

Secondary Organic Aerosol Formation from Reaction of 3-Methylfuran with Nitrate Radicals

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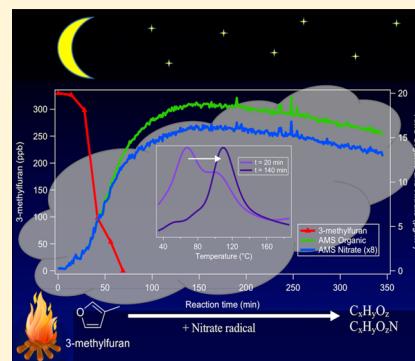
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Supporting Information

ABSTRACT: A significant amount of furan species is emitted from biomass burning. They are highly reactive to hydroxyl (OH) and nitrate radicals (NO_3), which can lead to the formation of secondary organic aerosol (SOA). Here, we investigate gas-phase oxidation and SOA formation from 3-methylfuran ($\text{C}_5\text{H}_6\text{O}$) via NO_3 reaction. Experiments are performed under dry conditions (RH < 5%) and with different initial concentrations of 3-methylfuran (from 95.9 to 562.8 ppb). We demonstrate that this reaction leads to SOA formation, with SOA yield ranging from 1.6 to 2.4% for organic mass loading ranging from 5.1 to 45 $\mu\text{g}/\text{m}^3$. More than half of the SOA mass is generated after complete depletion of 3-methylfuran, highlighting the importance of higher-generation or multiphase reactions to aerosol formation. Particle-phase organic nitrates contribute 39.4% of organics and their average volatility (average $\text{C}^* = 10^{-2.9} \mu\text{g}/\text{m}^3$) is higher than that of non-nitrate organic compounds (average $\text{C}^* = 10^{-3.3} \mu\text{g}/\text{m}^3$). A reaction mechanism is proposed based on the identified products, and $\text{C}_5\text{H}_5\text{NO}_5$ and $\text{C}_5\text{H}_6\text{O}_3$ are determined to be the major species in the gas and particle phases, respectively. Oligomer formation appears to determine the SOA composition and formation rate, and both gas-phase ROOR' formation via $\text{RO}_2 + \text{RO}_2$ (acylperoxy radical) reactions and particle-phase accretion reactions can lead to the formation of the dimeric (C_{10}) compounds observed. Results from this study provide detailed chemistry of 3-methylfuran oxidation that can improve our understanding of its impact on SOA and ozone formation in nighttime biomass burning plumes.

KEYWORDS: Furan, methylfuran, biomass burning, secondary organic aerosol, SOA, oligomers, organic nitrate, nighttime oxidation



INTRODUCTION

Biomass burning is a significant source of both gas- and particle-phase carbon in the atmosphere.^{1–3} It impacts air quality and climate and is expected to increase due to climate change.⁴ Biomass burning plumes contain thousands of nonmethane organic compounds (NMOCs) that can contribute to the formation of pollutants including ozone and secondary organic aerosol (SOA).^{5,6} There has been substantial effort to identify these species and understand their behavior as the plumes age.^{7–12} Recent studies have shown that furan compounds are an important class of NMOCs in biomass burning plumes, their contribution being higher in smoldering-type combustion than in flaming-type combustion.^{9,13–15} Given their large emission rates and rapid reaction rates with hydroxyl (OH) and nitrate radicals (NO_3),^{11,12,14,16} it is important to understand the atmospheric chemistry of furan derivatives. A few previous studies have shown the ability of these compounds to generate SOA.^{17,18} However, their gas-phase oxidation products and mechanisms have not been identified and their SOA formation potential remains poorly constrained.

Furans are formed from cellulose pyrolysis,¹⁹ and methylfurans are a major subset of furans, accounting for 20–25% of furan emissions.^{13,14} Nighttime biomass combustion has been associated with variations in organic aerosol composition in multiple field campaigns, and in particular with the formation of organic nitrates.^{20–23} Decker et al.¹² reported that furans can be responsible for 19–60% of initial NO_3 reactivity during nighttime biomass burning and around 60% of methylfurans are oxidized by NO_3 instead of OH or ozone. Tapia et al.²⁴ reported the rate coefficient and gas-phase products of 3-methylfuran oxidation by NO_3 radicals and proposed the gas-phase formation mechanism of first-generation organic nitrate products. However, the products were proposed based on the measurements of functional groups (rather than measurements of individual species) and SOA formation was not studied. For

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these reasons, there is a need to characterize the SOA formation potential and gas-phase chemistry of the 3-methylfuran + NO₃ reaction as it relates to nighttime aging in a biomass burning plume.

In this work, we investigate SOA formation from 3-methylfuran oxidation by NO₃ radicals in laboratory chamber experiments. SOA yields are determined from a series of experiments with different initial 3-methylfuran concentrations. Speciated gas-phase products and aerosol composition are monitored using online measurements in order to propose a mechanism for multigenerational oxidation and SOA formation. These results can be used to estimate SOA formation from biomass burning emissions and improve our understanding of oxidation in biomass burning plumes.

■ EXPERIMENTAL SECTION

Environmental Chamber Experiments. Experiments are performed in the Georgia Tech Environmental Chamber (GTEC) facility²⁵ at room temperature (~ 25 °C) and under dry conditions (<5% RH) (Table 1). Ammonium sulfate seed

Table 1. Experimental Condition and Aerosol Yield for All Experiments

Experiment	Condition	ΔHC (ppb)	ΔM_0 ($\mu\text{g}/\text{m}^3$)	Yield (%)
1	3-methylfuran + NO ₃	562.8 \pm 2.2	45.0 \pm 0.6	2.4 \pm 0.03
2	3-methylfuran + NO ₃	328.5 \pm 1.2	24.0 \pm 0.6	2.2 \pm 0.06
3	3-methylfuran + NO ₃	171.3 \pm 0.3	10.6 \pm 0.6	1.9 \pm 0.1
4	3-methylfuran + NO ₃	95.9 \pm 0.2	5.1 \pm 0.6	1.6 \pm 0.2

aerosol is introduced into the chamber by atomizing a 0.015 M solution. The initial particle number and volume concentrations are $\sim 16\,000$ particles cm^{-3} and $\sim 21\, \mu\text{m}^3 \text{ cm}^{-3}$, respectively. In each experiment, 3-methylfuran (ACROS, Geel, Belgium) is injected into a glass bulb, and the compound is introduced into the chamber by passing pure air through the glass bulb.

After the seed particle and 3-methylfuran injection, NO₂ (Matheson, 500 ppm) and O₃ (generated by passing purified air through a UV radiation cell, Jelight 610, ~ 250 ppm) are injected into the chamber through a flow tube (0.9 L min^{-1} flow rate, 102 s residence time) in order to generate NO₃ radicals and N₂O₅.^{25,26} NO₂ and O₃ react to form NO₃, which is followed by the reaction of NO₂ and NO₃ to form N₂O₅. N₂O₅ thermally decomposes back to NO₂ and NO₃ radicals establishing an equilibrium. To ensure >99% of 3-methylfuran reacts with NO₃ radicals instead of O₃ and to enhance the RO₂ + NO₃ channel, the concentration ratio of NO₂/O₃ is kept at $\sim 5:3$ and the 3-methylfuran/N₂O₅ ratio is targeted to be 1:4.

