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Nature-based approaches to managing climate change impacts in cities

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Managing and adapting to climate change in urban areas will become increasingly important as urban populations grow, especially because unique features of cities amplify climate change impacts. High impervious cover exacerbates impacts of climate warming through urban heat island effects and of heavy rainfall by magnifying runoff and flooding. Concentration of human settlements along rivers and coastal zones increases exposure of people and infrastructure to climate change hazards, often disproportionately affecting those who are least prepared. Nature-based strategies (NBS), which use living organisms, soils and sediments, and/or landscape features to reduce climate change hazards, hold promise as being more flexible, multi-functional and adaptable to an uncertain and non-stationary climate future than traditional approaches. Nevertheless, future research should address the effectiveness of NBS for reducing climate change impacts and whether they can be implemented at scales appropriate to climate change hazards and impacts. Further, there is a need for accurate and comprehensive cost-benefit analyses that consider disservices and co-benefits, relative to grey alternatives, and how costs and benefits are distributed across different communities. NBS are most likely to be effective and fair when they match the scale of the challenge, are implemented with input from diverse voices and are appropriate to specific social, cultural, ecological and technological contexts.

This article is part of the theme issue 'Climate change and ecosystems: threats, opportunities and solutions'.

1. Introduction

Devising strategies to manage and adapt¹ to the impacts of climate change in urban areas will become increasingly important as the global population becomes more and more concentrated in cities and climate continues to change in ways that have potentially severe impacts on urban populations and infrastructure. Some urban climate change hazards (physical manifestations of climate change that have detrimental consequences for human well-being) will be especially challenging to manage (i.e. to reduce) because urban environments interact with and influence climate in ways that can amplify (worsen) those hazards. This leaves urban residents with no choice but to adapt to those hazards, reducing exposure to hazards or their harmful effects, or to retreat. On the other hand, cities offer opportunities for managing and adapting to climate change hazards because most services and institutions are located within cities.

There is increasing interest among city practitioners in using 'nature-based' strategies [2], a broad suite of actions aimed at promoting human well-being in cities using approaches that restore aspects of 'natural' (non-urban) ecosystem structure and/or function. These strategies are seen as more flexible, multifunctional and adaptable to an uncertain climate future than traditional, more rigid, approaches [3–5]. Herein, we review the potential for nature-based strategies (NBS) to reduce climate change hazards in cities. NBS include parks and open space; intentional plantings; construction of structures that

restore natural hydrologic function such as stormwater ponds, bioswales, green roofs, riparian zones; and restoration and protection of natural protective habitats along coastlines [5]. We begin by reviewing the primary climate change hazards faced by urban residents, highlighting how cities can amplify many of those hazards (§2) and the impacts of climate change hazards on social, ecological and technical (SET) components of cities (§3). We then review opportunities to use NBS to reduce exposure to climate change hazards (§4) and manage and adapt to the SET impacts of those hazards (§5). We conclude by identifying research priorities (§6).

2. Exposure to and amplification of climate change hazards in cities

Cities often are at great risk from climate change because human settlements are concentrated in areas that increase the vulnerability of people and infrastructure to climate change hazards [6]. Furthermore, cities have unique features, such as high impervious and low pervious cover, that amplify some aspects of climate change. Thus, several aspects of climate change are expected to be acutely felt by urban populations, including sea-level rise, higher mean and night-time temperatures, reduced snowpack or rainfall leading to water scarcity, and increased frequency and magnitude of extreme events like heavy precipitation, heat waves, coastal storms, river flooding and drought (table 1). Indeed, extreme events are the aspect of climate change that will have the most direct and obvious impact on the greatest number people in their lifetimes.

(a) Sea-level rise and coastal storms

The rate of sea-level rise is increasing, with projections of 0.3–1.3 m by 2100 [8]. Globally, human populations and most major cities are concentrated in low-elevation coastal zones [9], increasing the exposure of people and infrastructure to inundation, storm surge, flooding, erosion and salt-water intrusion. For example, populations in nine Asian mega-deltas are greatest in areas less than 10 m above sea level, and many of these populations are growing rapidly [10,11]. In some sensitive, low-lying areas, such as much of South Florida (USA), groundwater levels are carefully managed to prevent intrusion of sea water into the water supply for greater than 5 million people. However, higher groundwater elevation reduces water storage capacity of surface soils, and may amplify overland flooding [12] and ultimately necessitate retreat from the coast (table 1).

Coastal cities also experience storms that exert impacts through flooding and sea surge, erosion and wind, and which are projected to increase in frequency and magnitude [13]. Sea surge can dramatically amplify coastal flood risk. In the case of superstorm Sandy (New York City, USA) in 2012, sea surges coincided with high tides, flooding subways and coastal infrastructure, and inundating an area equivalent to that projected to flood under a 2080 scenario of rapid icemelt [14]. Erosion caused by the storm also undermined infrastructure along much of the local coastline.

High winds from tropical cyclones can prove devastating to ecological and technological infrastructure on coasts, although wind speeds decline rapidly once making landfall. The coastal cities of East Asia, particularly those in the Pearl River Delta, Tokyo and Manila, are most at risk from this type of extreme event [15]. Inland, windthrow and infrastructure damage can result from other types of storms. Regardless of storm type, high winds also produce health risks because of entrainment of air pollutants, including disease organisms [16]; this risk is amplified in cities because they often have elevated concentrations of air pollutants, including particulate matter, ozone and toxins.

(b) Extreme heat

Global mean temperature has increased 0.8°C since 1880 [17] and will continue to rise. Warming is exacerbated in cities by the urban heat island (UHI) effect, whereby cities are warmer on average than surrounding areas [18]. For example, US cities are warming 1.5 times faster than rural areas [19]. A number of factors combine to cause UHI effects [20]: lack of vegetation and associated cooling effects of evapotranspiration, high capacity for heat storage in building materials, high aerodynamic resistance to heat dissipation and generation of waste heat from energy use by buildings and vehicles.

Hotter conditions drive impacts on cities through several mechanisms (table 1). Exceedance of thermal tolerances constrains activity of humans and other organisms, especially at low latitudes. Prolonged human exposure is exacerbated by the UHI, rising night-time minimum temperatures and longer warm seasons. The energy system is stressed by increased demand for cooling. Higher temperatures also exacerbate drought stress and worsen air pollution and episodes of air stagnation resulting from thermal inversions. The impacts of heat waves thus will be intensified in cities [21].

