

1    **Review: Molecular physiology of copepods - from biomarkers to transcriptomes**  
2    **and back again**

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9

10 **Abstract**

11 Planktonic copepods are a diverse and abundant group of small (~mm sized) aquatic  
12 animals that play a critical role in linking the base of the food chain with higher trophic  
13 levels. These invertebrates are a primary food source for marine fish larvae. Their  
14 ubiquitous presence is thus of vital importance for recruitment of fish stocks and also  
15 as promising live feed for finfish production in aquaculture. This paper reviews the  
16 application of molecular approaches to understanding copepod physiology, particularly  
17 in non-parasitic species. The review includes both targeted gene approaches and  
18 untargeted transcriptomic approaches, with suggestions for best practices in each case.  
19 Issues particularly relevant to studies of copepods include heterogeneity within species,  
20 morphologically cryptic species, experimental artifacts associated with sample  
21 handling, and limited annotation of copepod genes and transcripts. The emergence of  
22 high-throughput sequencing and associated increased availability of genomic and  
23 transcriptomic databases has presented a huge opportunity to advance knowledge of  
24 copepod physiology. The research community can leverage this opportunity through  
25 efforts to maintain or improve data accessibility, database annotation, and  
26 documentation of analytical pipelines.

27

28 **Key words**

29 best practices, biomarker, Copepoda, gene expression, non-model, review, RNA-seq,  
30 transcriptome

31

32 **1. Introduction**

33 Copepods (Subclass Copepoda) are a diverse and ecologically important group of  
34 crustaceans that reside freshwater, estuarine and marine environments. They inhabit the  
35 water column and the benthos, may be free-living or parasitic, and range from the  
36 intertidal to deep ocean basins. They have even been suggested as the most numerous  
37 multicellular organisms on earth (Walter and Boxshall, 2019). Several recent molecular  
38 studies have indicated that the traditional grouping of Crustacea is paraphyletic, and  
39 have suggested a new clade Pancrustacea or Tetraconata that includes hexapods (e.g.,  
40 insects; Dohle, 2001; Regier et al., 2005). The relationships of copepods to other  
41 lineages within the Pancrustacea is still under debate and is considered a matter of  
42 active research (Oakley et al., 2012; Regier et al., 2010; Rota-Stabelli et al., 2012).  
43 Copepoda is the second largest subclass in the Crustacea, including approximately  
44 12,000 copepod species described to date (Encyclopedia of Life, Accessed 8 June  
45 2018), with the greatest diversity in the marine environment (Boxshall and Defaye,  
46 2008; 11,443 species listed in the World Register of Marine Species, Accessed 8 June  
47 2018).

48 Within marine and freshwater food webs, small planktonic copepods serve a critical  
49 role linking phytoplankton and microzooplankton with larger predators (Sherr and  
50 Sherr, 2016; Turner, 2004; Zöllner et al., 2009). Thus, they channel energy from  
51 primary producers and the microbial loop up to higher trophic levels. Diel vertical  
52 migration of planktonic copepods is a significant conduit for the biological pump,  
53 which exports organic carbon below the euphotic zone (reviewed by Steinberg and  
54 Landry, 2017). Seasonal dormancy of many species enables efficient grazing of  
55 seasonally abundant phytoplankton populations, and within the Calanidae, creates an  
56 additional mechanism for export as lipids are respiration at depth over a prolonged period

57 (i.e., the "lipid pump"; Jónasdóttir et al., 2015). In coastal and freshwater ecosystems,  
58 many species produce quiescent or diapausing embryos that settle into the sediments,  
59 where they remain for months to years until hatching during favorable conditions  
60 (Holm et al., 2017). This "egg bank" enables species to adapt to seasonal variability,  
61 helps to smooth the effects of variable reproduction across years, and facilitates the  
62 coexistence of diverse species and genotypes (Hairston, 1996; Marcus et al., 1994).

63 In addition to their importance to natural ecosystems, copepods also have significance  
64 for aquaculture, due both to their impacts as parasites and to their promising  
65 contributions as a source of live feed. Copepods parasitize diverse species in the wild  
66 and within aquaculture. Among the parasitic species, caligid copepods ("sea lice",  
67 primarily in the *Lepeophtheirus* and *Caligus* genera) have infested salmonid cultures,  
68 where they can both compromise commercial production and spread to local wild  
69 populations. Treatments to reduce parasite load can have unintended consequences,  
70 such as the development of drug-resistant populations and impacts on non-target  
71 species (reviewed by Aaen et al., 2015). On the other hand, cultured free-living  
72 copepods can have significant benefits for aquaculture. Relative to more traditional  
73 food sources like brine shrimp and rotifers, copepod nauplii can result in increased  
74 survival and other quality metrics of fish larvae (Drillet et al., 2011; Nielsen et al.,  
75 2017). However, rearing copepods at high densities on a commercial scale requires  
76 optimization and is a subject of active research (Nilsson et al., 2017 and references  
77 therein; Vu et al., 2017).

78 Improved understanding of copepod physiology can refine our ability to predict how  
79 natural copepod populations will respond to environmental change. Furthermore,  
80 increased knowledge of copepod physiology can aid efforts to optimize live-feed  
81 production for aquaculture. Aspects of copepod physiology can be monitored using

82 gross organismal end-points like survival, development, growth, fecundity, respiration  
83 and swimming behavior. All of these can provide insight into physiological condition;  
84 however, to gain a comprehensive mechanistic understanding of copepod physiology a  
85 molecular approach is necessary.

86 To date, the application of molecular tools to study copepod physiology has been  
87 patchy, with different model species (Figure 1) used to study distinct sets of questions  
88 and relatively little integration between research communities. Ecologists have focused  
89 much attention on the impacts of climate change on species and ecosystems, including  
90 the potential for physiological plasticity to mitigate impacts. Elegant studies have  
91 demonstrated heritable and plastic components of thermal tolerance in the intertidal  
92 copepod *Tigriopus californicus* (Kelly et al., 2017; Lima and Willett, 2017; Pereira et  
93 al., 2014, 2017; Schoville et al., 2012; Tangwancharoen et al., 2018). In oceanic  
94 environments, investigations of the responses of *Calanus* spp. to the thermal  
95 environment have focused on understanding the interactions between physiological  
96 tolerances and range shifts in shaping future ecosystems (Ramos et al., 2015; Smolina  
97 et al., 2015). Ecotoxicological studies exploring the effects of diverse chemical  
98 stressors have most commonly focused on *Tigriopus japonicus* and *Calanus* spp., but  
99 a handful of other species have been used (Table 1). Estuarine species, including  
100 *Acartia tonsa* and *Eurytemora affinis*, have been studied in many contexts including  
101 characterization of responses to salinity changes and handling (Nilsson et al., 2018;  
102 Petkeviciute et al., 2015; Rahlff et al., 2017; Xuereb et al., 2012). Many copepods  
103 incorporate a dormant stage within their life history, which has consequences for  
104 developmental progression and energy utilization. Molecular approaches have been  
105 used to study embryonic dormancy in *Acartia tonsa* (Nilsson and Hansen, 2018),  
106 juvenile dormancy and lipid utilization in *Calanus finmarchicus* (Tarrant et al., 2008,

107 2014), and emergence from adult diapause in *Neocalanus flemingeri* (Roncalli et al.,  
108 2018b). Finally, molecular approaches have been extensively used to characterize the  
109 life history and stress responses of parasitic copepods, (e.g., Núñez-Acuña et al., 2016;  
110 Poley et al., 2015). In this case, many of the stress responses studied are intentionally  
111 induced with the aim of developing treatments to weaken or impair the propagation of  
112 these pests. Overall, most of the species discussed above and throughout this  
113 manuscript are calanoids (i.e., members of Order Calanoida), but studies have also been  
114 targeted toward a few species of harpacticoids (*Tigriopus* spp. and *Tisbe holothuriae*),  
115 siphonostomatiods (especially *Lepeophtheirus salmonis* and *Caligus rogercresseyi*),  
116 and cyclopoids (*Apocyclops royi* and *Paracyclops nana*).

117 This review seeks to provide a resource that summarizes previous studies of copepod  
118 molecular physiology, divided between targeted “candidate gene” approaches and  
119 untargeted transcriptomic approaches. To do this, we build upon earlier reviews  
120 focused on copepods that articulated the value of emerging genomic resources (Bron et  
121 al., 2011; Amato and Carotenuto, 2018) and characterized molecular stress responses  
122 (Lauritano et al., 2012). We focus primarily on non-parasitic species that are marine or  
123 euryhaline. In synthesizing the results from studies conducted with a wide variety of  
124 focal species and for diverse applications, we will both demonstrate how techniques  
125 have advanced over time and make suggestions for future study design and data  
126 analysis.

127

## 128 **2. Common Methodological Considerations**

129 Within this section, we discuss considerations common to both targeted and untargeted  
130 gene expression studies. For any gene expression study, it is essential to minimize

131 artifacts associated with handling of the animals and to maintain high-quality RNA  
132 throughout the molecular analysis. Also, in studying the responses of copepods to  
133 environmental stressors or other experimental conditions, developmental and sex-  
134 specific specificity should be considered. These topics are explored below.

135 **Handling**

136 Studies of copepod physiology have been conducted in a variety of contexts, including  
137 direct sampling of wild populations, short-term laboratory manipulations of field-  
138 collected animals, and genetically controlled experiments with animals that had been  
139 maintained in the laboratory over multiple generations. The potential effects of  
140 handling stress have rarely been assessed in copepods. In a limited example, *C.*  
141 *finmarchicus* expression of three small heat shock proteins was shown to increase  
142 between the time of collection and 2-3 hours post-collection (Aruda et al., 2011). More  
143 broadly, a transcriptomic study of *A. tonsa* demonstrated that intense handling stress  
144 created by holding adult copepods outside of water for 10 minutes on Nitex mesh led  
145 to substantial changes in gene expression 24 hours later (Nilsson et al., 2018). While  
146 the study did not profile handling-induced changes in gene expression under other  
147 conditions, elevated mortality was observed in copepods held out of the water for as  
148 little as 1 minute. As Nilsson et al. (2018) suggested, the stress imposed by field  
149 collections can be high and varies according to factors such as tow speed, mesh size,  
150 the temperature during retrieval and processing, and any additional manipulations  
151 associated with isolation and preservation. This study highlights the need to minimize  
152 stress effects during sampling and experimental manipulation. This can be done through  
153 the use of gentle towing methods, maintaining constant temperature and salinity, and  
154 minimizing the total time from collection until preservation. To the extent possible, the

155 effectiveness of any laboratory acclimatization periods should be experimentally  
156 validated, and controls for handling should be included.

157 **Selection of Developmental Stage**

158 Previous studies of copepod physiology have varied in the developmental stage(s) that  
159 were tested, along with the experimental duration, sampling times, season, and other  
160 environmental factors. This diversity in study designs is naturally driven by the diverse  
161 objectives of the individual studies, but such differences also make it difficult to  
162 compare across species or studies. Among these many factors, developmental stage  
163 merits additional discussion in the context of sensitivity to environmental stressors.

164 Within many groups of marine invertebrates, early life stages exhibit increased  
165 sensitivity to abiotic stressors, including hypercapnia, extreme temperatures (reviewed  
166 by Kurihara, 2008). Among crustaceans, studies in decapods have identified early  
167 developmental stages that are particularly vulnerable to thermal stress (e.g., Schiffer et  
168 al., 2014; Storch et al., 2011).

169 In copepods, relatively little research has been devoted to comparing the dynamics of  
170 gene expression across developmental stages, but among the available studies, there  
171 has been no consistent pattern in stage sensitivity to diverse stressors. For example,  
172 Nilsson et al. (2018) found that *A. tonsa* adults were much more sensitive to handling  
173 stress than nauplii, while nauplii were more sensitive to salinity stress than adults.  
174 Tangwancharoen and Burton (2014) showed that *T. californicus* adults were more  
175 sensitive to thermal stress than nauplii and copepodites. Jager et al. (2016) found that  
176 within *C. finmarchicus*, adult males were the most sensitive to exposure to fresh and  
177 weathered oil, followed by late copepodites. They reviewed numerous studies showing  
178 the differential sensitivity of copepod developmental stages to environmental toxicants,  
179 including several showing the increased sensitivity of nauplii (e.g., Lotufo and Fleeger,

180 1997; Saiz et al., 2009). The authors pointed toward the need for additional empirical  
181 observations in multiple species and also suggested that the molt from last naupliar  
182 stage to the first copepodite stage might be particularly energetically demanding and  
183 sensitive to external stressors. The physiological basis for stage-specific sensitivity of  
184 copepods to environmental stressors is not generally known, but transcriptional  
185 profiling can provide some insight. Using RNA-seq, Roncalli et al. (2017b) observed  
186 increased sensitivity of *C. finmarchicus* nauplii to saxitoxin, and also noted that, unlike  
187 adults, nauplii did not upregulate digestive enzymes in response to saxitoxin exposure.  
188 They hypothesized that upregulation of digestive enzymes by adults reduces  
189 assimilation of the toxin and provides increased tolerance.

190 For lipid-storing copepods, such as *Calanus* spp., accumulation of lipophilic  
191 compounds in the oil sac and potential mobilization of these contaminants into adult  
192 tissues and offspring may be a significant route of exposure. While studies in this area  
193 are just beginning, Hansen et al. (2016) showed that polycyclic aromatic hydrocarbons  
194 (PAHs) could be transferred from oil-exposed mothers to offspring and that maternal  
195 exposure resulted in mild but measurable effects on naupliar hatching and gene  
196 expression. Toxværd et al. (2018) found that exposure of females to pyrene during  
197 overwintering lead to a reduced rebuilding of lipid reserves, as well as decreased  
198 survival and egg production.