Gas-phase Measurements. The 3-methylfuran concentration is monitored using a gas chromatograph-flame ionization detector (GC-FID, Agilent 7890A) with a PLOT-Q column (Agilent). Oxidized reaction products are measured and reported as counts per second using a high-resolution time-of-flight chemical ionization mass spectrometer (HR-ToF-CIMS, Aerodyne Research Inc.) with iodide (I⁻) as a reagent ion, which selectively measures polar or acidic compounds.^{27–30} O₃ and NO_x are monitored using an O₃

Analyzer (Teledyne T400) and an ultrasensitive chemiluminescence NO_x monitor (Teledyne 200 EU), respectively.

Particle-phase Measurements. Size-dependent particle number and volume concentrations are monitored using a scanning mobility particle sizer (SMPS). The SMPS consists of a differential mobility analyzer (DMA) (TSI 3040) and a Condensation Particle Counter (CPC) (TSI 3775). A high-resolution time-of-flight aerosol mass spectrometer (HR-ToF-AMS, Aerodyne Research Inc.) measures bulk elemental composition of the particles (e.g., O/C and H/C ratios) and quantifies organics, nitrate, sulfate, ammonium, and chloride mass concentrations.^{31,32} A Filter Inlet for Gases and AEROSol (FIGAERO, Aerodyne Research Inc.) inlet system is coupled with the HR-ToF-CIMS to detect particle-phase molecular composition.^{26,33} FIGAERO allows particles to be collected onto a PTFE filter (Pall Corp, Zefluor 25 mm, 2 μm pore-size) for 20 min while measuring gas-phase species with the HR-ToF-CIMS. The particle-phase composition is analyzed through thermal desorption of the collected material (40 min total, including temperature ramping, soaking, and cooling cycle). FIGAERO-HR-ToF-CIMS provides insights into particle composition at the molecular level and aerosol volatility.^{33,34} The volatility distribution of bulk organic aerosol, as saturation mass concentration (C*, $\mu\text{g m}^{-3}$), is estimated based on the method used by Stark et al.³⁴ Briefly, a desorption temperature corresponding to the signal peak (T_{max}) is calibrated using compounds with known vapor pressures. The FIGAERO peak shape of the calibrated compounds is used to deconvolute the FIGAERO signal of the bulk aerosol using nonlinear multipeak fitting and to obtain the volatility distribution of the collected aerosol (Figure S-1).

■ RESULTS

SOA Formation from NO₃ Radical Oxidation of 3-Methylfuran. A series of experiments with different initial 3-methylfuran concentrations is performed to investigate the SOA formation potential of this system (Table 1). The initial 3-methylfuran concentrations ranges from 95.9 to 562.8 ppb. SOA formation is observed in all experiments and the particle wall loss-corrected³⁵ SOA mass yield (Y) ranges from 1.62% to 2.4% for an organic aerosol mass concentration (ΔM_0) of 5.1 to 45 $\mu\text{g}/\text{m}^3$ (Figure 1). A nucleation experiment is also conducted to determine SOA density. By comparing SMPS volume distribution and AMS mass distribution,³⁶ the SOA density is determined to be 1.24 g cm^{-3} . Although particle wall loss is corrected, the SOA yield reported here could represent the lower limit owing to vapor wall loss, where low volatility organic vapors could have partitioned to the chamber wall rather than to the aerosol.^{37–41}

There was no prior study on SOA formation from 3-methylfuran + NO₃ but two previous studies reported SOA formation from 3-methylfuran + OH reaction. Alvarez et al.¹⁷ reported a SOA yield of $8.5 \pm 2.5\%$ in the presence of 48 $\mu\text{g}/\text{m}^3$ of ΔM_0 . Strollo and Ziemann¹⁸ reported a similar average yield of 10%, but that corresponded to 2000–5000 $\mu\text{g}/\text{m}^3$ of ΔM_0 implying a much lower SOA yield than that reported by Alvarez et al.¹⁷ at 48 $\mu\text{g}/\text{m}^3$ of ΔM_0 . In the context of these previous studies, we find that the SOA yield from NO₃ radicals oxidation is about a factor of 4 lower than the SOA yield from photooxidation reported by Alvarez et al. at similar ΔM_0 values. In Figure 1a, SOA yield is parametrized as a function of organic mass produced using two types of fit: (1) a semiempirical model based on gas-to-particle partitioning of

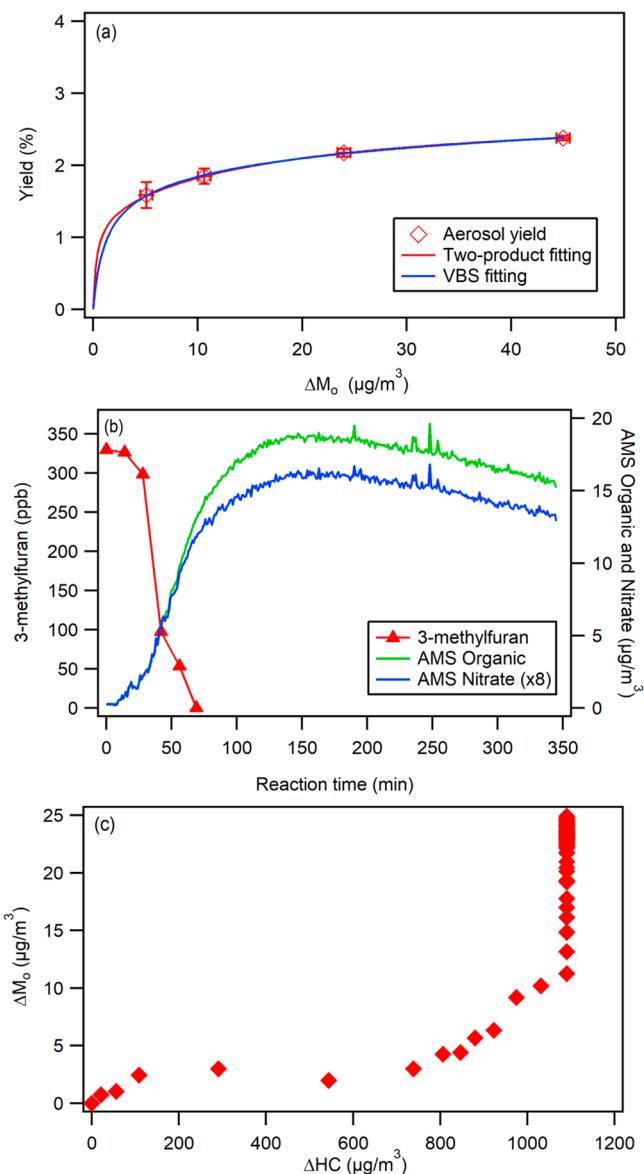


Figure 1. SOA yield and time profiles of 3-methylfuran oxidation and aerosol formation. (a) SOA yield and yield curve fitting using two-product model and VBS. (b) Reaction profile of an initial 3-methylfuran concentration of 328.5 ppb (Experiment #2 in Table 1). (c) Time-dependent growth curve. ΔM_o is calculated by the volume concentration measured with SMPS and SOA density, and ΔHC is obtained by interpolating GC-FID measurement to fit the time resolution between two measurements.

two semivolatile products^{42,43} and (2) the volatility basis set (VBS) (eqs 1 and 2, respectively).⁴⁴ The fitting coefficients for both approaches are shown in Table 2.

$$Y = \Delta M_o \left[\frac{\alpha_1 K_1}{1 + K_1 M_o} + \frac{\alpha_2 K_2}{1 + K_2 M_o} \right] \quad (1)$$

$$Y = \Delta M_o \sum_i \frac{\alpha_i}{C_i^* + M_o} \quad (2)$$

Typical temporal profiles of the oxidation of 3-methylfuran are shown in Figure 1b,c. In addition to organics measured by the HR-ToF-AMS, the concentration of nitrate also increases steadily over the course of the experiment (Figure 1b). Since