Heat waves will be especially likely to cause heat stress in hot and humid regions [20,22], since heat stress is a function of both temperature and humidity (along with physiological and behaviour factors) [23]. In a global analysis of heat waves and human mortality, temperatures of only 20°C were lethal (i.e. were associated with excess mortality) when relative humidity was high (80%), whereas the lethal temperature increased to 30°C at low relative humidity (20%) [24]. Based on these empirical relationships, Jakarta, Indonesia, is therefore predicted to experience anywhere from 117 to 365 d yr⁻¹ of lethal temperature–humidity combinations by 2100 (for a low-versus high-emissions scenario, respectively), compared with 9–50 d yr⁻¹ in New York City, USA [24].

(c) Water security, inland storms and pluvial flooding

Compared to temperature change, hydrologic impacts of climate change will be orders of magnitude more spatially and temporally variable. Because most cities rely on regional or distant watersheds for their water supply, or on ground-water sources that are replenished slowly, the impacts of climate change on hydrology over broad areas are important for urban water security. Changes in the total amount, intensity and seasonality of precipitation all have potential to influence urban areas. These changes will vary in space, with some regions seeing greater precipitation and some seeing less [8]. Cities amplify drought through increased water demand for maintenance of vegetation or cooling.

Of all natural disasters, river flooding exposes the most people (379 million) [15] to climate-related hazards. Much of urban infrastructure is located along rivers and even in riverbeds, where it is vulnerable to inundation and riverbank

Table 1. Climate change hazards that threaten cities. Included are the mechanisms by which these hazards impact social (S), ecological (E) and technical (E) aspects of cities, nature-based approaches to managing and adapting to these impacts, and, for comparison, alternative strategies that are not nature-based solutions (e.g. 'grey' infrastructure).

climate change hazards	mechanisms of impact	impacts (S, E, T)	nature-based adaptation and management strategies	alternative 'grey' strategies
sea-level rise	salt-water intrusion	S, E	none (retreat)	alternative water supplies desalinization
	tidal inundation	5, Т	none (retreat)	bulkheads sea gates dikes pumps
coastal storms	sea surge: flooding	S, E, T	mangroves, wetlands, dunes, reefs	sea walls sea gates dikes pumps
	coastal erosion	S, E, T	mangroves, wetlands, dunes, reefs	sea walls bank stabilization, armouring
wind storms (coastal or inland)	high winds: blow-down wind entrainment of particles	E, T S	wind-resistant plantings plants to filter air-borne pollutants	wind-resistant, buried infrastructure avoidance
inland storms—riverine flooding	overbank flooding: inundation	S, E, T	wetlands ponds reservoirs floodplain widening	levees
	erosion, bank failure	E, T	wetlands channel restoration tree planting floodplain widening	bank stabilization, armouring

(Continued.)

Table 1. (Continued.)

			nature-based adaptation and	
climate change hazards	mechanisms of impact	impacts (S, E, T)	management strategies	alternative 'grey' strategies
inland storms—pluvial flooding	damage from high flows	5, T	tree canopy cover to promote interception and slow flows	channelization storm drainage capacity
			green infrastructure ^a to slow flows and promote infiltration, especially in 'headwaters'	pervious pavement
	inundation of low-lying areas	5, T	wetland restoration, retention basins	pumps pervious pavement
	transport of pollutants from land to	S, E	green infrastructure ^a to slow flows and promote	water treatment
	stormwater system		infiltration	street narrowing pervious pavement
	combined sewer overflows	S, E, T	green infrastructure ^a to slow flows and promote	sanitary and storm sewer separation
			infiltration	storage tanks
			constructed treatment wetlands	
higher overall temperature	thermal intolerance (reduced recruitment, increased mortality), phenological shifts	ш	managed relocation	
	exacerbated urban heat island	S, E, T	tree canopy cover	air conditioning
			parks and open space	misting stations
			green roofs	swimming pools
				light-coloured building materials
				heat-resistant building materials
	drought stress	ш	plantings for drought tolerance	irrigation
higher night-time temperature	exacerbated urban heat island, prolonged	S, E	tree canopy cover	air conditioning
	human exposure		parks and open space	swimming pools
			green roofs	
				(Continued.)

Table 1. (Continued.)

			nature-based adaptation and	
climate change hazards	mechanisms of impact	impacts (S, E, T)	management strategies	alternative 'grey' strategies
heat waves	increased human exposure	S	tree canopy cover parks and open space green roofs	air conditioning misting stations swimming pools light-coloured building materials
	increased demand for cooling	7 S, E	tree canopy cover parks and open space green roofs tree canopy cover ^b	air conditioning misting stations swimming pools light-coloured building materials heat-resistant building materials remaining inside and reducing activity building to promote advection
extreme cold	increased demand for heating; increased human exposure	5, T	wind breaks	heating
drought	water use restrictions	S, E	green inflastructure ^a to slow flows and promote infiltration landscaping for drought tolerance runoff, grey water capture and reuse rainwater capture	replacing landscaping with hardscaping
	increased water scarcity	у, Б	green infrastructure ^a to promote infiltration and groundwater recharge landscaping for drought tolerance runoff, grey water capture and reuse rainwater capture	groundwater pumping water transfer systems
	water conflict	۲ ,۲	none	water transfer systems graded water pricing treaties
	reduced water quality	S	green infrastructure ^a to promote nutrient uptake restored wetlands	water treatment

^aStormwater ponds, raingardens, bioswales, parks and open space, green roofs and similar structures.

^bThe net effect of increased tree canopy cover on air pollution is unclear, as trees both remove and produce and trap pollutants [7].

erosion. Cities have traditionally sought to armour and strengthen their banks using built infrastructure rather than rely upon natural floodplains (table 1). This dependence upon levees, channel straightening and hardening can produce a false sense of security, especially given the changing probabilities of flooding [25,26].

High impervious cover in cities exacerbates the impacts of heavy rainfall and magnifies urban runoff and flooding. Roads, parking lots, buildings and other impervious surfaces prevent infiltration. Combined with dense drainage networks, low infiltration causes higher volume and more rapid runoff [27-29]. High impervious cover also promotes pluvial flooding, flooding that occurs when precipitation rates exceed the capacity of stormwater systems or infiltration [30]. Such flooding can occur far from coasts or rivers as a result of brief, intense rainstorms, and will become more common as the frequency and magnitude of extreme rainfall increase [8]. Pluvial flooding can damage infrastructure directly through the force of flowing water, and is often unpredictable; for example, US Federal Emergency Management Administration (FEMA) flood maps usually do not apply to pluvial flooding [30]. In general, low places in the urban landscape are most at risk and often inhabited by people with the lowest capacity to adapt or respond.

