

Using Spatially-Identified Effective Impervious Area to Target Green Infrastructure Retrofits: A Modeling Study in Knoxville, TN

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1 **Abstract**

2 There is a need for enhanced guidance in siting distributed, infiltrative green infrastructure (GI)
3 practices, especially in densely developed urban watersheds where retrofits come at a high cost. To
4 maximize the hydrologic benefit of GI practices on urban streams, the disconnection of Effective
5 Impervious Areas (EIA), or those impervious areas hydraulically connected to the stormwater network,
6 has been identified as a strategic management approach that is expected to have the greatest impact.
7 The overall effect of full disconnection of spatially-identified EIA on watershed hydrology is uncertain
8 because this type of full disconnection is rarely brought to full-scale implementation. In this study,
9 spatial EIA identification is used to parametrize an urban runoff model using the United States
10 Environmental Protection Agency's Storm Water Management Model (SWMM). The calibrated model is
11 used to assess runoff reductions resulting from GI practices distributed through the watershed via
12 different placement strategies, both spatially-informed and not. Full treatment of the spatially-
13 identified EIA using bioretention cells was compared to two scenarios treating the same area of
14 impervious surfaces, but with random placement either among all impervious areas or placement
15 focused in areas of higher imperviousness. Model results indicate that substantially higher runoff
16 reduction could be realized by targeting EIA, with a median runoff reduction of nearly 30% more than
17 other treatment scenarios across storm events ranging from 1.27 to 20.7 mm using this strategic siting.
18 Further improvements in optimizing distributed infiltrative GI practice placement are needed and
19 targeting of spatially-identified EIA appears to be a viable method for increasing the hydrologic
20 improvements realized through watershed scale implementations.

21 **Keywords:** Stormwater, green infrastructure, distributed restoration, effective impervious area

22 **1. Introduction**

23 In an effort to protect and restore ecological health and pre-development hydrology in urban streams
24 worldwide, watershed managers seek to reinstate more natural flow regimes in highly developed
25 watersheds. Streamflow has been recognized as a “master variable” defining ecological potential in
26 riverine systems (Poff et al. 1997), and the impacts of impervious surfaces on urban hydrology and a
27 range of other symptoms of degradation have been well established (Brabec et al. 2002, Leopold 1968,
28 Shuster et al. 2005, Walsh et al. 2005b) The value of healthy streams as both financial and cultural assets
29 (Assessment 2005) has led urban watershed managers to seek strategies that both maintain flood
30 protection from stormwater but also undo the effects of intense development on local hydrology.
31 Increasingly worldwide, this is accomplished with distributed, nature-inspired, infiltrative green
32 infrastructure (GI) retrofits. Specifically, these at-source, smaller-scale, infiltrative surface runoff
33 treatment practices are being used in urban watersheds in attempts to restore pre-development runoff
34 frequencies and volumes (Bernhardt and Palmer 2007). By shifting hydrology back to more natural
35 patterns and addressing runoff production at the source, it is believed that urban stream systems will
36 have a better chance of supporting thriving ecosystems without the need for ongoing active channel
37 restoration projects; that is, restoration efforts will better match the scale of the degrading process
38 (Walsh et al. 2005a).

39 Research on the effectiveness of distributed GI practices to collectively produce this hydrologic shift at
40 the watershed scale is limited and it remains as an important research focus area (Jefferson et al. 2017).
41 Several studies have begun to add to the knowledgebase and demonstrated the complexities of
42 managing urban runoff with distributed GI practices. Modeling has shown that site runoff dynamics are
43 sensitive to the location, type, and number of practices installed (Gilroy and McCuen 2009). For
44 instance, comparisons of hydrologic response to rainfall for developed catchments with different

45 stormwater management strategies (centralized vs. distributed) revealed that distributed GI practices
46 can yield more natural runoff dynamics than centralized stormwater systems (Loperfido et al. 2014). A
47 growing number of monitoring studies have also been conducted to measure the efficacy of distributed
48 GI retrofit strategies at several different scales. Notably, though not exhaustively, these studies include
49 investigation at the small residential scale (0.53 ha) (Page et al. 2015), larger development tracts (2-6 ha)
50 (Bedan and Clausen 2009), and catchment-wide studies with Shepherd Creek in Cincinnati, USA (180 ha)
51 (Roy et al. 2014, Shuster and Rhea 2013), and Little Stringybark Creek (450 ha) in Melbourne, Australia
52 (Walsh et al. 2015). Additional examples are presented by Jefferson et al. (2017). The results of these
53 studies indicate improvements in stated goals, though they are difficult to compare due to differences in
54 study approaches, stormwater control measures (SCMs) employed, the timeline of installation, and
55 siting strategies. These are nonetheless valuable studies, as the lack of adequate supporting evidence
56 for distributed SCMs to effectively address watershed-wide goals remains an impediment to widespread
57 use of this multi-benefit management strategy (Roy et al. 2008).

58 Another impediment to wider use of distributed SCM retrofits in existing urban areas is that these
59 practices carry a higher cost related to land value in densely developed areas. Costs are further
60 increased when you consider the opportunity costs of land used for stormwater management as
61 opposed to other uses in the urban setting (Roy et al. 2008). Because of this, distributed GI practices
62 have mostly been sited in an opportunistic, empirical manner where public land is available or where
63 redevelopment occurs. The high costs for urban retrofit projects warrants strategic siting of these
64 practices so that they impart the greatest good to stream hydrology, especially considering limited
65 water resource budgets and increased urbanization (Barbosa et al. 2012).

66 Research has shown that the effective impervious area (EIA), a subset of the total impervious area (TIA)
67 representing the impervious areas hydraulically connected to the stormwater network, has a much

68 greater effect on stream hydrology than TIA due to rapid surface drainage along hydraulically-efficient
69 stormwater infrastructure (Brabec et al. 2002, Shuster et al. 2005). Streams exhibit declining conditions
70 as TIA increases, but these declines have been shown to be even more tightly coupled to increases in
71 impervious connectivity (i.e., a higher EIA/TIA ratio) (Hatt et al. 2004, Lee and Heaney 2003, Leopold
72 1968, Walsh et al. 2005b). Because of this, the EIA should be leveraged to optimize the hydrologic
73 benefits of GI retrofits by intercepting runoff from these areas first and foremost. However, quantifying
74 and identifying EIA is more difficult than TIA because it can be subject to ambiguous urban drainage
75 patterns and variable conditions that influence runoff production (Alley and Veenhuis 1983, Boyd et al.
76 1993, 1994, Brabec et al. 2002, Chiew and McMahon 1999, Ebrahimi et al. 2016a, Han and Burian
77 2009).

78 Although disconnection of EIA by implementation of GI practices can theoretically maximize the
79 effectiveness of watershed restoration efforts, rapid and reliable identification of these areas is often
80 time-consuming (Roy and Shuster 2009). Recently developed methodology by Epps and Hathaway
81 (2018) provides spatially explicit identification of EIA that is informed by geospatial data and observed
82 runoff trends. The results of this method account for spatial differences in runoff pathways and land
83 cover along these pathways to identify areas most likely to be EIA in a GIS framework that can be used
84 to prioritize stormwater management projects for runoff reduction. This method (Epps and Hathaway
85 2018) uses EIA quantity estimates from rainfall-runoff data analysis that build upon the work of Boyd et
86 al. (1993) and Ebrahimi et al. (2016b). These quantity estimates for EIA are then spatially-identified
87 with GIS flowpath analysis that differentiates impervious areas based on the connectivity of runoff
88 produced by them via interceding surfaces and infrastructure towards the watershed outlet. Calibration
89 is accomplished using a runoff attenuation parameter, varying this until the spatially-identified EIA
90 quantity matches the quantity estimate from the observed data. Readers are encouraged to read Epps
91 and Hathaway (2018) for further methodology details and a discussion of the limits of the results. The

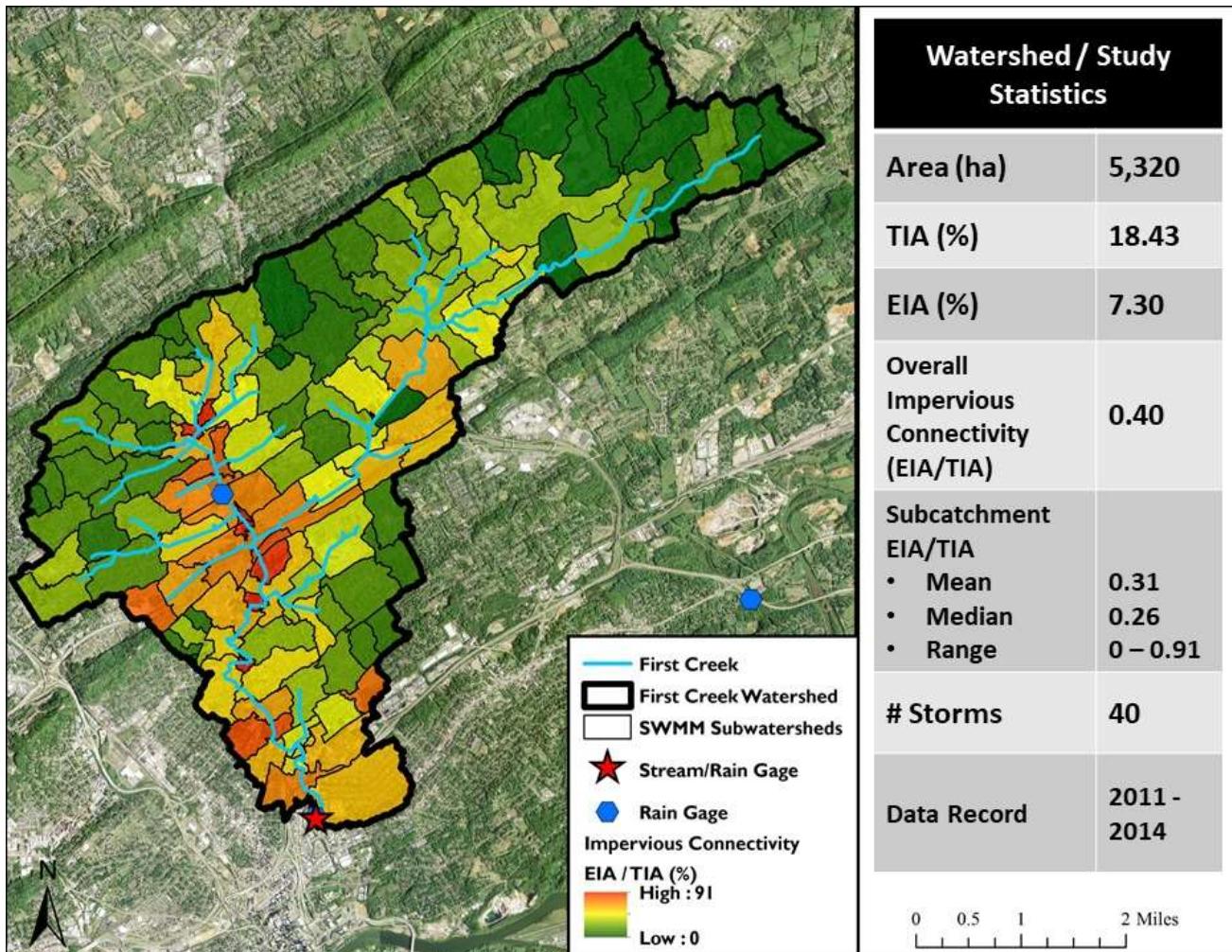
92 requirement of gaging a watershed to produce these results is a strong limitation on this methodology,
93 but efforts are underway to investigate how this approach can be adapted to be used on ungaged
94 watersheds. Determining the connectivity or lack thereof of rooftops is a concern with geospatial
95 approaches to EIA identification, another area for future research. A final limitation is that it is difficult
96 to assess the accuracy of the results of this new methodology or determine what the hydrologic effects
97 might be given implementation of GI retrofits (as full disconnection of the EIA with these would take
98 several years). Modeling can provide an initial means to assess the potential for this type of watershed-
99 wide deployment of GI retrofits, assess how targeting spatially-identified EIA for disconnection
100 compares to other less spatially guided strategies, and determine what level of changes in storm-event
101 response should be expected given watershed-wide implementation.