199 Sex is an additional consideration in studies of adult copepods. Female copepods  
200 usually have longer lifespans and higher stress resistances than males (e.g., Foley et al.,  
201 2019; Parrish and Wilson, 1978). This difference in resistance, and other physiological  
202 differences associated with reproduction may result in different transcriptional  
203 responses to stressors. When studying adult copepods, the selection of sex or the choice  
204 to include a mixture of sexes, should be carefully considered in light of the scientific

205 question to be answered. More broadly, understanding of developmental changes in  
206 environmental sensitivity is improving, but considerable work is needed to create an  
207 integrative view of molecular physiology during copepod development.

208 **Sample preparation**

209 Any measurement of gene expression requires that high RNA quality is maintained  
210 throughout sampling, storage, extraction, and subsequent analysis. Copepods have been  
211 successfully stored in liquid nitrogen for at least 10 years with high RNA yields, and  
212 no evidence on degradation (Hassett et al., 2010); however, preservation in liquid  
213 nitrogen is not always tractable. Obtaining liquid nitrogen in remote field locations can  
214 be difficult, liquid nitrogen levels must be maintained, and liquid nitrogen storage can  
215 pose problems during shipment. Alternative suitable storage methods include guanidine  
216 thiocyanate/phenol-based reagents and RNAlater. Zhang et al. (2013) reported that  
217 samples preserved in guanidine thiocyanate/phenol-based reagents (e.g. TRI Reagent,  
218 TRIzol), could be stored without degradation at 4°C for up to two weeks, or at -80°C  
219 for two years. They also reported satisfactory extraction of RNA following storage in  
220 RNAlater but noted that copepods stored in RNAlater sometimes become transparent  
221 and easy to lose during the extraction protocols. Asai et al. (2015) reported higher RNA  
222 yield and improved quality following storage in RNAlater compared with storage in  
223 TRIzol, but they did not specify the length of storage prior to extraction. In samples  
224 stored in RNAlater, Nilsson et al. (2018) obtained high-quality RNA following 1 week  
225 of storage at -20°C, but a noted decreased quality after transport on dry ice and a total  
226 of 3 months of storage at -20°C.

227 With copepod samples, several extraction protocols have been used to obtain total RNA  
228 that is of suitable quality for downstream measurements of gene expression (reviewed  
229 by Asai et al., 2015; Zhang et al., 2013). While DNase is frequently used during the

230 extraction protocol to remove residual genomic DNA before transcript quantification,  
231 this step can lead to RNA degradation, and effects on sample quality must be monitored  
232 (Nilsson, 2018; Zhang et al., 2013).

233 Assessment of RNA quality is a critical best practice for both targeted and untargeted  
234 gene expression studies; however, RNA quality metrics are not always reported.  
235 Spectrophotometry (e.g., NanoDrop<sup>TM</sup> by ThermoScientific) and fluorometry (e.g.,  
236 Qubit<sup>TM</sup> by Invitrogen) enable assessment of RNA yield, and spectrophotometric  
237 absorbance ratios indicate sample purity. RNA integrity can be assessed through  
238 visualization on denaturing agarose gels, or automated electrophoresis (via  
239 Bioanalyzer<sup>TM</sup> or TapeStation<sup>TM</sup>, both produced by Agilent Technologies). Automated  
240 electrophoresis has been used to assess copepod RNA quality for many years (e.g.,  
241 Voznesensky et al. 2004), and the approach is now common, particularly in association  
242 with RNA-seq studies. In automated electrophoresis of total RNA, the most commonly  
243 used single metric of quality is the RNA Integrity Number (RIN), which is derived  
244 from an algorithm that compares the relative proportions of 28S and 18S rRNA; values  
245 range from 1 to 10, with lower values indicating degradation. In copepods, like many  
246 other arthropods, the 28S band is fragile and can break (“the hidden break”) during  
247 sample preparation (Asai et al., 2015; McCarthy et al., 2015). Thus, many copepod  
248 studies disregard the RIN metrics and rely on a subjective visual analysis of an  
249 electropherogram trace, including the presence of a strong discrete 18S band, the  
250 absence of larger bands indicating contamination by genomic DNA, and limited  
251 smearing within the smaller size ranges (e.g., Almada and Tarrant 2016, Zhou et al.  
252 2018). In many cases 28S breakage occurs during heat denaturing of samples  
253 immediately before analysis, so integrity can sometimes be preserved by omitting this  
254 step (e.g., Figure 2 within Asai et al. 2015).

255

256 **3. Biomarkers**

257 Broadly, biomarkers are detectable molecular, biochemical, and tissue-level changes  
258 that indicate physiological effects (Smit et al., 2009). Compared to organismal metrics,  
259 biomarkers can provide increased sensitivity to detect changes and specific insights into  
260 their likely causes (reviewed by Hook et al., 2014). Quantitative real-time RT-PCR  
261 (qPCR) has been the primary approach to measure the expression of individual  
262 biomarker genes. Biomarker expression is typically normalized to the expression of one  
263 or more reference genes that exhibit stable expression. Below we first discuss criteria  
264 for selection of reference genes and methods of normalization. We then review the  
265 desirable characteristics of biomarkers and their historical application to studies of  
266 copepod physiology. Finally, we point toward additional considerations for future  
267 studies.

268 **Reference genes**

269 Relative changes in mRNA levels can be estimated through a variety of methods,  
270 including the comparative threshold cycle ( $2^{-\Delta\Delta C_t}$ ) (Livak and Schmittgen, 2001),  
271 Pfaffl (Pfaffl, 2001) and LinRegPCR (Ruijter et al., 2009) methods. With each of these  
272 methods, users typically account for systematic variation (e.g., differences in starting  
273 material, RNA quality, and PCR efficiencies) by normalizing expression against one or  
274 more reference genes (Chervoneva et al., 2010; Livak and Schmittgen, 2001).

275 Commonly used reference genes are often carry-overs from older studies that used  
276 semi-quantitative methods, e.g., Northern blots, RNase protection assays, and  
277 conventional reverse-transcription PCR assays (Huggett et al., 2005). Suitable  
278 reference genes should exhibit stable expression across experimental conditions or

279 groups to be compared, such as various developmental stages and tissue types.  
280 Unfortunately, the stability of reference genes is often insufficiently assessed and the  
281 requirement for stability is frequently violated (Dheda et al., 2005; Huggett et al., 2005;  
282 Kozera and Rapacz, 2013; Pfaffl et al., 2004; Svingen et al., 2015). Commonly used  
283 reference genes for copepods (Table 2) might exhibit stable expression within a set of  
284 experimental conditions, but will not necessarily be stable across a different set of  
285 conditions or in other species. Thus, it is important to carefully select and validate the  
286 reference genes to ensure optimal normalization.

287 Some recent studies with copepods have used only a single, or few, reference genes for  
288 normalization (e.g., Nilsson et al., 2017; Petkeviciute et al., 2015; Rahlff et al., 2017).  
289 The limited selection of references genes has historically been due to difficulties in  
290 generating suitable primers from species with no available genomic or transcriptomic  
291 resources. With the continued improvement of sequencing technologies and lower  
292 costs, copepod sequence resources are increasing, which makes it easier to identify and  
293 generate new primers for reference genes (e.g., Nilsson and Hansen, 2018). It is  
294 strongly recommended to normalize gene expression of target genes against the  
295 geometric mean of multiple reference genes (Vandesompele et al., 2002).

296 Several algorithms are available for selection of the most suitable reference genes,  
297 including geNorm (Perkins et al., 2012; Vandesompele et al., 2002), BestKeeper (Pfaffl  
298 et al., 2004) and NormFinder (Andersen et al., 2004). Of these, geNorm calculates a  
299 gene stability value (M), which is defined as the average pairwise variation of gene  
300 expression. The procedure is iterative, where the least-desirable reference gene is  
301 discarded, with subsequent recalculation of the M-values. The ranking of M-values is  
302 carried out in a step-wise manner starting with the two genes having the lowest pairwise  
303 variation. M-values lower than 1.5 are recommended for selecting stable reference

304 genes (Vandesompele et al., 2002). Furthermore, geNorm is able to estimate how many  
305 reference genes should be used for normalization in a given study (Perkins et al., 2012).  
306 BestKeeper assumes that reference genes have similar expression patterns; hence  
307 suitable reference genes should have highly-correlated expression patterns. From the  
308 geometric mean of Ct values and their standard deviation (SD), a “BestKeeper index”  
309 is estimated. Genes that are stably expressed have an SD below 1. The genes are  
310 compared pairwise, and those with the lowest SD values and highest coefficients of  
311 correlation (r) are assumed to exhibit the most stable expression among the candidate  
312 genes (Pfaffl et al., 2004). NormFinder uses a statistical linear mixed-effect model to  
313 estimate intra- and inter-group variation of gene expression and combines the two into  
314 a stability value. The genes with the lowest stability value are assumed to be the most  
315 stable across the experimental conditions (Andersen et al., 2004).

316 Across copepod species and conditions, *elongation factor 1 $\alpha$*  (*EFA*, Table 1) and  
317 *Histone H3* (*HIST*, Table 1), have been validated as some of the most stable reference  
318 genes (Christie et al., 2016; Hansen et al., 2008, 2010; Jeong et al., 2015; Lee et al.,  
319 2017). Another commonly used reference gene, that often is validated as stable, is the  
320 *18S ribosomal RNA* (*18S*) (e.g., Jeong et al., 2015). However, the expression of *18S* has  
321 been shown to be very high expression compared with other candidate reference genes  
322 and biomarker genes in studies with multiple copepod species and tested conditions  
323 (e.g., Lauritano et al., 2015; Nilsson and Hansen, 2018). This suggests that *18S* is not  
324 generally suitable as a reference gene for copepods. Where possible, reference genes  
325 should be selected from distinct functional groups to avoid co-regulation (Riemer et al.,  
326 2012).

327 **Biomarker selection and application in copepods**

328 Desirable characteristics in a biomarker include sensitivity, a large signal-to-noise ratio,  
329 consistency in responses, and known specificity for environmental stressor or other  
330 drivers of response. Numerous studies of copepods have reported the expression of  
331 small numbers of target genes, which were selected as putative biomarkers of processes  
332 of interest, including detoxification, antioxidant activity, apoptosis, and protein  
333 refolding (Table 1). Many of the individual genes that have been used as biomarkers  
334 belong to larger families (e.g., heat shock protein, cytochrome 450 oxidases). Within  
335 large gene families, gene function typically diverges and diversifies, with individual  
336 genes developing distinct expression patterns (developmental, tissue-specific and/or  
337 subcellular) and functionality (e.g., substrate specificity). These features contribute to  
338 the dynamic range of expression for each gene and the environmental conditions that  
339 affect that expression.

340 Initially, due to a lack of genomic resources, copepod genes needed to be individually  
341 cloned and sequenced, using degenerate primers based on known sequences in other  
342 animals. The genes selected for these studies were necessarily evolutionarily conserved  
343 and typically were widely used as biomarkers of similar processes in other animals. For  
344 example, heat shock proteins (HSPs) are a deeply conserved superfamily of molecular  
345 chaperones that enable proper three-dimensional folding of nascent proteins, help to  
346 repair or recycle damaged proteins, contribute to subcellular localization and prevent  
347 aggregation (reviewed by Kregel, 2002; Lanneau et al., 2010). While these proteins  
348 play essential roles in cellular maintenance, HSPs are also frequently up-regulated in  
349 response to diverse cellular stressors. As a biomarker, the best-studied form is the  
350 highly inducible cytosolic HSP70. In copepods, induced expression of HSP70 has been  
351 reported in response to elevated temperature, crowding, handling, embryonic transition  
352 between subitaneous and quiescence states, abnormal salinity, and various chemical

353 contaminants (e.g., Aruda et al., 2011; Nilsson et al., 2014; Petkeviciute et al., 2015;  
354 Rahlff et al., 2017; Rhee et al., 2009, VanderLugt, 2009). In addition to HSP70, several  
355 other HSP molecules display changes expression in response to temperature (Seo et al.,  
356 2006c, but see also Rhee et al. 2009), handling (Aruda et al., 2011), as well as exposure  
357 to endocrine disruptors (Seo et al., 2006b) or toxic diatoms (Lauritano et al., 2011b).

358 Another broad class of biomarkers is related to antioxidant activity. Reactive oxygen  
359 and reactive nitrogen compounds are produced through normal cellular metabolism,  
360 through exposure to and metabolism of environmental contaminants, and as a result of  
361 exposure to ultraviolet radiation or other physical stressors. Animals have developed  
362 several classes of antioxidant enzymes to neutralize these reactive compounds, and  
363 accordingly, studies in copepods have measured expression of antioxidant enzymes,  
364 including superoxide dismutases (Jiang et al., 2013; Kim et al., 2011), catalases  
365 (Hansen et al., 2008; Lauritano et al., 2011b, 2016), glutathione peroxidases (Zhuang  
366 et al., 2017) and peroxiredoxins (Zhuang et al., 2017). While exposure to cellular  
367 oxidants is broadly expected to lead to induction of antioxidant defenses, the observed  
368 patterns are complicated and dependent upon the duration, concentration, and type of  
369 stressor, as well as the specific genes, measured.