Table 2. Fit Parameters of Two-Product Model and Volatility Basis Set^a

(1) Two-product fitting				(2) VBS fitting		
α_1	K_1	α_2	K_2	α at C^* = 1	α at C^* = 10	α at C^* = 100
0.015	0.064	0.013	3.15	0.016	0.0067	0.01

^a α = gas-phase mass fraction, unitless. K = partitioning coefficient, $\text{m}^3 \mu\text{g}^{-1}$. C^* = saturation vapor pressure, $\mu\text{g m}^{-3}$

inorganic nitrate formation via hydrolysis of N_2O_5 is not expected owing to the low RH during the course of the experiment, the observed increase can be attributed to the formation of organic nitrates. Figure 1c shows the time-dependent “growth curve” (ΔM_o as a function of hydrocarbon consumed, ΔHC).^{45,46} The vertical section indicates that more than half of the SOA is formed after all the 3-methylfuran is consumed. These results demonstrate that there is a substantial contribution from multigenerational oxidation and/or heterogeneous/multiphase reactions to SOA formation.

Bulk Composition of SOA. A typical HR-ToF-AMS aerosol mass spectrum is shown in Figure S-2. The signals at m/z 39 (C_3H_3^+), m/z 68 ($\text{C}_4\text{H}_4\text{O}^+$), m/z 69 ($\text{C}_4\text{H}_5\text{O}^+$), m/z 82 ($\text{C}_5\text{H}_6\text{O}^+$), m/z 97 ($\text{C}_5\text{H}_5\text{O}_2^+$), and m/z 44 (CO_2^+) are particularly high. The signal at m/z 82 is a well-known isoprene SOA signature as determined by positive matrix factorization (PMF) analysis of AMS data.^{47–53} Here, m/z 82 constitutes 1% of total organics (Figure S-2). Structural similarities between 3-methylfuran + NO_3^- products and isoprene SOA could be the reason for the observed enhancement in m/z 82. Specifically, Lin et al.⁴⁹ proposed that 3-methyltetrahydrofuran-3,4-diols associated with IEPOX-SOA can form methylfuran-like structures when decomposed in the AMS. The enhancement of m/z 44 (CO_2^+) can be an indicator of the presence of organic acids in the SOA.^{54–57} HR-ToF-AMS nitrate-to-organics ratio is 0.11, where nitrate mainly comes from NO^+ and NO_2^+ fragments. The $\text{NO}^+/\text{NO}_2^+$ (NO_x^+) ratio ranges from 5.1 to 5.4 throughout all the experiments in this study, which is within the range that indicates the presence of organic nitrates as reported in previous studies.^{25,51,58–62}

Speciated Gas- and Particle-phase Chemical Composition. The mass spectrum of gas- and particle-phase compounds obtained from the HR-ToF-CIMS at the peak aerosol growth is shown in Figure 2, where both organic nitrates and non-nitrate organic species are detected. $\text{C}_5\text{H}_5\text{NO}_3\text{I}^-$ (m/z 286) is the strongest signal among the 18 major gas species, around 20 times higher than the other compounds (Figure 2a). $\text{C}_5\text{H}_6\text{O}_3\text{I}^-$ (m/z 241) is the next most abundant gas species, which is 5% of the $\text{C}_5\text{H}_5\text{NO}_3\text{I}^-$ signal, and the others are $\leq 2\%$ of the $\text{C}_5\text{H}_5\text{NO}_3\text{I}^-$ signal. In the particle phase, however, $\text{C}_5\text{H}_5\text{NO}_3\text{I}^-$ is a minor compound while $\text{C}_5\text{H}_6\text{O}_3\text{I}^-$ is the highest among the 224 organic ions detected. Compounds with five or fewer number of carbons are mostly distributed below 330 m/z (green region in Figure 2b), and compounds with six or higher number of carbons are mostly distributed over 330 m/z (blue region in Figure 2b), which are classified as the oligomer region. The six major species in each region (Figure 2b-i,b-ii) account for 68% of the total signal of FIGAERO-identified species in the particle phase. The particle-phase ions detected by the FIGAERO-HR-ToF-CIMS could potentially include the fragments resulting from thermal decomposition of particles.^{63,64} Therefore,

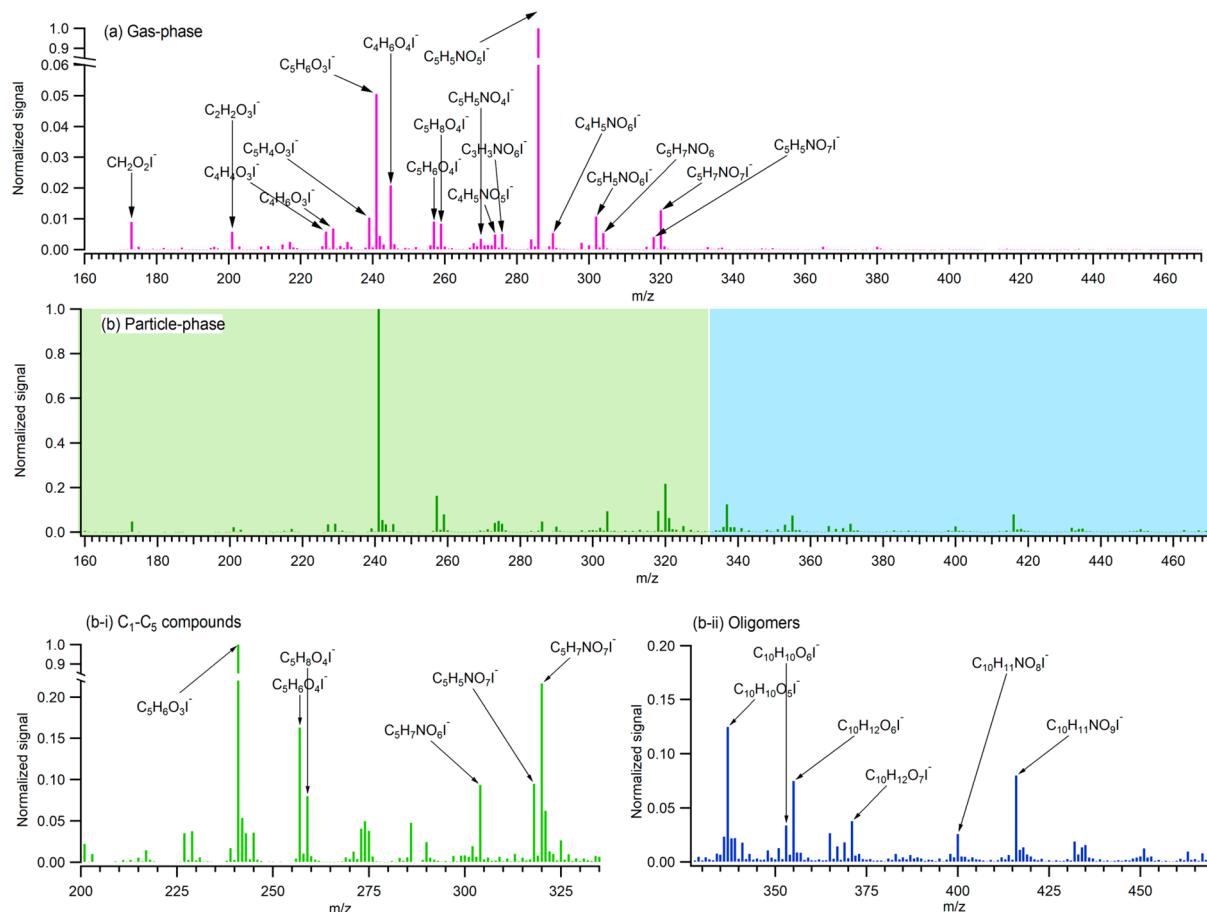


Figure 2. HR-ToF-CIMS mass spectra of (a) gas-phase and (b) particle-phase for a typical experiment (Experiment #2 in Table 1). The m/z include the mass of the I^- ion (i.e., $m/z = 126.905$). (b-i) Mass spectrum of $\text{C}_1\text{--C}_5$ compounds (green box region in (b)), and (b-ii) mass spectrum of oligomers in the particle phase (blue box region in (b)).

specific types of C_5 compounds detected can correspond to parent molecules and/or thermal decomposition of higher molecular weight oligomers.

