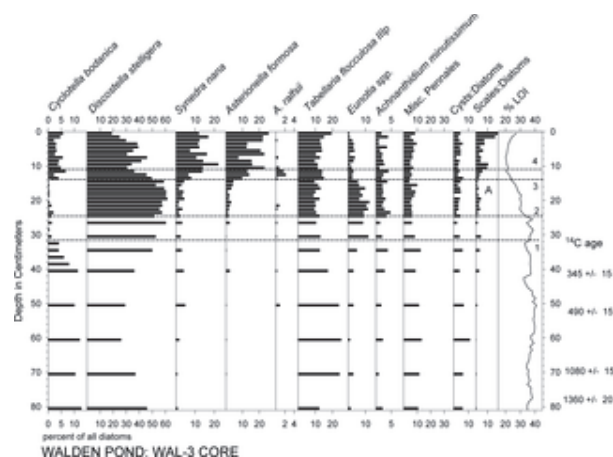


Fig 4. Percent abundances of the most common diatom taxa and %LOI in core WAL-3.

Horizontal dotted lines: (1) Onset of low *Cyclotella* percentages. (2) Decrease in %LOI indicating organic carbon content in the sediments. (3) Onset of increased *Asterionella* and *Synedra* percentages. (4) Last occurrence of *Mallomonas allorgei*. Letter "A" indicates last occurrence of *Mallomonas allorgei*.

<https://doi.org/10.1371/journal.pone.0191755.g004>

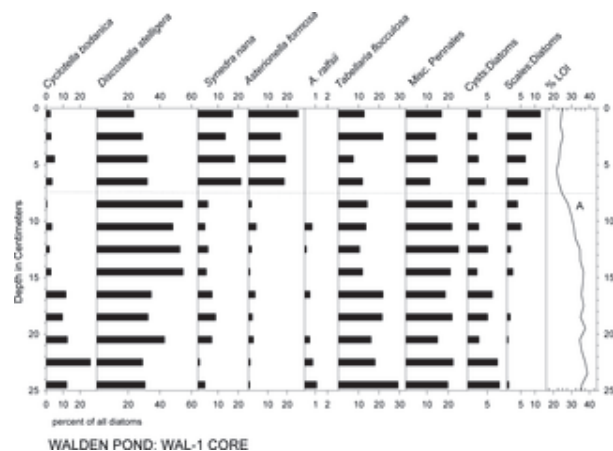


Fig 5. Percentages of the most common diatom taxa and %LOI in core WAL-1.
Horizontal dotted line indicates transition to high percentages of *Asterionella* and *Synedra*.
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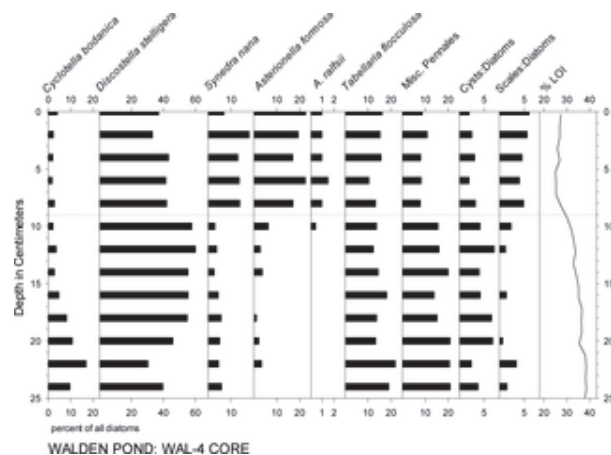


Fig 6. Percentages of the most common diatom taxa and %LOI in core WAL-4.
Horizontal dotted line indicates transition to high percentages of *Asterionella* and *Synedra*.
<https://doi.org/10.1371/journal.pone.0191755.g006>

