

1 **Rare Earth Element uptake during olivine/water hydrothermal interaction**

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Abstract

Ultramafic-hosted hydrothermal vent systems link the hydrosphere with the peridotitic mantle via serpentinization. Here, the fractionation and mobility of selected trace (Nd, Sm, Gd, Dy, Yb, Sr and Ba) and major cations (Si, Ca, Mg, Fe, Ni) during seawater-peridotite interaction was investigated through a series of experiments where natural olivine grains were reacted with an artificial seawater solution at a range of temperatures (15-90 °C) and grain size distributions. No evidence for any significant olivine dissolution or precipitation of carbonate and, Fe-oxyhydroxide phases was observed in these experiments. Experimental data show a strong decoupling of REE (Nd, Sm, Gd, Dy, and Yb) from Sr and Ba under all experimental conditions, with Sr and Ba remaining quantitatively in solution. The REE were removed from the solution and were adsorbed onto olivine surface with kinetic rate constants (i.e. uptake over time) that increase with increasing temperature and increasing surface area (i.e. decreasing particle size). Dysprosium and Yb (heavy REE; HREE) were removed from solution with a faster rate than Nd and Sm (light REE; LREE). Gadolinium is decoupled from this trend, with a slower kinetic rate constant than Sm. The activation energies (E_a) of REE adsorption on olivine were higher for Nd and Sm than Dy and Yb. This suggests that the adsorbance of LREE is generally more dependent on temperature than the HREE. The E_a correlates well with the summed 1st, 2nd and 3rd ionization energies of REE suggesting a link between kinetic rates of element adsorption and electron configuration of the 4*f*-orbitals. Gadolinium has higher E_a than the other analyzed REE, consistent with the electron configuration of Gd^{3+} where all 4*f*-orbitals are filled with one electron each. These experimental data suggest that REE are adsorbed on the surface of olivine via inner sphere complexes under low-temperature hydrothermal conditions, when alteration

34 processes are limited or extremely slow. Scavenging and fractionation of REE may occur within
35 the recharge zone of peridotite-hosted hydrothermal systems at relatively low temperatures
36 ($<100\text{ }^{\circ}\text{C}$), leading to fluids with progressively higher LREE/HREE which could impose
37 seawater-derived LREE enrichments in serpentinized peridotites during high temperature, high
38 pressure water/rock interaction deeper in the oceanic lithosphere.

1.0 Introduction

In oceanic basins, at subduction zones, axial, off-axis and rift systems seawater reacts with the peridotitic portion of the oceanic lithosphere linking the hydrosphere and the Earth's mantle. Hydrothermal circulation of seawater in peridotites results in the formation of reducing hydrothermal fluids enriched in dissolved volatiles (e.g., H₂, CH₄), and with a chemical composition distinctively different from basalt-hosted hydrothermal systems (Allen and Seyfried, 2005; Alt et al., 2013; Bach et al., 2004; Delacour et al., 2008; Evans et al., 2013; Foustoukos et al., 2008). Also, seawater - peridotite reaction at elevated temperatures (>180 °C) results in alteration of primary silicate minerals (e.g., olivine, clinopyroxene, and orthopyroxene) to secondary (or alteration) minerals (e.g., chrysotile, brucite, and magnetite) (Foustoukos et al., 2008; Janecky and Seyfried, 1986; Kadko et al., 1994; Paulick et al., 2006; Seyfried et al., 2007; Snow and Dick, 1995). However, high temperature basalt- and peridotite-hosted hydrothermal systems at oceanic spreading centers are thought to account for only 25-30 % of the total magmatic heat flow from the cooling lithosphere (Alt, 2003; Bach and Früh-Green, 2010; Johnson and Pruis, 2003; Mottl and Wheat, 1994; Rosenberg et al., 1993). The remaining 70-75 % of the total magmatic heat flow is thought to originate from more diffuse, lower-temperature hydrothermal advection at ridge flanks and within oceanic basins. Due to their larger areal coverage, such low temperature hydrothermal systems potentially result in much greater elemental exchange between oceanic crust, lithosphere and seawater than their on-ridge counterparts (Alt, 2003; Elderfield and Schultz, 1996; Johnson and Pruis, 2003; Mottl and Wheat, 1994). Yet, mantle peridotite-seawater interaction at relatively low temperature conditions (<100 °C) has been little studied.

In natural aqueous environments, rare earth elements (REE) have been proven useful tracers of both processes and pathways (Alibo and Nozaki, 1999; Elderfield and Greaves, 1982; Elderfield et al., 1990; German et al., 1990; Johannesson et al., 2011; Schijf and Marshall, 2011; Shiller, 2003; Sonke and Salters, 2006). Their effectiveness as tracers lies in their similar geochemical behavior (3^+ charge, with Ce and occasionally Eu as exceptions), and their gradual decrease in ionic radius from La to Lu (i.e., lanthanide contraction). Several studies (at ambient temperatures and pressures) have demonstrated that REE are highly particle reactive, e.g., adsorption of REE with marine particulates (De Baar et al., 1985a; De Baar et al., 1985b; Elderfield et al., 1990; Erel and Morgan, 1991; Sholkovitz et al., 1994), aquifer sands (Duncan and Shaw, 2003; Tang and Johannesson, 2005), mineral surfaces (silica beads, Schijf and Marshall, 2011; clays, Coppin et al., 2002; basaltic glass, Tertre et al., 2008) and organic macromolecules (Sonke and Salters, 2006; Stern et al., 2014; Stern et al., 2007). The relative depletion of light REE (LREE) over heavy REE (HREE) in seawater compared to an average continental crust shale rock (e.g. Alibo and Nozaki, 1999; Byrne and Kim, 1990; De Baar et al., 1985b; Sholkovitz et al., 1994) is generally attributed to the stronger complexation of HREE over LREE with carbonate complexes (e.g. Cantrell and Byrne, 1987; Luo and Byrne, 2004) leaving LREE more susceptible to scavenging by surface ligands (Sholkovitz et al., 1994). In the absence of strong organic and carbonate complexes, however, some studies have shown preferential removal of HREE over the LREE onto mineral substrates (Byrne and Kim, 1990; Coppin et al., 2002; Tertre et al., 2008). Rare earth elements also show a high affinity for hydrated oxide minerals (e.g., Fe-oxyhydroxide) that increases with increasing pH (Bau, 1999; Koeppenkastrop and De Carlo, 1993; Quinn et al., 2006b; Schijf and Marshall, 2011) and

temperature (Quinn et al., 2007). Although HREE generally have higher affinity than the LREE for Fe-oxyhydroxides, the available data does not always show a simple monotonic increase in partitioning with increasing atomic number and often exhibits maximum at Sm relative to the HREE (Bau, 1999; Quinn et al., 2004).