102 One of the most widely-used urban stormwater runoff models is the United States Environmental
103 Protection Agency's (EPA) Stormwater Management Model (SWMM) (Rossman 2004). Like most urban
104 runoff models, SWMM is very sensitive to impervious parameter inputs. The need for accurate
105 representation of impervious characteristics in urban runoff models has been noted, in particular in
106 relation to EIA (Alley and Veenhuis 1983, Lee and Heaney 2003). A recent review of research applying
107 SWMM to urban hydrologic investigations and the modeling of management paradigms for planning and
108 decision-making (Niazi et al. 2017) has identified several gaps and opportunities for future studies to
109 strengthen the use of this model. This review points to the gap in knowledge in accounting for
110 hydrologic continuity and GI spatial orientation as one of the most important foci for future research
111 due to the importance of assessing different GI retrofit configurations to support current urban
112 watershed management approaches. Incorporating spatially-identified EIA information into a SWMM
113 model to compare different siting strategies for GI retrofits forms the basis of the study presented here.
114 Results can be used to guide better approaches for assessing distributed restoration initiatives,
115 optimized GI planning, and the potential performance related to locational siting and configuration on a

116 watershed-scale basis. This information will be pursued in this study using SWMM to model First Creek,
117 a large urban watershed in Knoxville, TN. This will help determine what level of runoff reduction may be
118 possible by installing GI practices strategically to disconnect EIA and how this targeted strategy using
119 spatially-identified EIA compares to runoff reductions using less-informed spatial siting for the same
120 level of treatment with GI.

121 **Objectives**

122 The main goal of this study is to utilize the spatial identification of EIA areas from Epps and Hathaway
123 (2018) to assess the runoff volume reduction that might be realized given targeted application of GI
124 practices to these locations. Results will be compared to other scenarios, of similar treatment level, that
125 use different criteria for site selection not related to EIA location to help quantify the potential
126 advantage of using this spatial EIA information. A secondary goal is to verify the importance of the
127 results of spatial EIA identification using the methods of Epps and Hathaway (2018) pertaining to the use
128 of this high-resolution data. This will be accomplished by assessment of model results and discussion of
129 this important modeling parameter and its use in SWMM as it relates to surface runoff and GI models.



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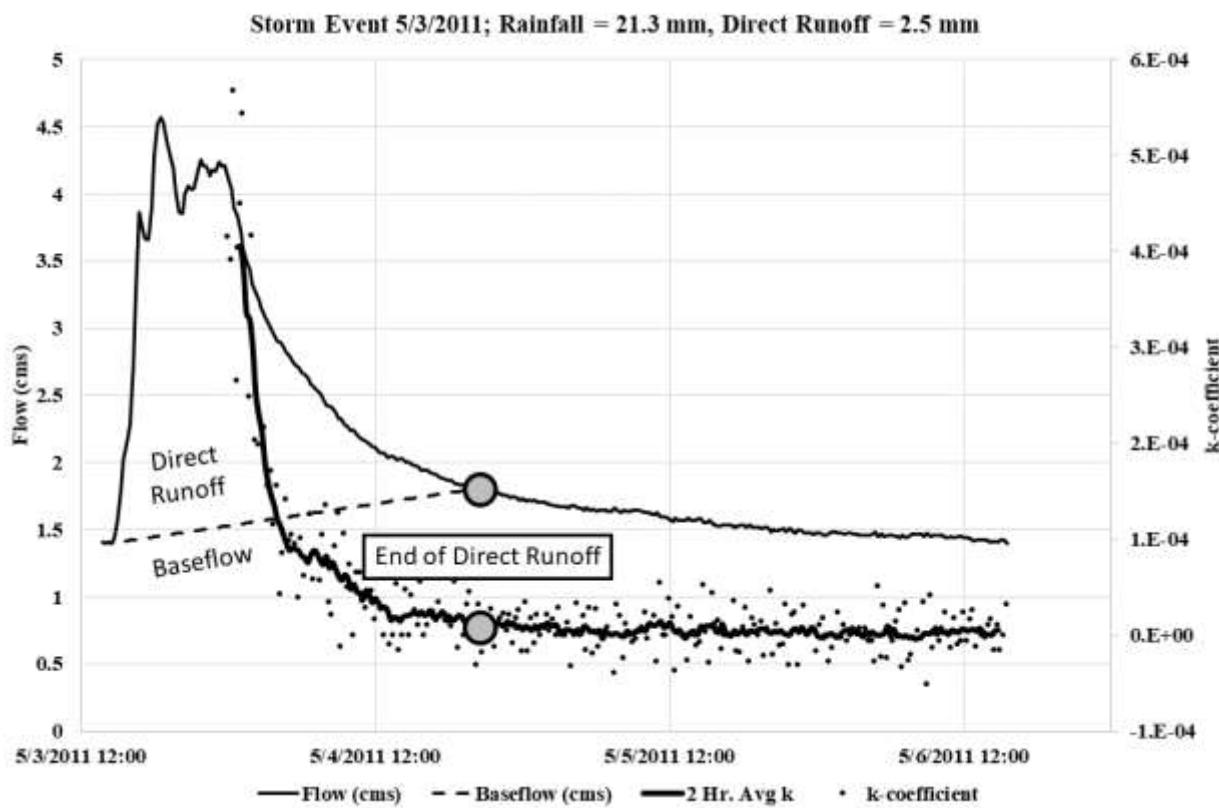
131 *Figure 1. Map of study watershed, First Creek in Knoxville TN, U.S.A. with summarized characteristics related to*
 132 *imperviousness and rainfall-runoff analysis used to estimate EIA. EIA spatial locations determined per Epps and*
 133 *Hathaway (2018).*

134 **2. Methods**

135 **2.1 Study Site and Data**

136 The focus of this study was the First Creek watershed in Knoxville, TN, United States (Fig. 1). First Creek
137 is a mixed-development watershed (5,320 ha) with an upland tributary that runs through rural areas
138 with sparse residential development and some agricultural lands and a main stem that flows through a
139 densely developed urban corridor following a main arterial road with flood-protection channelization.
140 Watershed statistics have been summarized in Figure 1. EIA quantity estimates were developed from
141 rainfall-runoff data using rainfall and streamflow records for the period of March 2011 – June 2014 from
142 the City of Knoxville Stormwater Engineering department (Epps and Hathaway 2018). Rainfall data (15
143 min. timestep) was obtained from three locations near the First Creek watershed and storm event
144 depths for the watershed were calculated based on a Thiessen polygon weighted-average. Storm events
145 with total rainfall depth above 1.27 mm and below 25.4 mm with at least 6 hours of separation from any
146 additional rainfall were considered in the analysis. Local stormwater management goals require capture
147 of runoff from the first 25.4 mm of rainfall for any size event. Focus on events of this size or smaller will
148 demonstrate how distinctions in GI retrofit placement impact runoff capture for the storm events of this
149 size or less. Additionally, this event size is approximately equal to the 90th percentile storm event size
150 for the region (Sylvester et al. 2016). Storm events were further screened for use in GI modeling using
151 best professional judgement to focus on events where all rainfall fell mostly in close temporal
152 succession (12 hours or less) to focus on simpler single-peaked hydrographs with a more uniform runoff
153 response over the watershed. While it has been shown that the effectiveness of GI practices can be
154 greatly reduced for subsequent peaks of complex storms (Versini et al. 2016), these dynamics are better
155 explored in studies of SCM design effectiveness. The stipulation for simpler single-peaked events is
156 made in this study to reduce temporal rainfall effects on SCM performance and focus on differences in

157 spatial distribution as much as possible in evaluating differences between placement scenarios. From
158 the full record of data, 68 distinct rain events were identified. Nine of these were removed because they
159 exceeded 25.4 mm in depth and 19 events were removed because they had multiple peaks. The largest
160 storm in the remaining 40 events to be analyzed was 20.7 mm. Of the 40 events, 20 were less than 6
161 mm in depth, 11 events were in the range of 6-12 mm, and 9 events were larger than 12 mm.