(d) Other climate-change-related hazards

In addition to temperature- and water-related hazards, cities are potentially exposed to extreme events and disturbances such as fires, tornados, hurricanes, landslides and seismic events, some of which may be exacerbated by climate change. Human decisions and settlement locations strongly influence the exposure of urban residents to these hazards; for example, fires that once levelled cities around the turn of the twentieth century no longer pose a threat, but fires at the urban-wildland interface are on the rise [31]. As people are increasingly settling at this interface [32], future increases in fire resulting from climate change [33] will likely increase human mortality. Even when fire does not directly threaten urban lives and settlements, wildfires can expose downwind urban residents to severe air pollution, worsening heatrelated mortality, as happened in Moscow during the Russian heat wave of 2010 [34].

3. Climate change impacts on social—ecological—technical components in cities

(a) Social impacts

In cities, extreme temperatures aggravated by UHI effects will cause a host of heat-stress-related physical health impacts [35], including heat stroke mortality and morbidity, dehydration and related illness, and heat exhaustion [36]. The most vulnerable include the elderly, young and socially isolated, those who lack air conditioning, outdoor workers and those experiencing homelessness [35–37]. Within cities, the poor may be more vulnerable to heat exposure because they often lack air conditioning and live in neighbourhoods with less vegetative cooling capacity [38,39]. Heat exposure will also reduce the capacity for physical activity, reducing worker productivity and exercise activities [36].

Heavy rains and associated pluvial flooding also can affect human health. Urban runoff entrains and transports

heavy metals, nutrients, salts and pathogenic bacteria from the landscape to stormwater systems [40,41], resulting in pulses of high pollution downstream and creating episodic risks to public health and aquatic ecosystem health. Combined sewer overflows (CSOs) occur when runoff from heavy rains overwhelms sewers carrying both stormwater runoff and human waste, leading to releases of untreated sewage into surface waters to prevent sewer backups [42].

Climate change can cause human mortality, injury and displacement directly, by posing immediate hazards, and can also drive displacement by influencing economic, political, demographic and social drivers of migration [43,44], with the greatest impacts in low- to middle-income countries [45]. Exposure of many urban residents to hazards associated with sea-level rise, storms and river flooding may mean the damage or loss of homes, safety concerns from erosion undermining buildings or roads, and contamination of drinking water supplies. The poor, such as those in informal settlements, are at greatest risk because they lack adequate housing and other infrastructure, clean water, and access to healthcare and emergency services [46]. Extreme heat will make parts of the world without significant capacity for adaptation uninhabitable. Even holding average global warming to 1.5°C since pre-industrial times, 40% of mega-cities, concentrated in Africa and Asia, will experience periods of deadly heat indices each year [47]. Warming of 4°C will translate into periods of deadly heat indices in nearly 80% of the world's mega-cities.

Exposure to climate-change-related hazards will have mental as well as physical health impacts, including in cities. Weather-related disasters have been related to post-traumatic stress disorder, depression and anxiety [48]. Extreme heat has been linked to aggression, criminal behaviour, suicides, mood disorders and dementia. Disruptions to livelihoods and health caused by climate-change-related damage of infrastructure and forced migration will disrupt social systems and social ties, reducing adaptive capacity of climate change migrants [48,49].

(b) Ecological impacts

Ecological systems in cities are particularly vulnerable to climate change because of interactive effects with disturbances, pests and pathogens, and stressors such as road deicers and soil compaction. In addition, rapid urban expansion in many parts of the world likely will alter regional biodiversity [50].

Correlations between climate factors and plant diversity suggest that urban biodiversity will be sensitive to climate change. For example, across cities, those with higher maximum temperatures had lower plant species richness of both cultivated and spontaneously occurring plants in residential yards [51]. On the other hand, richness of spontaneous species increased where winters were relatively mild, a pattern also seen for urban trees across North America [52]. Thus, climate change may reduce plant species richness in regions that are already hot, but increase it in colder regions.

Because of UHI effects, limited plant rooting volumes, compacted soils and contamination by pollutants such as heavy metals, pesticides, herbicides and salts [53–56], climate change likely will worsen heat- and drought-related stress on urban ecological communities [57], particularly where impervious cover and local temperatures are highest [58–60]. Under these stresses, species are likely to be more vulnerable

to pests and pathogens, whose populations and ranges may expand with climate change [61,62].

Urban tree cover may be reduced by climate change because many species of trees currently being planted in cities will not be suited to future climates [63,64]. In addition, cities are at risk from accidental introductions of novel pests and pathogens by the nursery trade [62], and low diversity of urban forests may promote establishment and spread of these pests and pathogens [65]. Urban trees are susceptible to damage and mortality from windthrow associated with more intense storms, as the root systems of urban trees often are inadequate to support trees during high winds [66]. This is because urban soils often are compacted, waterlogged, drought-stressed or contain fills, and because trenching, construction or mowing can damage root systems.

In urban lakes and ponds, higher water temperature combined with high nutrient loads will promote blooms of algae and harmful cyanobacteria [67,68] and extend the period of stratification [69]. Blooms combined with reduced exchange of deep waters with oxygenated surface waters can result in hypoxic events, which kill fish and benthic biota and stimulate anaerobic biogeochemical processes that release sediment nutrients into the water column, further contributing to urban lake eutrophication [69]. More thermally stable and warmer surface waters, among other factors, will promote harmful cyanobacteria [67], increasing exposure of urban residents and pets to cyanobacterial toxins [70].

Most streams in urban areas are drastically altered through channelization, burial and impoundment [71], which reduces their capacity to withstand climate-related impacts such as rising water temperature [72], increased peak flow associated with extreme events [29] and increased nutrient and pollutant loading from storms and CSOs. Both urbanization and climate change interact to adversely affect stream fish populations and communities, according to models manipulating the two drivers individually and in combination [73].

(c) Technical impacts

Technical or infrastructure systems are designed and managed to deliver specific services or protect urban populations. Services provided include electrical power, water delivery, mobility (i.e. transportation) and waste management; stormwater infrastructure, seawalls and levees, and climate-controlled buildings (for shelter) provide protection for urban populations [74,75]. Climate change will have strong impacts on infrastructure because of its location, age, design and exceedance of design limits.

Infrastructure is more dense and is co-located in cities, amplifying technical impacts of climate change hazards such as storms, as a storm of a given extent and intensity will cause more damage in urban than in rural areas. Furthermore, urban infrastructure such as power lines, sewer pipes and water delivery networks often occurs along public rights of way like roads or streams, which may be vulnerable to flooding. Thus, the co-location of multiple infrastructure systems can result in simultaneous failures. Waste removal infrastructure in some cities is combined with stormwater conveyance, with both water streams merging at wastewater treatment plants where pollutants are removed. During flooding, however, CSOs occur. Thus, there are benefits of co-location but also potentially severe consequence when systems fail or their capacity is exceeded.