Despite these data, the behavior of REE during peridotite-seawater interaction is less understood. Experiments on the reaction of serpentinized peridotite with fluids at temperatures greater than 300 °C show LREE enrichment in the resulting hydrothermal fluid that was attributed to a complex interplay between dissolved REE species in the Na-Mg bearing fluid and precipitation of secondary mineral phases, such as talc and chrysotile (Allen and Seyfried, 2005). Enrichments in LREE in hydrothermal fluids were further attributed to mobilization of REE from the interstitial secondary minerals of serpentinized peridotite (Bach and Irber, 1998). In contrast, field studies of Paulick et al. (2006), Frisby et al. (2016a) and Frisby et al. (2016b) suggested that serpentinization under high integrated water/rock mass ratios results in REE uptake from seawater and enrichments in altered peridotites. This was particularly exemplified by the overprint of the mantle-like Nd isotope composition of peridotites by a seawater component (Frisby et al., 2016a). The apparent REE mobilization during peridotite serpentinization (Allen and Seyfried, 2005) appears at odds with the strong adsorption of REE onto mineral surfaces, and field studies discussed above. It is unclear whether this discrepancy is driven by precipitation or dissolution of mineral species (e.g. serpentine group minerals, carbonates, etc.; Allen and Seyfried, 2005; Zhong and Mucci, 1995) or surface related reactions controlled by temperature and fluid pH conditions (Marmier et al., 1999; Ridley et al., 2005).

In order to gain further insights into the extent of elemental fractionation between seawater and natural peridotite, and especially test whether there is REE uptake or not during olivine – seawater interaction, we present time-series reaction experimental data for Sr, Ba, Nd, Sm, Gd, Dy and Yb between fresh, natural forsteritic olivine and Na-Mg-Ca-Cl-bearing aqueous solutions at vapor-saturated pressure conditions, and as a function of temperature (15, 60 and 90 °C) and reactive surface area. The starting solutions were of seawater composition for the major cations, however, without SO_4^{2-} which may act as strong ligand for REE aqueous complexation (Migdisov et al., 2016) and open to the atmosphere so bicarbonate, carbonate ions are present but at lower concentrations than typical seawater. The experiments were designed to constrain the relative fractionation and mobility of selected REE at conditions reflecting low temperature peridotite-seawater interaction at hydrothermal environments (i.e., off-axis, ridge flank settings, diffuse flow) that are thought to dominate the global oceanic hydrothermal heat flux (Bach and Früh-Green, 2010; Elderfield and Schultz, 1996; Johnson and Pruis, 2003; Kadko et al., 1995; Mottl and Wheat, 1994; Stein and Stein, 1994). Strontium and Ba were determined in addition to the REE to contrast their generally more conservative behavior relative to REE in seawater (in the absence of precipitating carbonates). The experimental data are used to derive the activation energy of the surface area-normalized kinetic constants for the REE adsorption to olivine and the implications of these data towards seawater – peridotite interaction.

2.0 Methods

2.1 Starting Material

Olivine mineral separates with typical upper mantle peridotite composition ($Mg\# = 0.91$, where $Mg\# = Mg/(Mg+Fe)$ cation mole; Table 1) were obtained from a large fresh spinel peridotite xenolith (88SAL 1-1) from Salt Lake Crater, Hawaii (e.g. Bizimis et al., 2004 and references therein). The peridotite was crushed and hand sieved to obtain three size fractions ($<200\ \mu m$, $200-350\ \mu m$ and $500-850\ \mu m$). The two larger fractions were magnetically separated with a Frantz isodynamic separator to concentrate olivine. Olivine grains were handpicked for clarity under a binocular microscope, and to be devoid of alteration or inclusions. In order to remove any surface alteration or adsorbed metals, the picked olivine grains were leached in a 10 wt% HCl solution and sonicated for ten minutes, and subsequently rinsed and sonicated several times in $18\ M\Omega\ H_2O$. The grains were dried at $90\ ^\circ C$ and were sieved again with a clean polypropylene sieve to better constrain the size fraction by separating any smaller grains that may have mechanically broken off during sonication. To obtain the “fine” fraction of olivine used in the experiments, a hand-picked olivine fraction was processed through an agate ball-mill until a $<30\ \mu m$ powder was achieved. Average grain size for the three fractions was determined using a Beckman Coulter LS 100Q laser particle size analyzer using the diffractometry laser method (Beuselinck et al., 1998). Surface area measurements are based on the assumption of sphere shape of the counted particles, and reflect the sum of particles’ surface area divided by the sum of particles’ volume (Foustoukos and Stern, 2012). Repeated analyses of the material were reproducible to better than 1% relative. The geometric surface area (GSA) was determined by accounting for olivine’s specific gravity of 3.33 (Klein and Dutrow, 2008).

Aqueous solution with seawater-like composition was prepared by dissolving NaCl (99.99%), $MgCl_2$ (99.99%), and $CaCl_2$ (99.99%) (Alfa Aesar) in $18\ M\Omega\ H_2O$ to obtain the

concentrations of dissolved Na^+ (0.42 mol/L), Mg^{2+} (0.05 mol/L), Ca^{2+} (0.01 mol/L), and Cl^- (0.54 mol/L) found in average seawater (Table 2; Wilson, 1975). We did not added sulfate for simplicity and because it is not thought to be a major REE ligand in seawater (Millero et al., 2009). An aliquot of this solution was spiked with Sr, Ba, Nd, Sm, Gd, Dy, and Yb, to a concentration of 100 ppb each while maintain the pH at ~ 7.5 .

At the water/rock mass ratios of these experiments (~ 40) the total dissolved REE content (2.5 μg) in the starting solution is about 100 times higher than the total REE content of the olivine used for this investigation (< 37 ng, Table 1 and Supplementary Table S1). Therefore olivine contributions to the REE concentrations in solution (even if all olivine was to dissolve) were relatively insignificant.

2.2 Experiment Protocols

2.2.1 Olivine Time-series experiments

All experiments were performed in a trace metal free, HEPA filtered air clean lab at the Center for Elemental Mass Spectrometry, University of South Carolina. In these experiments 125 mg of olivine grains with GSA ranging from 28.2 to 3339 cm^2/g were added in 5ml of reactant aqueous solution, resulting in a water-to-rock mass ratio (W/R) of ~ 40 (Table 3). The experiments were performed in Savillex 7ml PFA (Teflon) vials. The beakers were set on an aluminum block, capped to ensure even heat transfer. Temperatures were set at 15 $^{\circ}\text{C}$ (ambient lab temperature, controlled within 1 degree), 60 $^{\circ}\text{C}$ and 90 $^{\circ}\text{C}$. The actual solution temperatures were found constant and reproducible to better than 2 $^{\circ}\text{C}$ for the 90 $^{\circ}\text{C}$ experiments. Before adding the mineral grains to the reaction PFA vial, the solution was heated to the temperature of

the experiment. An aliquot of the solution was taken to establish the initial composition of the solution, then the olivine was added to the vial and the vial was capped. Time series samples were collected during the course of the experiment (a total of 131 analyzed aliquots) and in varied intervals: 30-minute intervals for the first hour, every 3 hours for the next 24 hours, then every 6 hours over 3 days followed by every 24 hours out to 14 days total experiment time. Sampling of the fluid consisted of opening the vials, withdrawing 50 μ l of the solution with a pre-cleaned polypropylene pipette tip, and recapping the vial. Therefore during sampling the experiments were exposed to atmosphere. The 50 μ l sample extraction was chosen to minimize change in the volume of the solution in the experiment (1% of the experiment volume per analysis).