Figure 3 shows the time evolution of selected gas- and particle-phase species measured by the FIGAERO-HR-ToF-CIMS. Gas-phase compounds detected by the HR-ToF-CIMS showed two different patterns in the time traces. $\text{C}_5\text{H}_5\text{NO}_5\text{I}^-$ and $\text{C}_5\text{H}_7\text{NO}_6\text{I}^-$ in Figure 3a increase rapidly once N_2O_5 is injected and stay constant after all the 3-methylfuran is consumed, which is characteristic of first-generation gas-phase oxidation products. The signal for other species increases slowly and steadily until the end of the experiment (Figure 3b), which is characteristic of later generation products. $\text{C}_5\text{H}_6\text{O}_3\text{I}^-$ is an exception as seen in Figure 3b, where the signal is stabilized after all the 3-methylfuran is consumed. The time series data suggest that this compound is a multigenerational product: (1) Unlike $\text{C}_5\text{H}_5\text{NO}_5\text{I}^-$ and $\text{C}_5\text{H}_7\text{NO}_6\text{I}^-$ (first-generation products), $\text{C}_5\text{H}_6\text{O}_3\text{I}^-$ does not show an immediate increase upon onset of oxidation and (2) it continues to increase after N_2O_5 injection is stopped. These two observations are consistent with the proposed mechanism that $\text{C}_5\text{H}_6\text{O}_3$ is a second-generation product (see discussion in later sections). The six major C_5 and C_{10} particle-phase compounds show similar time evolutions as the bulk aerosol mass concentration (Figure 3c,d). In general, C_{10} compounds reach the maximum with a faster rate than the most of the C_5 compounds. For both C_5 and C_{10} compounds, the non-nitrate

organics increase at a faster rate than the nitrate compounds. Since low-volatility species will readily partition into the particle phase,^{43,44,65} this observation suggests that non-nitrate organic compounds could be less volatile than nitrate compounds.

Saturation Vapor Pressure Evolution and SOA Contribution by Carbon Number. The time evolution of FIGAERO thermograms for SOA products with different carbon number is shown in Figure 4. The thermograms of $\text{C}_1\text{--C}_5$, $\text{C}_6\text{--C}_{10}$, and $\text{C}_{11}\text{--C}_{18}$ compounds show a similar pattern throughout the experiment. $\text{C}_1\text{--C}_5$ compounds contribute most of the total particle-phase signal in Figure 4, and $\text{C}_5\text{H}_6\text{O}_3$ (green triangles) alone contributes ~40% of the $\text{C}_1\text{--C}_5$ signal throughout the experiment. All thermograms in Figure 4 show a multiple-peak pattern, indicating the presence of both monomers and oligomers in SOA.^{33,64,66} The multiple-peak pattern is more obvious at $t = 20$ min. The T_{\max} of the largest peak appears at ~ 70 °C during the first FIGAERO cycle ($t = 20$ min), shifts to ~ 110 °C at the maximum aerosol mass ($t = 140$ min), and remains the same until the end of the experiment ($t = 320$ min). The evolution of the SOA volatility distribution is also shown in Figure 4. At $t = 20$ min, the SOA includes both semivolatile and low-volatility species. The volatility distribution of SOA shows a peak at $\text{C}^* = 10^{-1} \mu\text{g m}^{-3}$ and is fairly evenly spread out across other bins. The peak shifts to $\text{C}^* = 10^{-4} \mu\text{g m}^{-3}$ at $t = 140$ min (maximum aerosol mass) and stays the same afterward.

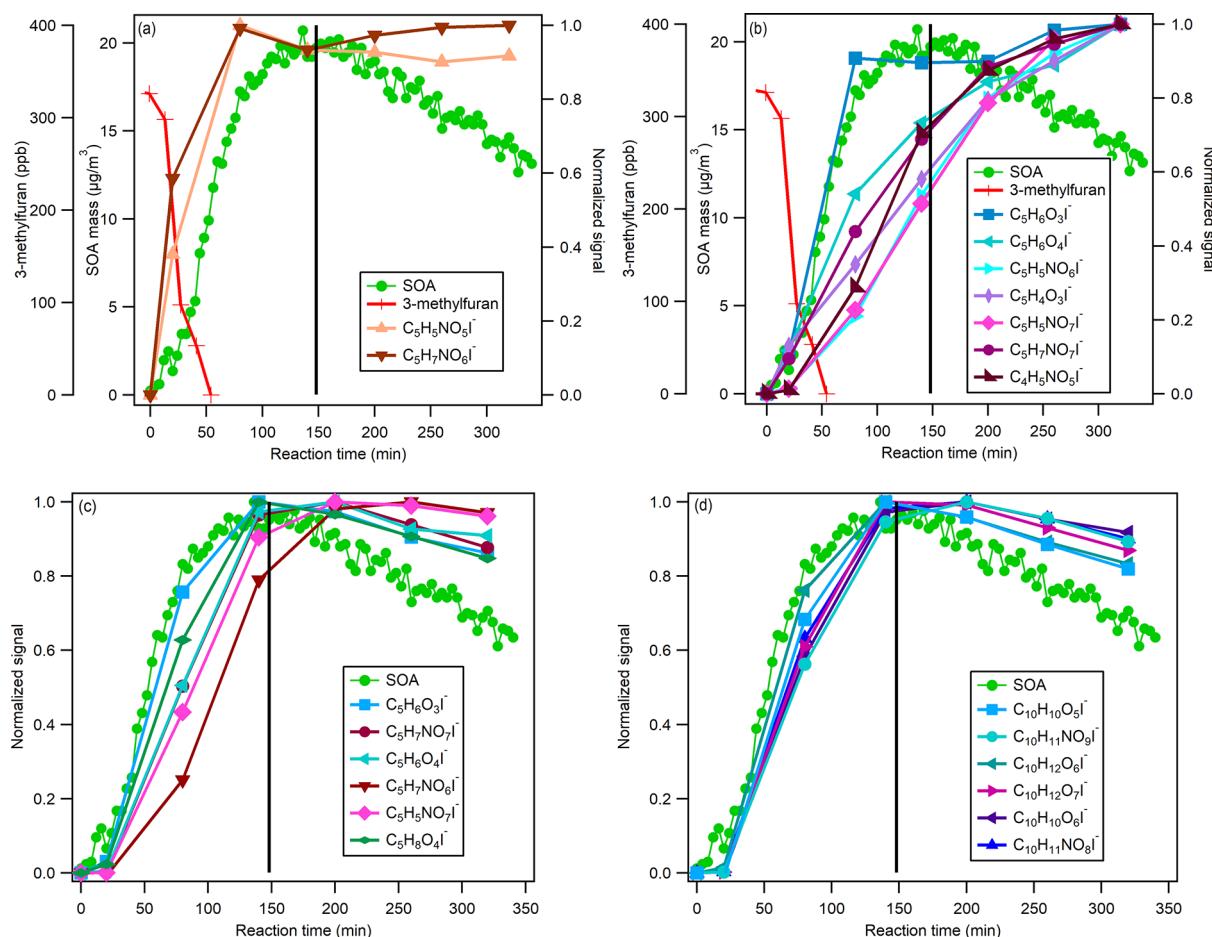


Figure 3. Time series of gas- and particle-phase species from HR-ToF-CIMS (Experiment #2 in Table 1). The top two panels are gas-phase time series and the bottom two panels are particle-phase time series. (a) Time series of first-generation products and (b) Time series of later generation products. (c) Time series of major C_5 compounds and (d) Time series of major C_{10} compounds. The time series of 3-methylfuran and SOA formation (calculated from SMPS volume concentration and SOA density) are also shown for reference. The black line represents the time when N_2O_5 injection is terminated. The signals are normalized by the maximum signal of each species throughout the experiment.