Infrastructure built to withstand historical climate is beginning to fail as extreme events become more common, and building for the past will be inadequate in a non-stationary world. Many systems are not robust to extreme events that are happening with ever-greater frequency, not just because of old age and deteriorating condition [76], but because they are built to design standards based on past probabilities (e.g. 1% is a common standard) [77]. This practice is inadequate in a non-stationary, uncertain world where the future probability of an event cannot be predicted from knowledge of the past. Recent efforts to employ non-stationary flood-frequency analyses [78] show promise for rethinking infrastructure design.

4. Opportunities for reducing exposure to climate change hazards with nature-based strategies

NBS make use of living organisms, soils and sediments, and/ or landscape features to reduce climate change hazards or the amplifying effects of urban features on those hazards. Such strategies can provide alternative or complementary approaches to technical strategies and delay the need for human relocation, buying time to accelerate climate change mitigation (table 1) [3,79,80]. Diverse NBS, also referred to as green infrastructure or low-impact development [81], range from highly engineered structural stormwater control measures to parks and open space, building materials and designs incorporating natural elements, conservation and restoration of natural ecosystems particularly on coastlines, and intentional plantings (table 1) [5,80]. (Pauleit et al. [7] provide a useful lexicon for nature-based solutions, ecosystem-based adaptation, ecosystem services and urban green infrastructure.)

(a) Sea-level rise and coastal storms

NBS can reduce the growing risks of coastal flooding and erosion from higher storm surges accompanying sea-level rise. Conserving and restoring near shore habitats such as barrier islands, coral and oyster reefs, and kelp and seagrass beds, and coastal habitats such as dunes, mangrove forests and saltmarshes can reduce erosion and protect human settlements [82]. These habitats dissipate wave energy, attenuate wave height, reduce storm surge, and trap and stabilize soils and sediments [5,83], and are thus more resistant (i.e. they sustain less damage) and more resilient (i.e. they can potentially self-recover) to damaging effects of storms than grey infrastructure, such as bulkheads and sea walls [79,83,84].

No nature-based strategy can entirely prevent the gradual march of seas inland because of sea-level rise. Costly engineered structures such as sea gates, dikes and pumps, along with development of alternative water supplies, can delay the need for human resettlement. Yet, managing inundation and salinization of drinking and irrigation water supplies ultimately will necessitate retreat from coastal areas and relocation of coastal residents [85]. Nevertheless, NBS such as mangrove forests and salt marshes in some instances may establish shoreward through sediment accretion to keep pace with sea-level rise [83], if there is adequate space between any coastal development and the waterline.

(b) Extreme heat

Urban green space, in the form of parks and open space, green roofs and tree canopy, has received a great deal of attention regarding its potential to reduce the UHI and provide relief from climate-change-induced heat waves [19,86,87]. A meta-analysis of the cooling effects of urban green space in cities throughout the world found that parks were 1°C cooler on average than non-park areas, presumably because of high evaporative cooling and low heat storage, and these cooling effects extended well beyond the park boundaries. Local-scale cooling was the only nature-based benefit rated as 'high' among ecosystem services (including climate regulation, air quality regulation and carbon sequestration) assessed for the Barcelona, Spain metropolitan area [88]. On hot summer days in Phoenix (Arizona, USA), vegetated surfaces can be as much as a 25°C cooler than bare surfaces [38]. Studies of the cooling effects of green roofs in cities from tropical, subtropical and temperate regions found inconsistent temperature differences between green and nearby non-vegetated roofs [86]. Nevertheless, a modelling study of climate change in US cities found that 100% deployment of green roofs offset projected climate warming [89], although reflective roofs (technical strategy) were more effective at cooling than green roofs. Trees provide shade and evaporative cooling and are generally cooler than nearby areas without trees [86]. Therefore, increasing tree canopy in cities is projected to offset some climate warming and UHI effects [90]. However, in one city, nonlinear effects of tree canopy were found, whereby increasing the canopy cover did not yield significant cooling effects until relatively high canopy cover was reached [91]. Furthermore, cooling effects of tree canopy were lower at night and when impervious cover was higher.

(c) Inland storms, pluvial flooding and droughts

NBS such as green roofs, stormwater ponds, bioswales, raingardens and retention basins can promote infiltration and groundwater recharge and/or evapotranspiration, thereby reducing runoff volumes and flow rates during heavy rain storms [92,93]. If placed strategically in the landscape, i.e. dispersed throughout the landscape and placed adjacent to roads, such strategies can decrease risks of pluvial flooding [94] and the damaging effects of high-velocity runoff. For example, in simulations of a Chicago (IL, USA) watershed, having 10% of the landscape area in green infrastructure minimized flood risk associated with moderate storms. However, increasing the storm intensity to that expected under climate change (today's 1% probability), required increasing areal extent of green infrastructure by a factor of two or more to manage flooding [94]. Indeed, many nature-based interventions in stormwater systems are implemented at too small a scale to have any effect on large-scale, catastrophic events [95]. Thus, significant land area of stormwater green infrastructure will be needed to manage increased flood risk in regions predicted to experience more severe storms.

Many cities have been focused on increasing urban tree canopy cover, in part to reduce stormwater runoff volumes [96,97]. By intercepting rainfall, which is stored in the canopy and eventually evaporated [98], trees may reduce stormwater runoff volumes and delay peak flows during low-intensity rainstorms [98]. Transpiration by trees

can potentially reduce runoff by providing greater soil volume for water storage [98].

Many cities are expanding use of green stormwater infrastructure to reduce runoff volumes during heavy rain events, reduce the risk of CSOs and potentially improve water quality [99–101]. Green infrastructure provides some capacity for pollutant removal from stormwater [101] and for receiving and treating CSOs (e.g. constructed wetlands) [102]. Stormwater green infrastructure traps particulate nutrients and promotes sorption, biotic uptake or gaseous losses of soluble nutrients [5,103,104]. Thus, increasing green infrastructure in the watershed may reduce stormwater export of nutrients to streams [93,105]. On the other hand, increasing tree canopy cover near streets contributes litter-derived nutrients to stormwater [106,107] and green roofs show inconsistent water quality benefits [108].

5. Managing and adapting to social, ecological and technical impacts of climate change using nature-based strategies

Besides reducing climate change hazards and the amplification of these hazards in cities to minimize climate change impacts, nature-based approaches can help cities manage and adapt to the SET impacts of climate change when they occur (table 1). The occurrence of disasters may represent opportunities for city governments to go beyond impact and effect real change or adapt by 'building back better' [109,110]—a social response that may incorporate NBS. In rapidly urbanizing areas, such as in sub-Saharan Africa and many parts of Asia, cities have the opportunity to address climate change as they develop, implementing NBS in the most effective combinations with grey infrastructure and in ways that maximize locally valued co-benefits [111,112].