The recovered solutions were filtered by centrifugation in a pre-cleaned micro-centrifugal filter (<0.2 μ m; Fisher Scientific part #UFC30GV0S with a low hold volume of < 5 μ l), to filter any submicron suspended material. Thirty (30) μ l of the filtered aliquots were then diluted ~80x with sub-boiling, Teflon-distilled 2% (wt/wt) HNO₃ with 2 ppb Indium added as internal standard for elemental analysis. Strontium, Ba, Nd, Sm, Gd, Dy, Yb, Ni, Si, Fe, Mg, and Ca concentrations for all starting materials and reaction fluid products were determined using a combination of external standard calibration with the USGS BIR-1 basalt, and internal drift correction using Indium on a Thermo-ELEMENT 2 HR-ICPMS following established protocols for this lab (Table 4) (Das et al., 2013; Frisby et al., 2016b; Sen et al., 2011). The isotopes ⁸⁸Sr, ¹³⁷Ba, ¹⁴⁶Nd, ¹⁴⁷Sm, ¹⁵⁷Gd, ¹⁶³Dy, and ¹⁷²Yb were analyzed in low-resolution mode and the isotopes ⁶⁰Ni, ²⁹Si, ⁵⁷Fe, ²⁵Mg, and ⁴³Ca in medium-resolution to maximize the signal to background ratio by avoiding isobaric interferences (e.g., ⁴⁰Ar¹⁶OH on ⁵⁷Fe). Detection limits for

the experimental solutions were approximately 25 ppt for Nd and Sm, 30 ppt for Gd, and 2 to 5 ppt for Dy and Yb. As the starting solution was 100 ppb for each REE we resolved the REE depletion in solution down to a 0.03% to 0.002% of the starting solution. Precision of concentrations as judged by internal statistics of the measurements is from 1 to 6% for the lowest concentrations. However, elemental ratios in each aliquot are precise to better than 3% because some uncertainties in the calculation of the absolute concentrations (e.g. drift correction, plasma noise), effectively cancel out during measurement.

The pH of the solutions was measured in-situ at each sampling interval using a micro-pH electrode (Thermo Scientific 8220BNWP) with the exception of the CF-23 experiment (15 °C) (Supplementary Table S1), where the pH of the solution was measured at termination. The analytical uncertainty of the pH measurements was within 0.05 log units (1 σ standard deviation).

Least-squares fitting of the experimental data was performed utilizing the technical graphic and data analysis software IGOR Pro (Wavemetrics) and Microsoft Excel.

2.2.2 Analysis of Solid Samples

To examine the presence of secondary alteration phases, the reactant solids of the high surface area (< 30 μm , i.e. powder) experiment HP (Table 3, Supplementary Table S1) were retrieved at the termination of the experiment, and then dried down at 90 °C overnight in an enclosed HEPA filtered dry down box. Once dry, the solids were rinsed with ~5 ml of 18 M Ω H₂O; to remove remnant chlorides. The solids were analyzed by X-ray diffraction on Shimadzu LabX XRD-6000 (at the University of South Carolina) and for Fe⁺³/Fe⁺² by Mößbauer spectrometry (Carnegie Institution of Washington: Geophysical Lab). Mößbauer spectroscopy

analysis were conducted with a ~ 40 mCi ^{57}Co flat source with the drives (Austin Science) operating in constant acceleration mode between -4 and 4 mm/s, and calibrated with metallic Fe at the operational temperature conditions of 25°C . Data were collected over 1024 channels, which on folding resulted in 512 channels. All Mößbauer spectra were accumulated to a low statistical uncertainty (typically $5 - 7 \times 10^6$ counts per channel). Samples were prepared by homogeneously mixing the starting olivine and the final reaction-products (15 mg) with transoptic powder (120 mg), and then converted to a 12.7-mm diameter and 1 mm thick pellet at 350 bar and 120°C (Foustoukos and Stern, 2012).

The solids from the high temperature (90°C) coarser olivine experiment CF-19 (Table 3, Supplementary Table S1) were inspected by SEM on a Tescan Vega 3 SBU variable pressure SEM at the Electron Microscopy Center, University of South Carolina (Supplementary Fig. SF1). The grains were compared to the unreacted fractions to test for the presence of newly formed secondary mineral phases.

2.2.3 Thermodynamic modeling

Assessment of phase equilibria during the experiments including calculation of mineral solubility and distribution of aqueous species at experimental conditions was performed using Geochemist's WorkBench (GWB) software package (Bethke, 1996). Mathematical derivations and numerical techniques used by the GWB code involve a Newton-Raphson approach to simultaneously solve the grid of non-linear equations that describe chemical equilibria and mineral saturation states, while assessing aqueous speciation through equations of mass and charge balance. This formulation of geochemical modeling is in accordance with previous

numerical algorithms developed to estimate aqueous speciation and approximate mineral-fluid equilibria at elevated temperatures and pressures (e.g. SOLVEQ, EQ3/6) (Reed, 1982; Wolery and Daveler, 1992). The activity coefficients of ionic species were calculated with an extended form of the Debye-Hückel equation (“*B-dot*”) (Helgeson et al., 1981) (additional details in (Foustoukos, 2016). To ensure internal consistency, thermodynamic data on mineral hydrolysis, dissociation reactions of aqueous species and H₂O properties were derived from SUPCRT92 (Johnson et al., 1992) and the 1998 database update (Shock et al., 1997; Sverjensky et al., 1997). The standard state for end-member minerals and liquid H₂O is unit activity of the pure phase at the temperature/pressure conditions of this study (Ding and Seyfried, 1992).

For the thermodynamic properties of chloro- and hydroxyl- bearing REE aqueous complexes we utilized the thermodynamic data of Migdisov et al. (2016) (REE-Cl²⁺, REE-Cl₂⁺, Nd(OH)_{3(aq)}) along with the dataset of Haas et al. (1995) (REE-OH²⁺, other chloro-complexes, REE³⁺). Unfortunately, the recent advances in the thermodynamic properties of REE-carbonate aqueous species (Luo and Byrne, 2004) are limited to ambient temperature/-pressure conditions. Thus, to model the aqueous REE-carbonate species in our hydrothermal experiments, use of the theoretical predictions of Haas et al. (1995) was adopted (see more in (Migdisov et al., 2016)).

3.0 Results

To better illustrate and model the change in dissolved REE concentrations during olivine/fluid interaction, we present the natural logarithm of the solution concentration, C , at any given time, normalized to the starting solution concentration, C_0 (i.e., $\ln(C/C_0)$) (Supplementary

Table S1). Results are grouped with temperature. In all figures circle symbols depict experimental data at 15 °C, squares data at 60 °C, and triangles data at 90 °C.