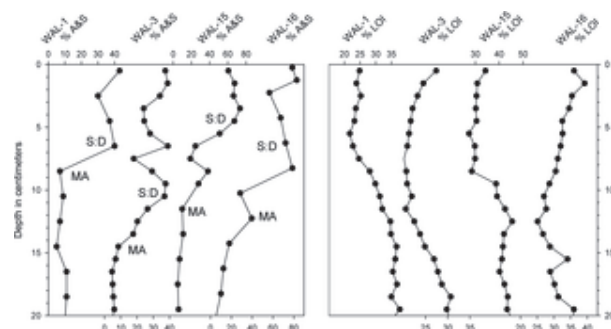


Fig 7. Comparison of microfossil and %LOI records in the uppermost 20 cm of four cores from Walden Po
Left panel: Combined percentages of *Asterionella* and *Synedra* (A&S) were interpreted as re during the last century. The most recent presence of the oligotrophic indicator species, *Mallo* a pronounced increase in the abundances of chrysophyte scales relative to diatoms are indic respectively. Right panel: Organic content in the core sediments as represented by weight lo
<https://doi.org/10.1371/journal.pone.0191755.g007>

Percentages of diatom taxa that are typically found in planktonic habitats (%PLANK) fluctuated WAL-3, with minimal %PLANK in the 52–63 cm interval and maximal values in the most recent (Figs 4 and 8). In WAL-15, %PLANK typically ranged between ca. 80% and 90% in the lower pc values within the 42–50 cm interval and maximal values above the 10 cm interval (Fig 8). Perce highest in the upper ca. 20 cm of WAL-1 (Fig 5), WAL-3 (Fig 4), WAL-4 (Fig 6), and WAL-15 (Fi

upper half of deep-water core WAL-16. Percentages of planktonic taxa also increased significantly in the upper half of the sediment transect ($r^2 = 0.86$, $P < 0.0001$).

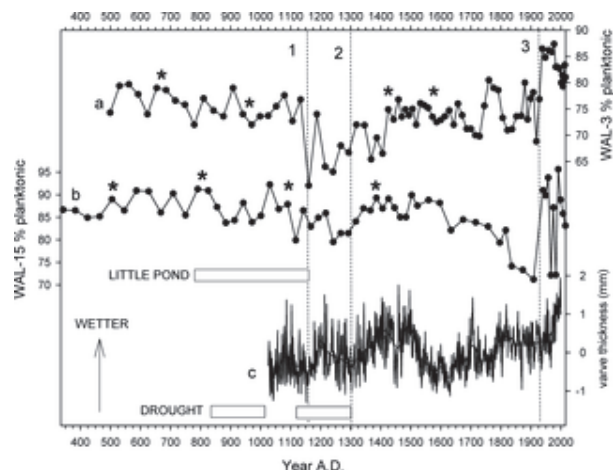


Fig 8. Comparison of sediment core records representing hydroclimate at Walden Pond and elsewhere in a, b. Percent planktonic diatoms in Walden Pond cores WAL-3 and WAL-15, respectively. Data represent increased cultural eutrophication rather than climate alone. c. Composite hydroclimate sediments of a lake and estuary in New York and Massachusetts, respectively [38]. Horizontal bars indicate Little Pond, MA (upper)[36] and the New England region (lower)[34]. Vertical dotted lines 1, 2, and 3 indicate specific intervals, line #3 indicates approximate onset of cultural eutrophication. Asterisks indicate core radiocarbon ages.

<https://doi.org/10.1371/journal.pone.0191755.g008>

In all of the cores, abundances of chrysophyte scales relative to diatoms (S:D) were highest within the upper 4–6 cm. Common chrysophyte taxa in those recent assemblages included *M. crassiusquama*, *M. elaeagnifolia*, and *M. allorgei*. In WAL-3, WAL-15, and WAL-16, S:D values approximately doubled again within the uppermost 4 cm. Scales of *M. allorgei* were common in the older sediments but that were deposited after the rise in *Asterionella* and *Synedra* (7–12 cm range; Figs 3 and 7).