DISCUSSION

Mechanism of 3-Methylfuran Oxidation by NO_3 Radical. Figure 5 shows a proposed mechanism for the reaction of 3-methylfuran with NO_3 radicals based on the observed oxidation products. The formation of first- and higher-generation products in the proposed mechanism are consistent with the time evolutions of the species shown in Figure 3. The reaction starts with the addition of $-ONO_2$ to the 2 or 5 position, which are favored due to the stability of the resulting radical.^{18,24,67} The cyclic nitrooxyalkyl radical can either react with O_2 to form a cyclic nitrooxyperoxy radical (RO_2 , PR1) or lose NO_2 and become 2-methylbutenedial (1), a first-generation product that has been shown to form in high yield (38–83%) from both OH radical and NO_3 radical oxidations.^{17,24,68,69} The other first-generation products 2, 3, and 4 proposed in Figure 5 are formed via reaction of PR1. PR1 reacts with HO_2 to form nitrooxy hydroperoxide ($R(ONO_2)OOH$, C₅H₇NO₆) (4) or reacts with $NO_3/HO_2/RO_2$ to form nitrooxyalkoxy radical ($R(ONO_2)O$, AR1). Tapia et al.²⁴ proposed the formation of peroxydinitrate ($R(ONO_2)-(OONO_2)$), nitrooxy carbonyl ($R(ONO_2)(O)$), and nitrooxy alcohol ($R(ONO_2)(OH)$) from AR1, but these reaction pathways are not included owing to their low signals in the HR-ToF-CIMS (Figure 2a) or their slow formation rates

relative to other pathways as shown in Figure S-4. AR1 can undergo further reaction to form C₅H₅NO₅, which is the major gas-phase product detected by the HR-ToF-CIMS (Figure 2a). There are two possible isomers of C₅H₅NO₅ from AR1: ring-opened nitrooxy dicarbonyl ($R(ONO_2)(O)(O)$) (2) and cyclic nitrooxy carbonyl ($R(ONO_2)(O)$) (3) via decomposition and reaction of AR1 with O_2 , respectively. The formation of isomer 2 is expected to be favored because the decomposition pathway is estimated to be eight orders of magnitude faster than the reaction with O_2 (Figure S-4).

A suite of second-generation products is formed from reactions of compounds 1 and 2. Compound 1 reacts with NO_3 and forms acylperoxy radical ($R(O)O_2$, PR2) or tertiary nitrooxyperoxy radical (PR3) via H-abstraction from aldehyde or addition to the C=C double bond, respectively. Among these two pathways, the formation of PR2 will be favored because the aldehyde groups reduce the reactivity of the double bond and Rayez et al.⁷⁰ reported that the H-abstraction from an aldehyde is always dominant relative to the addition to the double bond in an alkenal. Further oxidation of compound 2 results in the formation of nitrooxy carbonyl acylperoxy radical (PR4). In general, the fates of the PR1–4 depend on their reaction partners: $RO_2 + NO_3$, $RO_2 + RO_2$, $RO_2 + HO_2$, $RO_2 + NO_2$, $RO_2 + NO$, and RO_2 isomerization. Under the reaction conditions in this study, the $RO_2 + NO_3$ channel is

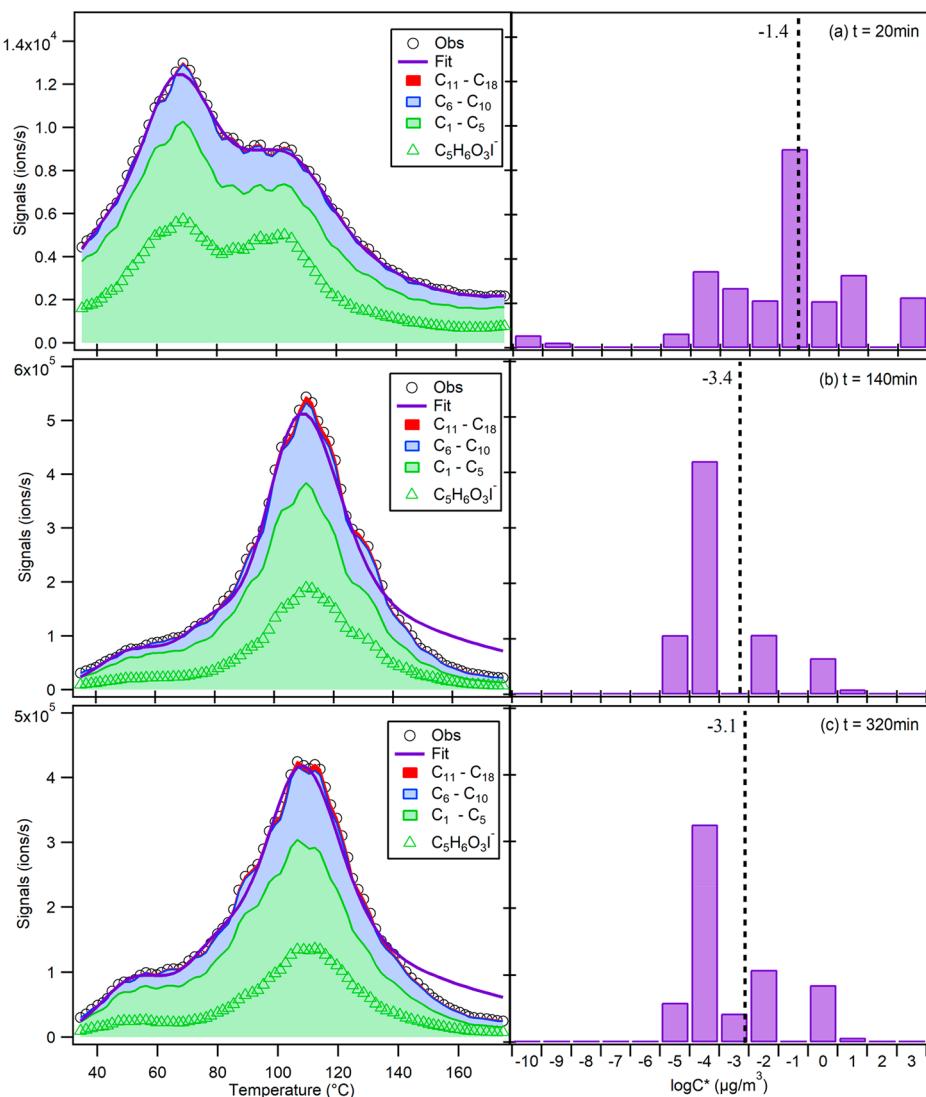


Figure 4. Time evolution of thermogram for total SOA, C_1 – C_5 compounds, C_6 – C_{10} compounds, and C_{11} – C_{18} compounds over the course of a typical experiment (Experiment #2 in Table 1). The dotted line represents the log-weighted averaged C^* . (a) Beginning of the oxidation. (b) Peak aerosol growth. (c) End of the experiment. Saturation vapor pressure is calculated from the FIGAERO calibration. The difference between observation and fitting (tail) in the middle and bottom panels is due to the shape of thermogram obtained during FIGAERO calibration (Figure S-1). Signals at $\log C^*$ of -10 and $-9 \mu\text{g}/\text{m}^3$ at $t = 20$ min are due to the interference from the background.

expected to dominate, though $\text{RO}_2 + \text{RO}_2$ and $\text{RO}_2 + \text{HO}_2$ reactions and the corresponding products are also detected by the HR-ToF-CIMS.