(a) Managing social impacts using nature-based strategies

Many NBS to manage and adapt to climate change (table 1) have potential social co-benefits. For example, access and exposure to green space in cities improves aspects of mental and physical health [5,113,114]. Urban green space also has been linked to social benefits such as reduced violence and crime, although conflicting results indicate a need to understand mechanisms underlying potential links as well as for standardized approaches to quantifying them [113,115]. Whether NBS to addressing climate change in cities will have additional social benefits by countering the detrimental social effects of climate change, including impaired mental health, reduced social cohesion and increased violence, remains unknown. What is certain, however, is that the success of NBS is in large part dependent on public acceptance [116].

(b) Managing ecological impacts using nature-based strategies

Myriad opportunities exist to manage urban ecosystems to facilitate transitions to species more suited to a changing climate [64]. For example, forest managers in Chicago, IL, USA, identified a number of management actions to reduce urban forest vulnerability to climate change [64]. Increasing the

Table 2. Research priorities related to implementing NBS for addressing climate change hazards and their impacts.

Ten priority research questions

Effectiveness

- 1. What kinds, amounts and arrangements of nature-based strategies mitigate different climate-change-related hazards?
- 2. How transferable are nature-based strategies among cities?
- 3. Are there nonlinearities (e.g. thresholds or limits) that influence the effectiveness of nature-based strategies for mitigating particular climate-change-related hazards?
- 4. Can the amounts and arrangements of green infrastructure be designed to mitigate multiple hazards simultaneously?
- 5. Under what conditions will climate change hazards overcome the capacity for nature-based solutions to reduce SET impacts in cities?
- 6. What approaches are most effective for designing for future rather than past climate conditions?

Costs and benefits

7. What are the costs (including disservices) and benefits (including co-benefits) of nature-based strategies relative to grey alternatives, in different social, environmental and technical contexts?

Equity and environmental justice

- 8. How can the costs and benefits of nature-based solutions be distributed equitably across different communities within cities?
- 9. How can cities avoid 'green gentrification' and other unintended outcomes of implementing nature-based strategies?
- 10. How can the implementation of nature-based strategies accelerate improvement of living conditions in the world's poorest, but fastest-growing, cities?

diversity of plantings in developed areas and installing drainage systems to reduce flooding could reduce vulnerability of urban forests to extreme events. Planting more drought-, flood- or heat-adapted species and more pest- and pathogen-resistant varieties could facilitate the transition to species better suited to future climate.

In theory, many of the NBS for reducing climate change hazards have co-benefits for biodiversity in a changing environment, by providing habitat reserves and corridors for species migration in the face of climate change. However, NBS for reducing climate change hazards are not automatically optimal for conserving biodiversity and facilitating ecological adaptation to climate change. A number of factors might impede efforts to achieve biodiversity goals [117-120]. These factors include the challenges of coordinating management of distributed (often private) small patches of green space to create habitat patches or corridors at a scale appropriate to promote species of interest; goals for green space that conflict with biodiversity goals; management practices that might be outright detrimental to achieving biodiversity goals (e.g. mowing, trimming, use of herbicides and pesticides, planting of ornamentals); and misunderstanding of biodiversity benefits and negative perceptions of urban green space managed for biodiversity that might generate dissatisfaction among urban residents. Thus, enhanced ecological adaptation to climate change impacts will not occur as an inevitable outcome of implementing NBS to reduce climate change hazards, but potentially could be helpful if pursued with intent and accompanied by education efforts [118].

(c) Managing technical impacts using nature-based strategies

The use of nature to modify the built infrastructure of a city may seem antithetical to managing technical impacts in a city, but in fact many engineers are seeing the value of incorporating a systems view and redefining risk in the

context of a non-stationary world [42,74]. Designs that move away from emphasis solely on the probability of an event to also consider its consequences allow for more flexible, adaptable infrastructure that is safe-to-fail, not just fail-safe [77]. For example, the use of protective wetlands along coasts [82] may reduce the hazard of erosion or wave damage from storm surges to coastal infrastructure. NBS can help to counter technical impacts when used alone or in combination with technical strategies (also known as hybrid strategies) along an eco-technical spectrum [121] or green—grey gradient [122,123].

6. Research priorities

Below, we articulate three priority research areas related to using NBS to manage and adapt to climate change in cities, articulating specific research questions in table 2. We propose that in planning for investment in NBS to address climate change hazards, a city needs to consider the effectiveness of NBS to reduce climate change hazards and their impacts, both now and in the future; the costs and benefits of implementing nature-based versus alternative strategies; and equity and environmental justice issues related to the distribution of costs and benefits across different cities globally and across communities within cities. Answers will depend on the particular social, environmental and technical context of each city; in other words, there is no *a priori* reason to expect 'one size fits all' solutions to cities' climate change hazards.

(a) How effective are nature-based strategies at reducing the impacts of climate change hazards?

Research should address the degree to which NBS can be implemented at scales that match the scale of hazards and impacts caused by climate change [116] (table 2). Such scaling questions need to be addressed in the contexts of individual

cities: their local biome, climate and hydrogeology (e.g. sources of surface and groundwater); the magnitude and types of anticipated climate change hazards (e.g. droughts, extreme rainfall, sea-level rise); their specific SET characteristics (e.g. spatial segregation of risks, the age, type and distribution of green and grey infrastructure, and the social barriers to implementation of NBS); and opportunities to combine green with grey infrastructure. An important question is how NBS scale with implementation area, and how their effectiveness alone or in combination with grey infrastructure depends on their physical position in the landscape [84,93,94,116]. For example, a study of green roofs in Beijing, China, found that the reduction in UHI scaled linearly with green roof area: while 100% green roof coverage resulted in significant city-wide cooling (1.5°C), 10% green roof coverage led to only 0.1-0.2°C cooling [124]. On the other hand, a study in Madison (Wisconsin, USA) found that temperature was related to tree canopy cover in nonlinear ways that depended on impervious cover, spatial scale and time of day (daytime versus night-time) [91]. Given that space is limited in most cities, or its use for public good is complicated by private ownership, implementing NBS at sufficient scales and in appropriate configurations to allow meaningful adaptation may not be feasible given projected climate change impacts. Furthermore, many NBS require significant development time to reach maximum efficacy [125], or efficacy may increase and then decrease over time, owing to growth of plants, changes in substrate (e.g. soils and sediments) and deposition of litter or particulate matter. Thus, scaling in time as well as in space must be considered to fully understand the benefit-cost ratio of NBS.

(b) What are the costs and benefits of nature-based strategies relative to alternatives?