Geochemical analyses

Profiles of organic matter as represented by %LOI revealed a decline of varying magnitude and timing in the upper sediments of all cores (Figs 4–7). In WAL-3, %LOI values declined above the 24 cm level before the top (Fig 4). In most of the other cores, %LOI also declined and then increased within the uppermost 4 cm. Core WAL-15, from a deeper part of the central basin, had lower %LOI values than the cores from shallower sites that were situated closer to shore (WAL-1, WAL-3, and WAL-16). Values were found in core WAL-16 from the deep western basin (Figs 1 and 7).

Discussion

Consistency among the microfossil records

%LOI In general, the diatom stratigraphies of our cores were similar to those reported by Winkler et al. [34]. We confirm that major changes in the phytoplankton community occurred during the 20th century. At Walden Pond we have recorded a pronounced increase in the prevalence of *A. formosa* and *S. na*

large, sustained increase in the abundance of chrysophyte scales during the late 20th century.

The WAL-15 diatom record appears to indicate a slightly later increase in the eutrophication-indicator *Synedra* in the core analyzed by Köster et al. [7] (Fig 9). However, its age-depth model lacks support from radiocarbon profiles that were available for WAL-3, and we therefore interpret the timing of the recent microphyte increase with greater caution.

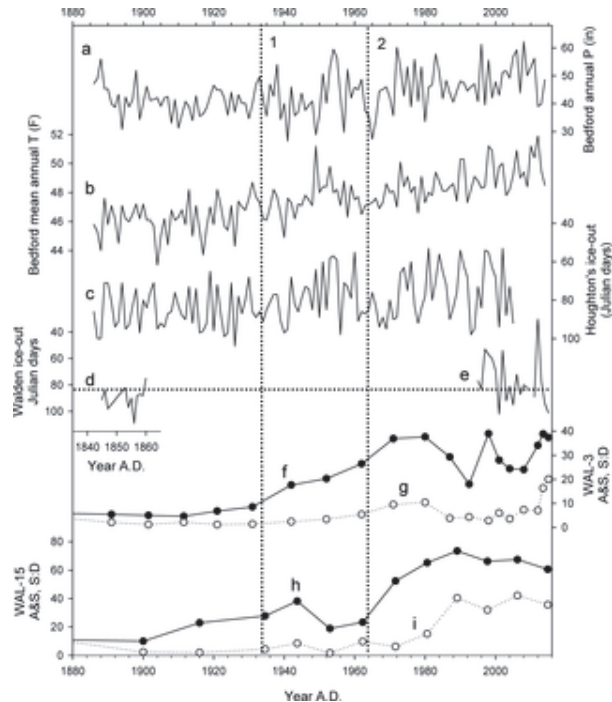


Fig 9. Comparison of temperature and precipitation in eastern Massachusetts since A.D. 1880 with ice-out dates at Walden Pond.

(a, b) Precipitation and temperature, respectively, at Bedford, MA. (c) Ice-out dates at Houghton's Pond recorded by Thoreau [1],[4],[48] and Primack [4], respectively. (d, e) Ice-out dates at Walden Pond recorded by Thoreau [1],[4],[48] and Primack [4], respectively. (f, g) *Synedra* (A&S) in cores WAL-3 and WAL-15, respectively. (h, i) Ratios of chrysophyte scales to total diatoms in cores WAL-3 and WAL-15, respectively. Horizontal dotted line indicates mean ice-out date at Walden Pond. Vertical dotted lines (1,2) indicate approximate onset of main increase in A&S percentages in cores WAL-3 and WAL-15, respectively. <https://doi.org/10.1371/journal.pone.0191755.g009>

The relative abundances of *C. bodanica* declined to near zero across the A&S transition in all cores [7], but the timing of the onset of the decline was not consistent. Köster et al. [7] dated the change in the WAL-3 and WAL-15 records it apparently occurred during the 17th and 18th centuries, respectively. *C. bodanica* rose again within the A.D. 1960s intervals of the Köster et al. [7] core, WAL-3, and WAL-4. In addition, %PLANK values declined within the A.D. 1600–1900 interval of the WAL-15 core. The reasons for these stratigraphic inconsistencies are unclear, but they show that single cores cannot provide a precise history of an entire lake with perfect precision.