In Figure 5, compounds 5–12 are formed via further reactions of PR2 or PR3, and compounds 13 and 14 are formed via further reactions of PR4. PR2 can react with NO_2 to form organic acyl peroxy nitrate ($\text{R}(\text{O})\text{OONO}_2$, $\text{C}_5\text{H}_5\text{NO}_6$) (5) and with HO_2 to form peroxy acid ($\text{R}(\text{O})\text{OOH}$, $\text{C}_5\text{H}_6\text{O}_4$) (6) or organic acid (RCOOH , $\text{C}_5\text{H}_6\text{O}_3$) (7). Compound 7 can also be produced through $\text{RO}_2 + \text{RO}_2$ reactions.^{71,72} PR2 can also react with NO_3 , HO_2 , or RO_2 to form acylalkoxy radical ($\text{R}(\text{O})\text{O}$, AR2). AR2 can then decompose to form cyclic dicarbonyl ($\text{R}(\text{O})(\text{O})$, $\text{C}_5\text{H}_4\text{O}_3$) (8) or isomerize to form carboxylic acyl peroxy radical (PR5) that reacts further to form dicarboxylic acid ($\text{R}(\text{COOH})\text{COOH}$, $\text{C}_5\text{H}_6\text{O}_4$) (9) and carboxylic peroxyacetyl nitrate ($\text{R}(\text{COOH})(\text{O})\text{OONO}_2$, $\text{C}_5\text{H}_5\text{NO}_7$) (10). Among these two reaction branches, it is expected that the isomerization pathway would be favored because the H-abstraction rate from the aldehyde is two orders

of magnitude faster than the decomposition rate (Figure S-4). As for PR3, it can react with HO_2 to form nitrooxy hydroperoxide (11) or with $\text{RO}_2/\text{NO}_3/\text{HO}_2$ to form nitrooxy alkoxy radical (AR3) and subsequently decompose to C_4 nitrooxy dicarbonyl ($\text{R}(\text{ONO}_2)(\text{O})(\text{O})$, $\text{C}_4\text{H}_5\text{NO}_5$) (12). Lastly, further reactions of PR4 lead to the formation of nitrooxy peracetic acid ($\text{R}(\text{ONO}_2)(\text{O})\text{OOH}$, $\text{C}_5\text{H}_5\text{NO}_7$) (13) and nitrooxy carboxylic acid ($\text{R}(\text{ONO}_2)\text{COOH}$, $\text{C}_5\text{H}_5\text{NO}_6$) (14).

Multigenerational Chemistry in SOA Formation. As seen in Figure 1b,c, there is continuous SOA formation after the complete depletion of 3-methylfuran. The particle mass spectrum measured by FIGAERO-HR-ToF-CIMS indicates the presence of dimers (Figure 2) and various gas- and particle-phase species continue to increase over the course of the experiment (Figure 3). The continued SOA formation can be due to gas-particle partitioning of higher-generation gas-phase products and/or further particle-phase reactions. We highlight the species $\text{C}_5\text{H}_6\text{O}_3$ here as it is the most abundant

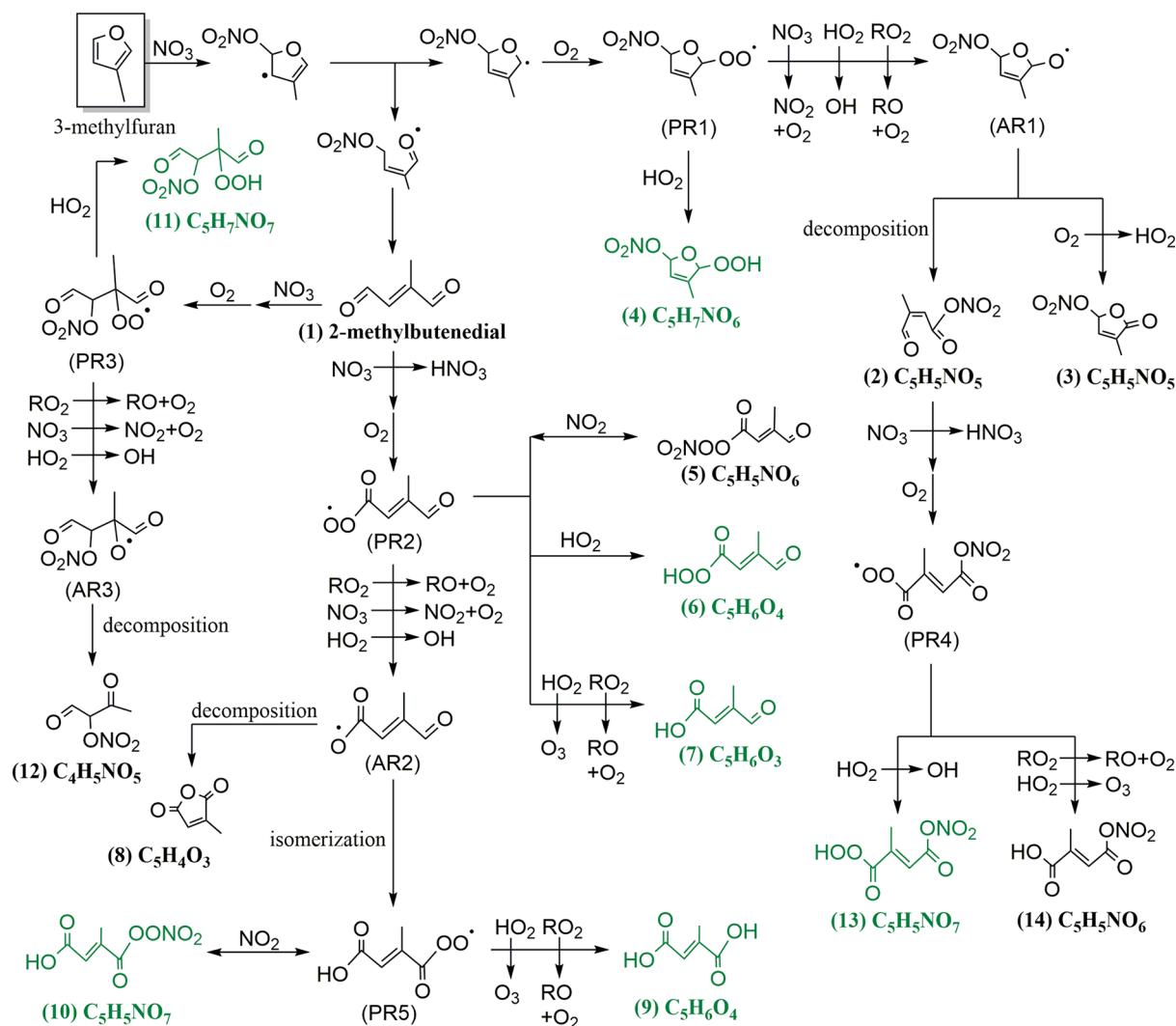


Figure 5. Proposed mechanism for 3-methylfuran oxidation by NO_3 radicals. Selected particle species in Figure 2 is colored in green. NO_3 can be added to 2 or 5 position, but addition to 5 position is shown here as the representative case.