370 Genes are often selected as biomarkers based on their specific mode of action to  
371 indicate the disruption of a process process or exposure to a specific stressor. For  
372 example, xenobiotic metabolizing enzymes often indicate exposure to chemical  
373 stressors. Among these, cytochrome P450 oxidases and glutathione S-transferases have  
374 been most widely studied in copepods (see Table 2). Vitellogenins are precursors to  
375 major egg yolk proteins and have been proposed as markers of the reproductive  
376 condition in copepods. Studies in *L. salmonis* have demonstrated that vitellogenins are  
377 produced in subcuticular tissues of adult females, secreted into the hemolymph, and

378 deposited in the maturing oocytes (Dalvin et al., 2011). Vitellogenins have been  
379 identified in other copepod species, with measurable expression in late copepodid  
380 stages that greatly increases in adult females (e.g., Hwang et al., 2009). Two studies  
381 have demonstrated induction of vitellogenin expression in response to metal exposure,  
382 but the mechanism of disruption and links to reproductive endpoints are still unclear  
383 (Hwang et al., 2010b; Lee et al., 2008a).

384 **Challenges and opportunities**

385 To date, there have been limited instances where biomarkers developed for copepods  
386 have been adopted for studies by distinct research groups or across species. As with  
387 many physiological studies, experimental differences in factors such as handling  
388 protocols, the nutritional status and developmental stage of the animals, and the  
389 duration and intensity of any experimental exposure complicate direct comparisons  
390 across studies. Two additional consideration merit additional discussion: homologous  
391 relationships of biomarkers and genetic complexity of study species.

392 Homologous relationships

393 Because full copepod transcriptomic databases have only recently become available,  
394 earlier biomarker studies frequently required cloning and sequencing of individual  
395 genes. In the case of multi-gene families, this could lead to an analysis of paralogous  
396 genes that might not be directly comparable to one another. For example, in studies of  
397 *Calanus finmarchicus*, Voznesensky et al. (2004) reported the induction of HSP70  
398 following thermal stress, but Hansen et al. (2008) found no effect of naphthalene  
399 exposure on HSP70 expression. While it is entirely plausible that the two different  
400 stressors would induce distinct physiological responses, it would not be evident to a  
401 casual reader that the two studies measured different HSP70 family members (Aruda

402 et al., 2011). This issue of homology becomes even more complex in cross-species  
403 comparisons, in which the roles of various gene family members may have diverged.

404 High-throughput sequencing and the increased availability of copepod transcriptomes  
405 and genomes have also provided an opportunity to place biomarkers within a gene  
406 family, and more broadly to study gene diversification and loss. As an example, Porter  
407 et al. (2017) characterized the evolutionary relationships among phototransduction  
408 genes in 10 copepod species from diverse lineages and identified four primary groups  
409 of copepod opsins, two of which were broadly distributed, and two of which were  
410 restricted to a subset of species.

411 Genetic complexity

412 It is becoming increasingly apparent that morphological identification may be  
413 insufficient for many copepod species. This is particularly problematic for studies of  
414 natural populations, where cryptic species may co-occur. For example, it has recently  
415 been demonstrated that morphological characters do not reliably discriminate *C.*  
416 *finmarchicus* and *C. glacialis* and that these species widely co-occur, particularly  
417 within fjord environments (Choquet et al., 2017, 2018). Smolina et al. (2015)  
418 incorporated this consideration into their methodology, using genetic techniques to  
419 verify species identity before pooling RNA from multiple individuals and measuring  
420 gene expression.

421 Heritable physiological variability has been described in both *Acartia tonsa* and *A.*  
422 *hudsonica* (Avery, 2005; Cournoyer, 2013), along with substantial genetic diversity  
423 within each group. It has been suggested that the major genetic lineages represent  
424 cryptic species, which share broadly overlapping ranges (Chen and Hare, 2011;  
425 Milligan et al., 2011). No studies have yet compared these physiological

426 characteristics of these *Acartia* lineages in a controlled genetic context. In contrast,  
427 studies conducted in *Eurytemora affinis* and *Tigriopus californicus*, have frequently  
428 incorporated genetic variation into their experimental design, comparing expression in  
429 genetically distinct populations or genetically-controlled lineages to study gene by  
430 environment interactions, adaptation to novel environments, and predicted responses  
431 to climate change (e.g., Kelly et al. 2017; Lee et al. 2011; Pereira et al. 2014, 2017).

432 Thus, moving forward with natural populations, it will be necessary in many cases to  
433 conduct molecular species identifications alongside any other biomarker analyses. This  
434 has consequences for experimental design. For example, addressing this concern  
435 requires nucleic acid extraction and analysis before pooling any material from  
436 individual animals.

437

#### 438 **4. Transcriptomes**

439 The development of high-throughput sequencing and associated bioinformatic  
440 pipelines has revolutionized our understanding of copepod physiology. As copepod  
441 physiologists have adopted these new methods, best practices for experimental design,  
442 analysis, documentation and data availability have also developed. This section reviews  
443 transcriptional profiling approaches that have been applied to copepods, including  
444 technical recommendations and suggestions for best practices.

##### 445 **Early untargeted approaches to expression profiling**

446 An advantage of whole-transcriptome profiling is that it enables candidate biomarkers  
447 to be identified based on observed expression patterns, without requiring *a priori*  
448 selection of candidate genes. Before the development of high-throughput sequencing,  
449 gene sequences were typically determined by targeted amplification and cloning using

450 degenerate primers, or by mining libraries of expressed sequence tags (ESTs) produced  
451 through Sanger sequencing. EST libraries also facilitated *de novo* identification of  
452 candidate biomarkers, both within suppressive subtractive hybridization (SSH)  
453 experiments and by enabling probe design for microarray analysis (e.g., Lenz et al.  
454 2012). Because ESTs are derived from longer Sanger sequences, they have sometimes  
455 provided a useful check on the accuracy of transcripts predicted from the assembly of  
456 shorter reads produced by high-throughput sequencing (e.g., Christie et al., 2013).  
457 Overall, while these older methods are becoming less common, microarrays, in  
458 particular, are still useful for some applications and continue to be used.

459 With SSH, complementary DNA fragments from two different libraries (e.g., treatment  
460 and control) are hybridized, and fragments overrepresented within one of the libraries  
461 are amplified, cloned and sequenced. This method has been used to identify biomarkers  
462 associated with copepod energetics and diapause (*Calanus finmarchicus*, Tarrant et al.,  
463 2008), as well as with exposure to nickel (*Pseudodiaptomus annandalei*, Jiang et al.,  
464 2013), diethanolamine (*C. finmarchicus*, Hansen et al., 2010), toxic diets (*C.  
465 helgolandicus*, Carotenuto et al., 2014), an organophosphate (*Lepeophtheirus salmonis*,  
466 Walsh et al., 2007) and multiple stressors (*C. finmarchicus*, Hansen et al., 2007). This  
467 approach has produced useful biomarkers: for example, genes identified during an SSH  
468 screen of active lipid-storing copepods were subsequently shown to parallel changes in  
469 oil sac volume during juvenile copepod development (Tarrant et al., 2014). Limitations  
470 of the method include a high rate of false positives and a bias toward genes with high  
471 overall expression. For example, Jiang et al. (2013) tested 8 randomly-selected genes  
472 from an SSH screen and was able to verify consistent expression patterns for only 5 of  
473 them using quantitative PCR (qPCR). In bi-directional screens comparing active and  
474 dormant copepods, Tarrant et al. (2008) found that ribosomal genes represented 18%

475 of the clone libraries, and the highly expressed myosin transcripts represented 32% of  
476 the annotated mRNA sequences.

477 For microarray analysis of gene expression, DNA probes are arrayed onto glass slides  
478 and hybridized with fluorescently-labeled cDNA. The probes may be prepared from  
479 cDNA libraries or synthetic oligonucleotides, and the hybridization may be with one  
480 sample or two distinctly labeled samples. Microarrays have been extensively used to  
481 characterize responses of the intertidal copepod *Tigriopus japonicus* to environmental  
482 stressors including copper (Ki et al., 2009), manganese (Kim et al., 2013a), ultraviolet  
483 radiation (Rhee et al., 2012), and  $\beta$ -naphthoflavone (Rhee et al., 2015). Lee et al. (2011)  
484 used a custom cDNA microarray to study changes in Na<sup>+</sup>K<sup>+</sup> ATPase expression  
485 associated with invasion of freshwater habitats by *Eurytemora affinis*. A targeted  
486 “physiological” microarray was used to measure changes in *Calanus finmarchicus* gene  
487 expression associated with food availability, lipid storage, development and vertical  
488 migration (Lenz et al., 2012; Unal et al., 2013). A distinct broad-scale array was used  
489 to evaluate the effects of components of petroleum extraction on *C. finmarchicus*  
490 (Jensen et al., 2016). For the parasitic *Lepeophtheirus salmonis*, microarrays have been  
491 used to characterize adult female maturation (Eichner et al., 2008) and abiotic stress  
492 responses of free-living larvae (Sutherland et al., 2012). For some species, multiple  
493 arrays have been designed, generally increasing in complexity over time. Successive *T.*  
494 *japonicus* arrays produced by the same research group increased from ~6000 35-mer  
495 probes (Ki et al., 2009) to ~55,000 60-mer probes (Rhee et al., 2012). For *C.*  
496 *finmarchicus* and *L. salmonis*, earlier arrays were built using cDNA amplified from  
497 normalized clone libraries, which were enriched for rare genes (Eichner et al., 2008;  
498 Lenz et al., 2012). Later arrays constructed for both of these species used larger  
499 numbers of oligonucleotide probes (Jensen et al., 2016; Sutherland et al., 2012).

500 While microarray is a powerful technique, it requires a substantial initial investment in  
501 array design and synthesis. In addition, sample preparation, hybridization, scanning and  
502 downstream data processing all require careful optimization and quality control. The  
503 popularization of microarray approaches eventually led to the development of  
504 analytical pipelines, standards for data reporting, and best practices for data analysis  
505 (e.g., Brazma et al., 2001; Knapen et al., 2009; Shi et al., 2008). This increased attention  
506 toward data management provided some of the initial frameworks for data analysis and  
507 storage associated with high-throughput sequencing approaches. Within NCBI, the  
508 Gene Expression Omnibus (GEO) is a public repository for curated gene expression  
509 datasets produced using either microarray or sequencing-based platforms (Barrett et al.,  
510 2013).

511 ***De novo Transcriptome Assembly***

512 To identify current and recent practices in transcriptome assembly, we searched the  
513 NCBI transcriptome shotgun assembly (TSA) in August 2018 for available copepod  
514 transcriptomes. From each of these, we compiled details of the source biological  
515 material, methods of sequencing and assembly, and statistics regarding the number of  
516 contigs, and BUSCO score (Table 3, Figure 2). We also included the *Apocyclops royi*  
517 transcriptome, which had not yet been publicly released. In total, we identified 19  
518 transcriptome assemblies, corresponding to 13 non-parasitic marine species. These  
519 efforts were not uniformly distributed across taxonomic groups; for example, five of  
520 the transcriptomes corresponded to two species of *Calanus*, and four of the  
521 transcriptomes corresponded to three species of *Tigriopus*. In addition to the TSA's,  
522 we identified 6 whole-genome shotgun assemblies (WGS) (*Eurytemora affinis*:  
523 GCA\_000591075.2, GCA\_000591075.2; *Calanus finmarchicus*: GCA\_002740975.1,

524 GCA\_002740985.1; *Oithona nana*: GCA\_900157175.1; *Acartia tonsa*:  
525 GCA\_900241095.1) for 4 copepod species.

526 Sequencing technologies included Illumina-based methods (HiSeq, MiSeq, and  
527 NextSeq), 454 pyrosequencing (e.g., 454 GS FLX), and Ion-Torrent. Assembly  
528 methods included Trinity, CLC, Mira, Newbler, and CAP3. Of the most recently  
529 published assemblies (9 transcriptomes published or posted from 2016-2018), the  
530 majority used Illumina-based methods (HiSeq and NextSeq). All of these used paired-  
531 end reads. In most of these, (7 of 9) the read length was 150 bp, and in the other cases,  
532 the read length was 100 or 125 bp. Trinity was used for all of these assemblies, and one  
533 of the studies included both Trinity and CAP3 assemblies.

534 In the 18 published transcriptomes, the number of contigs ranged dramatically from  
535 28,954 to 554,991. Some of this range represents variation in the extent of coverage  
536 and completeness of the transcriptome assembly. One approach to assess the  
537 completeness of an assembled transcriptome is the comparison with a curated set of  
538 well-conserved single copy eukaryotic genes. To apply this method, we used BUSCO  
539 ver. 2 (Benchmarking Universal Single-Copy Orthologs; Simão et al., 2015) to assess  
540 the completeness of the same 18 copepod transcriptomes shown in Table 3. Of these,  
541 three (*C. finmarchicus*, GBXU; *C. glacialis* GBXT and HACJ) had a large proportion  
542 of missing or fragmented genes (79-87% combined). These three assemblies were  
543 produced from relatively shallow sequencing (0.7-5 Mb total) and using either Ion-  
544 Torrent or 454 GS FLX methods. While these smaller assemblies miss a large number  
545 of genes, they can still be useful for targeted applications, such as profiling expression  
546 of highly expressed genes and identification of biomarker sequences for qPCR studies  
547 (e.g., Smolina et al., 2015). In contrast to the smaller transcriptome assemblies, nine  
548 transcriptomes had less than two percent missing BUSCO genes, but also had 20-54%

549 of the genes represented by more than one transcript. While some of these may  
550 represent true lineage-specific duplications, the BUSCO gene set was curated to include  
551 genes represented by single orthologs in a vast majority of the diverse animal taxa  
552 studied. Thus, a large proportion of the apparent duplications are thought to represent  
553 incorrectly assembled haplotypes (Simão et al., 2015). Of the two *C. finmarchicus*  
554 transcriptomes with nearly complete BUSCO sets (GAXK and GBFB), the proportion  
555 of duplicated genes was lower in the assembly by Lenz et al. (2015; GAXK, 26%  
556 duplicated BUSCO genes), which included 206,012 contigs, compared with the  
557 assembly by Tarrant et al. (2014; GBFB, 36% duplicated BUSCO genes), which  
558 contained 241,140 contigs. Thus, the BUSCO analysis suggests that many of the  
559 additional contigs in the second transcriptome may represent duplicates. A likely  
560 explanation for this difference is that the smaller transcriptome was filtered post-  
561 assembly to retain only the longest contig associated with each clustered component  
562 (“comp”) produced by the Trinity assembler. While this filtering approach undoubtedly  
563 removes many duplicates, it can also result in the removal of some distinct genes (e.g.,  
564 Lenz et al., 2014). Finally, it should also be noted that the BUSCO score does not  
565 directly indicate the completeness of the transcriptome; transcriptome assemblies with  
566 high BUSCO scores may still be missing large proportions of rare or conditionally  
567 expressed transcripts.