There is a need to accurately and comprehensively analyse costs and benefits, including disservices and co-benefits, of NBS relative to grey (or non-nature-based) alternatives (table 2) [5], recognizing that cities may not have mechanisms for valuing non-market services contributed by NBS. Two examples of this research approach include studies of green roofs and studies of nature-based coastal defences. Studies of the costs and benefits of green roofs have compared green roofs against those of reflective 'cool' roof alternatives, where costs included installation, maintenance and replacement costs, along with heating costs during cool times of the year, while benefits included reductions in energy use, avoided morbidity and mortality, and other health benefits from cooling, as well as reductions in stormwater runoff and air pollution [126,127]. Considering environmental context is critical, since cooling roofs will increase heating costs more in regions with colder climates and green roofs will increase irrigation costs more in regions with drier climates [126]. A synthesis of cost-benefit analyses of coastal defences worldwide showed that coastal habitats have high potential for protecting coastlines from flooding and erosion [128], and that salt marshes and coral reefs can be two to five times more cost-effective than engineered structures in protecting coastlines. These examples provide models for research analysing costs and benefits of implementing NBS to address climate change impacts.

(c) Are the benefits of nature-based strategies distributed equitably within and across cities?

Those who are most likely to experience climate change impacts within and across cities may also be those with the least access to nature-based relief from those impacts. Poorer communities within cities often are more vulnerable to climate change hazards because their members live or work in high-exposure areas and because they may lack the resources to adapt to climate change (e.g. air conditioning, adequate shelter or drainage). In addition, urban green spaces such as parks are more accessible to wealthy, white, able-bodied urban residents [129]. Similarly, tree canopy cover is often concentrated in wealthier neighbourhoods [130]. For example, many cities in the Amazon delta are experiencing rapid urbanization; these communities are highly vulnerable to flooding and contaminated water. Informal settlements, in particular, concentrate the urban poor in flood-prone areas that lack basic sanitation and water infrastructure [6]. This pattern repeats globally, with rapidly expanding cities in the decolonized world surrounded by unplanned settlements that develop an informal infrastructure that is rarely adequate to protect their inhabitants, while wealthy urban residents enjoy benefits of higher ground, adequate infrastructure or access to green space [131].

There is some hope that small-scale NBS can provide solutions to large disparities in vulnerability, if appropriate to the particular location [132]. However, a lurking challenge is the green gentrification that can accompany nature-based strategy implementation. New inequities often arise in urban greening projects when underprivileged people are denied access or representation (in planning and decisionmaking on siting of nature-based projects), or are edged out by deliberate attraction of high-income clients to newly constructed green space [133,134]. To ensure that all voices are heard in decision-making about NBS and that access and distribution are fair, deliberate policies that avoid green gentrification need to be developed in concert with urban planning. Kabisch et al. [135] outline four recommendations to ensure environmental justice in planning for urban green spaces: that the distribution of green spaces ensures equitable access; that efforts are made to give voice to all members of the population, including underrepresented groups; that interactions and social exchanges are open and safe; and that local characteristics are considered.

Across cities, NBS may be less effective in cities that will be most exposed to climate change hazards. Many cities in the Global South are equatorial, where rising temperature is more likely to cross tolerance thresholds. Many of these cities are also located in low-lying coastal areas and thus are vulnerable to storms and sea-level rise. Finally, cities in the Global South are rapidly growing and many lack financial means to keep up with construction of basic critical infrastructure, let alone combat the impacts of climate change through the construction of adequate protective infrastructure or design or preservation of NBS [74,75]. Without major acceleration of climate change mitigation, these cities may become unliveable and no magnitude of nature-based strategy implementation can sufficiently reduce the hazards they face. Yet, in others, NBS may provide a means to accelerate improvement of living conditions (table 2). Goal 11 of the Sustainable Development Goals adopted by the United Nations [136] pertains to sustainable cities and communities.

Of the 10 targets for Goal 11, three relate specifically to NBS: increase access to green space, reduce loss of lives and livelihoods from disaster, and increase city planning to create safe, inclusive, resilient and sustainable cities.

7. Conclusion

In this paper, we have considered how cities may amplify the hazards of climate change to SET components of cities. NBS have potential to reduce climate-change-related hazards directly; to dampen, rather than amplify, their effects on cities; and to minimize SET impacts. Reaching that potential requires fundamental research to understand the mechanisms that lead to amplification of hazards in cities and to quantify the conditions under which NBS will be effective. Researchers and practitioners must also understand fully the costs and benefits of green-infrastructure approaches to adapting cities to a changing climate, including both disservices and co-benefits, relative to traditional, grey approaches. We need a much-improved awareness of and commitment to ensuring that the most vulnerable populations within and among cities are not neglected, as proactive plans are developed to build socialecological-technological resilience to the challenges posed by a rapidly changing climate. Finally, we conclude that NBS to meet climate-related challenges are most likely to be effective when they match the scale of the challenge and are appropriate to a specific place, in terms of its social, cultural, ecological and technological milieu.

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Endnote

¹Throughout, we use adaptation, hazards, impacts, vulnerability and exposure according to the definitions of Field *et al.* [1].

References

- Field CB et al. 2014 Technical summary. In Climate change 2014: impacts, adaptation, and vulnerability. Part A: global and sectoral aspects. Contribution of working group II to the fifth assessment report of the intergovernmental panel on climate change (eds CB Field et al.), pp. 35–94. Cambridge, UK: Cambridge University Press.
- Shandas V, Matsler A, Caughman L, Harris A. 2019
 Towards the implementation of green stormwater infrastructure: perspectives from municipal managers in the Pacific Northwest. J. Environ. Planning Manage. Online. (doi:10.1080/09640568. 2019.1620708)
- Kabisch N, Korn H, Stadler J, Bonn A. 2017 Naturebased solutions to climate change adaptation in urban areas: linkages between science, policy and practice. Berlin, Germany: Springer Open.
- Gilrein EJ, Carvalhaes TM, Markolf SA, Chester M, Allenby BR, Garcia M. 2019 Concepts and practices for transforming infrastructure from rigid to adaptable. Sustain. Resilient Infrastruct. Online. (doi:10.1080/23789689.2019.1599608)
- Keeler BL et al. 2019 Social-ecological and technological factors moderate the value of urban nature. Nat. Sustain. 2, 29–38. (doi:10.1038/ s41893-018-0202-1)
- Mansur AV, Brondízio ES, Roy S, Hetrick S, Vogt ND, Newton A. 2016 An assessment of urban vulnerability in the Amazon Delta and Estuary: a multi-criterion index of flood exposure, socioeconomic conditions and infrastructure. Sustain. Sci. 11, 625–643. (doi:10.1007/s11625-016-0355-7)
- 7. Pauleit S, Zölch T, Hansen R, Randrup TB, van den Bosch CK. 2017 Nature-based solutions and climate