species detected in the particle phase. There are multiple pathways that could lead to the formation of $\text{C}_5\text{H}_6\text{O}_3$, which include multigenerational oxidation and particle-phase reaction: (1) further gas-phase oxidation of the first-generation product 2-methylbutenedial, (2) decomposition from oligomers, and (3) functional group conversion through particle-phase reactions. The 2-D thermogram⁷³ in Figure S-3 shows similar T_{max} for species with a wide range of molecular weights, which could indicate that lower molecular weight compounds are thermal decomposition products of high molecular weight compounds. For instance, $\text{C}_5\text{H}_6\text{O}_3$, which is detected as the major compound in the particle phase, is estimated to have $\text{C}^* = 10^{4.4} \mu\text{g m}^{-3}$ by simple group contribution method (SIMPOL),⁷⁴ which is much more volatile than the C^* value estimated from the thermogram in Figure 4 ($\text{C}^* = 10^{-4} \mu\text{g m}^{-3}$). It is likely that the C^* estimated by the group contribution method is biased high and that $\text{C}_5\text{H}_6\text{O}_3$ arises from thermal decomposition of lower volatility oligomers during the FIGAERO cycle.³⁴

We propose a set of particle-phase accretion reactions from the six C_5 compounds with the highest signal intensities that can lead to the formation of C_{10} dimers (Figure S-5). The particle-phase accretion reactions require functionalized

compounds, which are either 2-methylbutenedial or its reaction products as shown in Figure 5.^{75,76} The C_{10} compounds can also be generated via gas-phase ROOR' formation ($\text{RO}_2 + \text{RO}_2$) from acylperoxy radicals. Although it has been reported that ROOR' formation is uncertain⁷⁷ and is a minor pathway,⁷⁸⁻⁸⁰ the branching ratio of ROOR' formation could be higher here because the larger RO_2 in this study would have more vibrational modes that can distribute collisional energy to prevent breakage of ROOR' .⁸¹ All proposed dimers in Figure S-5 share similar backbone with $\text{C}_5\text{H}_6\text{O}_3$, which could lead to the enhancement in $\text{C}_5\text{H}_6\text{O}_3$ signal during the thermal desorption cycle of the FIGAERO as dimers break down to from $\text{C}_5\text{H}_6\text{O}_3$. The proposed dimers can also further react and undergo Baeyer–Villiger reactions that can lead to the formation of carboxylic acids,⁸² which could contribute to $\text{C}_5\text{H}_6\text{O}_3$ signal.

The time evolution of the thermograms of particle-phase species (Figure 4) also indicate that multigenerational or multiphase reactions are contributing to SOA formation. As the variation in the shape of the thermogram is related to the evolution in the saturation vapor pressure of SOA, this indicates that volatile species ($\text{C}^* \approx 10^{-1} \mu\text{g m}^{-3}$) are contributing to SOA formation at the beginning of the

experiment, and further reactions continue to generate less-volatile species ($C^* \approx 10^{-4} \mu\text{g m}^{-3}$) over the course of the experiment. This is in agreement with the changes in O/C ratio measured by the HR-ToF-AMS in Figure S-6. It has been reported that O/C ratio and volatility of SOA are anticorrelated in general,^{83–86} thus increasing O/C ratio over the course of the experiment is consistent with the presence of more low-volatility species in SOA over time. After peak aerosol growth, however, the thermogram shows relatively small changes (slight decrease of signal and peak broadening) until the end of the experiment (Figure 4c). The O/C ratio also shows relatively small changes compared to the variation before reaching the peak aerosol growth (Figure S-6). This may indicate that the aging processes over the course of the experiment do not alter bulk SOA characteristics and O/C to a large extent.

Organic Nitrates Formation. As mentioned in the previous section, the HR-ToF-AMS mass spectrum in Figure S-2 indicates the presence of nitrate species. Farmer et al.⁵⁹ suggested that NO_x^+ is higher for organic nitrates than inorganic nitrates. In all the furan experiments in this study, the NO_x^+ for organic nitrates (R_{ON}) ranges from 5.1 to 5.4, which is higher than the NO_x^+ that is measured by atomizing pure ammonium nitrate into the AMS ($R_{\text{AN}} = 3.8$) in our study. The R_{ON} observed in the furan experiments is comparable to the isoprene + NO_3 system ($R_{\text{ON}} = 5$) but smaller than the R_{ON} of monoterpene + NO_3 system ($R_{\text{ON}} = 6.5\text{--}15$).^{25,26,58,60,62} However, the corresponding $R_{\text{ON}}/R_{\text{AN}}$ ratio (~1.4 in this study), which was proposed by Fry et al.⁸⁷ to constrain instrumental bias, is lower than both isoprene + NO_3 ($R_{\text{ON}}/R_{\text{AN}} = 2.1$) and monoterpene + NO_3 ($R_{\text{ON}}/R_{\text{AN}} = 3.7\text{--}4.2$) systems in previous studies.^{25,60}

We estimate the fraction of particle-phase organic nitrates in organics to be 39.4% using the observed nitrate-to-organics mass ratio (0.11) measured by the HR-ToF-AMS. The molecular weight of particle-phase organic nitrates needs to be considered in order to estimate the fraction of organic nitrates in organic aerosol. Based on the speciated particulate organic nitrates measured by the FIGAERO-HR-ToF-CIMS, the signal-weighted average organic nitrate molecular weight is approximated to be 222 g mol⁻¹.

The fraction of organic aerosol composed of organic nitrates is generally lower than previously reported values from monoterpene + NO_3 (56–80%).^{25,26,61,88} However, since furans are highly reactive toward NO_3 radicals,^{11,12} this could indicate that furan + NO_3 reactions during biomass burning can be an important source of particulate organic nitrates, which is a NO_x reservoir in the atmosphere.⁶¹

Volatility of SOA. As shown in the FIGAERO thermogram in Figures 4 and S-7, about 55% of the SOA at $t = 140$ min and $t = 320$ min has a $C^* \leq 10^{-4} \mu\text{g m}^{-3}$. This observed volatility range corresponds to extremely low-volatility organic compounds (ELVOCs) or low-volatility organic compounds (LVOCs), which can readily partition into the aerosol.^{89,90} The volatility of the total 3-methylfuran + NO_3 SOA is lower than that of the ambient biomass burning OA (BBOA), which is reported to mostly consist of semivolatile organic compounds (SVOCs) or LVOCs.^{91–94} Ambient BBOA can consist of primary and/or secondary OA, which can have different volatilities. The lower volatility in this study is likely due to the higher contribution from secondary chemistry to OA concentration. We believe C_{15} or higher compounds contribute to SOA composition since the C^* of selected C_5

and C_{10} compounds, estimated using SIMPOL,⁷⁴ ranges from 10^2 to $10^5 \mu\text{g m}^{-3}$ and 10^{-2} to $10^1 \mu\text{g m}^{-3}$, respectively. This is much more volatile than the observed volatility distribution for C_5 compounds as shown.