## 568 **Transcriptome-wide Differential Gene Expression via RNA-seq**

569 To evaluate recent methods for analyzing differential gene expression in copepods, we  
570 identified 18 RNA-seq-based studies published between 2012 and 2019 (Table 4; new  
571 studies continue to emerge, and this list is not comprehensive). The studies were  
572 conducted in a total of 8 species, and they investigated responses to a range of  
573 environmental or experimental conditions, including salinity shock, handling stress,

574 pH, and temperature, as well as experimental exposure to cultured bacteria, toxic  
575 dinoflagellates, or organic contaminants. In addition, some studies addressed variation  
576 among geographically isolated populations or across developmental stages. All of these  
577 studies utilized pools of copepods to construct each library, with the specific number  
578 ranging from 3 to 500. Most of the studies (13 of 18 in Table 4) included biological  
579 replication, typically with 3 or 4 replicates per treatment. Replication occurred at  
580 different levels. For example, Kelly et al. (2017) produced only one library per genetic  
581 line and treatment, but the experimental conditions were fully replicated across three  
582 genetic lines (i.e., providing full biological replication). Bailey et al. (2017) included 5-  
583 6 replicates per treatment, including three experimental replicates (separate aquaria),  
584 each with 1-2 libraries. Five studies had only one biological replicate per treatment;  
585 however, these often used the differential expression analysis in a more exploratory  
586 manner. For example, Smolina et al. (2015) conducted a small-scale transcriptomic  
587 study of responses in *Calanus* spp. to different temperature conditions and used the  
588 results to develop hypotheses regarding thermal sensitivity and to select genes for more  
589 detailed expression profiling by qPCR.

590 The studies in Table 4 used a variety of methods for read-alignment, calculation of  
591 counts per read and identification of differentially expressed sequences; however, not  
592 all studies indicated the method used for each step. For alignment, methods included  
593 bowtie, bowtie2, bwa, Rsubread, and the proprietary CLC genomics workbench. Two  
594 studies used the Kallisto alignment-free method (described below). From the mapping  
595 results, matrices of counts per transcript can be generated using RSEM, Rsubread, and  
596 other custom scripts; in these analyses, transcripts with very low levels of expression  
597 are frequently removed prior to differential expression analysis. Differentially  
598 expressed genes are typically identified using R-based packages including

599 DESeq/DESeq2, edgeR, limma, and Sleuth (Anders and Huber, 2010; Love et al., 2014;  
600 Pimentel et al., 2016; Ritchie et al., 2015; Robinson et al., 2010).  
601 Overall, most of these pipelines have included *de novo* transcriptome assembly  
602 followed by read mapping, abundance estimation, and differential expression analysis.  
603 These methods are time-consuming, require high computational capacity, and can be  
604 limited by the quality of the reference genome or transcriptome (Bray et al., 2016;  
605 Pimentel et al., 2016). “Alignment-free” methods (e.g., Sailfish, Kallisto, Salmon) have  
606 recently been developed, in which the reference transcriptomes are shredded into kmers  
607 (Conesa et al., 2016). The kmers from experimental reads can then be matched to the  
608 kmers from the transcriptome, resulting in fast and accurate estimations of abundance.  
609 Kallisto (e.g., used by Nilsson et al., 2018), constructs a *de Bruijn* graph from  
610 transcriptome kmers (abbreviated as t-DBG), and then a pseudo-alignment is  
611 constructed in which the kmers from experimental reads are evaluated for compatibility  
612 with the t-DBG (Bray et al., 2016). By skipping kmers for which compatibility does  
613 not change with the t-DBG, the process is accelerated. Because Kallisto only accounts  
614 for exact k-mer matches, most sequencing errors are discarded (Bray et al., 2016).  
615 Expression of genes or transcripts is determined by quantifying the k-mers associated  
616 with each component of the indexed reference transcriptome. The R package Sleuth is  
617 designed for processing the output from Kallisto and analyzing differential expression  
618 at the transcript or gene level (Pimentel et al., 2016). Analysis methods continue to  
619 develop, and future studies will most likely include a balance between investigators  
620 using established pipelines and those choosing to incorporate new methods.

621 **Best Practices and Challenges**

622 Sequencing method and experimental design

623 Most considerations of sequencing depth and replication are not unique to copepods.  
624 Specific concerns are mostly related to working with heterogeneous samples and  
625 lacking a well-annotated reference genome for scaffolding. Illumina-based sequencing  
626 is the most widely used technique for *de novo* transcriptome assembly, and paired-end,  
627 stranded reads of at least 100 bp are recommended (Haas et al., 2013). Studies in diverse  
628 animals have shown that *de novo* assemblies derived from 20-40 M reads can typically  
629 recover most transcripts (Francis et al., 2013; MacManes, 2016). In evaluating a  
630 *Calanus finmarchicus* transcriptome assembly, Lenz et al. (2014) progressively  
631 subsampled their data set and found steep increases in the number of assembled contigs  
632 obtained when increasing from 6 M to 50 M reads (100 bp, paired-end). They report  
633 that a good quality assembly can be constructed from as few as 50 M reads, but suggest  
634 based on rarefaction analysis that rare transcripts may be missing even with up to 400  
635 M reads.

636 In comparing transcript expression among groups of samples, choices must be made to  
637 optimize the statistical power of the study within the constraints of funding and sample  
638 availability (reviewed by Todd et al., 2016). Several studies have demonstrated that  
639 increased replication is more important than sequencing depth in maximizing statistical  
640 power. Depending on the type of sample, once 10-20 M reads are obtained, it is  
641 generally much more beneficial to increase the number of replicates rather than  
642 sequencing depth (Ching et al., 2014; Liu et al., 2013). Another possible element of the  
643 experimental design is the use of paired samples or blocking factors. For example,  
644 copepods from replicate cultures (or collection sites) may each be split into treatment  
645 and control groups that are treated as pairs. In this case, such a design would account  
646 for variability among cultures (or collection sites), can increase the signal-to-noise  
647 ratio, and thereby can increase statistical power in gene expression studies (Ching et

648 al., 2014). Paired or blocked study designs should be considered in future studies with  
649 copepods.

650 Assessment of Transcriptome Quality

651 The best approach for transcriptomic or targeted gene expression studies depends on  
652 the available resources as well as the goals of the study. While a variety of methods can  
653 be successfully used, most recent high-throughput sequencing studies have used paired-  
654 end Illumina-based sequencing coupled with *de novo* assembly using the Trinity  
655 software suite. Until recently, few tools were available to assess and compare the  
656 quality of various transcriptome assemblies, with many investigators citing the number  
657 of contigs along with some transcript size metrics. The most common size metric has  
658 been the N50, which was developed for genome assemblies where very long contigs  
659 (full chromosomes) are desirable, and sequence representation should be uniform.  
660 More recently, the ExN50, which is weighted toward the most abundant transcripts, has  
661 been proposed as a more relevant metric for transcriptome assemblies  
662 (<https://github.com/trinityrnaseq/trinityrnaseq/wiki>, Accessed 8 March 2019). Overall,  
663 such general assembly statistics provide a useful metric of comparison, but they do not  
664 necessarily provide direct insight into the completeness or accuracy of the assembly.  
665 BUSCO analysis, as we have used here, provides one means to assess transcriptome  
666 completeness and duplication. Other methods include DETONATE (Li et al., 2014)  
667 and TransRate (Smith-Unna et al., 2016). The selection of specific tools will depend on  
668 factors such as the sequencing platform and the availability of a sequenced reference  
669 genome (e.g., Moreton et al., 2016). As sequencing technologies and assembly  
670 algorithms continue to develop and improve, these tools can and should be used to  
671 inform choices about sequencing platforms, assembly parameters, and post-assembly  
672 filtering.

673 Variability

674 For eventual comparison across samples, it has been recommended to pool all sequence  
675 reads from all samples prior to transcriptome assembly (Haas et al., 2013). However,  
676 sequence polymorphism increases the complexity of the *de Bruijn* graph and can  
677 negatively affect the assembly (Iqbal et al., 2012; Studholme, 2010). To account for  
678 this, MacManes (2016) recommended assembling sequences derived from a single  
679 individual. In cases where distinct sets of transcripts may be present in different groups  
680 of animals (e.g., developmentally restricted or sex-specific transcripts), sequences  
681 should be assembled from one individual per group. For larger-bodied copepods and  
682 later developmental stages, it is possible to obtain sufficient RNA for library  
683 construction from individual animals, and this relatively new recommendation has been  
684 adopted into some of the most recent copepod transcriptome assemblies (e.g., Nilsson  
685 et al., 2018; Roncalli et al., 2018a).

686 In studies that encompass divergent populations, investigators have sometimes utilized  
687 the alternative approach of independently assembling population-specific  
688 transcriptomes. In several studies comparing isolated and divergent populations of the  
689 intertidal copepod *Tigriopus californicus*, investigators have utilized custom analysis  
690 pipelines to integrate the transcriptomic databases and enable analysis of orthologous  
691 transcript sets (Barreto et al., 2014; DeBiasse et al., 2018; Kelly et al., 2017; Lima and  
692 Willett, 2017).

693 Genomic resources

694 As with other non-model organisms, a challenge with molecular physiology studies of  
695 copepods is the limited availability of genomic resources, including genome  
696 assemblies, annotation, and integration of data types. Aspects of this situation are

697 rapidly improving, particularly with recent successes in genome assembly. At the time  
698 of writing, sequenced genomes are available for at least 7 non-parasitic copepod  
699 species: *Acartia tonsa*, *Apocyclops royi*, *Eurytemora affinis*, *Oithona nana*, *Tigriopus*  
700 *californicus*, *T. japonicus* and *T. kinsejongensis* (Barreto et al., 2018; Eyun et al., 2017;  
701 Jørgensen et al., in press; Kang et al., 2017; Lee et al., 2010; Madoui et al., 2017).  
702 Among these, the *T. californicus* genome assembly is of particularly high quality, with  
703 >94% complete predicted transcripts when compared to the BUSCO arthropod gene  
704 set, and >99% of assembled sequence contained within 12 chromosomal scaffolds. The  
705 relatively compact genome size of *T. californicus*, thought to be around 200 Mb,  
706 undoubtedly facilitated its assembly (Barreto et al. 2018).

707 In other cases, assembly of copepod genomes has proven more challenging, for reasons  
708 including low GC content, small organism size, and sometimes large genomes with  
709 high proportions of repetitive DNA (Bron et al., 2011; Jørgensen et al., in press). For  
710 example, the complete *A. tonsa* genome is estimated to be nearly 2.5 Gb, with only ~0.5  
711 Gb assembled and non-repetitive sequence (Jørgensen et al., in press). Concerning  
712 genome size, another potentially complicating factor is chromatin diminution, the  
713 process by which selected heterochromatin regions are eliminated from somatic cells  
714 during early embryogenesis. Chromatin diminution has been identified in 23 species  
715 of freshwater cyclopoid copepods (reviewed by Grishanin, 2014). The process creates  
716 large differences in genome size between the somatic and germ lineages and differences  
717 in somatic genome structure among related copepod species (reviewed by Bron et al.,  
718 2011). Other considerations affecting genome assembly have included variation  
719 associated with pooling heterogeneous individuals and contamination with prey or  
720 epibiont sequences. A variety of approaches are being used to address these challenges,

721 including low-input sequencing, rearing of inbred stocks, careful selection of input  
722 material, and improved bioinformatic pipelines for filtering of foreign sequences.

723 The utility of these newly available genomes and transcriptomes will increase as efforts  
724 continue to improve their annotation. In January 2016, a symposium “Tapping the  
725 Power of Crustacean Transcriptomes to Address Grand Challenges in Comparative  
726 Biology” was convened as part of the annual meeting of the Society of Integrative and  
727 Comparative Biology. A key recommendation emerging from an associated workshop  
728 was to improve integration of genomic and transcriptomic assemblies to facilitate  
729 visualization of gene-specific expression patterns, cross-species comparisons, and  
730 identification of novel genes (Mykles et al., 2016). Toward this end, if appropriately  
731 leveraged, RNA-seq studies can help to improve gene predictions, reveal alternative  
732 splicing and allelic variants, and provide functional insights. Annotation of genomic  
733 and transcriptomic databases would benefit from integration with ongoing efforts  
734 toward annotating other crustacean databases (e.g., RNA-seq studies in decapods,  
735 reviewed by Nguyen et al., 2018) and incorporating emerging results from functional  
736 studies conducted in crustaceans (e.g., knockout and knockdown approaches).