- change—four shades of green. In *Nature-based* solutions to climate change adaptation in urban areas: linkages between science, policy and practice. Theory and practice of urban sustainability transitions (eds N Kabisch, H Korn, J Stadler, A Bonn), pp. 15–28. Berlin, Germany: Springer Open.
- Hayhoe K, Wuebbles DJ, Easterling DR, Fahey DW, Doherty S, Kossin JP, Sweet W, Vose R, Wehner M. 2018 Chapter 2: Our changing climate. In *Impacts, risks, and adaptation in the United States: fourth national climate assessment,* vol. II (eds DR Reidmiller, CW Avery, DR Easterling, KE Kunkel, KLM Lewis, TK Maycock, BC Stewart), pp. 72–144. Washington, DC: US Global Change Program.
- Small C, Nicholls RJ. 2003 A global analysis of human settlement in coastal zones. *J. Coast. Res.* 19, 584–599.
- Small C, Sousa D, Yetman G, Elvidge C, MacManus K. 2018 Decades of urban growth and development on the Asian megadeltas. *Glob. Planet. Change* 165, 62–89. (doi:10.1016/j.gloplacha.2018.03.005)
- Güneralp B, Güneralp İ, Liu Y. 2015 Changing global patterns of urban exposure to flood and drought hazards. *Global Environ. Change* 31, 217–225. (doi:10.1016/j.gloenvcha.2015.01.002)
- Czajkowski J, Engel V, Martinez C, Mirchi A, Watkins D, Sukop MC, Hughes JD. 2018 Economic impacts of urban flooding in South Florida: potential consequences of managing groundwater to prevent salt water intrusion. Sci. Total Environ. 621, 465–478. (doi:10.1016/j.scitotenv.2017.10.251)
- Kossin JP, Hall T, Knutson T, Kunkel KE, Trapp RJ, Waliser DE, Wehner MF. 2017 Extreme storms. In Climate science special report: fourth national climate

- assessment, vol. I (eds DJ Wuebbles, DW Fahey, KA Hibbard, DJ Dokken, BC Stewart, TK Maycock), pp. 257–276. Washington, DC: US Global Change Research Program.
- Rosenzweig C, Solecki W. 2014 Hurricane Sandy and adaptation pathways in New York: lessons from a first-responder city. *Global Environ. Change* 28, 395–408. (doi:10.1016/j.gloenvcha.2014.05.003)
- Sundermann L, Schelske O, Hausmann P. 2013 Mind the risk: a global ranking of cities under threat from natural disasters. Zurich, Switzerland: Swiss Reinsurance Company Ltd.
- Sprigg WA et al. 2014 Regional dust storm modeling for health services: the case of valley fever. Aeolian Res. 14, 53–73. (doi:10.1016/j.aeolia. 2014.03.001)
- NASA Earth Observatory. World of change: global temperatures. Online. See https://earthobservatory. nasa.gov/world-of-change/DecadalTemp.
- Arnfield AJ. 2003 Two decades of urban climate research: a review of turbulence, exchanges of energy and water, and the urban heat island. *Int. J. Climatol.* 23, 1–26. (doi:10.1002/joc.859)
- Stone B, Vargo J, Habeeb D. 2012 Managing climate change in cities: will climate action plans work? Landscape Urban Plan. 107, 263–271. (doi:10.1016/ j.landurbplan.2012.05.014)
- Zhao L, Lee X, Smith RB, Oleson K. 2014 Strong contributions of local background climate to urban heat islands. *Nature* 511, 216. (doi:10.1038/ nature13462)
- Luber G, McGeehin M. 2008 Climate change and extreme heat events. *Am. J. Prev. Med.* 35, 429–435. (doi:10.1016/j.amepre.2008.08.021)

- Oleson KW, Monaghan A, Wilhelmi O, Barlage M, Brunsel N, Feddema J, Hu L, Steinhoff DF. 2015 Interactions between urbanization, heat stress, and climate change. *Clim. Change* 129, 525–541. (doi:10.1007/s10584-013-0936-8)
- Epstein Y, Moran DS. 2006 Thermal comfort and the heat stress indices. *Ind. Health* 44, 388–398. (doi:10.2486/indhealth.44.388)
- 24. Mora C *et al.* 2017 Global risk of deadly heat. *Nat. Clim. Change* **7**, 501. (doi:10.1038/nclimate3322)
- Milly PCD, Betancourt J, Falkenmark M, Hirsch RM, Kundzewicz ZW, Lettenmaier DP, Stouffer RJ. 2008 Climate change. Stationarity is dead: whither water management? Science 319, 573–574. (doi:10.1126/ science.1151915)
- Markolf SA, Chester MV, Eisenberg DA, Iwaniec DM, Davidson CI, Zimmerman R, Miller TR, Ruddell BL, Chang H. 2018 Interdependent infrastructure as linked social, ecological, and technological systems (SETSs) to address lock-in and enhance resilience. Earth's Future 6, 1638–1659. (doi:10.1029/ 2018EF000926)
- Leopold LB. 1968 A guidebook on the hydrologic effects of urban land use. US Geological Survey Circular Contract No. 554. Washington, DC: United States Department of the Interior, Geological Survey.
- 28. Kaushal SS, Belt KT. 2012 The urban watershed continuum: evolving spatial and temporal dimensions. *Urban Ecosyst.* **15**, 409–435. (doi:10. 1007/s11252-012-0226-7)
- 29. Walsh CJ, Fletcher TD, Burns MJ. 2012 Urban stormwater runoff: a new class of environmental flow problem. *PLoS ONE* **7**, e45814. (doi:10.1371/journal.pone.0045814)
- Rosenzweig BR, McPhillips L, Chang H, Cheng C, Welty C, Matsler M, Iwaniec D, Davidson Cl. 2018 Pluvial flood risk and opportunities for resilience. Wiley Interdiscip. Rev. Water 5, e1302. (doi:10.1002/ wat2.1302)
- Grimm NB, Pickett ST, Hale RL, Cadenasso ML. 2017
 Does the ecological concept of disturbance have utility in urban social—ecological—technological systems? *Ecosyst. Health Sustain.* 3, e01255. (doi:10. 1002/ehs2.1255)
- 32. Radeloff VC *et al.* 2018 Rapid growth of the US wildland-urban interface raises wildfire risk. *Proc. Natl Acad. Sci. USA* **115**, 3314. (doi:10.1073/pnas. 1718850115)
- Liu Y, Stanturf J, Goodrick S. 2010 Trends in global wildfire potential in a changing climate. For. Ecol. Manage. 259, 685–697. (doi:10.1016/j.foreco.2009. 09.002)
- 34. Shaposhnikov D *et al.* 2014 Mortality related to air pollution with the Moscow heat wave and wildfire of 2010. *Epidemiology* **25**, 359–364. (doi:10.1097/EDE.00000000000000000)
- 35. Watts N *et al.* 2018 The 2018 report of the *Lancet* Countdown on health and climate change: shaping the health of nations for centuries to come. *Lancet* **392**, 2479–2514. (doi:10.1016/S0140-6736(18)32594-7)
- 36. Kjellstrom T, Briggs D, Freyberg C, Lemke B, Otto M, Hyatt O. 2016 Heat, human performance, and