15 °C Experiments: Considering the very sluggish kinetics of olivine hydrolysis at such low temperatures, these experiments served as control experiments to constrain the REE olivine/fluid partitioning and possibly decouple the effects of REE mineral uptake during alteration from surface adsorption at the higher temperatures of 60 and 90 °C. Time series samples collected at 15 °C (ALG, CF-11, CF-20, CF-22, CF-23, AP) (Fig. 1, Supplementary Table S1) showed that Sr and Ba concentrations remained essentially unchanged in the solution for all size fractions. In contrast, the REE are increasingly removed from solution with increasing GSA (i.e. decreasing particle size), with the HREE being preferentially removed over the LREE. For example, as the GSA approximately doubles (from 257 to 577 cm²/g), the amount of Nd remaining in solution after 1920 minutes decreased from ~30 % to ~4 %, respectively, while the Yb/Nd_(n) in solution decreased from 0.39 to 0.35 (where n = concentration normalized to C₀) (Fig. 1, Supplementary Table S1). In two extreme cases, experiment ALG (Fig. 1A) with the largest size olivine grains (500 to 850 micron size fraction, GSA = 28.16 2 cm²/gr) showed no appreciable change in REE and Sr, Ba concentrations in solution, even after 22 days of reaction time; while in experiment AP with the highest GSA = (3339 cm²/gr) (Fig. 1F) the REE concentrations in solution decreased to 1-2 % of the starting solution concentration within only 6 hours (Fig. 1F, Supplementary Table S1). We note that there was no significant difference in the composition of solutions collected from static experiment CF-11, and a continuously agitated experiment CF-20 using the same sized olivine starting solid material (Supplementary Table S1, Fig. 1 A; B), particularly for early on in the experiments where kinetic constants are calculated

(discussed later). We take this to suggest that these static experiments adequately represent the reactivity of REE on olivine.

60 °C Experiments: The 60 °C time-series experiments at a GSA of 28.2, 57.8, 176 and 3339 cm²/g (CF-18, CF-12, CF-13, and CF-21; Fig. 2, Supplementary Table S1) also showed that Sr and Ba remain in solution while REE are increasingly removed over time. Again, with increasing GSA the REE are progressively depleted in solution, with preferential depletion of HREE over LREE (Fig. 2). Experiment CF-18 solution concentration with the smallest GSA of 28.2 cm²/g, remained constant within 5% of the starting solution for the last 3 days of the experiment (from 2800 to 7200 minutes; Fig. 2A).

90 °C Experiments: As in the 15 °C and 60 °C experiments, the 90 °C time-series experiments at a GSA of 28.2, 57.8, 924 and 3339 cm²/g (CF-17, CF-19, CF-24, and HP; Fig. 3, Supplementary Table S1) showed that Sr and Ba remain quantitatively in solution while REE are progressively removed with increasing GSA, and the HREE are removed preferentially over the LREE (Fig. 3). Experiment CF-17 (Fig. 3A) showed a step-wise decrease in REE concentration at 1200 and 3600 minutes. We speculate that this may be due to breakage of some grains that resulted to an increase in surface area.

Temperature effect: Experiments conducted at a given olivine surface area and at the 3 temperatures (GSA = 57.8 cm²/g) allow to constrain the effect of temperature on the reactivity of REE (Fig. 4). With increasing temperature the amount of REE removed from the solution increases, while Sr and Ba (as well as Ca²⁺ and Mg²⁺) remain quantitatively in solution (Fig. 4). Also, REE fractionation increases with increasing temperature, with greater removal of the HREE over the LREE. For example, as the temperature increases from 60 to 90 °C and after

1400 minutes, the Nd concentration in solution decreases from 75 % to 46 % of the starting solution and the Yb/Nd_(n) ratio decreased from 0.65 to 0.42, respectively. Note that we cannot perform the same quantitative analysis with the highest GSA (3339 cm²/g) experiments due to the near quantitative removal of REE from solution at all temperatures, approaching analytical detection limits.

A common observation in all the above experimental data is that the REE solution concentrations change quickly in the beginning of the experiment and the change becomes progressively slower over time, with some experiments (e.g. Figures 1A, 2C, 3C) approaching an apparent steady-state condition. Importantly, the HREE Yb and Dy are removed to a greater extent than the LREE Nd and Sm, implying greater reactivity for the HREE than the LREE.

Olivine Leaching: To confirm the REE adsorption on olivine surface we performed blank experiments where the starting aqueous solution was placed in the PFA vials without olivine. The data showed no change in REE concentrations over time, confirming that REE were not adsorbed on the vial walls, as is also generally known for polyethylene bottles (e.g., Stern et al. 2007). We also performed leaching tests on mineral fractions recovered from experiments CF-12 and CF-13 at 60 °C. The recovered olivine grains were rinsed twice with 18 MΩ H₂O to remove salts, and then dried on HEPA-filtered hot plates at 60 °C. Once dry, a 10 wt% HCl solution was added to each beaker and after an hour the solution was subsampled and analyzed. The total content of the REE, as the sum of those remained in solution at experiment termination and those recovered from the olivine match the REE composition of the starting solution to within 5 % of the starting amount (Table 5). This confirms that the depletion of REE in solution is due to scavenging and adsorption on reactant olivine.

Olivine dissolution and alteration: Under the temperature conditions of our experiments, olivine dissolution can occur with the potential formation of secondary Fe(III)-(hydr)oxides (Hänchen et al., 2006; Mayhew et al., 2013; Oelkers, 2001; Pokrovsky and Schott, 2000; Stopar et al., 2006). In addition, experimental data from Chen and Brantley (2000) show stoichiometric olivine dissolution at pH=2.95 to 5 at 65 °C in HCl media, at conditions generally similar (albeit far more acidic) to ours.

Our experimental results, however, do not show stoichiometric dissolution of olivine, or any dissolution for that matter. In all experiments the Mg/Ca in the solution (to account for any instrumental drift) remained effectively unchanged and within measurement uncertainty (Supplementary Figure SF4), of the concentrations in the starting solution (e.g., ~1 % relative), therefore no significant Mg addition for possible olivine dissolution was observed, stoichiometric, or otherwise.

Moreover, we monitored the concentration of dissolved Ni and Fe in the experimental solution as a proxy for olivine dissolution. Under the circumnatural pH (~7 to 7.8) of the experimental solutions, Fe is highly insoluble and will precipitate from solution as Fe oxyhydroxide while Ni remains soluble (Ji and Cooper, 1996; Takeno, 2005). Ni and Fe are found in high concentrations in olivine but absent in the starting solution (Table 1, 2). The use of medium resolution mode ($\Delta m/m = 4000$) on the ELEMENT2 eliminates isobaric interferences on the analyzed ^{60}Ni and ^{57}Fe isotopes, resulting in low detection limits (typically 2 to 5 ppt). This allows the use of Ni in solution as a sensitive tracer for the amount of olivine dissolution in the experiments. Results show that while Ni in some experiments slowly increased over time (e.g., CF-18), the Ni concentrations in most experimental solutions (for the time range where kinetic

rate constants were calculated, see below) ranged from <1 ppb to 10 ppb with most being less than 3 ppb (Supplementary Table S1). Occasional “spikes” in Ni, Fe and Mn concentrations (e.g., experiment CF-23 at 120 minutes) are likely <0.2 μ m particles that were not filtered. Considering the high end of Ni concentrations for the filtered experiments (~10 ppb, experiment CF-23, Supplementary Table S1) and assuming stoichiometric dissolution of olivine, our estimations suggest that up to 1.2 μ g of olivine could have dissolved which is 0.013 % of starting olivine mass.