737 Data Availability and Analytical Reproducibility

738 As the application of high-throughput sequencing has matured, expectations for data  
739 availability and analytical reproducibility have increased. These expectations are not  
740 unique to studies of copepods (Conesa et al., 2016; Das et al., 2016), but studies  
741 published by the community (e.g., those within Tables 3 and 4) have shown increasing  
742 documentation of workflows and improved availability of data. Minimally,  
743 publications must provide access to raw sequence data and reference databases. Each  
744 step of the analysis pipeline must be clearly described, including names and versions  
745 of software programs, along with details of options specified within the analysis.

746 Expectations for sharing of custom code have varied. In some cases, the code is  
747 provided only for complex analyses, but increasingly investigators have documented  
748 complete analytical pipelines and referenced version-controlled scripts and workflows  
749 in repositories such as GitHub.

750

## 751 **5. Back Again**

752 Now that both targeted and untargeted approaches at expression profiling have become  
753 widely accessible, it is possible to reflect on some lessons learned and directions for  
754 future research. In the following sections, we first consider how high-throughput  
755 sequencing has greatly accelerated our ability to identify candidate biomarkers that can  
756 be used in targeted physiological studies. We then provide some “food for thought” as  
757 to how additional physiological insight can be gained from targeted functional studies,  
758 characterization of taxonomically restricted genes, and integration of databases.

## 759 **Building a Better Biomarker**

760 The increasing availability of annotated transcriptomes and transcriptome-wide  
761 expression data has greatly informed the selection of biomarkers. For example, Tarrant  
762 et al. (2014) compared transcriptomic patterns between two times within a copepodite  
763 developmental stage. From this dataset, they identified variable genes related to  
764 development and molting and then conducted detailed profiling of a small number of  
765 genes during progression through the stage. Similarly, Roncalli et al. (2016) used both  
766 transcriptome-wide expression profiling and targeted qPCR to characterize the effects  
767 of consumption of toxic algae on the expression of glutathione-S-transferase (GST)  
768 enzymes. From the 41 predicted GSTs in *C. finmarchicus*, three were profiled over time  
769 in two independent experiments; expression of one of these three genes was induced by

770 exposure to toxic algae. RNA-seq analysis of a subset of time points corroborated the  
771 qPCR analysis and revealed two other GSTs that were induced by exposure to toxic  
772 algae. These additional genes could serve as useful biomarkers in future experiments.

773 **From Sequences to Functional Characterization**

774 Beyond selecting biomarkers based on their expression pattern and the known functions  
775 of homologous genes in other organisms, new physiological understanding will come  
776 from the functional characterization of genes in copepods. Among the possible  
777 knockdown/knockout approaches, RNAi has been successfully used for targeted  
778 knockdown of copepod transcripts. To date, this approach has primarily been applied  
779 to parasitic species (e.g., Eichner et al., 2015; Trölöö et al., 2014), but RNAi was also  
780 used to confirm the role of HSPb1 in conferring thermal tolerance to the free-living *T.*  
781 *californicus* (Barreto et al. 2014). Direct functional studies provide a much deeper  
782 understanding of the roles played by biomarker genes, and have the potential to  
783 transform our current understanding of copepod physiology.

784 **Where next?**

785 As curated sequence databases (e.g., Swiss-Prot/Uniprot) have grown, the ability to  
786 annotate copepod transcriptomes has improved. As a rough example (significance  
787 thresholds and other methodological details varied among studies), early efforts to  
788 annotate *Calanus* transcriptomes reported 33-40% of transcripts with positive BLAST  
789 hits and only 11-28% of these homologous sequences associated with GO terms (Lenz  
790 et al., 2014; Tarrant et al., 2014). More recently, Roncalli et al. (2018a) annotated 57%  
791 of coding transcripts against Swiss-Prot, and over 90% of the annotated transcripts were  
792 associated with GO terms. Still, the GO annotations are based primarily on knowledge  
793 of gene function in model organisms, and many genes remain unannotated. This limited

794 annotation of copepod genes and transcripts has represented a challenge in harnessing  
795 the full power of transcriptomic approaches. Homology-based annotation is by  
796 definition biased toward evolutionarily conserved genes, yet within eukaryotes,  
797 taxonomically restricted genes (TRGs, also called lineage-specific genes) provide an  
798 important source of developmental, physiological and regulatory diversity (Lespinet et  
799 al., 2002). TRGs can evolve in association with lineage-specific traits, such as honeybee  
800 sociality (Johnson and Tsutsui, 2011) or coral calcification (Moya et al., 2012). In  
801 addition, TRGs are strongly associated with adaptation to changing environments  
802 (Schlötterer, 2015). For example, TRGs are overrepresented in responses by *C. elegans*  
803 to extreme environments (Zhou et al., 2015) and *Daphnia magna* to a suite of  
804 environmental perturbations (Orsini et al., 2018). Weighted gene co-expression  
805 network analysis (WGCNA) methods can provide powerful insight into regulatory  
806 patterns of TRGs by identifying clusters or modules of genes with similar expression  
807 patterns. The annotation of conserved genes within individual modules can provide  
808 some clues as to potential functions for unannotated genes and can facilitate the  
809 selection of biomarkers. For example, investigators can use WGCNA results to identify  
810 sets of candidate genes that belong to distinct modules and reflect different  
811 physiological processes, or they can select multiple genes within a module to enable  
812 more detailed analysis of the association. WGCNA-based methods are increasingly  
813 being applied to non-model organisms (e.g., Fuess et al., 2018; Johnson et al., 2018),  
814 and will likely become prevalent within future studies in copepods.

815 Mining underutilized data from older studies, including data generated with older  
816 technologies (e.g. microarrays), with machine learning approaches (reviewed by  
817 Golestan Hashemi et al., 2018) may enable researchers to further leverage existing data  
818 to unlock new insights into copepod physiology. In many cases, data was generated for

819 answering questions within a targeted study and used in a relatively narrow context.  
820 Re-analyzing publicly available data in new ways with new questions has the potential  
821 to entrain a broader community of scientists and could lead to novel discoveries without  
822 the need for additional sequencing costs and labor. A major challenge in this approach  
823 is the harmonization of heterogeneous data from multiple platforms into a unified  
824 computational framework to extract the signals. When appropriately harmonized,  
825 combining multiple datasets can aid in the construction of expression atlases and  
826 identification of regulatory relationships. Expression atlases are frequently built to  
827 synthesize spatiotemporal expression patterns within an organism or tissue (e.g.,  
828 Papatheodorou et al., 2017; Zhang et al., 2014). To date, little is known regarding  
829 spatial gene expression patterns within copepods, but a great deal of data is being  
830 amassed regarding the environmental, developmental, and experimental conditions  
831 associated with expression. Using methods such as WGCNA, described above, genes  
832 with similar expression profiles within expression atlases can be associated with a  
833 shared function and provisionally annotated (Carnielli et al., 2015; Oliver, 2000). With  
834 increasing availability of ‘omics data (e.g., transcriptomic, proteomic and  
835 metabolomic) for copepods, another possibility will be to concatenate the information  
836 across distinct types of ‘omics analysis (i.e., multi-omics) to gain more information  
837 about biological processes, to identify regulatory networks, and to search for robust  
838 biomarkers across datasets (Bersanelli et al., 2016).

839 In conclusion, the tools and databases available for expression profiling studies in  
840 copepods have radically advanced over the past ten years. These technological  
841 advances are being matched with increased sophistication in analytical approaches as  
842 well as improved practices for experimental design, documentation, and data

843 accessibility. Leveraging these rich datasets will lead to a greatly improved  
844 understanding of copepod physiology and copepod responses to environmental change.

845

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850

851 **Figure Legends**  
852

853 **Figure 1.** Examples of diverse marine copepods utilized in physiological studies. (A)  
854 The calanoid *Acartia tonsa* male, (B) The calanoid *Calanus glacialis* C5 copepodite  
855 with prominent oil sac, (C) The cyclopoid *Apocyclops royi* egg-bearing female (D)  
856 The harpacticoid *Tigriopus japonicus* egg-bearing female. Of these, *Tigriopus spp.*  
857 (particularly *T. californicus*, not shown) have been extensively developed as a model  
858 for studies of molecular evolution and plasticity. The others represent a growing  
859 diversity of species for which molecular physiology studies are being driven by their  
860 ecological importance. Photos courtesy of Dr. Minh Thi Thui Vu (A), A.M.T. (B),  
861 Dr. Hans van Someren Gréve (C), and Professor Hans Uwe Dahms (D).

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863

864 **Figure 2.** BUSCO analysis of the following copepod transcriptomes (with NCBI  
865 accession numbers, alphabetized by scientific name as in Table 3): **GFWY**: *Acartia*  
866 *tonsa* (GFWY00000000.1, 27-sep-2017); **HAGX**: *Acartia tonsa* (HAGX00000000.1,  
867 29-sep-2017); **GAXK**: *Calanus finmarchicus* (GAXK00000000.1, 14-may-2018);  
868 **GBFB**: *Calanus finmarchicus* (GBFB00000000.1, 30-jan-2015); **GBXU**: *Calanus*  
869 *finmarchicus* (GBXU00000000.1, 13-jan-2015); **GBXT**: *Calanus glacialis*  
870 (GBXT00000000.1, 13-jan-2015); **HACJ**: *Calanus glacialis* (HACJ00000000.1, 29-  
871 sep-2017); **GBGO**: *Eurytemora affinis* (GBGO00000000.1, 07-jul-2015); **GEAN**:  
872 *Eurytemora affinis* (GEAN00000000.1, 16-nov-2016); **GFWO**: *Labidocera madurae*  
873 (GFWO00000000.1, 14-may-2018); **GFUD**: *Neocalanus flemingeri*  
874 (GFUD00000000.1, 14-may-2018) ; **GCJT**: *Paracyclopina nana* (GCJT00000000.1,  
875 20-jul-2015); **GFCI**: *Pleuromamma xiphias* (GFCI00000000.1, 18-dec-2017); **GBSZ**:  
876 *Tigriopus californicus* (GBSZ00000000.1, 02-feb-2015); **GBTC**: *Tigriopus*  
877 *californicus* (GBTC00000000.1, 02-feb-2015); **GCHA**: *Tigriopus japonicus*  
878 (GCHA00000000.1, 20-jul-2015) ; **GDFW**: *Tigriopus kingsejongensis*  
879 (GDFW00000000.1, 18-apr-2016); **HAHV**: *Tisbe holothuriae* (HAHV00000000.1,  
880 23-jan-2018).

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**Table 1.** Commonly used reference genes in copepods.

| Gene   | Species                               | Reference   | Validation |
|--|---------------------------------------|---|------------|
| $\beta$ -actin (ACT)                             | <i>Acartia tonsa</i>                  | (Nilsson et al., 2018)  | Yes        |
|  |                                       | (Aguilera et al., 2016)   | No         |
|  |                                       | (Nilsson et al., 2014, 2017; Petkeviciute et al., 2015; Rahlf et al., 2017)   | No         |
|  | <i>Calanus finmarchicus</i>           | (Hansen et al., 2008; Roncalli et al., 2016b)   | Yes        |
|  |                                       | (Lauritano et al., 2011a, 2013, 2016)   | Yes        |
|  |                                       | (Lauritano et al., 2015)  | Yes        |
|  |                                       | (Rahlf et al., 2017)  | No         |
|  |                                       | (Hwang et al., 2010b)   | No         |
|  |                                       | (Jeong et al., 2015)  | Yes        |
|  |                                       | (Jiang et al., 2013)  | No         |
|  |                                       | (Zhuang et al., 2017)   | Yes        |
|  |                                       | (Chan et al., 2014)   | No         |
|  | <i>Tigriopus californicus</i>         | (Lee et al., 2007, 2008b; Rhee et al., 2009; Seo et al., 2006b)   | No         |
|  |                                       | (Jeong et al., 2014; Lee et al., 2017)  | Yes        |
| Elongation factor 1 $\alpha$ (EFA)               | <i>Acartia tonsa</i>                  | (Nilsson et al., 2014, 2017; Petkeviciute et al., 2015)   | No         |
|  | <i>Calanus finmarchicus</i>           | (Hansen et al., 2008, 2010; Roncalli et al., 2016b)   | Yes        |
|  | <i>Calanus helgolandicus</i>          | (Lauritano et al., 2011a, 2013, 2016)   | Yes        |
|  | <i>Calanus sinicus</i>                | (Zhou et al., 2016)   | No         |
|  | <i>Lepeophtheirus salmonis</i>        | (Lauritano et al., 2015)  | Yes        |
|  |                                       | (Tribble et al., 2007)  | No         |
|  |                                       | (Borchel et al., 2018; Park et al., 2017)   | Yes        |
|  |                                       | (Jeong et al., 2015)  | Yes        |
|  |                                       | (Zhuang et al., 2017)   | Yes        |
|  |                                       | (Park et al., 2017)   | No         |
|  |                                       | (Jeong et al., 2015; Lee et al., 2017)  | Yes        |
| 18S rRNA (18S)                                   | <i>Calanus helgolandicus</i>          | (Lauritano et al., 2011a, 2013, 2016)   | Yes        |
|  | <i>Calanus sinicus</i>                | (Lee et al., 2017)  | Yes        |
|  |                                       | (Lauritano et al., 2015; Lee et al., 2017)  | Yes        |
|  | <i>Paracyclops nana</i>               | (Han et al., 2015a; Lauritano et al., 2015; Lee et al., 2012, 2016; Puthumana et al., 2017)   | No         |
|  |                                       | (Jeong et al., 2015; Lauritano et al., 2011a, 2013, 2016)   | Yes        |
|  | <i>Tigriopus japonicus</i>            | (Han et al., 2015a, 2015a; Hwang et al., 2016; Jeong et al., 2014, 2015, 2016; Kim et al., 2011, 2013a, 2013b, 2014, 2015a; Lee et al., 2012, 2016; Puthumana et al., 2017; Rhee et al., 2013; Yi et al., 2014) | No         |
| Ribosomal protein S16 (S16)                      | <i>Calanus finmarchicus</i>           | (Aruda et al., 2011; Roncalli et al., 2016b)  | No         |
|  | <i>Calanus sinicus</i>                | (Roncalli et al., 2016b)  | Yes        |
| Ribosomal protein S20 (S20)                      |                                       | (Zhou et al., 2016)   | No         |
| <i>Calanus helgolandicus</i>                     | (Lauritano et al., 2011a, 2013, 2016) | Yes   |            |
| <i>Calanus sinicus</i>                           | (Lauritano et al., 2015)              | Yes   |            |
| Ribosomal protein S7 (S7)                        | <i>Pseudodiaptomus poplesia</i>       | (Zhuang et al., 2017)   | Yes        |
|  | <i>Calanus helgolandicus</i>          | (Lauritano et al., 2011a, 2013, 2015, 2016)   | Yes        |
|  | <i>Calanus sinicus</i>                | (Lauritano et al., 2015)  | Yes        |
| Glyceraldehyde-3-phosphate dehydrogenase (GAPDH) | <i>Calanus helgolandicus</i>          | (Lauritano et al., 2011a, 2013, 2016)   | Yes        |
|  | <i>Calanus sinicus</i>                | (Zhou et al., 2016)   | No         |
|  |                                       | (Jeong et al., 2015; Lauritano et al., 2015)  | Yes        |
|  | <i>Paracyclops nana</i>               | (Jeong et al., 2015)  | Yes        |
|  | <i>Pseudodiaptomus annandalei</i>     | (Jiang et al., 2013)  | No         |
|  | <i>Tigriopus californicus</i>         | (Barreto et al., 2015)  | No         |
|  | <i>Tigriopus japonicus</i>            | (Lee et al., 2006, 2017; Seo et al., 2006a, 2006c)  | No         |
| Ubiquitin (UBI)                                  | <i>Calanus helgolandicus</i>          | (Lauritano et al., 2011a, 2013)   | Yes        |