Although the age-depth relationships reported by Köster et al. [7] are consistent with those in our study, the age of 530 $\pm$ 70 yr BP obtained by Winkler [6] for a sediment sample from 75–81 cm depth in Walden Pond is younger than those determined for corresponding intervals in WAL-3 and WAL-15 (Table 1, Fig 10).

age model for the microfossil records reported by Winkler [6] might therefore require re-evaluati

The lack of rising values of %LOI and % *C. bodanica* along with the presence of rare but distinct top of core WAL-4 suggest that the upper section of its sediment record was missing and/or phy with problematic radiometric age discrepancies in core WAL-2, the causes of such stratigraphic with certainty but could include disturbances by boat anchors, contact with the bottom by free-d turtles which are common in the pond.

The upper sediments of WAL-15 (0–18 cm) and WAL-16 (0–14 cm) were dark brown to black, r brown sediments below, but there was little to no apparent color variability in the other cores co darkening is likely due to higher concentrations of dissolved oxygen at their shallower collection depth limit of summer hypoxia in the lake and the lower depth limit of oxygen-producing *Nitella* |

Hydroclimate history

Climatic interpretation of %PLANK in sediment cores can be problematic where anthropogenic also influence diatom communities [7],[12]. However, historical and paleolimnological records in on Walden Pond before the late 19th century [1–7]. We therefore interpret %PLANK in the older a qualitative indicator of hydroclimate variability before the late 19th century.

We focus here on an interval of reduced %PLANK within the WAL-3 and WAL-15 cores (Fig 8) reduced effective moisture ca. A.D. 1150–1300 during the Medieval Climate Anomaly (MCA; A. record also includes a secondary decline that might extend the period to ca. A.D. 1420, but it is WAL-3 was collected from a site that was both closer to shore and shallower than that of WAL- pre-19th century sediments contain higher percentages of benthic diatoms than WAL-15 and gr 20–40% and 8–20%, respectively). Köster et al. [7] did not provide data on benthic taxa in their compared to our own records, but the relatively low percentages of benthic taxa in WAL-16 (ca. assemblages in the deeper sectors of the lake may be less sensitive indicators of hydroclimate coring sites in the shallower central basin.

Other paleoecological records have documented aridity in the northeastern United States during of its timing and duration vary. Drought conditions were inferred from pollen and charcoal data f around A.D. 800–1300 [29], and Sidney Bog, ME, registered relatively dry conditions ca. A.D. 1 A.D. 900–1400 have also been inferred from records in northwest Ontario [31], the Great Lakes United States [33],[34].

In contrast, lake level records from Deep, Davis, and New Long Ponds, MA, revealed no evide instead placed a shorter period of low lake levels earlier, ca. A.D. 650–750 [35]. At Little Pond, I from ca. A.D. 750–1150 as indicated by increased abundances of spores of the benthic plant *Is* content (Fig 8) [36]. A high-resolution diatom record from Wolf Lake, NY, registered a low stand 1150–1300 [12]. Dendroclimatological records registered severe droughts throughout much of 1 A.D. 1122–1299 (Fig 8) [37] but relatively little change in New England then, a conclusion that ti support. A composite time series based on varved sediment records from Rhode Island and Ne wetter conditions during the A.D. 1100–1300 interval (Fig 8)[38]. That increase in precipitation r not strongly registered in the New York record and could presumably represent a maritime micr rather than the region as a whole. The Walden core data lack sufficient temporal resolution to ri decadal-scale variability during that period, but the WAL-3 diatom record more closely resemble than WAL-15 (Fig 8).