Solid reaction products were analyzed at the termination of the experiment for selected experiments (CF-19, HP) using SEM, x-ray diffraction (XRD) and Mößbauer spectrometry (Supplementary Figs. SF1, SF2, SF3). There were no distinct textural or compositional variations detected with the SEM images to suggest alteration or precipitation of secondary mineral (e.g., chrysotile, brucite, magnetite, Fe-oxides) (Supplementary Figs. SF1). Similarly, no secondary minerals or Fe³⁺ contributions in both the XRD and Mößbauer spectra were identified in solid reaction products (Supplementary Figs. SF2, SF3). These results are consistent with the very sluggish kinetic rates of olivine alteration and serpentinization reported in previous low-temperature (<100 °C) experimental studies, in which trace amounts of alteration phases such as Fe³⁺-(hydro)oxides, talc and serpentine were developed after at least 800 hours of reaction (Mayhew et al., 2013; Neubeck et al., 2014; Miller et al. 2017).

4.0 Discussion

Our data show that REE are readily adsorbed on to the surface of olivine, while Sr and Ba remain quantitatively in solution. These findings are consistent with earlier studies (Byrne and

Kim, 1990; Tang and Johannesson, 2005; Tertre et al., 2008) that showed in the absence of strong organic and carbonate ligands, REE are particle reactive with HREE being preferentially removed from both seawater and fresh water over LREE by adsorption on to mineral surfaces. The low reactivity of the alkaline earths seen here is consistent with experimental data that show little adsorption of Sr, Ba, Ca, and Mg at circumneutral pH for a variety of mineral reactive surfaces (rutile, goethite, amorphous silica, quartz; Sverjensky, 2006 and references therein for a compilation of experimental data on alkali metal reactivity). These findings are also qualitatively consistent with the well-known long residence time (~4 million years) of Sr and the short residence time (hundreds to thousand years) of REE in seawater (Bruland, 1983; Goldberg, 1965). However, our data allows us to quantify REE reaction rates at conditions relevant to low temperature hydrothermal conditions and seawater / rock interaction. In the following, we explore the kinetic rates of REE reaction and the role of carbonates, Fe-hydroxides, and REE aqueous speciation on REE uptake and fractionation during low-temperature olivine/fluid interaction.

4.1 Adsorption kinetic rate constants

Experimental data indicate that the concentration of REE in solution (as $\ln(C/C_0)$) changes rapidly at the beginning of the experiments (Figs. 1 to 3), but slows with time, i.e., the kinetics of REE adsorption changes with time from fast to slow. In some cases, especially in the coarser grain (small GSA) experiments (e.g., Fig. 4C), the concentrations approach steady-state conditions where little additional REE removal is observed. This change in adsorption kinetics is clearly not that of a first order reaction, where the natural logarithm of the normalized REE

concentrations is expected to show strong linear correlation with reaction time. Thus, we evaluated the possibility of a second order reaction by plotting the $(C/C_0)^{-1}$ for Nd and Yb vs. reaction time for selected experiments (Fig. 5). Some experiments show a strong linearity particularly early in the experiment (Fig. 5E, and F), whereas others show characteristic changes in kinetic rates, as a function of the reaction time (Figs. 5A, B, D, and C; concave down curves).

From this, it becomes apparent that the kinetics of REE adsorption on to olivine surface do not follow a first or second order reaction pattern. More complex kinetics with multiple rates were not investigated. However, the essential observation is that during the early stages of reaction conditions of enhanced kinetics are established (Figures 1 through 4). For simplicity, therefore, the adsorption kinetic rate constants were approximated as two pseudo first-order reaction regimes using linear regression between $\ln(C/C_0)$ and reaction time (an example is shown Fig. 6). The transition point between these two reaction regimes was approximated by manually optimizing the fit of a linear regression to the data in the early stage of the experiments to increase the r-squared values of the fitted model. Our only limit to the fit was to include a minimum of 4 points in the first reaction regime, so that the calculated rate constants represent at least 2 hours of reaction time. The initial kinetic rate constants of REE scavenging, k^{REE} , calculated for the experiments are given in Table 6.

This approach explicitly assumes that k^{REE} are constant for the selected time intervals. While this is obviously not entirely valid as discussed above, nevertheless, the k^{REE} effectively represents an “average” REE adsorption kinetic rate constant over that time interval. As a result of this approximation, some uncertainty in the selection of the transition point between the two reaction regimes (Fig.6) does not significantly affect the estimated values k^{REE} (Table 6). The

uncertainties of the estimated k^{REE} values are calculated assuming a 5-minute sampling time error, whereas the error on the $\ln(C/C_0)$ value is calculated by a full error propagation of the analytical uncertainties (1 standard deviation, typically 2-3 % relative). Note that for experiments with the highest GSA (3339 cm²/g), we were unable to quantify kinetic constants due to the relatively long sampling intervals compared to the rapid rate of REE removal from solution (Figures 1F, 2D, 3D).

4.2. REE adsorption kinetic rate constants and activation energy

The adsorption kinetic rate constants (k^{REE}) increase with both increasing temperature and GSA. This is shown in Figure 7, where the experimentally-derived k^{Nd} and k^{Yb} (only Nd and Yb are shown for clarity as the rest of the analyzed REE fall between the two) are plotted against total olivine surface area (SA: mass, multiplied by the GSA of the particular olivine fraction) and as a function of temperature. At a given temperature, the k^{REE} increase approximately linearly with the surface area (Fig. 7). Based on this linearity, we quantify the effect of temperature on k^{REE} by normalizing k^{REE} to the mineral surface area using a least-squares regression, and for each temperature. The slope of this line defined as k^{REE}/SA , is effectively the “surface area normalized adsorption kinetic rate constant” for a given REE at a given temperature (Fig. 7, Table 7).

Figure 8 shows that the natural *log* of surface-normalized REE adsorption kinetic rate constants, $\ln(k^{REE}/SA)$, decrease with the reciprocal of temperature, i.e., the highest k^{REE}/SA are found at the highest temperature experiments. The approximately linear relationship of

$\ln(k^{REE}/SA)$ with $1000/T$ suggests that the kinetics of REE olivine surface adsorption may be described through an Arrhenius relationship:

$$\ln\left(\frac{k^{REE}}{SA}\right) = \ln(A) - E_a^{REE}/RT \quad (1)$$

where $R = 8.314$ (J/(K*mol)), E_a is the activation energy and A is the frequency factor. In Figure 8, the E_a is defined as the slope of $\ln(k^{REE}/SA)$ vs. $1000/T$, and the frequency factor, A , is the intercept. The calculated E_a for each REE are reported in Table 7 and range from 16.3 ± 0.6 to 13.1 ± 0.6 kJ/mol. The HREE have lower activation energies and frequency factors than the LREE (Table 7). This suggests that adsorption of HREE is less temperature dependent than the LREE. In other words, the HREE are more readily adsorbed on the surface of olivine than the LREE at the temperature conditions of our experimental study.