|  |                              |   |     |
|--|------------------------------|---|-----|
|  | <i>Calanus sinicus</i>       | (Lauritano et al., 2015)                    | Yes |
| <i>Histone H3</i><br>( <i>HIST</i> )   | <i>Acartia tonsa</i>         | (Nilsson and Hansen, 2018)                  | Yes |
|  | <i>Calanus helgolandicus</i> | (Lauritano et al., 2011a, 2013, 2015, 2016) | Yes |
|  | <i>Calanus sinicus</i>       | (Lauritano et al., 2015)                    | Yes |
| <i>ATP synthase</i><br>( <i>ATPS</i> ) | <i>Acartia tonsa</i>         | (Nilsson and Hansen, 2018)                  | Yes |
|  | <i>Calanus helgolandicus</i> | (Lauritano et al., 2011a, 2013, 2015)       | Yes |
|  | <i>Calanus sinicus</i>       | (Lauritano et al., 2015)                    | Yes |

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**Table 2.** Commonly-used biomarkers for transcriptional analysis with real-time quantitative PCR in copepods. Copepod species abbreviations: *Acartia tonsa* (*A. tonsa*); *Calanus finmarchicus* (*C. finmarchicus*); *Calanus glacialis* (*C. glacialis*); *Calanus helgolandicus* (*C. helgolandicus*); *Eurytemora affinis* (*E. affinis*); *Paracyclops nana* (*P. nana*); *Pseudodiaptomus annandalei* (*P. annandalei*); *Pseudodiaptomus poplesia* (*P. poplesia*); *Tigriopus californicus* (*T. californicus*); *Tigriopus japonicus* (*T. japonicus*);

| Biomarker  | Stressors   | Effect                                  | Species                 | References                      |
|--|---|---|-------------------------|---------------------------------|
| <i>Aldehyde dehydrogenases (ALDHs)</i>                 | Diatom toxins   | Isoform dependent                       | <i>C. helgolandicus</i> | (Lauritano et al., 2011b, 2016) |
| <i>Catalases (CATs)</i>                                | Naphthalene   | No significant change                   | <i>C. finmarchicus</i>  | (Hansen et al., 2008)           |
|  | Toxic diatoms   | No significant change                   | <i>C. helgolandicus</i> | (Lauritano et al., 2011b)       |
|  |   | Elevated expression                     | <i>C. helgolandicus</i> | (Lauritano et al., 2016)        |
| <i>Cell cycle and apoptosis regulatory 1 protein</i>   | Toxic diatoms   | No significant change                   | <i>C. helgolandicus</i> | (Lauritano et al., 2011b)       |
| <i>Cellular apoptosis susceptibility protein</i>       | Toxic diatoms   | Decreased expression                    | <i>C. helgolandicus</i> | (Lauritano et al., 2016)        |
| <i>Corticotropin Releasing Hormone Binding Protein</i> | Salinity  | Elevated expression                     | <i>T. japonicus</i>     | (Lee et al., 2008c)             |
| <i>COI</i>   | Copper  | Elevated expression                     | <i>T. japonicus</i>     | (Weaver et al., 2016)           |
|  | Diatom toxins   | Isoform dependent                       | <i>C. helgolandicus</i> | (Lauritano et al., 2011b, 2016) |
|  | Naphthalene   | Concentration dependent down regulation | <i>C. finmarchicus</i>  | (Hansen et al., 2008)           |
|  | Diethanolamine  | Decreased expression                    | <i>C. finmarchicus</i>  | (Hansen et al., 2010)           |
| <i>Cytochromes P450 (CYPs)</i>                         | Water-soluble fractions of crude oil and oil droplets                       | Stressor dependent                      | <i>C. finmarchicus</i>  | (Hansen et al., 2009)           |
|  | Polycyclic aromatic hydrocarbons, water accommodated fractions of crude oil | Elevated expression                     | <i>P. nana</i>          | (Han et al., 2015a)             |
| <i>Delta-1 pyrroline-5-carboxylase reductase</i>       | Salinity  | No significant change                   | <i>T. californicus</i>  | (Willett and Burton, 2002)      |
| <i>Delta-pyrroline-5-carboxylate synthase</i>          | Salinity  | No significant change                   | <i>T. californicus</i>  | (Willett and Burton, 2002)      |
| <i>DnaJ homolog</i>                                    | Increasing temperature  | No significant change                   | <i>C. finmarchicus</i>  | (Smolina et al., 2015)          |
|  | Crowding  | No significance                         | <i>A. tonsa</i>         | (Nilsson et al., 2017)          |
| <i>Ferritin</i>  | Quiescence  | Time dependent peaks                    | <i>A. tonsa</i>         | (Nilsson et al., 2014)          |
|  | Epibiont infestation  | Elevated expression                     | <i>A. tonsa</i>         | (Petkeviciute et al., 2015)     |

|   |   |   |   |   |
|---|---|---|---|---|
|   | CO <sub>2</sub> pressure  | Acclimatization dependent (elevated expression for coastal copepods compared to estuarine copepods) | <i>A. tonsa</i>   | (Aguilera et al., 2016)   |
|   | Nickel  | Elevated expression   | <i>P. annandalei</i>  | (Jiang et al., 2013)  |
| <i>Glucose-Regulated Protein, 78kDa</i>                         | Temperature; salinity shock   | Elevated expression   | <i>E. affinis</i>   | (Xuereb et al., 2012)   |
| <i>Glutamate Dehydrogenase</i>                                  | Increasing temperature  | Elevated expression, time dependent   | <i>C. finmarchicus</i>  | (Smolina et al., 2015)  |
|   | Salinity  | No significant change   | <i>T. californicus</i>  | (Willett and Burton, 2003)  |
| <i>Glutathione Peroxidases (GPxs)</i>                           | Pyrene; naphthalene   | No significant change   | <i>P. poplesia</i>  | (Zhuang et al., 2017)   |
| <i>Glutathione Reductase (GR)</i>                               | Salinity  | Elevated for high salinities, decreased for low   | <i>T. japonicus</i>   | (Seo et al., 2006b)   |
|   | Heavy metals (Cu, Mn)   | Elevated expression   | <i>T. japonicus</i>   | (Weaver et al., 2016)   |
|   | Hydrogen peroxide   | Concentration and time dependent  | <i>T. japonicus</i>   | (Seo et al., 2006b)   |
|   | Endocrine disrupting chemicals  | Concentration and stressor dependent  | <i>T. japonicus</i>   | (Lee et al., 2006)  |
|   | Toxic diatoms and dinoflagellates   | Time and/or isoform dependent   | <i>C. helgolandicus</i> ,<br><i>C. finmarchicus</i>   | (Lauritano et al., 2011b., 2016; Roncalli et al., 2016b)  |
|   | Heavy metals (Cu, Mn, Ag, As, Cd)   | Time, metal-type and isoform dependent  | <i>T. japonicus</i>   | (Lee et al., 2007, 2008b)   |
| <i>Glutathione S-transferases (GSTs)</i>                        | Naphthalene   | Elevated expression   | <i>C. finmarchicus</i>  | (Hansen et al., 2008)   |
|   | Hydrogen peroxide   | Elevated expression (peak)  | <i>P. poplesia</i>  | (Zhuang et al., 2017)   |
|   | Pyrene  | Time dependent  | <i>T. japonicus</i>   | (Lee et al., 2007)  |
|   | Water-soluble fractions and water accommodated fractions of crude oil, oil droplets | Time and isoform dependent  | <i>P. poplesia</i>  | (Zhuang et al., 2017)   |
|   | Diethanolamine  | Concentration - and time dependent  | <i>C. finmarchicus</i> ,<br><i>C. glacialis</i>   | (Hansen et al., 2009, 2011)   |
|   | Triphenyltin chloride   | Concentration and time dependent  | <i>C. finmarchicus</i>  | (Hansen et al., 2010)   |
|   | Decreased expression  |   | <i>T. japonicus</i>   | (Yi et al., 2014)   |
| <i>Glutathione Synthase</i>                                     | Toxic diatoms   | No significant change   | <i>C. helgolandicus</i>   | (Lauritano et al., 2011b)   |
|   | Diethanolamine  | Concentration dependent   | <i>C. finmarchicus</i>  | (Hansen et al., 2010)   |
|   | Temperature   | No significant change   | <i>T. japonicus</i>   | (Rhee et al., 2009)   |
|   |   | Elevated expression for high and, decreased expression for low temperatures                         | <i>T. japonicus</i>   | (Seo et al., 2006c)   |
| <i>Heat-shock protein 10, 20, 21, 22, 40, 60, 94 or 105 kDa</i> | Salinity  | No significant change   | <i>T. japonicus</i>   | (Seo et al., 2006b)   |
|   | Endocrine disrupter chemicals   | Stressor dependent  | <i>T. japonicus</i>   | (Seo et al., 2006b)   |
|   | Handling  | Elevated expression   | <i>C. finmarchicus</i>  | (Aruda et al., 2011)  |
|   | Diapause  | Elevated expression   | <i>C. finmarchicus</i>  | (Aruda et al., 2011)  |
|   | Toxic diatoms   | Decreased expression  | <i>C. helgolandicus</i>   | (Lauritano et al., 2011b)   |
| <i>Heat-shock protein 70kDa (HSP70)</i>                         | Heat-shock / increasing temperature   | Elevated expression (acclimatization result in lower expression)                                    | <i>A. tonsa</i><br><i>E. affinis</i><br><i>T. japonicus</i><br><i>T. californicus</i><br><i>C. finmarchicus</i><br><i>C. finmarchicus</i> | (Petkeviciute et al., 2015; Rahlf et al., 2017)<br>(Rahlf et al., 2017)<br>(Rhee et al., 2009)<br>(Chan et al., 2014)<br>(Voznesensky et al., 2004)<br>(Smolina et al., 2015) |
|   |   | No significance   |   |   |