163 *Figure 2. Demonstration of hydrograph separation used to develop rainfall-runoff pairs for the estimation of EIA.*

164 Direct runoff depths for these storm events were estimated by hydrograph separation using the
165 constant-k method of Blume et al. (2007). This method assumes that baseflow discharges from a linear
166 reservoir with exponential decline such that the point on the receding limb of the hydrograph when the
167 recession coefficient (k) becomes nearly constant represents the end of direct runoff and the

168 streamflow's return to solely baseflow sources (Fig. 2). The recession coefficient was calculated for all
169 points on the receding limb of the storm-event hydrographs via the equation from Blume et al. (2007):

170
$$k = -\frac{dQ}{dt} * Q(t) \quad (1)$$

171 Flow data were assessed against the k-coefficient data using a 2-hour moving average to visually select
172 the point in the receding limb of the hydrograph at which k stabilized to a near constant value. This
173 point in the hydrograph was then connected by straight-line to the point just before hydrograph rise to
174 complete hydrograph separation (Fig. 2). The volume of water above this line was then summed and
175 converted to a depth through division by the watershed area so that rainfall-runoff trends could be used
176 to quantify EIA. The quantity of EIA in the First Creek watershed was estimated in the same manner
177 used in Epps and Hathaway (2018), employing regression analysis of rainfall-runoff data by the methods
178 of Boyd et al. (1993) and Ebrahimian et al. (2016) as a guide. This EIA quantity was used as a target for
179 calibration of spatial EIA models presented in Epps and Hathaway (2018). This spatial EIA data both: (1)
180 forms the impervious connectivity input for a SWMM model (as opposed to using impervious
181 connectivity as a calibration parameter), and (2) is the focal point of various SCM siting methodologies
182 that are compared herein to evaluate how volumetric runoff reduction varies based on restoration
183 approach. Spatial differences in the ratio of EIA/TIA (as a measure of impervious connectivity) for the
184 First Creek watershed are shown in Figure 1, summarized at the subcatchment level used in the SWMM
185 model.

186 **2.2 Surface Runoff Model**

187 A previously developed SWMM model for First Creek was provided by the City of Knoxville Stormwater
188 Engineering for this study. This model had been developed for flooding analysis and did not include a
189 groundwater component. Because the goal of the study was to assess the representation of impervious

190 connectivity and the disconnection of EIA using GI retrofits as they pertain to surface runoff, the
191 development of a groundwater component was deemed unnecessary. The existing SWMM model
192 disaggregated the First Creek watershed into 125 subcatchments based on infrastructure location and
193 grouping of homogeneous surface cover and topography. All surveyed channel information, stormwater
194 infrastructure, hydraulic parameterization, and internal storage portions of the model were preserved
195 for this study. Subcatchment boundaries were also preserved and the impervious area for each
196 subcatchment was updated to ensure model parameters were consistent with high-resolution data used
197 by the spatial EIA models.

198 **2.3 SWMM Model Setup and Parameterization**

199 Aside from subcatchment geospatial data, the SWMM model parameters were adapted based on
200 guidance from SWMM documentation (Rossman and Huber 2016). The Green-Ampt method was used
201 to represent infiltration in the model and runoff routing was modeled using the kinematic wave
202 approximation. Rainfall observations (15 min. timestep) from the three tipping bucket gages (Fig. 1)
203 used in EIA quantity estimation were used in the model, being assigned to subcatchments based on
204 which representative Thiessen polygon area for the three gages that the subcatchment was located in.

205 Model parameters were developed from literature values in most cases, though a few were based on
206 the assumptions of the spatial EIA model (Table 1). Infiltration parameters for the Green-Ampt model
207 were taken from literature values based on subcatchment soil properties obtained from the National
208 Resources Conservation Service's SSURGO soil survey database (SSURGO 2017). The distribution of each
209 subcatchments soils among different Hydrologic Soil Groups (HSG) was used to produce weighted values
210 for the literature-based Green-Ampt parameters. Average monthly evaporation values were used for
211 evapotranspiration modeling and derived from pan evaporation measurements for local gages in
212 eastern Tennessee (Farnsworth and Thompson 1983). Impervious routing parameters were parsed

213 between TIA and EIA according to the amount of EIA identified in each subcatchment from the
 214 previously described spatial models (Epps and Hathaway 2018). For each subcatchment, the EIA was
 215 routed directly to the outlet while the balance of TIA was routed to pervious areas representing the
 216 disconnected portion of impervious surfaces. The original SWMM model obtained from the City of
 217 Knoxville had routed impervious and pervious areas directly to the outlet, utilizing model calibration of
 218 other surface runoff parameters to achieve adequate model fitting. The inclusion of the spatially-
 219 identified EIA routing is an improvement over this representation that reflects site-specific impervious
 220 connectivity data and addresses a concern for SWMM models as described by Niazi et al. (2017).

221 *Table 1. Summary of pertinent SWMM model parameters and guidance used to develop the uncalibrated base*
 222 *models. Bolded values represent final parameters values for the calibrated model. Starred parameters indicate*
 223 *those that were varied during simple manual calibration.*

Parameter	Description	Initial Value	Range Lo	Range Hi	Guidance/Source
Subcatchment Parameters					
Nimp*	Manning's n for Impervious	0.015 (avg. Impervious)	0.01	.018	Yen (2001)
Nperv*	Manning's n for Pervious	0.05 (avg. grass)	0.038	0.12	Yen (2001)
Simp (mm)	Depress. Storage for Impervious	0.3	---	---	(Ebrahimian et al. 2016b)
Sperv* (mm)	Depress. Storage for Pervious	6.4	2.5	10.2	ASCE (1992)
%Zero	Portion of IMP w. no Simp	0	---	---	Accounted for in EIA Spatial Model
%Routed	Portion of IMP routed to PERV/Outlet	EIA %	---	---	EIA % within each Subcatchment
Infiltration Parameters					
Suction (mm)	Suction head at wetting front	$\psi_s=0.127*K_{sat}^{-0.328}$			Brakensiek et al. (1981)

K _{sat} * (mm/hr)	Sat. hydraul. conductivity	Avg. value from soil composition	Low value from soil composition	High value from soil composition	Musgrave (1955); based on soil HSG comp.
IMD*	Max moisture deficit avail.	Based on K _{sat} /soil classification	Low value from table range	High value from table range	Rawls et al. (1983)

224

225 **2.4 Model Calibration and Assessment**

226 Surface runoff was modeled continuously and reported at a 15-minute timestep over the period from
 227 March 2011 to June 2014 during which records for streamflow and three rain gages existed. A subset of
 228 40 of the identified storm events that had been used to estimate the watershed EIA were selected for
 229 analysis of potential runoff volume reduction. Analysis of overall GI practice runoff reduction was
 230 constrained to these storms since larger storms would likely be associated with runoff from more than
 231 just EIA areas. This focus on storm events most likely producing runoff from just EIA areas provides a
 232 greater assurance that different GI retrofit placement scenarios (described more below) will capture the
 233 differences in runoff reduction possible under strategic siting using spatial EIA identification versus other
 234 less-targeted guidance. Rainfall depths for these 40 storms ranged from 1.27 mm to 20.7 mm with an
 235 average storm size of 8.3 mm. While more than 40 valid storm-events occurred during the period of
 236 record, only storms that had been identified to derive from predominantly EIA in regression analysis
 237 were used. Modeled runoff from the equivalent timeframe of each of these observed storm events was
 238 summed to produce a runoff volume that was normalized to a depth over the watershed area for
 239 comparison to the observed runoff depth over the same equivalent timeframe. Model fit for storm-
 240 event runoff depths was used for calibration and assessed using the Nash-Sutcliffe efficiency index (NSE)
 241 (Nash and Sutcliffe 1970) and the Kling-Gupta efficiency (KGE) (Gupta et al. 2009).

242 Simple manual calibration for selected storm event runoff volumes was performed to improve the
243 SWMM model fit by varying the starred parameters in Table 1 between their high, mean, and low values
244 identified in the literature to determine model sensitivity and provide a suitable model of surface runoff.
245 This calibration was simple in that the parameters for all 125 subcatchments were assigned the same
246 value class (low, median, high) for the parameter of interest for each calibration iteration. Further
247 calibration was not deemed necessary because the focus of the study was on GI practice placement. A
248 reasonably accurate runoff model provided a consistent baseline for these different GI practice
249 placement scenarios to be studied and compared. Final model parameters are noted in bold (Table 1).
250 Calibration was assessed for the storm event runoff depths during 2011 – 2012 (23 events) and verified
251 by the remaining event runoff depths during 2013 – 2014 (17 events).

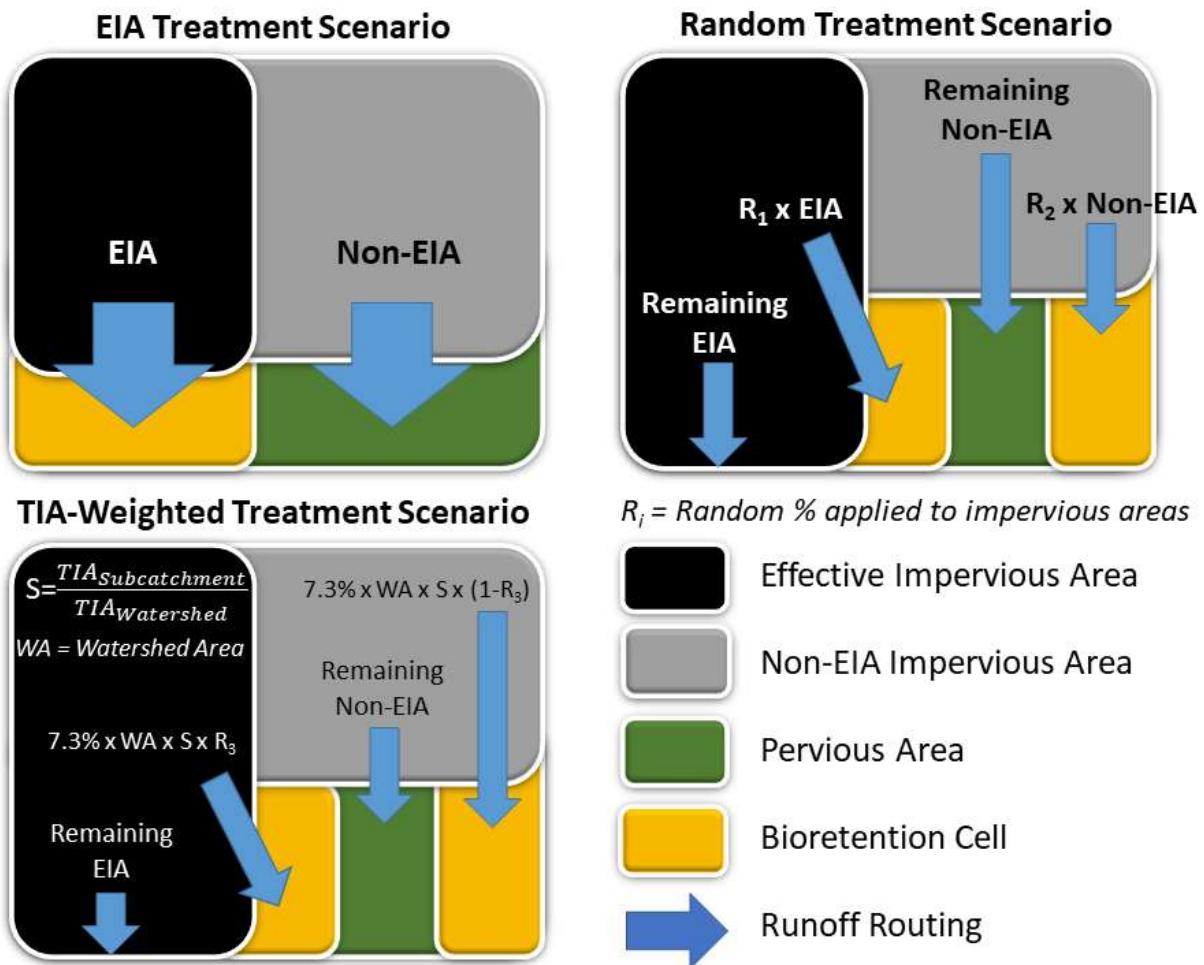
252 **2.5 Green Infrastructure Scenarios**