4.3 Gd decoupling and REE electron configuration

The calculated kinetic rate activation energies decrease with decreasing atom radius (or increasing atomic number) from LREE to HREE (Fig. 9a; i.e., from Nd to Yb). However, Gd falls off this trend having characteristically higher E_a than the neighboring Sm and Dy and higher than all analyzed REE, implying another property of REE, in addition to ionic radius, controls their reactivity as a function of temperature. Instead, the activation energy, E_a^{REE} , yields a strong linear relationship with the sum of the 1st, 2nd, and 3rd ionization potentials (referred herein as total ionization potential) of the REE (Fig. 9b). Gadolinium and Nd with the highest E_a both have the lowest total ionization potential, whereas Yb with the lowest E_a has the highest total ionization potential. The total ionization potential represents the amount of energy required to remove the three outer electrons from the atom and is a function of the electron configuration of

the REE. Gadolinium +3 has an $[\text{Xe}]4f^7$ electron configuration where each of the seven f -orbitals are filled with one unpaired electron (i.e., half-filled f orbitals). In contrast, the +3 cations of Nd, Sm, Dy and Yb either have some unfilled f -orbitals (Nd, Sm) or some of the f -orbitals with two electrons (Dy, Yb). This electron configuration likely results in a more symmetric and less polarized electric field for Gd^{+3} (De Baar et al., 1985b), than the other analyzed REE and may account for its relatively lower reactivity (and higher activation energy) compared to Sm and Dy with the nearest atomic number and radius (Bau, 1999)). The relative enrichment of Gd relative to other REE in some Atlantic and Pacific seawaters (Alibo and Nozaki, 1999; De Baar et al., 1985b) has also been attributed to a lower reactivity due to its electron configuration.

4.3 Mechanisms of adsorption on olivine.

4.3.1 REE-complexation by carbonates?

As our experiments were open to the atmosphere, the presence of dissolved CO_2 may stabilize carbonate-bearing REE aqueous species and promote the precipitation of carbonate minerals. However, thermodynamic calculations conducted by accounting for the chemical composition of the sampled aliquots, suggest that calcite is not stable, whereas dolomite and magnesite are. Formation of secondary dolomite could alter the $\text{Ca}^{2+}/\text{Mg}^{2+}$ of the solution due to preferential uptake of Ca^{2+} over Mg^{2+} in the dolomite crystalline structure. With $\text{Ca}^{2+}/\text{Mg}^{2+}$ ratio of 0.21 in reactant fluids and the ideal dolomite mineral with $\text{Ca}^{2+}/\text{Mg}^{2+}$ ratio of 1.65, dolomite precipitation would decrease the $\text{Ca}^{2+}/\text{Mg}^{2+}$ ratio of the experimental solutions. The $\text{Ca}^{2+}/\text{Mg}^{2+}$ in our solutions remain unchanged within error (Supplementary Figure SF4) providing no evidence for significant carbonate formation. These data are also consistent with the lack of evidence for

carbonates in the experimental solids in the SEM and XRD analyses. Ca-bearing phases, such as dolomite and calcite, should also remove Sr (Foustoukos et al., 2008; Haggerty et al., 1992; Menzies et al., 1993; Scambelluri et al., 2001), and the lack of evidence for Sr removal in our experiments further argues against calcite precipitation.

The overall HREE enriched patterns of seawater (Alibo and Nozaki, 1999; Nozaki and Alibo, 2003) are attributed to the greater reactivity of LREE than HREE in marine particles, which is opposite to what we observe in our experiments. This has been attributed to the stronger complexation of HREE to carbonate ligands, REE-CO_3^+ and $\text{REE-(CO}_3)_2$ which leaves proportionally more LREE than HREE as free 3^+ cations to hydrolyze and react with particle surfaces (Johannesson et al., 2017; Luo and Byrne, 2004; Schijf et al., 2015). This further implies that REE-carbonate ligands do not control the REE reactivity observed in our experiments. We speculate that with higher carbonate activity in solution, more typical of seawater, the stronger complexation of HREE over LREE with carbonate ligands may result in increased LREE uptake, as observed in amorphous ferric hydroxide experiments with increasing carbonate concentration (Quinn et al., 2006b).

4.3.1 REE⁺³ scavenging by Fe-oxyhydroxides?

REE have very high affinity for freshly precipitated hydrous ferric oxides (HFO) (Koeppenkastrup and De Carlo, 1993; Quinn et al., 2004; Quinn et al., 2006a; Quinn et al., 2007; Schijf and Marshall, 2011). In our experiments, olivine dissolution would result in the release of Fe into the solution, which would likely form HFO under these experimental conditions (pH~7, open to atmosphere) and precipitate. For example, Fe-(hydr)oxide precipitation has previously

reported in low-temperature olivine dissolution (Chen and Brantley, 2000) and peridotite alteration experiments under slightly acidic to neutral pH conditions (pH = 5-7) (Mayhew et al., 2013). However, based on the low amount of dissolved Ni, and the constant Mg concentrations in the experimental solutions (see Results section), olivine dissolution would have released much less than 500 ng of Fe during the course of the experiments. This is far less than the starting amount of REE in solution (2.5 µg total) and we deem unlikely that this amount of Fe could control REE precipitation. Furthermore, the chemical and spectroscopic analysis did not reveal any Fe-bearing secondary minerals or Fe³⁺ contributions in both the XRD and Mößbauer spectra in the solid reaction products (Supplementary Figures SF2, SF3).

We can further evaluate the role of possible HFO precipitation by comparing our results with HFO-induced REE scavenging experiments (Koeppenkastrop and De Carlo, 1992; Koeppenkastrop and De Carlo, 1993; Quinn et al., 2004; Quinn et al., 2006b; Quinn et al., 2007). Some of our experiments approached a steady state in REE dissolved concentrations (e.g., Figs. 1B, E; Figs. 2A,C,D; Figs. 3B,C), and thus, we use these experimental data to calculate apparent distribution coefficients as: $appD^{REE} = C_{oli}^i / C_{sol}^i$, where C_{oli}^i is the amount of element i removed from solution normalized to the mass of olivine in the experiment, and C_{sol}^i is the concentration of REE remaining in solution (Bau, 1999) (Table 8). For experiments where more than 99% of the REE were removed (AP, CF-21, HP, CF-24), $appD^{REE}$ values are less reliable, because $appD^{REE}$ becomes very sensitive to the diminishing C_{sol}^{REE} (Table 8).