|   |   |   |   |  |
|---|---|---|---|--|
|   | Handling                                    | Elevated expression   | <i>C. finmarchicus</i><br><i>E. affinis</i> | (Aruda et al., 2011; Rahlf et al., 2017)<br>(Rahlf et al., 2017) |
|   | Salinity                                    | Concentration dependent, but in general elevated expression outside acclimatization range           | <i>A. tonsa</i>                             | (Petkeviciute et al., 2015)                                      |
|   | Crowding                                    | No significance   | <i>A. tonsa</i>                             | (Nilsson et al., 2017)   |
|   | Quiescence                                  | Time dependent  | <i>A. tonsa</i>                             | (Nilsson et al., 2014)   |
|   | CO <sub>2</sub> pressure                    | Acclimatization dependent (elevated expression for coastal copepods compared to estuarine copepods) | <i>A. tonsa</i>                             | (Aguilera et al., 2016)  |
|   | Shallow active vs. deep diapausing copepods | Elevated expression for active copepods in shallow waters   | <i>C. finmarchicus</i>                      | (Aruda et al., 2011)   |
|   | Naphthalene                                 | No significant change   | <i>C. finmarchicus</i>                      | (Hansen et al., 2008)  |
|   | Toxic diatoms                               | Elevated expression   | <i>C. helgolandicus</i>                     | (Lauritano et al., 2016)   |
|   |   | No significant change   | <i>C. helgolandicus</i>                     | (Lauritano et al., 2011b)  |
|   | Heavy metals (Cu, Ag, Zn)                   | Elevated expression   | <i>T. japonicus</i>                         | (Rhee et al., 2009)  |
|   | Endocrine disrupting chemicals              | Concentration and /or stressor dependent  | <i>T. japonicus</i>                         | (Rhee et al., 2009; Yi et al., 2014)                             |
| <i>Heat-shock protein 90kDa (HSP90)</i>           | Heat shock / increasing temperature         | No significant change   | <i>T. japonicus</i>                         | (Rhee et al., 2009)  |
|   |   | Elevated expression   | <i>E. affinis</i>                           | (Xuereb et al., 2012)  |
|   |   | Elevated expression   | <i>A. tonsa</i>                             | (Petkeviciute et al., 2015)                                      |
|   | Salinity shock                              | Concentration dependent, elevated expression outside acclimatization range                          | <i>A. tonsa</i>                             | (Petkeviciute et al., 2015)                                      |
|   | Naphthalene                                 | No significant change   | <i>E. affinis</i>                           | (Xuereb et al., 2012)  |
|   | Crowding                                    | No significant change   | <i>C. finmarchicus</i>                      | (Hansen et al., 2008)  |
|   |   |   | <i>A. tonsa</i>                             | (Nilsson et al., 2017)   |
| <i>Inhibitor of apoptosis protein</i>             | Toxic diatoms                               | Decreased expression  | <i>C. helgolandicus</i>                     | (Lauritano et al., 2011b)  |
| <i>Methylmalonate-semialdehyde dehydrogenase</i>  | Pyrene                                      | Elevated expression   | <i>P. poplesia</i>                          | (Zhuang et al., 2017)  |
| <i>Myohemerythrin-1</i>                           | Nickel                                      | Elevated expression   | <i>P. annandalei</i>                        | (Jiang et al., 2013)   |
| <i>Nucleosome Assembly Protein 1</i>              | Increasing temperature                      | Elevated expression, time dependent   | <i>C. finmarchicus</i>                      | (Smolina et al., 2015)   |
| <i>p53 tumor suppressor protein</i>               | Endocrine disrupting chemicals              | Elevated expression   | <i>T. japonicus</i>                         | (Hwang et al., 2010a)  |
| <i>Peroxiredoxin-6</i>                            | Naphthalene                                 | Elevated expression   | <i>P. poplesia</i>                          | (Zhuang et al., 2017)  |
| <i>Ras-related C3 botulinum toxin substrate 1</i> | Naphthalene                                 | Elevated expression   | <i>P. poplesia</i>                          | (Zhuang et al., 2017)  |
| <i>Retinoid X receptor</i>                        | Triphenyltin chloride                       | Decreased expression  | <i>T. japonicus</i>                         | (Yi et al., 2014)  |

|                                     |                                |  |                         |                           |
|-------------------------------------|--------------------------------|--|-------------------------|---------------------------|
| <i>Ribosomal protein L13</i>        | Nickel                         | Elevated expression                        | <i>P. annandalei</i>    | (Jiang et al., 2013)      |
| <i>Ribosomal Protein S11</i>        | Increasing temperature         | No significant change                      | <i>C. finmarchicus</i>  | (Smolina et al., 2015)    |
| <i>Separase</i>                     | Nickel                         | Elevated expression                        | <i>P. annandalei</i>    | (Jiang et al., 2013)      |
|                                     | Heavy metals (Cu, Zn, Ag)      | Elevated expression at high concentrations | <i>T. japonicus</i>     | (Kim et al., 2011)        |
|                                     | Endocrine disrupting chemicals | Concentration and stressor dependent       | <i>T. japonicus</i>     | (Kim et al., 2011)        |
| <i>Superoxide dismutases (SODs)</i> | Naphthalene                    | No significant change                      | <i>C. finmarchicus</i>  | (Hansen et al., 2008)     |
|                                     | Toxic diatoms                  | No significant change                      | <i>C. helgolandicus</i> | (Lauritano et al., 2011b) |
|                                     |                                | Decreased expression                       | <i>C. helgolandicus</i> | (Lauritano et al., 2016)  |
|                                     | Nickel                         | Elevated expression (dose response)        | <i>P. annandalei</i>    | (Jiang et al., 2013)      |
| <i>Toll-like receptor</i>           | Increasing temperature         | Elevated expression                        | <i>T. californicus</i>  | (Chan et al., 2014)       |
| <i>Tubulins</i>                     | Toxic diatoms                  | Decreased expression                       | <i>C. helgolandicus</i> | (Lauritano et al., 2011b) |
|                                     | Nickel                         | Elevated expression                        | <i>T. japonicus</i>     | (Jiang et al., 2013)      |
| <i>Ubiquitin</i>                    | Naphthalene                    | No significant change                      | <i>C. finmarchicus</i>  | (Hansen et al., 2008)     |
|                                     | Diethanolamine                 | Concentration dependent                    | <i>C. finmarchicus</i>  | (Hansen et al., 2010)     |
| <i>Vitellogenin</i>                 | Heavy metals (Cd, Ag, As, Cu)  | Metal dependent                            | <i>T. japonicus</i>     | (Lee et al., 2011)        |
|                                     |                                | Elevated expression                        | <i>P. nana</i>          | (Hwang et al., 2010b)     |

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**Table 3.** Copepod transcriptome assemblies. Type is the type of Illumina sequencing in terms of read length (bp) and if it is paired-end (PE), or single-end (SE) sequencing. Reads used in assembly are reported in millions; the number of pairs is reported when paired reads were generated. Software gives the used assembly software (“CLC” indicates CLC Genomics Workbench, Trinity versions indicated if reported). Contigs is the number of resulting contigs within the assembly. BUSCO analyses were performed in August 2018 using publicly accessible NCBI Transcriptome Shotgun Assemblies (TSA); the rounded percent of complete (C, includes both single and duplicated) and complete single-copy (S) transcripts is shown.

| Species                      | Sample   | Platform             | Type   | Reads Used                    | Software                             | Contigs | BUSCO %         | NCBI accession      | Resources   |
|------------------------------|--|----------------------|--------|-------------------------------|--------------------------------------|---------|-----------------|---------------------|---|
| <i>Acartia tonsa</i>         | 1 adult female   | Illumina NextSeq     | 150 PE | ~350 M                        | Trinity 2.3.2                        | 60,662  | 99% C<br>45% S  | <b>GFWY00000000</b> | (Nilsson et al., 2018)  |
| <i>Acartia tonsa</i>         | Multiple eggs, nauplii, copepodites, and adults                                | Illumina Next/Mi-Seq | 150 PE | 111 M                         | Trinity 2.5.1                        | 119,439 | 88% C<br>45% S  | <b>HAGX00000000</b> | (T.S. Jørgensen et al., unpublished data)   |
| <i>Apocyclops roysi</i>      | Multiple eggs, nauplii, copepodites, and adults                                | Illumina NextSeq     | 150 SE | 203 M                         | Trinity v. 2.5.1                     | 75,477  | na              | <b>GHAJ00000000</b> | (Jørgensen et al., in press),   |
| <i>Calanus finmarchicus</i>  | Multiple eggs, nauplii, copepodites, and adults                                | Illumina HiSeq       | 100 PE | 28 M per stage<br>640 M total | Trinity r2012-03-17-<br>IU ZIH TUNED | 206,012 | 99% C<br>72% S  | <b>GAXK00000000</b> | (Lenz et al., 2014)   |
| <i>Calanus finmarchicus</i>  | 3 CV stage copepodites   | Illumina HiSeq       | 100 PE | 93 M                          | Trinity r2012-06-08                  | 241,140 | 97% C<br>61% S  | <b>GBFB00000000</b> | (Tarrant et al., 2014)  |
| <i>Calanus finmarchicus</i>  | 5 individuals exposed to short – and long-term thermal stress (3 temperatures) | Ion-Torrent          | N/A    | 5 M                           | Trinity r2013-08-14                  | 28,954  | 13% C<br>10% S  | <b>GBXU00000000</b> | (Smolina et al., 2015)  |
| <i>Calanus glacialis</i>     | 5 individuals short – and long-term stress exposed to 3 temperatures.          | Ion-Torrent          | N/A    | 3.5 M                         | Trinity r2013-08-14                  | 36,880  | 21% C<br>14% S  | <b>GBXT00000000</b> | (Smolina et al., 2015)  |
| <i>Calanus glacialis</i>     | 10 CV stage copepodites  | 454 GS FLX           | N/A    | 720 K                         | Mira v 3.0                           | 54,344  | 17% C<br>12% S  | <b>HACJ00000000</b> | (Ramos et al., 2015)  |
| <i>Eurytemora affinis</i>    | Females exposed to two strains of <i>Vibrio</i> and females not exposed        | Illumina HiSeq       | 100 PE | 300 M                         | Trinity r2013-08-14                  | 138,088 | 100% C<br>65% S | <b>GBGO00000000</b> | (Almada and Tarrant, 2016)  |
| <i>Eurytemora affinis</i>    | Pooled adult males and females   | Illumina HiSeq       | 100 PE | Not reported                  | Trinity r2013-11-10                  | 107,445 | 71% C<br>29% S  | <b>GEAN00000000</b> | Unpublished, Munro, J.B., Posavi, M., Brady, A., Orvis, J., Nadendla, S., Abolude, K., Kumari, P., Shetty, A. Lee, C.E. and Silva, J.C. |
| <i>Labidocera madurae</i>    | 5-6 adult females<br>15-26 CIII-CV stage copepodites                           | Illumina NextSeq     | 150 PE | 530 M                         | Trinity v. 2.0.6                     | 211,002 | 99% C<br>59% S  | <b>GFWO00000000</b> | (Roncalli et al., 2017a)  |
| <i>Neocalanus flemingeri</i> | 1 adult female   | Illumina NextSeq     | 150 PE | 150 M                         | Trinity v. 2.0.6 & CAP3              | 140,841 | 98% C<br>62% S  | <b>GFUD00000000</b> | (Roncalli et al., 2018a)  |

|                                  |   |                  |        |       |                  |         |                |                     |                           |
|----------------------------------|---|------------------|--------|-------|------------------|---------|----------------|---------------------|---------------------------|
| <i>Paracyclops nana</i>          | Unknown - adults  | Illumina HiSeq   | 100 PE | 200 M | Trinity          | 60,687  | 85% C<br>64% S | <b>GCJT00000000</b> | (Lee et al., 2015)        |
| <i>Pleuromamma xiphias</i>       | Pooled adult males and females                            | Illumina HiSeq   | 125 PE | 267 M | Trinity v. 2.1.1 | 554,991 | 85% C<br>76% S | <b>GFCI00000000</b> | (Maas et al., 2018)       |
| <i>Temora longicornis</i>        | Pooled adult males and females (7:10 males:females ratio) | Illumina HiSeq   | 150 PE | 460 M | Trinity v. 2.1.1 | 179,569 | 80% C<br>67% S | <b>GGQN01000000</b> | (Semmouria et al., 2019)  |
| <i>Tigriopus californicus</i>    | 4-500 mixed developmental stages. San Diego population    | Illumina HiSeq   | 100 PE | 128 M | CLC v. 5.1       | 36,620  | 63% C<br>61% S | <b>GBSZ00000000</b> | (Barreto et al., 2014)    |
| <i>Tigriopus californicus</i>    | 4-500 mixed developmental stages. Santa Cruz population   | Illumina HiSeq   | 100 PE | 49 M  | CLC v. 5.1       | 43,077  | 62% C<br>61% S | <b>GBTC00000000</b> | (Barreto et al., 2014)    |
| <i>Tigriopus japonicus</i>       |   | Illumina HiSeq   | 100 PE | 108 M | Trinity          | 54,758  | 99% C<br>55% S | <b>GCHA00000000</b> | (Kim et al., 2015b)       |
| <i>Tigriopus kingsejongensis</i> | 200 adults  | Illumina HiSeq   | 150 PE | 140 M | Trinity v. 2.0.6 | 38,250  | 98% C<br>77% S | <b>GDFW00000000</b> | (Kim et al., 2016)        |
| <i>Tisbe holothuriae</i>         | Multiple eggs, nauplii, copepodites, and adults           | Illumina NextSeq | 150 PE | 162 M | Trinity v. 2.2.0 | 218,711 | 46% C<br>28% S | <b>HAHV00000000</b> | Roskilde University, 2017 |

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1603**Table 4.** Overview of sequencing depth and number of replicates for differential gene expression assessment in copepods by RNA sequencing. M: million reads. # rep: replicates used per treatment. Type: type of used animals, field-caught or culture-reared. Seq. depth: sequencing depth.