253 To study differences in runoff reduction that may be realized given alternative approaches to SCM
254 application, the calibrated SWMM base model was used to assess runoff reduction under different siting
255 strategies for GI retrofits in the First Creek watershed. Since EIA was estimated as 7.3% of the
256 watershed, this forms the targeted amount of impervious area to be treated by GI practices in all
257 scenarios. Three scenarios of GI practice siting were compared and a schematic representation of these
258 is shown in Figure 3.

259 • The first scenario (“EIA”) applied GI practices only to the spatially-identified EIA areas (treating
260 7.3% of the watershed). This represents the targeted scenario where spatial models of EIA are
261 used to identify locations for GI practices to disconnect EIA runoff from the stream network.
262 • The second scenario (“Random”) distributed GI practices randomly by first using a random
263 number generator to apportion the 7.3% treatment level amongst the subcatchments, and then
264 using a second random number generator to divide that apportioned percentage for a given

265 subcatchment between EIA and non-EIA areas (subject to availability of each of these
266 impervious types). In this scenario, it is assumed that the equivalent area of impervious surfaces
267 (7.3% of the watershed) are treated for disconnection, but that the placement of practices
268 among the TIA is not guided by any spatial information (i.e., random).

269 • The third scenario (“TIA-weighted”) was guided by subcatchment imperviousness wherein the
270 7.3% treatment level was apportioned amongst subcatchments using a weighted average of the
271 TIA in each subcatchment to the overall watershed TIA. This TIA-weighted impervious
272 percentage was then divided between EIA and non-EIA areas in each subcatchment based on a
273 random number generator. This scenario represents an intermediate guidance strategy in
274 which the treatment amount in each subcatchment is proportional to the TIA distribution over
275 the watershed, but that there is no further spatial differentiation between EIA and non-EIA
276 areas receiving treatment. This scenario was chosen as an intermediate level of spatial guidance
277 since distributed GI retrofits are typically placed more often in the most highly impervious areas
278 of a watershed.



279

280 *Figure 3. Schematic representation of GI siting scenarios used for comparison.*

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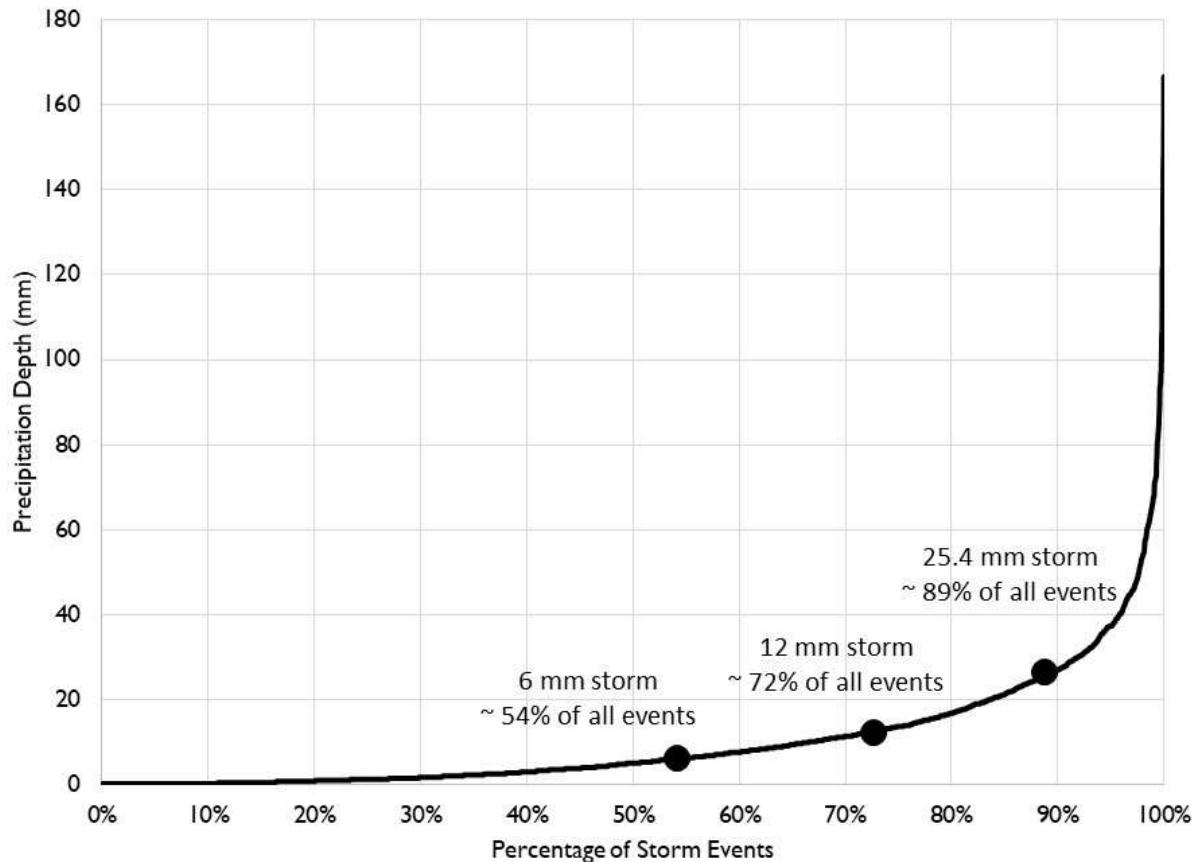
282 This process was iterated for each alternative siting strategy (Random and TIA-weighted) with different
 283 randomization to produce 30 different applications of GI practices for each to be compared to the EIA-
 284 focused siting strategy. Each of these scenarios was then modeled in SWMM by placing practices
 285 appropriately within the model to identify differences in runoff reduction between each model. Median
 286 and maximum reductions over the 30 model iterations for each alternative siting strategy were then
 287 compared to the EIA scenario results to assess differences in performance between the three. This was
 288 done to compare both the average and the best performance for less spatially-informed siting methods

289 to quantify the advantage (or lack thereof) that utilizing spatial EIA data might provide over the
290 alternative siting scenarios. Runoff volume reductions were compared using a storm-event pairwise
291 Wilcoxon signed rank test (Wilcoxon et al. 1970) due to non-normality to assess whether GI siting
292 scenarios exhibited dissimilar runoff reductions on average and thus demonstrate whether a given siting
293 strategy might be advantageous.

294 **2.6 Green Infrastructure Modeling**

295 To apply the appropriate GI treatment in each scenario, a generic bioretention cell was modeled to treat
296 both EIA and non-EIA impervious areas separately in each subcatchment. Bioretention cells were
297 modeled using the LID editor in SWMM (version 5.1). A single, lumped bioretention cell receiving runoff
298 from each impervious area subset was modeled for a total of two bioretention cells per subcatchment
299 and these were sized distinctly with enough storage to capture the runoff from a 25.4 mm storm for the
300 given impervious area each was determined to treat. Parameterization of these bioretention cells is
301 summarized in Table 2. Bioretention cells were sized vertically per guidance from municipal stormwater
302 management manuals (Table 2). For each GI practice scenario, the impervious surface area identified
303 for treatment (EIA and non-EIA separately) was used to calculate the runoff volume for a storm event of
304 25.4 mm. The required surface area of bioretention cells to treat this amount of runoff was then
305 calculated based on the total storage available in the bioretention cell to accommodate this volume.
306 SWMM allows a given bioretention cell to receive runoff from a subset of the subcatchment impervious
307 areas. Thus, two bioretention cells were defined and sized for each subcatchment (one for EIA, one for
308 non-EIA), and the appropriate percentage of the impervious area was routed to each based on the EIA
309 and non-EIA impervious areas to be treated in each scenario. SWMM also allows underdrain flow and
310 excess water from the bioretention cell to be routed back to pervious areas or directly to the outlet.
311 This option was utilized to maintain the model representation of surface routing, with the bioretention

312 cell treating non-EIA routed to pervious areas and that treating EIA routed to the outlet. Storm event
313 runoff depths were calculated from model outputs and the percent runoff reduction was calculated by
314 comparison to surface runoff depths for the base model with no GI practices. The percent reduction of
315 runoff for the 40 storm events was calculated for the 30 iterations in each alternative scenario to
316 compare to the EIA-focused runoff reductions. Further, pairwise comparison of storm event runoff
317 reduction between the three siting strategies was then used to assess whether there were statistical
318 differences. This was done for all storm events as well as by three groupings of storm event size (< 6
319 mm, 6 – 12 mm, and 12+ mm) to aid in discussion of differences in runoff reduction by storm size.
320 These grouping represent storm sizes for the area of approximately 54%, 72%, and 89% frequency ranks,
321 respectively, (Fig. 4) based on rain data from the Tyson-McGhee airport station for the years 1981 –
322 2010 and are consistent with the results presented in Sylvester et al. (2016). The EIA scenario was
323 compared to the median and maximum runoff reduction scenarios (of the 30 randomizations) for each
324 of the Random and TIA-weighted scenarios to demonstrate the range of performance and sensitivity to
325 placement for these alternate siting strategies in comparison.