Figure 11 compares the relative fractionation of the $appD^{REE}$ as D^{Yb}/D^{Nd} , with the magnitude of the $appD^{Gd}$ anomaly, here defined as $\log(appD^{Gd})/Gd^*$, where Gd^* is the expected value for $appD^{Gd}$ if it were to fall between those of $appD^{Sm}$ and $appD^{Dy}$, i.e.: $\log[(appD^{Dy} +$

app D^{Sm})/2]. Our olivine adsorption data show greater adsorption of HREE over LREE and larger negative Gd anomalies than the D^{REE} from HFO precipitation experiments. Based on the greater HREE over LREE partitioning on olivine surface than HFO, the low amount of Fe released from olivine dissolution and the lack of spectroscopic evidence for secondary minerals evidence, we conclude that precipitation of HFO is unlikely the dominant mechanism of REE removal from solution in our experiments.

4.3.2 Insights from REE speciation in solution

Additional insights can be gained by modeling the speciation of REE in solution as a function of temperature and under atmospheric $\text{CO}_{2(g)}$ fugacity (Fig. 12). We compare the results with the olivine experiments CF-12 and CF-19 at low GSA ($57.8 \text{ cm}^2/\text{g}$) where the Gd anomaly is more prominent (e.g. Fig. 10, Fig. 11). At 60°C and slightly acidic pH conditions REE in solution exist dominantly as 3^+ free species, however at near neutral pH conditions the REE- OH^{2+} species become increasingly abundant and approximately equal to the REE^{3+} . Across the same pH range but at 90°C the dominant species are the REEOH^{2+} . The REE-CO_3^+ species appear to be of less importance in these experimental conditions. This is also consistent with our observations described above where REE-CO_3^+ complexes do not seem to control the REE reactivity.

The decrease in the proportion of REE^{3+} to REEOH^{2+} species with increasing temperature coincides with the decrease in the magnitude of the D^{Gd} anomaly with temperature (Fig. 11). In other words, when REE^{3+} are the dominant species in solution, their reactivity to particle surfaces must be dictated by their electron structure. The apparent effect of REE electron

configuration is also most prominent in the coarser olivine grain experiments (lowest GSA). This may suggest that under low GSA conditions the available reactive surfaces are naturally limited, and REE outcompete each other for the few available adsorptive sites. At the higher GSA experiments, as in HFO precipitation experiments (Koeppenkastrop and De Carlo, 1993; Quinn et al., 2004; Quinn et al., 2006a; Quinn et al., 2007; Schijf and Marshall, 2011), the excess of adsorptive sites that result in near complete removal of REE from solution can explain the smaller Gd anomaly.

Tertre et al., (2008) suggested that REE adsorption on basalt is controlled by two parallel processes, cation exchange at the surface along Na-saturated sites ($>XNa$). and surface complexation ($>SOH$) on hydroxylated sites (Si, Al, Fe), respectively. In their experiments adsorption is enhanced under neutral pH and high ionic strength conditions leading to the preferential adsorption of HREE on mineral surfaces relative to LREE, qualitatively similar to our observations.. Their calculated surface “sorption constants” for REE on to basalt surface hydroxylated sites show greater Nd/Yb fractionation than the HFO precipitation experiments equilibrium constants discussed above and better match with the observed REE systematics (e.g. D^{Yb}/D^{Nd} vs. $\log(appD^{Gd})/Gd^*$, see Figure 11) of our experiments. Following Tertre et al., (2008), we surmise that the apparent link between REE electron configuration and their adsorption systematics on olivine in our experiments points to surface complexation through inner-sphere type complexes of the REE^{3+} ions on hydroxylated surfaces (Stumm, 1995; Sverjensky and Fukushi, 2006), and in the case of olivine likely those of silicon or iron, as Mg is more soluble and would be removed in solution and Al is typically in very low abundance.

5.0 Summary and Implications

During reaction of olivine and seawater in the absence of strong complexing organic ligands (carbonate, sulfate) and precipitation of secondary minerals, Sr and Ba remain in solution while REE are removed from solution on to olivine with increasing kinetic rate constants as both temperature and reactive surface area increase. Based on the behavior of Nd, Sm, Dy and Yb as a function of their ionic radii, the kinetics and the apparent partitioning of Gd on to the olivine surface are lower than expected, but correlate best with the total ionization potential of the REE, which is a function of their electron configuration. As temperature increases the extent of the Gd anomaly coincides with changes in the REE speciation in solution from dominantly REE^{3+} to REE^{3+} and REEOH^{2+} at 60 °C and REEOH^{2+} at 90 °C. The dominance of electron configuration on the reactivity and adsorption of REE^{3+} on olivine implies that inner sphere type surface complexes dominate the adsorptive characteristics under these experimental conditions. In our experimental set up REE-carbonate speciation or Fe-oxyhydroxide precipitation have a minor, if any, effect.

During progressive reaction of seawater with peridotite, the enhanced reactivity of HREE will progressively generate LREE-enriched fluids that may ultimately impart a LREE enriched signature in serpentinized peridotites (Frisby et al., 2016b). We speculate that the variable REE systematics seen in secondary phases in serpentinites (e.g. Gillis et al., 1992; Rouméjon et al., 2015) and the highly variable REE patterns (from LREE depleted to LREE enriched) in hydrothermal vent fluids (Douville et al., 1999) may also be controlled in part by adsorption of REE on to mineral surfaces (at least at near neutral pH conditions) and subsequent incorporation of the adsorbed elements within the structures of the crystallizing secondary minerals. Seawater

has REE patterns that are enriched in HREE over the LREE, but they are nonetheless more LREE enriched than typical depleted peridotites (Frisby et al., 2016b). We envision that preferential scavenging of HREE from seawater may take place within the recharge zone at relatively low temperature conditions (<100 °C). As the fluids become progressively LREE enriched, they reach high temperature water/rock interaction zones deep in oceanic crust, imparting seawater-like Nd isotope systematics in peridotites exhibiting high water / rock mass ratio alteration (Delacour et al., 2008; Frisby et al., 2016a). Additional detailed trace element and isotopic studies in serpentinized peridotites, especially a comparison between serpentine and relic fresh minerals, are needed to decipher the origin of their flat LREE patterns, either as melt infiltration (Deschamps et al., 2013), seawater addition (Frisby et al., 2016a; Frisby et al., 2016b), or some combination therefore (Rouméjon et al., 2018) that is depended on tectonic setting. Increasingly complex experiments are also needed, where the role of serpentine formation on REE and other trace element uptake (partitioning) can be distinguished from that of surface adsorption during growth.

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Figure Captions

Figure 1:

Sr, Ba, REE normalized concentrations in solution, $\ln(C/C_0)$, vs. time for experiments at 15 °C. **A:** ALG, with GSA = 28.2 cm²/g; **B:** CF-11, with GSA = 57.8 cm²/g; **C:** CF-20, with GSA 57.8 cm²/g agitated; **D:** CF-22, with GSA 257cm²/g; **E:** CF-23, with GSA 577 cm²/g and **F:** AP with GSA = 3339 cm²/g. Y-axis scale is fixed for every pair of panels for clarity, but changes in the Y-axes between the panel pairs. X-axes change between panels corresponding with total experiment time. Detection limit (DL) range is indicated in panel F for these analyses. In all cases Sr and Ba concentrations remain constant while REE concentrations decrease over time. Towards the end of the experiments, REE concentrations level off, approaching conditions of steady state. At a given time, REE concentrations are lower in experiments with greater surface area (e.g. compare panels C and D, E and F). Moreover, inter-element REE fractionation increases with increasing surface area.