The yield-based VBS fitting parametrizes mass fraction at $C^* = [1, 10, 100 \mu\text{g m}^{-3}]$ (Figure 1a and Table 2), which is different from the FIGAERO-based values in Figure S-7 and Table S-1 (C^* from 10^{-5} to $10^1 \mu\text{g m}^{-3}$). We note that the VBS-fitting corresponds to data from all experiments, which include compounds of differing volatility in experiments with lower and higher organic loadings, while the FIGAERO-based estimation uses data from one specific experiment (Experiment 2 in Table 1). Nevertheless, one can still obtain some insights from the difference in aerosol volatility derived using VBS-fitting vs FIGAERO. In general, the mass concentrations in the higher-volatility bins (10 and $100 \mu\text{g m}^{-3}$) from the yield-based VBS fitting are higher than those from FIGAERO-based estimations. It is possible that this is a result of FIGAERO preferentially detecting compounds of relatively lower volatility with the use of I^- as reagent ion. We find that the mass concentration that corresponds to the least volatile bin $C^* = 1 \mu\text{g m}^{-3}$ ($17.64 \mu\text{g m}^{-3}$, Table 2) from VBS-fitting is not substantially different from the integrated mass from C^* of 10^{-10} to $10^{-3} \mu\text{g m}^{-3}$ ($16.97 \mu\text{g m}^{-3}$, Figure S-7 and Table S-1) in FIGAERO data. Thus, it appears that ELVOC and LVOC masses measured by FIGAERO have been allocated into the $C^* = 1 \mu\text{g m}^{-3}$ bin in the yield-based VBS fitting. Our observations are consistent with Saha and Grieshop,⁸⁵ who reported that yield-based VBS fitting may overestimate the contribution from SVOCs compared to thermal desorption method and lead to underestimation of the SOA mass yield particularly at low organic mass loading.

The volatility distributions of particulate organic nitrates and non-nitrate organics are compared in Figure S-8. The C^* of organic nitrates is more evenly distributed across all volatility bins, especially with more contribution from higher volatility bins compared to non-nitrate organics. About 55% and 66% of organic nitrates and non-nitrate organics have $C^* \leq 10^{-4} \mu\text{g m}^{-3}$, respectively. Overall, the average C^* of organic nitrates and non-nitrate organics is $10^{-2.9}$ and $10^{-3.3} \mu\text{g m}^{-3}$, respectively. Such volatility difference is consistent with the time evolutions in Figure 3c,d where nitrated compounds reach peak aerosol growth at a later time than non-nitrate organic compounds.

A few prior studies have reported on the volatility of organic nitrates. The volatility of bulk organic nitrates from 3-methylfuran + NO_3 is generally lower than organic nitrates from β -pinene + NO_3 oxidation (~ $10^{1.7} \mu\text{g m}^{-3}$).⁵⁸ However, comparison of the volatility of particle-phase organic nitrates should be carefully considered since the volatility in Fry et al.⁵⁸ was estimated using a partitioning and group contribution method, which is different from this study. Field studies reported that polymerized organic nitrates could exist in the low-volatility region,⁹⁶ but further studies are required to understand the underlying chemistry leading to these species.

■ ATMOSPHERIC IMPLICATIONS

To our knowledge, this work is the first study to demonstrate SOA and particle organic nitrates formation from the 3-methylfuran + NO_3 reaction. There have been a few studies focusing on SOA formation from gas-phase oxidation of biomass burning emissions but studies that reported nighttime oxidation of biomass burning precursors are much more

limited even though nighttime oxidation also has the potential to form SOA. Hartikainen et al.¹¹ observed an inverse correlation between furans and organic aerosol mass in the flame chamber experiments and proposed that furans can potentially be a major group of SOA precursors. Here, we report a series of chamber experiments with an initial 3-methylfuran concentration of 95.9 to 562.8 ppb under dry conditions at room temperature. SOA yield ranges from 1.6 to 2.4% for 5–45 $\mu\text{g m}^{-3}$ of organic aerosol mass. Previous studies reported that furans can account for 10–50% of the potential SOA formed during daytime oxidation of biomass burning plume when an assumed 10% SOA yield is applied for all furans.^{18,97} Considering the relatively large emission rates of furans during biomass burning, a 2% SOA yield from this study suggests that furans can potentially be an important nighttime SOA precursor. If typical nighttime RH conditions are considered, which are higher than in this study, the SOA yield may increase owing to oligomerization or reactive uptake of water-soluble compounds.^{48,75,98} Future studies are warranted to investigate SOA formation from furan oxidation under various reaction conditions.

Higher-generation oxidation products and possibly multiphase accretion reactions contribute significantly to SOA formation from the 3-methylfuran + NO_3 reaction. There are two lines of evidence to support this: (1) SOA continues to increase after 3-methylfuran is completely oxidized and (2) the volatility bin that contributes the most to the SOA is $\text{C}^* = 10^{-4} \mu\text{g m}^{-3}$, which is much lower than the volatility expected from C_5 compounds. One of the main gas-phase products expected from the 3-methylfuran + NO_3 reaction is 2-methylbutenedial.^{17,24,69} The results from the HR-ToF-CIMS indicate that further oxidation of 2-methylbutenedial can contribute to SOA significantly. This is likely due to the formation of acylperoxy radicals when 2-methylbutenedial reacts with NO_3 . The slow N_2O_5 injection used in this study as well as the fact that these radicals have some of the fastest $\text{RO}_2 + \text{RO}_2$ reaction rates can lead to considerable dimer formation (ROOR') and multifunctional species. The aldehyde groups can also participate in accretion reactions or be oxidized to carboxylic acids in the condensed phase. The molecular formulas of the observed products imply that the compounds in the particle phase are multifunctional and likely to engage in accretion reactions. If other VOCs in biomass burning plumes have similar products, it is expected that these compounds will form SOA efficiently through multigeneration and multiphase chemistry.

3-Methylfuran + NO_3 reaction can impact O_3 formation as well. In addition to ROOR' formation, acylperoxy radicals can form acylperoxy nitrates via reaction with NO_2 , which can act as NO_x reservoirs during biomass burning.⁹⁹ Another NO_x reservoir that is formed during the 3-methylfuran + NO_3 oxidation are particle-phase organic nitrates, which contribute 39.4% of the organics. Both acylperoxy nitrates and particle organic nitrates can participate in NO_x recycling via thermal dissociation and hydrolysis/photochemical loss, respectively.^{25,26,61} Recent studies have shown that particulate organic nitrates are ubiquitous in the atmosphere.⁶¹ Considering 3-methylfuran is highly reactive with NO_3 radicals, this work highlights the potential importance of organic nitrates sources outside of traditional VOCs.

Although it has been reported in some previous studies that net OA mass addition in biomass burning plumes can be small, the degree of oxidation of aerosol has been found to increase with photochemical evolution, indicating the strong influence

and importance of secondary chemistry.^{100–105} The results in this study provide detailed nighttime chemistry and SOA formation from 3-methylfuran oxidation and improve our understanding of the chemical evolution of biomass burning plumes.

■ ASSOCIATED CONTENT

Supporting Information

The Supporting Information is available free of charge on the ACS Publications website at DOI: [10.1021/acsearthspacechem.9b00068](https://doi.org/10.1021/acsearthspacechem.9b00068).

FIGAERO volatility calibration curve, AMS organic mass spectrum, 2-D thermogram from FIGAERO-HR-ToF-CIMS, description of the mechanism and reaction rate of radicals, possible dimer formation reactions, variation of O/C ratio, organic mass concentration for each volatility bin, and thermogram of nitrate and non-nitrate organic compounds and their saturation vapor pressure estimation (PDF)

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Notes

The authors declare no competing financial interest.

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