326

327 *Figure 4. Storm depth frequency distribution for McGhee-Tyson airport, Knoxville, TN.*

328 *Table 2. Summary of bioretention cell parameterization and source of information for application of SCMs in*
 329 *different siting scenarios.*

BIORETENTION CELL SCM (Type = BC)			
Parameter	Description	Value Used	Rationale
Surface Layer Parameters			
<i>StorHt</i>	Max depth water can pond (cm)	15.24	County (2008)
Soil Layer Parameters			
<i>Thick</i>	Thickness of soil layer (cm)	60.96	County (2008)
<i>Por</i>	Soil porosity (pore space/total volume)	0.44	Committee (2005)

<i>FC</i>	Soil field capacity (volume pore water/ total volume when fully-drained)	0.09	Committee (2005)
<i>WP</i>	Soil wilting point (vol. pore water/ total volume for well-dried soil)	0.04	Committee (2005)
<i>Ksat</i>	Saturated hydraulic conductivity (cm/hr)	1.27	County (2008)
<i>Kcoeff</i>	Slope of curve of log(conductivity) vs soil moisture content (dimensionless)	50	Rossman (2004)
<i>Suct</i>	Soil capillary suction (cm)	10.31	Brakensiek et al. (1981)
Storage Layer Properties			
<i>Height</i>	Thickness of storage layer (cm)	30.48	County (2008)
<i>Vratio</i>	Void ratio (porosity = vr/(1+vr))	0.4	Miller (1978)
<i>Seepage</i>	Rate of drainage into native soil (cm/hr)	<i>Ksat</i>	Model parameter for subcatchment
<i>Vclog</i>	Clogging parameter	0	Clogging ignored
Drain System Properties			
<i>Coeff</i>	Determines rate of flow through drain as a function of hydraulic head (C)	0.6	County (2008)
<i>Expon</i>	Determines rate of flow through drain as a function of hydraulic head (n)	0.5	County (2008)
<i>Offset</i>	Ht. of drain above bottom of storage layer (cm)	30.48	Miller (1978)

330

331 **3. Results and Discussion**

332 **3.1 SWMM Model Results**

333 The base model performed well in matching storm event runoff volume given simple manual calibration.

334 Model performance results over the entire period and the calibration and verification periods are

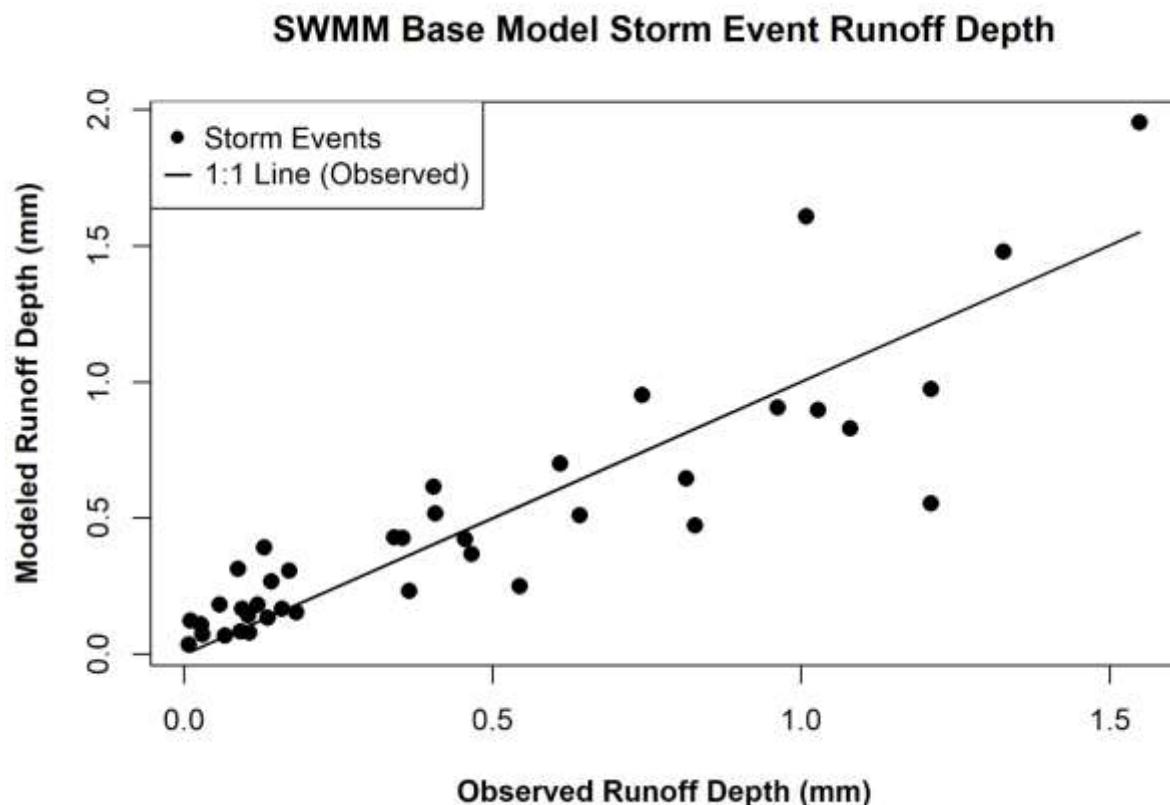
335 summarized in Table 3, and modeled runoff depth versus the observed runoff depth has been plotted in

336 Figure 5. The model typically performs well, but shows some underprediction of larger storm events, an
337 expected outcome given the lack of a groundwater component within the model.

338 Hydrograph separation distinguishes total runoff between a baseflow component and direct runoff, the
339 latter which may contain some portion of interflow, or shallow subsurface return flow not related
340 entirely to surface runoff dynamics (Beven 1989). This portion is not easily discernible through graphical
341 analysis and may over-estimate true surface runoff for any given event. Manual parameter adjustment
342 for this SWMM model used mean literature values as a starting point and adjusted parameters to
343 minimum and maximum literature values for all subcatchments to identify the greatest sensitivities in
344 the model for surface runoff to arrive at the final calibrated base model. Parameters that were adjusted
345 in this process away from the mean value were all related to soils and pervious area runoff, and the
346 calibrated values were all on the low end of the literature ranges. It is possible that this reflects the
347 condition of urban soils which typically have poorer infiltration and thus contribute to greater surface
348 runoff. However, this may also be due to the potential overestimate of surface runoff from the inclusion
349 of interflow portion which is not included in the SWMM model. Overall, these results demonstrate that
350 the model is well-suited to serve as a basis for comparing runoff reduction between SCM placement
351 scenarios.

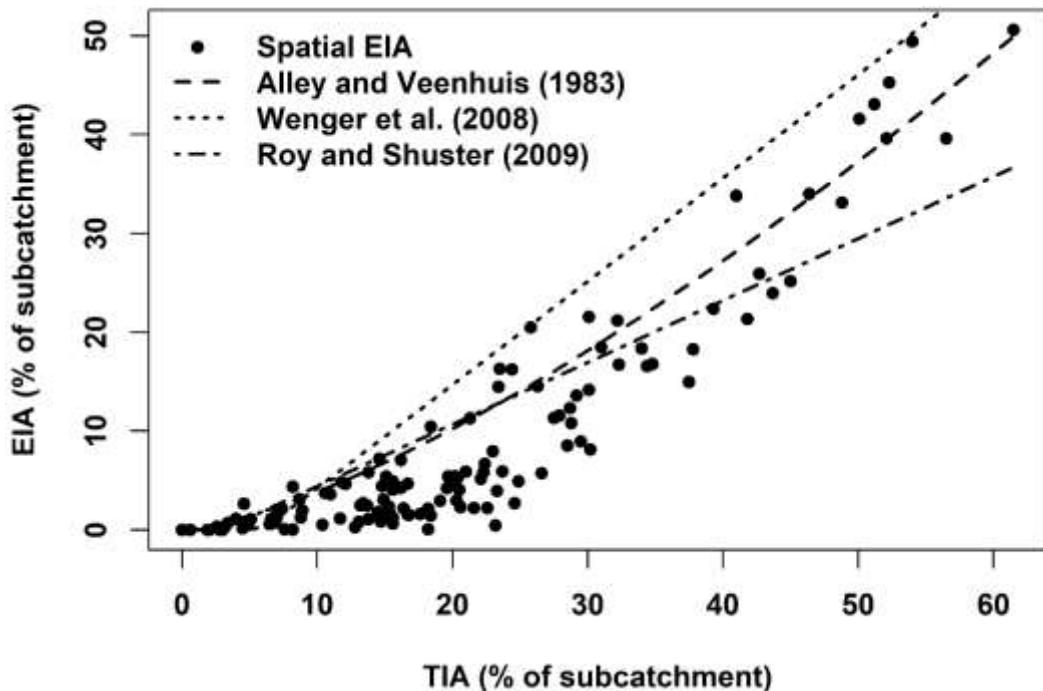
352 *Table 3. Summary of SWMM base model performance for prediction of storm event runoff depth.*

Model Period	Nash-Sutcliffe Efficiency	Kling-Gupta Efficiency
Overall (2011 – 2014)	0.77	0.88
2011/2012 (23 events)	0.79	0.72
2013/2014 (17 events)	0.74	0.73



364 subcatchments (Fig. 6) derived from spatial analysis in Epps and Hathaway (2018) demonstrates a very
365 different relationship for the First Creek watershed than predicted by literature regression equations
366 sometimes used for EIA estimation (Alley and Veenhuis 1983, Roy and Shuster 2009, Wenger et al.
367 2008). This plot indicates that EIA would be mostly overpredicted by regression equations for the First
368 Creek watershed, especially for intermediate values of TIA.