At present, the limitations of radiocarbon dating and a need for more high-resolution paleolimnc

make it difficult to determine whether these differences represent true regional variability or not. Precipitation is spatially variable, particularly in complex terrain such as that which characterizes lakes of different forms and hydrological settings can respond differently to it, so more recognition of hydroclimate variability in this region during the late Holocene can be more fully characterized.

The large increase in %PLANK that occurred in cores WAL-3 and WAL-15 within the uppermost result of cultural eutrophication than climate [3],[7], and is discussed below.

Recent changes in the phytoplankton community

Phytoplankton productivity in Walden Pond has increased notably since the A.D. 1920s, following shoreline [2],[3],[5],[7]. Other recent nutrient inputs included erosion of sediment from the lakeshore also contributed to declines in %LOI within the cores through dilution with inorganic mineral grains such as stabilization of the shoreline and trails and improvement of public sanitary facilities have reduced nutrient inputs to the lake, but their more recent effects on the lake remain to be evaluated with data provided by sediment cores.

The previous core studies at Walden Pond documented an abrupt decline in the abundance of diatoms in the 20th century, probably due to shading from enhanced planktonic productivity [6],[7]. Köster et al [7] also starting around A.D. 1930 in which *C. bodanica* and *D. stelligera* were replaced by abundant *A. fuscum*, commonly associated with cultural eutrophication in North American lakes [39–43]. That transition in abundances of chrysophytes during the late 20th century.

The diatom and chrysophyte records of our cores corroborate those previous studies and extend them. In WAL-1, WAL-3, WAL-15, and WAL-16, scales of the chrysophyte *M. allorgei* were present throughout the cores but were not found above the A&S transition (Fig 7). This species is typically found in oligotrophic lakes and its disappearance from the sediment assemblages probably reflects the eutrophication that is so apparent in the records.

We further examine the microfossil records of Walden Pond here in qualitative terms in order to assess the condition of the lake in recent times: What caused the recent changes in the phytoplankton community returning to its pre-impact condition or developing a new "no-analog" state?

Climatic warming can affect plankton communities by enhancing thermal stratification which, in turn, reduces sediment-bound nutrients through hypoxia at the bottom of the lake [45]. Increased precipitation through erosion of nearby soils and lake-shore deposits [12],[46]. However, the rise in A&S during the last century by several decades the notable increases in temperature and precipitation that occurred in Massachusetts (Fig 9). No significant changes are evident in the duration of lake ice cover, an indirect indicator of winter severity from Houghton's Pond, MA [47], or Walden Pond [4],[48] that could also help to explain change in the lake. We therefore concur with previous authors that recreational use of the lake was the primary trigger for the changes in Pond [2],[6],[7], although climatic changes could also be contributing to plankton productivity in the lake.

The darkening of sediments in the upper sections of cores WAL-15 and WAL-16 probably reflects anoxic conditions in the hypolimnion within the last century or so because iron compounds in sediment environments are less likely to produce the pale, rusty coloration that is typical of oxidized ferric iron. Darker sediments reflect longer stratification seasons and/or increased phytoplankton productivity. The WAL-15 anoxic boundary corresponds to the mid-18th century, but in WAL-16 the change occurred only slightly later in the early 20th century. The lower boundaries of the darkened zones in those cores could reflect different conditions in addition to deposition under hypoxia so they do not necessarily indicate precise timing. Nonetheless, they do suggest that cultural eutrophication has likely caused cause dissolved oxygen depletion.

hypolimnion to decline.

Comparison of A&S percentages in four cores (Fig 7) reveals no consistent, significant trends. We interpret this to mean that cultural eutrophication in Walden Pond has more or less stabilized since the 1980s. The cores also indicate that the phytoplankton community and dissolved oxygen conditions have improved. It is likely that Walden Pond remains in a state of enhanced productivity due to a combination of direct anthropogenic nutrient enrichment [3], could also include atmospheric nitrogen deposition [4], [50–53] and internal loading of nutrients from bottom sediments [43] in the deepest portions of the lake. *Nitella* meadows to oxygenate the surrounding water.