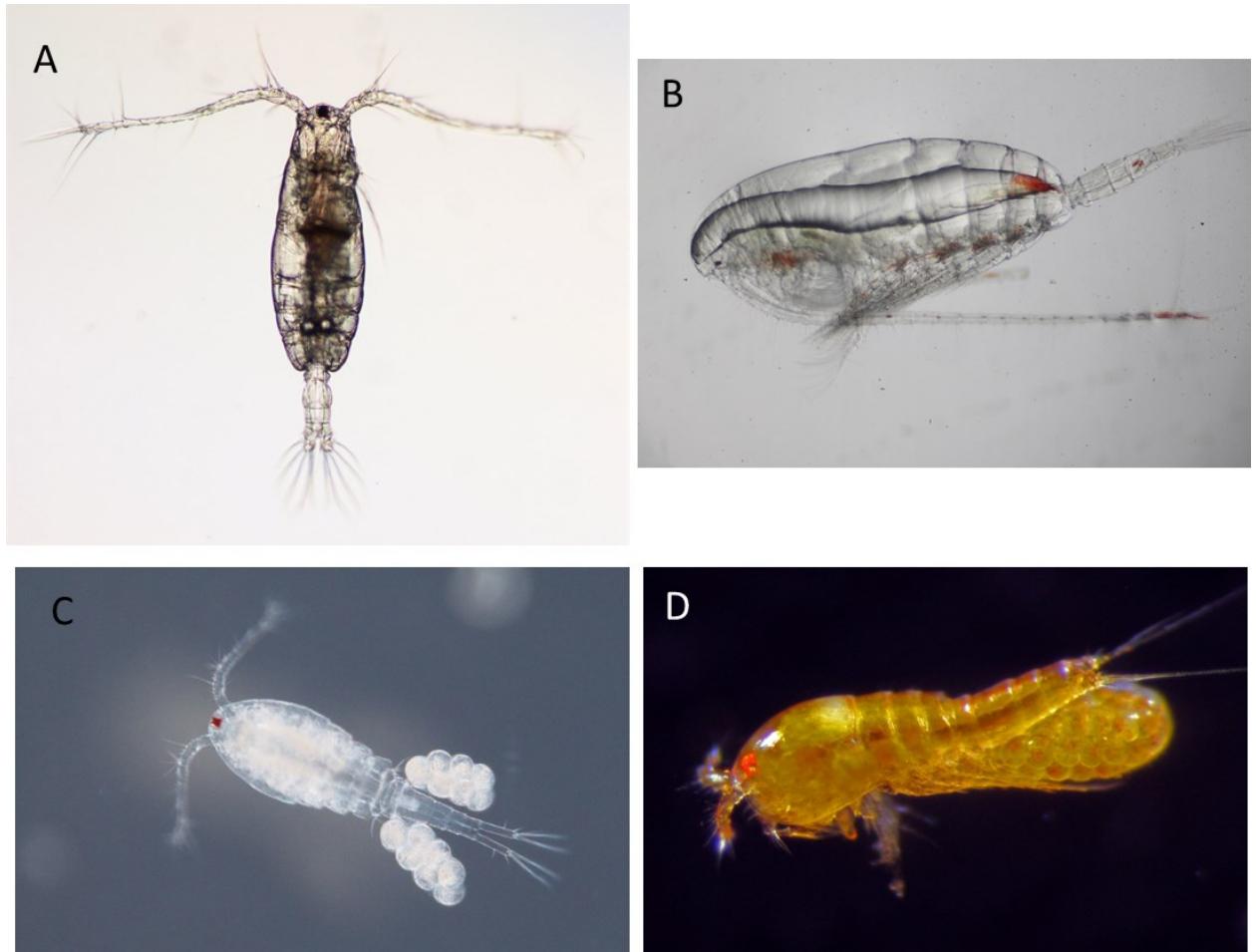
| Species                       | Treatments  | # Individuals                                     | Seq. depth | Method  | # rep.     | Type                   | Reference                  |
|-------------------------------|---|---|------------|---|------------|------------------------|----------------------------|
| <i>Acartia tonsa</i>          | Control, salinity shock, handling stress  | 10 adults   | ~25 M      | Kallisto/Sleuth                                 | 3          | Culture                | (Nilsson et al., 2018)     |
| <i>Calanus finmarchicus</i>   | Control and feeding on the neurotoxic <i>Alexandrium fundyense</i>  | ~74-80 pooled nauplii                             | ~20 M      | Bowtie, Custom script, edgeR                    | 3          | Field                  | (Roncalli et al. 2017b)    |
| <i>Calanus finmarchicus</i>   | Control, low-dose – and high-dose of dinoflagellate, two times  | 10 adult females                                  | ~26 M      | Bowtie, edgeR                                   | 3          | Field                  | (Roncalli et al., 2016a)   |
| <i>Calanus finmarchicus</i>   | Short – and long-term stress exposed to 3 temperatures  | 5 individuals                                     | ~0.8M      | Rsubread, featureCounts, DESeq2                 | 1          | Field                  | (Smolina et al., 2015)     |
| <i>Calanus finmarchicus</i>   | CIV to CV molt (collected on day 3 and 10)  | 3 CV stage copepodites                            | ~12M       | Bowtie, RSEM, edgeR                             | 4          | Field/Culture          | (Tarrant et al., 2014)     |
| <i>Calanus glacialis</i>      | Four pH treatments  | 10 nauplii  | ~28 M      | BWA, Custom script (De Wit et al. 2012), DESeq2 | 5-6        | Field                  | (Bailey et al., 2017)      |
| <i>Calanus glacialis</i>      | Control and 5 temperature treatments  | 10 CV stage copepodites                           | ~60-180K   | IDEG6   | 1          | Field                  | (Ramos et al., 2015)       |
| <i>Calanus glacialis</i>      | Short – and long-term stress exposed to 3 temperatures  | 5 individuals                                     | ~0.6M      | Rsubread, featureCounts, DESeq2                 | 1          | Field                  | (Smolina et al., 2015)     |
| <i>Eurytemora affinis</i>     | Control, exposure to acetone, pyriproxyfen and chlordanone. 3 experimental replicates pooled for sequencing | 400-500 females or males                          | ~11- 50 M  | Bowtie2, RSEM, EBseq                            | 1          | Field                  | (Legrand et al., 2016)     |
| <i>Eurytemora affinis</i>     | Control, exposure to <i>Vibrio</i> sp. (F10) and <i>V. ordalii</i>  | 20 females  | ~25 M      | Bowtie, RSEM, edgeR                             | 4          | Culture                | (Almada and Tarrant, 2016) |
| <i>Neocalanus flemingeri</i>  | Diapause, emergence phase from diapause (10 weeks)  | 6 females (diapause), 4 females (emergence phase) | ~10-22 M   | Kallisto/ edgeR                                 | 3 per week | Field                  | (Roncalli et al., 2018b)   |
| <i>Pseudocalanus acuspes</i>  | 7 pCO <sub>2</sub> exposure conditions  | 28-74 adults                                      | ~12-25M    | Custom ANCOVA analysis                          | 2          | Field                  | (De Wit et al., 2016)      |
| <i>Temora longicornis</i>     | 4 temperature treatments  | 50 adults ( 7:10 males:females ratio)             | Unknown    | Bowtie, TopHat, Cuffmerge, HTseq, edgeR         | 3          | Field                  | (Semmouria et al., 2019)   |
| <i>Tigriopus californicus</i> | 2 populations, 2 temperatures   | >300 copepods, mixed stages                       | 2.5-7 M    | CLC genomics workbench, Z-test                  | 1          | Field                  | (Schoville et al., 2012)   |
| <i>Tigriopus californicus</i> | 2 distinct populations and 2 interpopulation crosses  | 400-500 pooled                                    | ~3-40 M    | CLC genomics workbench, edgeR                   | 2-3        | Field                  | (Barreto et al., 2014)     |
| <i>Tigriopus californicus</i> | 2 populations, 3 salinities   | 50 pooled   | ~24 M      | RSEM, DESeq2                                    | 3          | Cultured 2 generations | (DeBiasse et al., 2018)    |

|                               |   |            |          |                               |                 |               |                          |
|-------------------------------|---|------------|----------|-------------------------------|-----------------|---------------|--------------------------|
|                               |   |            |          |                               |                 |               |                          |
| <i>Tigriopus californicus</i> | Control (parent populations) and F4 crossed population exposed control and high-temperature selection conditions. | 30 adults  | Unknown  | RSEM, limma                   | 3 genetic lines | Field/Culture | (Kelly et al., 2017)     |
| <i>Tigriopus californicus</i> | 4 sites, 2 thermal regimes, 2 times   | 100 pooled | ~7.6-29M | CLC genomics workbench, edgeR | 2               | Field         | (Lima and Willett, 2017) |

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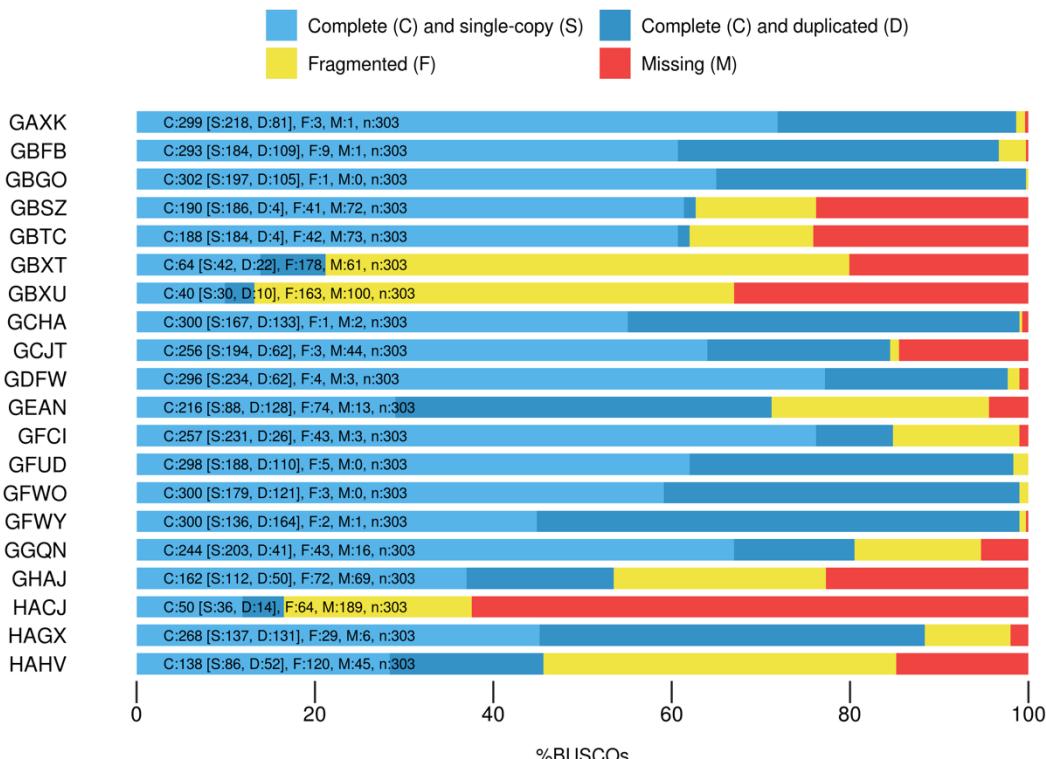
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1610 **Figure 1.** Examples of diverse marine copepods utilized in physiological studies. (A)  
1611 The calanoid *Acartia tonsa* male, (B) The calanoid *Calanus glacialis* C5 copepodite  
1612 with prominent oil sac, (C) The cyclopoid *Apocyclops royi* egg-bearing female (D)  
1613 The harpacticoid *Tigriopus japonicus* egg-bearing female. Of these, *Tigriopus spp.*  
1614 (particularly *T. californicus*, not shown) have been extensively developed as a model  
1615 for studies of molecular evolution and plasticity. The others represent a growing  
1616 diversity of species for which molecular physiology studies are being driven by their  
1617 ecological importance. Photos courtesy of Dr. Minh Thi Thui Vu (A), A.M.T. (B),  
1618 Dr. Hans van Someren Gréve (C), and Dr. Greg Rouse (D).  
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## BUSCO Assessment Results



**Figure 2.** BUSCO analysis of the following copepod transcriptomes (with NCBI accession numbers, alphabetized by scientific name as in Table 3): **GFWY**: *Acartia tonsa* (GFWY00000000.1, 27-sep-2017); **HAGX**: *Acartia tonsa* (HAGX00000000.1, 29-sep-2017); **GAXK**: *Calanus finmarchicus* (GAXK00000000.1, 14-may-2018); **GBFB**: *Calanus finmarchicus* (GBFB00000000.1, 30-jan-2015); **GBXU**: *Calanus finmarchicus* (GBXU00000000.1, 13-jan-2015); **GBXT**: *Calanus glacialis* (GBXT00000000.1, 13-jan-2015); **HACJ**: *Calanus glacialis* (HACJ00000000.1, 29-sep-2017); **GBGO**: *Eurytemora affinis* (GBGO00000000.1, 07-jul-2015); **GEAN**: *Eurytemora affinis* (GEAN00000000.1, 16-nov-2016); **GFWD**: *Labidocera madurae* (GFWD00000000.1, 14-may-2018); **GFUD**: *Neocalanus flemingeri* (GFUD00000000.1, 14-may-2018); **GCJT**: *Paracyclops nana* (GCJT00000000.1, 20-jul-2015); **GFCI**: *Pleuromamma xiphias* (GFCI00000000.1, 18-dec-2017); **GGQN**: *Temora longicornis* (GGQN00000000, 12-jun-2018); **GBSZ**: *Tigriopus californicus* (GBSZ00000000.1, 02-feb-2015); **GBTC**: *Tigriopus californicus* (GBTC00000000.1, 02-feb-2015); **GCHA**: *Tigriopus japonicus* (GCHA00000000.1, 20-jul-2015); **GDFW**: *Tigriopus kingsejongensis* (GDFW00000000.1, 18-apr-2016); **HAHV**: *Tisbe holothuriae* (HAHV00000000.1, 23-jan-2018).