369 The use of these regression equations to predict EIA in catchments other than where they were
370 developed has been demonstrated to poorly reflect actual EIA measurements (Roy and Shuster 2009).
371 While this is not the focus of this paper, it demonstrates that the method utilized to estimate EIA can
372 vary greatly, and because runoff modeling results are sensitive to this parameter, it makes sense to use
373 the most site-specific information available. Spatial EIA estimates used in this study were conducive to
374 incorporation into SWMM model subcatchment discretization for more accurate runoff production and
375 routing information. Implemented within typical GIS analysis for model parameterization, they were
376 easily summarized by subcatchment in the same way as TIA typically is. Estimations of EIA using
377 regression equations developed elsewhere or by using EIA as a model calibration parameter (that is,
378 when it is one of many calibration parameters) may provide a less accurate representation of
379 impervious connectivity and urban runoff processes.



380

381 *Figure 6. The spatial distribution of EIA among SWMM model subcatchments as a function of TI are not well*
 382 *represented by literature regression equations.*

383 *Table 4. Summary of mean/median percent runoff reductions for different treatment scenarios for all storms and by*
 384 *size groupings.*

Percent Runoff Reduction	EIA Treatment		Random Treatment		TIA-weighted Treatment	
	Mean	Median	Mean	Median	Mean	Median
All Storms	37.3	39.8	27.2	28.7	27.0	28.6
< 6 mm	49.2	49.4	34.7	33.4	34.6	33.6
6 – 12 mm	32.2	34.6	25.0	27.9	24.4	27.4
12+ mm	16.9	17.8	13.3	12.5	13.1	12.6

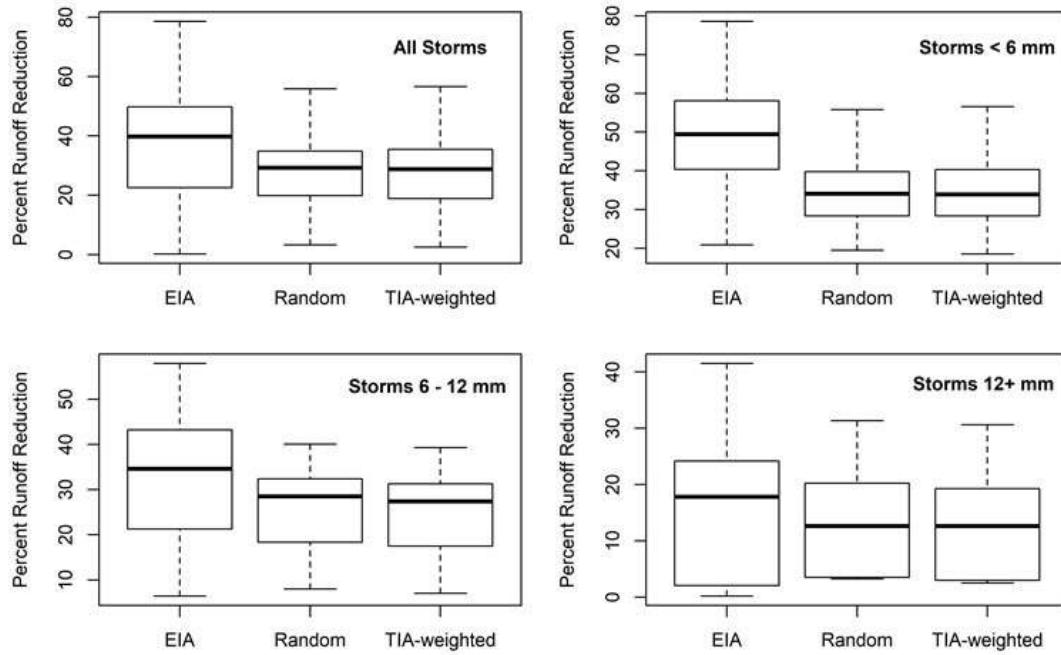
385

386 3.3 Runoff Reduction for GI Placement Scenarios

387 Average percent runoff reductions achieved via different GI placement scenarios are summarized overall
 388 and by storm size groupings in Table 4. These values represent the mean and median reductions by

389 scenario and storm size for all modeled events (40 events for EIA and the 30 randomization scenarios of
390 40 events for the alternative siting scenarios). Runoff reduction for the 40 storms are presented as
391 boxplots to demonstrate differences in the range of values between GI siting strategies. These have
392 been summarized for both the median runoff reduction scenario for the Random and TIA-weighted
393 treatments over the 30 randomizations (Fig. 7) and the maximum runoff reduction scenario over the 30
394 randomizations as well (Fig. 8). The results of pairwise comparison for each grouping by the Wilcoxon
395 signed-rank test (Wilcoxon et al. 1970) are presented in Table 5 for EIA runoff reduction and the
396 maximum runoff reductions under the 30 randomizations of the alternative strategies. Results indicate
397 that focusing GI applications in locations identified as EIA can result in greater runoff reduction,
398 especially for smaller storms up to 12 mm. Over all storm events, GI placement focused on EIA resulted
399 in nearly 4-10% greater runoff reductions (Table 5) than those for less spatially guided strategies with an
400 overall mean runoff reduction of 37.3% (Table 4). This is in comparison to mean reductions of 27.2%
401 and 27.0% for the random and TIA-weighted placement scenarios over all storm events, respectively.
402 The differences in runoff reduction when storms are partitioned by size offers insight on modeled runoff
403 dynamics and the performance of GI treatment over a range of conditions. This information can inform
404 future model parameterization and utilization of this information for application of distributed GI
405 practices in terms of runoff reduction potential and management strategies.

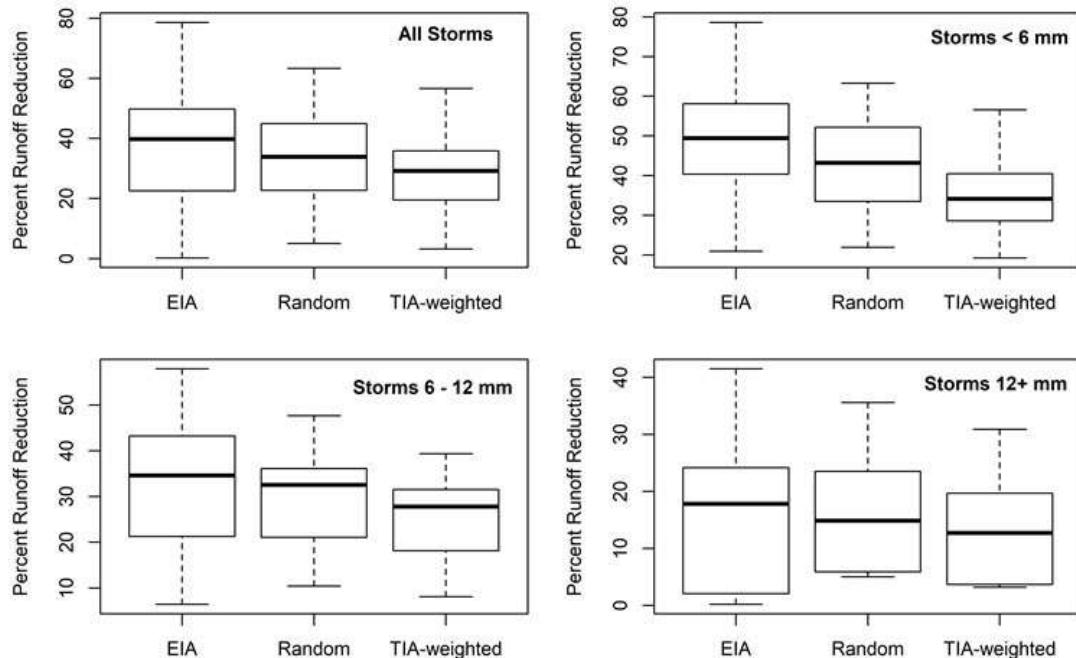
RUNOFF REDUCTION COMPARISON – MEDIAN PERFORMANCE SCENARIO



406

407 *Figure 7. Comparison of EIA-siting to median values of runoff reduction for alternative strategies.*

RUNOFF REDUCTION COMPARISON – MAXIMUM PERFORMANCE SCENARIO



408

409 *Figure 8. Comparison of EIA-siting to maximum values of runoff reduction for alternative strategies.*

410 For smaller events (less than 6 mm), runoff reduction is highest when practices are placed according to
 411 spatially-identified EIA with a mean runoff reduction of 49.2%. Runoff from connected impervious
 412 surfaces during such events is well-handled by the modeled bioretention cells and precipitation is
 413 sufficiently small that pervious areas do not likely produce substantial runoff, if any. This is important to
 414 note when you consider that the higher frequency of runoff from smaller storm events is one of the
 415 indicators of urban hydrologic regime shift. Substantially greater runoff reduction for small storm
 416 events using EIA-focused placement may have a large impact on decreasing runoff frequency overall.
 417 While runoff would not be eliminated entirely, the amount would be negligible considering the size of
 418 the events.

419 *Table 5. Summary of Wilcoxon signed-rank tests for maximum runoff reductions between GI siting strategy overall*
 420 *and for storm size groupings.*

Storm Events	Comparison	Statistical Difference (p-value)	Pseudo-Median Difference	95 % Confidence Interval	
All Storms	EIA to Random	< 0.001	4.24	2.37	5.40
	EIA to TIA-weighted	< 0.001	10.05	7.67	12.48
	Random to TIA-weighted	< 0.001	5.70	4.37	6.98
Storms < 6 mm	EIA to Random	< 0.001	5.76	4.85	8.78
	EIA to TIA-weighted	< 0.001	14.72	12.18	16.74
	Random to TIA-weighted	< 0.001	7.80	6.23	9.59
Storms 6–12 mm	EIA to Random	0.15	No statistical difference at $\alpha = 0.05$		
	EIA to TIA-weighted	0.010	7.09	3.07	11.70
	Random to TIA-weighted	< 0.001	4.23	3.21	6.59

Storms 12+ mm	EIA to Random	0.820	No statistical difference at $\alpha = 0.05$		
	EIA to TIA-weighted	0.055	No statistical difference at $\alpha = 0.05$		
	Random to TIA-weighted	0.004	3.19	2.07	4.41

421

422 For intermediate events greater than 6 mm but less than 12 mm, runoff reduction using EIA-focused
 423 siting was greater than the maximum TIA-weighted scenarios with an average runoff reduction of 32.2%,
 424 but not statistically higher overall than the maximum runoff reductions for Random siting scenarios.
 425 While results were inconclusive for this intermediate range of storm sizes, a comparison of EIA runoff
 426 reductions and median runoff reductions for alternate strategies indicate that on average, less-spatially
 427 informed strategies resulted in approximately 7% less runoff reduction ($\alpha < 0.01$) on average. Runoff
 428 reductions for these intermediate storm events were moderately greater for the EIA treatment scenario,
 429 though with diminishing returns relative to the smallest storms. Events in this size range are typically
 430 frequent, especially in more humid climates like the southeastern United States. Full treatment of all
 431 EIA with distributed practices could thus reduce runoff by greater than 30% annually for storms less
 432 than 12 mm in size. While the magnitude of this reduction overall would be dependent on local
 433 precipitation frequencies, it could represent a substantial amount of annual runoff reduction if a larger
 434 percentage of storm events for a given urban watershed were in this range.