The recent increase in scaled chrysophytes in Walden Pond resembles similar increases in other lakes that have been attributed to the competitive advantages that longer summer stratification seasons give to them [53], [54]. However, its causes at Walden Pond are difficult to ascertain. Much of the increase in American lakes has been among colonial taxa such as *Synura* [55], [56] rather than the unicellular taxa common in Walden Pond, and multiple local anthropogenic impacts on the lake during the last century make it difficult to identify the cause.

Although the reasons for the recent rise in abundance of chrysophytes remain unclear, the long-term change is unprecedented during the last 1800 years at Walden Pond. It is not readily related to temperature, or lake ice-cover (Fig 9), and therefore seems likely to be more the result of increasing nutrient cascades than climatic factors. Regardless, the recent large increase in scaled chrysophytes, and the decline of *Isoetes* and *M. allorgei*, also means that Walden Pond has not been restored to the conditions of many other lakes around the world [54], Walden Pond has instead moved to a new ecological state. The next 50 years if not the entire post-glacial history of the lake.

Increasing %LOI within the uppermost intervals of WAL-3, WAL-15, and WAL-16 since the 1980s may reflect increased inorganic sediment inputs from stabilization of shorelines and foot trails, but it could also reflect declining oxygen concentrations in the hypolimnion. The lack of such an increase of %LOI at the bottom of the cores likely reflects the absence of the youngest sediments from that core.

Lake management in a changing future

The widespread occurrence of low lake levels during the MCA has implications for the interpretation of the data that anticipate rising temperatures and heavier precipitation in New England during this century. The warming has become evident in Massachusetts in recent decades (Fig 9). The warm MCA was not caused by increased precipitation as much of the warming of the last half-century was [58], but as an imperfect historical record, lake levels might not rise significantly in a warmer future even if total annual precipitation increases. Reduced precipitation, for example, could reduce spring and summer groundwater supplies, and higher temperatures could reduce effective moisture.

Paleolimnological evidence such as that from Crawford Lake, Ontario [43], [59], demonstrates that cultural eutrophication can persist for centuries after the causal nutrient inputs have ceased, and that it is easier to maintain eutrophication than to reverse it once it occurs. The sediment darkening and high percentages of organic matter in Walden Pond therefore indicate not only that the lake ecosystem is now quite different from that of the 19th century, but also that it may be primed for more severe reductions in water clarity in a warming future. Increased warming in the 21st century [57], for example, are likely to encourage heavier summer recreational use of the lake, and increased summer stratification season through warming of the lake surface could also further enhance internal nutrient loading.

Major reductions in the extent of *Nitella* meadows due to decreasing water transparency under a warming future could represent an ecological tipping point that could severely and permanently reduce the clarity of the lake and d

However, the taxonomic composition and ecology of the aquatic macrophyte communities of W. characterized. Although the benthic meadows have been described as being composed primarily of *Fontinalis* moss, not *Nitella*, being dredged from 15.7 m depths, and the decline of *Isoetes* sporophytes in the 20th century [6],[7] suggests that *Nitella* might not have been as prevalent in the past as it is now. Changes in vegetation to the oxygen and nutrient dynamics in Walden Pond, such information could be useful for lake management. For example, different species of *Nitella* and other benthic macrophytes can have different effects on the lake. So changes in species composition within the benthic meadows could serve as biotic indicators of lake health.

As Jeppeson, et al. [63] noted in a review of lake management practices in a changing climate, from precipitation mean that nutrient inputs to many lakes must be reduced if they are to maintain water quality as they are in today. It will therefore be prudent to further reduce the flow of anthropogenic nutrients to lakes. Under wetter conditions that most climate models project for New England during the 21st century [57] the source of such nutrients now, and demand for the beach facilities is likely to increase in a warm climate. Lake management programs or construction of a separate swimming pool facility nearby to relieve pressure on the lake.

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