Figure 2:

Sr, Ba, REE normalized concentrations in solution, $\ln(C/C_0)$ vs. time for experiments at 60 °C. **A:** CF-18, with GSA = 28.2 cm²/g; **B:** CF-12, with GSA = 57.8 cm²/g; **C:** CF-13, with GSA = 176 cm²/g; **D:** CF-21, with GSA = 3339 cm²/g. Y-axis scale is fixed for every pair of panels for clarity, but changes in the Y-axes between the panel pairs. X-axes change between panels corresponding with total experiment time. Detection limit (DL) range is indicated in panel D for these analyses. In all cases Sr and Ba concentrations remain constant while REE concentrations decrease over time. Towards the end of the experiments, REE concentrations level off, approaching conditions of apparent equilibrium. At a given time, REE concentrations

are lower in experiments with greater surface area (e.g. compare panels C and D). Moreover, inter-element REE fractionation increases with increasing surface area.

Figure 3:

Sr, Ba, REE normalized concentrations in solution, $\ln(C/C_0)$ vs. time for experiments at 90 °C. **A:** CF-17, with GSA = 28.2 cm²/g; **B:** CF-19, with GSA = 57.8 cm²/g; **C:** CF-24, with GSA = 924 cm²/g; **D:** HP, with GSA = 3339 cm²/g. Y-axis scale is fixed for every pair of panels for clarity, but changes in the Y-axes between the panel pairs. X-axes change between panels corresponding with total experiment time. Detection limit (DL) range is indicated in panel D for these analyses. In all cases Sr and Ba concentrations remain constant while REE concentrations decrease over time. Towards the end of the experiments, REE concentrations level off, approaching conditions of apparent equilibrium. At a given time, REE concentrations are lower in experiments with greater surface area (e.g. compare panels C and D, E and F). Moreover, inter-element REE fractionation increases with increasing surface area.

Figure 4:

Sr, Ba, REE normalized concentrations in solution, $\ln(C/C_0)$ vs. time for experiments at 57.8 cm²/g. **A:** CF-11, at 15 °C; **B:** CF-12, at 60 °C; **C:** CF-19, at 90 °C. Y-axis scale is fixed for all panels, x-axis scale changes between panels reflecting total time of the experiment. Sr and Ba concentrations remain constant while REE concentrations decrease in solution over time. At a given time, REE concentrations are lower in experiments as temperature increases. Moreover, inter-element REE fractionation increases with increasing temperature.

Figure 5:

(C/C_0)⁻¹ against time of representative experiments (**A**: CF-11, **B**: CF-12, **C**: CF-19, **D**: CF-23, **E**: CF-13, **F**: CF-24) for Nd and Yb to test for second order reaction processes. Experiments B, C, and F display a near linear trend while A, D, and E have a distinct concave down appearance. The lack of a distinctive linear trend is indicative of these associated reactions to not be purely second order in characteristic.

Figure 6:

$\ln(C/C_0)$ of Nd and Yb against time for representative experiments **A**: CF-11 (15 °C) and **B**: CF-19 (90 °C) with GSA of 57.8 cm²/g. Experimental data for each element are treated as the combination of two first order reactions, an initial fast reaction (bold black line) followed by a slower reaction (grey dashed line). Each segment represents an average k . Transition point selection (Δk) is further explained in the text, represented here as a dashed arrow.

Figure 7:

Reaction rates for representative elements **A**: Nd and **B**: Yb at 15, 60 and 90 °C against surface area (m² x 1000). Surface area is determined from GSA times the mass of olivine in each experiment. The slope of the line connecting these data points effectively represents the surface area normalized kinetic constants (k/SA). The k/SA is used to remove the influence of surface area related reactions imposed on the experimental kinetic constant. Represented errors are smaller than data points on figure.

Figure 8:

$\ln(k/SA)$ for each element plotted against 1000/T (K), to describe the Arrhenius function ($\ln(k/SA) = \ln(A) - (E_a/RT)$) where E_a is the activation energy of the reaction, R is the gas constant 8.134 JK⁻¹mol⁻¹ and A is the frequency factor. Represented errors are smaller than data points on

figure. Furthermore the figure shows that reaction rate constants are faster for the HREE than LREE.

Figure 9:

The estimated activation energy (kJ/mol) plotted against **A:** Ionic radius (Å) (Sonke and Salters, 2006) and **B:** sum of the first, second and third ionization potentials (eV) for Nd, Sm, Gd, Dy and Yb. Plot A has HREE on the left and LREE on the right. Near linear correlation is displayed between E_a and ionic radius with the exception of Gd showing a positive anomaly. Plot B displays a near linear correlation for E_a and total ionization potentials for all REE where Gd has the higher E_a and second lowest total ionization potential, and Yb has the lowest E_a and highest total ionization potential.

Figure 10:

Reaction rate constants (k) plotted against apparent distribution coefficient (D_{REE}) for CF-11, CF-12, CF-19 with GSA 57.8 cm²/g at 15, 60 90 °C respectively. Experiments with GSA 57.8 are shown that cover all three temperatures and allow reliable D^{REE} calculations The k and D_{REE} show a linear correlation for the three temperatures. Gd displays anomalous behavior falling between Nd and Sm for both the determined k and D_{Gd} rather than between Sm and Dy.

Figure 11:

REE fractionation as $appD_{Yb}/appD_{Nd}$ for select experiments at 15 °C (blue circles), 60 °C (red squares), and 90 °C (green diamonds) as a function of the Gd anomaly $logD_{Gd}/Gd^*$. Symbol size increases with increasing GSA. Data is compared with equilibrium distribution coefficients for REE sorbed on Fe(OH)₃ from Quinn et al., 2004 (Q'04) in the absence of dissolved carbonate; Quinn et al., 2006b (Q'06b) in the presence of dissolved carbonate; Quinn et al., 2007

in the absence of carbonate at 39.3°C and 10°C (Q'07 a, Q'07 b, respectively); Koeppenkastrop and De Carlo, 1992 (K&DeC'92) for REE sorbed on vernadite and hydroxyapatite; Tertre et al., 2008 (T'08) for REE sorption constants on to basalt surface as inner sphere complexes. All data is reported in Table 8.

Figure 12:

REE speciation of the experimental aqueous solutions in equilibrium with atmospheric fCO_2 at 60 and 90 °C. At the experimental pH range of 7-7.5 the REE exist predominantly as $REE(OH)^{2+}$. At higher pH conditions and with increase in temperature, the speciation of dissolved REE transitions from a) carbonate dominated (60 °C) to b) hydroxyl dominated (90 °C). The distribution of Nd hydroxyl species includes the single complex $Nd(OH)_{3(aq)}$ by adopting the thermodynamic properties developed by Migdisov et al. (2016). The depicted REE chloride complexes (e.g. Nd^{3+} , Yb^{3+} Cl^-) correspond to the total concentrations of the dominant $REECl^{2+}$ and $REECl_2^+$. Species reported are those with concentrations greater than 1 μ molal.