435 When rainfall exceeded 12 mm, runoff reductions appear to be greater when using EIA-focused siting
 436 based on mean and median percent reduction statistics (Table 4). However, there was not a statistically
 437 significant difference in storm event runoff reduction between siting strategies at the $\alpha = 0.05$
 438 confidence level. There was greater variability in runoff reductions for EIA treatment scenarios for these
 439 larger storm events than for the other storm size groupings, and this is also evident in the Random and
 440 TIA-weighted treatment scenarios. One possible explanation for this could be the influence of

441 antecedent moisture conditions and pervious areas on runoff production for larger storm events in the
442 SWMM model. When rainfall exceeds infiltrative capacity, pervious areas begin to produce runoff, and
443 this is more likely to occur for larger storm events. Further, infiltrative capacity for pervious areas
444 ranges based on soil moisture conditions and thus varies by storm event depending on recent rainfall.
445 This could also be related to performance of the bioretention cells as well for the same reasons since
446 previous saturation may cause these practices to handle less runoff for storms falling in close temporal
447 succession

448 Runoff from impervious areas routed onto pervious areas (as in the case of non-EIA portions of the
449 watershed) would additionally influence soil saturation and further increase runoff from pervious areas
450 for larger events. For the EIA treatment, all runoff from non-EIA is routed to pervious areas without
451 treatment. For larger storms, and especially those falling when soil saturation is higher due to recent
452 rainfall, all runoff from these non-EIA areas is routed to pervious areas in the SWMM model, which may
453 produce runoff for these events. This effect would be lessened in the random and TIA-weighted
454 scenarios which treat a portion of non-EIA runoff and thus would contribute less to intra-event pervious
455 saturation. For larger storms, surface runoff thus becomes the sum of interactions between pervious
456 conditions and surface runoff routing from different portions of the TIA that are subject to a range in
457 antecedent moisture conditions. Treatment location may become secondary to simply the level of
458 treatment for maximum runoff reductions as pervious influences impact surface runoff more for these
459 larger events. This is also influenced by model parameterization and the scale of subcatchment
460 disaggregation. While the results of the study are specific to the specific watersheds, they do indicate
461 some advantage in targeting EIA with GI practices over other siting strategies, though the magnitude in
462 other watersheds may be unclear. Despite this, the results highlight important foci for research
463 development in urban stormwater management planning and areas where improved data and model
464 representation may be warranted.

465 **3.4 Limitations and Sources for Error**

466 Table 6 provides a summary of important assumptions, uncertainties, and sources of error in this study.

467 Acknowledgement of the limitations of these findings serves to point to important opportunities for

468 future research that this study hoped to illuminate. While many of these have already been discussed

469 throughout the text, it is important to consider them together as they can have a compounding effect

470 on the results. Each of the items in Table 6 point to further lines of inquiry as watershed managers

471 continue to improve data collection and modeling efforts in support of urban watershed management.

472 *Table 6. Summary of relevant sources for error in this study.*

Potential Source of Error	Description	Impacts on Results
Spatial Rainfall Variability	Thiessen-polygons may not accurately capture the distribution of rainfall over the watershed based on differences in observed rainfall at 3 gage locations	Rainfall estimates used in EIA estimation may bias foundational numerical targets of study
Hydrograph Separation	Baseflow and interflow components are difficult to separate from surface derived runoff	Runoff depths used for EIA estimation and model fitting may not accurately represent surface runoff well
Storm Event Selection	Focused screening of storm events in the record based on size and compact temporal distribution of rainfall and single-peaked hydrologic response	May bias results and conclusions to a smaller subset of realized storm events
EIA Quantity Estimate	Graphical analysis is subject to accuracy of data (rainfall and runoff, as discussed above)	Results may be biased based on data accuracy
EIA Spatial Identification	Spatial EIA determination by new methodology is not well-vetted or easily verified	Results and Conclusions may be influenced by incorrect spatial distribution of EIA
SWMM Parameterization	Literature values may not apply to specific site hydrology or land cover	Model may not reflect realistic hydrologic characteristics of watershed
Subcatchment Scale	The size and resolution of subcatchments may not match process scale of impervious runoff connectivity and GI applications	Simple model calibration may be associated with some level of equifinality
Lack of Groundwater in SWMM model	Shallow surface runoff and groundwater interactions are not accounted for	Runoff reductions focusing on surface runoff may not provide the full hydrologic effect of GI and pervious runoff processes
Uniform Bioretention Cell Parameters	Certain GI types may not be applicable or feasible in all locations; multiple practices may not behave ideally in series or parallel compared to lumped	Runoff reductions might be very different given more fine-resolution of system drainage, GI type, and treatment train

SWMM LID representation	SWMM model representation of GI may not be adequate to capture GI practice performance	Results may demonstrate less or more runoff reduction and impacts on hydrology than actual practices may attain in-situ
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473

474 **3.5 Modeling Considerations and Further Opportunities**

475 While the results of this study indicate that utilizing spatial EIA information for GI practice placement
 476 can result in greater runoff reduction, some considerations should be made regarding model
 477 representation in Table 6 that will point to areas for further study. This modeling exercise utilized a
 478 single idealized bioretention cell for EIA and non-EIA areas to represent runoff reduction potential given
 479 watershed wide application. However, the routing of runoff deriving from EIA and non-EIA areas needs
 480 to be further studied considering the results for runoff reduction from larger storms. Model results
 481 suggests interactive effects in surface runoff processes (as modeled) that may be at play in urban
 482 hydrology, however it could also be a function of the modeling structure itself or possibly the scale of
 483 subcatchment disaggregation and representation. Realistic application of GI practices in this watershed
 484 would likely necessitate the disaggregation of the idealized bioretention cells into smaller units that
 485 were further distributed within each subcatchment, i.e., modeling at a finer scale for both the GI
 486 practices and their contributing watersheds. To holistically model this interactivity at the watershed
 487 scale, investigation of the relationship between groundwater conditions and surface runoff may also be
 488 warranted. Additionally, bioretention may not always be the most applicable or feasible type of GI for
 489 all areas. Detention-based GI may be more effective than infiltrative practices given certain site
 490 conditions and objectives, and vice versa. The use of a single type of GI practice with uniform design
 491 might be expanded in the future to assess how different GI types and designs might perform in
 492 conjunction with the additional information on runoff routing that is offered by the high-resolution EIA
 493 data.

494 Another consideration is the feasibility for GI retrofits in areas identified as EIA. Placement of GI where
495 EIA has been identified may not always be feasible due to site constraints not identified by the spatial
496 model. This could be related to poor infiltration rates, difficult topography, or adequate space. For the
497 First Creek watershed, 3.1% of the watershed was identified as EIA in the public domain, less than half of
498 the total EIA. Targeting these areas first for any GI retrofit efforts in the watershed would be a good
499 strategy to begin implementation of distributed restoration efforts based on these modeling results.
500 This would represent the intersection of opportunistic GI applications with spatially-informed siting that
501 would provide watershed managers with the best hydrologic benefit for the watershed using readily
502 available spaces for distributed restoration. Spatial EIA models could then be revised given the added GI
503 to reassess impervious connectivity and further identify areas where runoff reductions would be
504 greatest given the disconnection of those where GI has been established over the timeline of
505 implementation.

506 **4. Conclusions**

507 There is a need to prioritize areas for GI retrofits in urban watersheds that enable cities to place (often
508 limited) resources in areas where they will have the greatest impact on urban streams. Watershed
509 assessment using high-resolution geospatial data can provide robust information concerning priority
510 areas for distributed watershed restoration when coupled with hydrologic information. Modeling that
511 incorporates the best information from both of these sources can be used to assess management
512 options and refine approaches in order to optimize results for urban watershed improvement. The
513 results of this study indicate that spatial EIA information can be applied to a SWMM model to
514 adequately predict surface runoff using literature-based parameters with simple manual calibration.
515 The calibrated model was used to investigate three GI placement scenarios. The strategy specifically
516 targeting EIA within the watershed showed moderate differences in runoff reduction compared to more

517 random GI placement strategies which treated both EIA and less-connected portions of TIA. The EIA
518 based strategy was estimated to average approximately 37% runoff reduction over all storm events at
519 about 10% more than the other two methods. However, the difference between the siting strategies
520 weakened as storm size increased, indicating that treating effective impervious area is most
521 advantageous for mitigating small, frequent rainfall events. Further study is warranted, in particular on
522 watersheds with varying impervious connectivity, and especially those with a very high EIA/TIA ratio or
523 where GI opportunities are constrained by dense development. These results highlight the importance
524 of focusing GI applications to areas identified as EIA to optimize urban hydrologic benefits and point to
525 how this methodology can help watershed managers prioritize restoration efforts. Further investigation
526 with higher-resolution runoff modeling should lead to management recommendations that fully account
527 for the spatial variability of urban runoff production, but this initial study demonstrates that the
528 utilization of spatially-identified EIA data can be used as a basis for GI retrofit siting when runoff
529 reduction is the primary objective.

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534

535 **5. References**

536

537 Alley, W.M. and Veenhuis, J.E. (1983) Effective impervious area in urban runoff modeling. *Journal of*
538 *Hydraulic Engineering* 109(2), 313-319.

539 ASCE (1992) *Design and Construction of Urban Stormwater Management Systems*, ASCE.

540 Assessment, M.E. (2005) *Millennium Ecosystem Assessment Findings*, Millennium Ecosystem
541 Assessment.

542 Barbosa, A.E., Fernandes, J.N. and David, L.M. (2012) Key issues for sustainable urban stormwater
543 management. *Water Res* 46(20), 6787-6798.

544 Bedan, E.S. and Clausen, J.C. (2009) Stormwater runoff quality and quantity from traditional and low
545 impact development watersheds1, Wiley Online Library.

546 Bernhardt, E.S. and Palmer, M.A. (2007) Restoring streams in an urbanizing world. *Freshwater Biology*
547 52(4), 738-751.

548 Blume, T., Zehe, E. and Bronstert, A. (2007) Rainfall—runoff response, event-based runoff coefficients
549 and hydrograph separation. *Hydrological Sciences Journal* 52(5), 843-862.

550 Boyd, M.J., Bufill, M.C. and Knee, R.M. (1993) Pervious and impervious runoff in urban catchments.
551 *Hydrological Sciences Journal* 38(6), 463-478.

552 Boyd, M.J., Bufill, M.C. and Knee, R.M. (1994) Predicting pervious and impervious storm runoff from
553 urban drainage basins. *Hydrological Sciences Journal* 39(4), 321-332.

554 Brabec, E., Schulte, S. and Richards, P.L. (2002) Impervious surfaces and water quality: a review of
555 current literature and its implications for watershed planning. *Journal of planning literature* 16(4), 499-
556 514.

557 Brakensiek, D., Engleman, R. and Rawls, W. (1981) Variation within texture classes of soil water
558 parameters. *Trans. ASAE* 24(2), 335-339.

559 Chiew, F. and McMahon, T. (1999) Modelling runoff and diffuse pollution loads in urban areas. *Water*
560 *Science and Technology* 39(12), 241-248.

561 Committee, M.S.S. (2005) The Minnesota Stormwater Manual. Developed by Emmons and Olivier
562 Resources for the Stormwater Steering Committee, Minnesota Pollution Control Agency, St. Paul, MN.
563 Paul, MN.

564 County, K. (2008) Knox County Tennessee Stormwater Management Manual. Knox County, Knoxville,
565 TN.

566 Ebrahimian, A., Gulliver, J.S. and Wilson, B.N. (2016a) Effective Impervious Area for Runoff in Urban
567 Watersheds. *Hydrological Processes* 30(20), 3717-3729.

568 Ebrahimian, A., Wilson, B.N. and Gulliver, J.S. (2016b) Improved methods to estimate the effective
569 impervious area in urban catchments using rainfall-runoff data. *Journal of Hydrology* 536, 109-118.

570 Epps, T. and Hathaway, J. (2018) Establishing a Framework for the Spatial Identification of Effective
571 Impervious Areas in Gauged Basins: Review and Case Study. *Journal of Sustainable Water in the Built*
572 *Environment* 4(2), 05018001.

573 Farnsworth, R. K., & Thompson, E. S. (1983). Mean monthly, seasonal, and annual pan evaporation for
574 the United States. US Department of Commerce, National Oceanic and Atmospheric Administration,
575 National Weather Service.

576 Gilroy, K.L. and McCuen, R.H. (2009) Spatio-temporal effects of low impact development practices.
577 *Journal of Hydrology* 367(3), 228-236.

578 Gupta, H.V., Kling, H., Yilmaz, K.K. and Martinez, G.F. (2009) Decomposition of the mean squared error
579 and NSE performance criteria: Implications for improving hydrological modelling. *Journal of Hydrology*
580 377(1-2), 80-91.

581 Han, W.S. and Burian, S.J. (2009) Determining effective impervious area for urban hydrologic modeling.
582 *Journal of Hydrologic Engineering* 14(2), 111-120.

583 Hatt, B.E., Fletcher, T.D., Walsh, C.J. and Taylor, S.L. (2004) The influence of urban density and drainage
584 infrastructure on the concentrations and loads of pollutants in small streams. *Environ Manage* 34(1),
585 112-124.

586 Jefferson, A.J., Bhaskar, A.S., Hopkins, K.G., Fanelli, R., Avellaneda, P.M. and McMillan, S.K. (2017)
587 Stormwater management network effectiveness and implications for urban watershed function: A
588 critical review. *Hydrological Processes* 31(23), 4056-4080.

589 Lee, J.G. and Heaney, J.P. (2003) Estimation of urban imperviousness and its impacts on storm water
590 systems. *Journal of Water Resources Planning and Management* 129(5), 419-426.

591 Leopold, L.B. (1968) Hydrology for urban land planning: A guidebook on the hydrologic effects of urban
592 land use.

593 Loperfido, J.V., Noe, G.B., Jarnagin, S.T. and Hogan, D.M. (2014) Effects of distributed and centralized
594 stormwater best management practices and land cover on urban stream hydrology at the catchment
595 scale. *Journal of Hydrology* 519, 2584-2595.

596 Mancipe-Munoz, N. A., Buchberger, S. G., Suidan, M. T., & Lu, T. (2014) Calibration of rainfall-runoff
597 model in urban watersheds for stormwater management assessment. *Journal of Water Resources
598 Planning and Management*, 140(6), 05014001.

599 Miller, R. (1978) The hydraulically effective impervious area of an urban basin, Broward County, Florida.

600 Musgrave, G. (1955) How much of the rain enters the soil. *Water: US Department of agricultural
601 yearbook*, 151-159.

602 Nash, J.E. and Sutcliffe, J.V. (1970) River flow forecasting through conceptual models part I—A
603 discussion of principles. *Journal of Hydrology* 10(3), 282-290.

604 Niazi, M., Nietch, C., Maghrebi, M., Jackson, N., Bennett, B.R., Tryby, M. and Massoudieh, A. (2017)
605 Storm Water Management Model: Performance Review and Gap Analysis. *Journal of Sustainable Water
606 in the Built Environment* 3(2), 04017002.

607 Page, J.L., Winston, R.J., Mayes, D.B., Perrin, C. and Hunt, W.F. (2015) Retrofitting with innovative
608 stormwater control measures: Hydrologic mitigation of impervious cover in the municipal right-of-way.
609 *Journal of Hydrology* 527, 923-932.

610 Poff, N.L., Allan, J.D., Bain, M.B., Karr, J.R., Prestegaard, K.L., Richter, B.D., Sparks, R.E. and Stromberg,
611 J.C. (1997) The natural flow regime. *BioScience*, 769-784.

612 Rawls, W.J., Brakensiek, D.L. and Miller, N. (1983) Green-Ampt infiltration parameters from soils data.
613 *Journal of Hydraulic Engineering* 109(1), 62-70.

614 Rossman, L. (2004) Storm Water Management Model (SWMM version 5.0) user's manual. United States
615 Environment Protection Agency.

616 Rossman, L. and Huber, W. (2016) Storm Water Management Model Reference Manual Volume I–
617 Hydrology (Revised). US Environmental Protection Agency, Cincinnati, OH. Revised January.

618 Roy, A.H., Rhea, L.K., Mayer, A.L., Shuster, W.D., Beaulieu, J.J., Hopton, M.E., Morrison, M.A. and Amand,
619 A.S. (2014) How much is enough? Minimal responses of water quality and stream biota to partial retrofit
620 stormwater management in a suburban neighborhood. *PLoS one* 9(1), e85011.

621 Roy, A.H. and Shuster, W.D. (2009) Assessing impervious surface connectivity and applications for
622 watershed management1. *JAWRA Journal of the American Water Resources Association* 45(1), 198-209.

623 Roy, A.H., Wenger, S.J., Fletcher, T.D., Walsh, C.J., Ladson, A.R., Shuster, W.D., Thurston, H.W. and
624 Brown, R.R. (2008) Impediments and solutions to sustainable, watershed-scale urban stormwater
625 management: lessons from Australia and the United States. *Environ Manage* 42(2), 344-359.

626 Shuster, W. and Rhea, L. (2013) Catchment-scale hydrologic implications of parcel-level stormwater
627 management (Ohio USA). *Journal of Hydrology* 485, 177-187.

628 Shuster, W.D., Bonta, J., Thurston, H., Warnemuende, E. and Smith, D.R. (2005) Impacts of impervious
629 surface on watershed hydrology: A review. *Urban Water Journal* 2(4), 263-275.

630 SSURGO (2017) Soil Survey Geographic (SSURGO) Database. Natural Resources Conservation Service,
631 United States Department of Agriculture (Available online at <https://sdmdataaccess.sc.egov.usda.gov>).

632 Sylvester, L., Omitaomu, O., & Parish, E. (2016) Analyzing the Implications of Climate Data on the Rainfall
633 Frequency Spectrum: Case Study of Knoxville, Tennessee and Surrounding Region (No. ORNL/TM--
634 2016/485). Oak Ridge National Laboratory (ORNL), Oak Ridge, TN (United States).

635 Versini, P.A., Gires, A., Tchiguirinskaia, I., Schertzer, D., (2016) Toward an operational tool to simulate
636 green roof hydrological impact at the basin scale: a new version of the distributed rainfall-runoff model
637 Multi-Hydro. *Water Science and Technology* 74(10), 1845-1854.

638 Walsh, C.J., Fletcher, T.D., Bos, D.G. and Imberger, S.J. (2015) Restoring a stream through retention of
639 urban stormwater runoff: a catchment-scale experiment in a social–ecological system. *Freshwater
640 Science* 34(3), 1161-1168.

641 Walsh, C.J., Roy, A.H., Feminella, J.W., Cottingham, P.D., Groffman, P.M. and Morgan, R.P. (2005a) The
642 urban stream syndrome: current knowledge and the search for a cure. *Journal of the North American
643 Benthological Society* 24(3), 706-723.

645 Walsh, C.J., Fletcher, T.D. and Ladson, A.R. (2005b) Stream restoration in urban catchments through
646 redesigning stormwater systems: looking to the catchment to save the stream. *Journal of the North*
647 *American Benthological Society* 24(3), 690-705.

648 Wilcoxon, F., Katti, S. K., & Wilcox, R. A. (1970). Critical values and probability levels for the Wilcoxon
649 rank sum test and the Wilcoxon signed rank test. *Selected tables in mathematical statistics*, 1, 171-259.

650 Yen, B.C. (2001) *Hydraulics of sewer systems. Stormwater collection systems design handbook*.