- occupational health: a key issue for the assessment of global climate change impacts. *Annu. Rev. Public Health* **37**, 97–112. (doi:10.1146/annurev-publhealth-032315-021740)
- Hondula DM, Davis RE, Saha MV, Wegner CR, Veazey LM. 2015 Geographic dimensions of heatrelated mortality in seven US cities. *Environ. Res.* 138, 439–452. (doi:10.1016/j.envres.2015.02.033)
- Jenerette GD, Harlan SL, Stefanov WL, Martin CA.
 2011 Ecosystem services and urban heat riskscape moderation: water, green spaces, and social inequality in Phoenix, USA. *Ecol. Appl.* 21, 2637–2651. (doi:10.1890/10-1493.1)
- Harlan SL, Chakalian P, Declet-Barreto J, Hondula D, Jenerette GD. 2019 Pathways to climate justice in a desert metropolis. In *People and climate change:* vulnerability, adaptation, and social justice (eds LR Mason, J Rigg), p. 23. Oxford, UK: Oxford University Prace
- Grimm NB, Faeth SH, Golubiewski NE, Redman CL, Wu J, Bai X, Briggs JM. 2008 Global change and the ecology of cities. *Science* 319, 756–760. (doi:10. 1126/science.1150195)
- 41. Kaushal SS, McDowell WH, Wollheim WM. 2014
 Tracking evolution of urban biogeochemical cycles:
 past, present, and future. *Biogeochemistry* **121**,
 1–21. (doi:10.1007/s10533-014-0014-y)
- 42. Pandit A *et al.* 2017 Infrastructure ecology: an evolving paradigm for sustainable urban development. *J. Clean. Prod.* **163**, S19–S27. (doi:10. 1016/j.jclepro.2015.09.010)
- 43. Warner K, Hamza M, Oliver-Smith A, Renaud F, Julca A. 2010 Climate change, environmental degradation and migration. *Nat. Hazards* **55**, 689–715. (doi:10.1007/s11069-009-9419-7)
- Black R, Adger WN, Arnell NW, Dercon S, Geddes A, Thomas D. 2011 The effect of environmental change on human migration. *Global Environ. Change* 21, S3–S11. (doi:10.1016/j.qloenvcha.2011.10.001)
- 45. Revi A et al. 2014 Urban areas. In Climate change 2014: impacts, adaptation, and vulnerability. Part A: global and sectoral aspects. Contribution of working group II to the fifth assessment report of the intergovernmental panel on climate change (eds CB Field et al.), pp. 535–612. Cambridge, UK: Cambridge University Press
- 46. Lankao PR, Qin H. 2011 Conceptualizing urban vulnerability to global climate and environmental change. *Curr. Opin. Environ. Sustain.* **3**, 142–149. (doi:10.1016/j.cosust.2010.12.016)
- 47. Matthews TK, Wilby RL, Murphy C. 2017
 Communicating the deadly consequences of global warming for human heat stress. *Proc. Natl Acad. Sci. USA* **114**, 3861–3866. (doi:10.1073/pnas. 1617526114)
- 48. Berry HL, Bowen K, Kjellstrom T. 2010 Climate change and mental health: a causal pathways framework. *Int. J. Public Health* **55**, 123–132. (doi:10.1007/s00038-009-0112-0)
- Torres JM, Casey JA. 2017 The centrality of social ties to climate migration and mental health. *BMC Public Health* 17, 600. (doi:10.1186/s12889-017-4508-0)

- Seto KC, Güneralp B, Hutyra LR. 2012 Global forecasts of urban expansion to 2030 and direct impacts on biodiversity and carbon pools. *Proc. Natl Acad. Sci. USA* 109, 16 083–16 088. (doi:10.1073/ pnas.1211658109)
- 51. Padullés Cubino JP *et al.* 2019 Drivers of plant species richness and phylogenetic composition in urban yards at the continental scale. *Landscape Ecol.* **34**, 63–77. (doi:10.1007/s10980-018-0744-7)
- Jenerette GD *et al.* 2016 Climate tolerances and trait choices shape continental patterns of urban tree biodiversity. *Global Ecol. Biogeogr.* 25, 1367–1376. (doi:10.1111/geb.12499)
- 53. Pouyat RV, Groffman P, Yesilonis I, Hernandez L. 2002 Soil carbon pools and fluxes in urban ecosystems. *Environ. Pollut.* **116**, 5107–5108. (doi:10.1016/S0269-7491(01)00263-9)
- Hall S, Ahmed B, Ortiz P, Davies R, Sponseller R, Grimm N. 2009 Urbanization alters soil microbial functioning in the Sonoran Desert. *Ecosystems* 12, 654–671. (doi:10.1007/s10021-009-9249-1)
- Zhuo X, Boone CG, Shock EL. 2012 Soil lead distribution and environmental justice in the Phoenix metropolitan region. *Environ. Justice* 5, 206–213. (doi:10.1089/env.2011.0041)
- Cheon J-Y, Ham B-S, Lee J-Y, Park Y, Lee K-K. 2014 Soil temperatures in four metropolitan cities of Korea from 1960 to 2010: implications for climate change and urban heat. *Environ. Earth Sci.* 71, 5215–5230. (doi:10.1007/s12665-013-2924-8)
- Ordóñez C, Duinker P. 2015 Climate change vulnerability assessment of the urban forest in three Canadian cities. *Clim. Change* 131, 531–543. (doi:10.1007/s10584-015-1394-2)
- Meineke EK, Frank SD. 2018 Water availability drives urban tree growth responses to herbivory and warming. *J. Appl. Ecol.* 55, 1701–1713. (doi:10. 1111/1365-2664.13130)
- Meineke E, Youngsteadt E, Dunn RR, Frank SD. 2016 Urban warming reduces aboveground carbon storage. *Proc. R. Soc. B* 283, 20161574. (doi:10. 1098/rspb.2016.1574)
- Savi T, Bertuzzi S, Branca S, Tretiach M, Nardini A.
 2015 Drought-induced xylem cavitation and hydraulic deterioration: risk factors for urban trees under climate change? *New Phytol.* 205, 1106–1116. (doi:10.1111/nph.13112)
- 61. Meineke EK, Dunn RR, Sexton JO, Frank SD. 2013 Urban warming drives insect pest abundance on street trees. *PLoS ONE* **8**, e59687. (doi:10.1371/journal.pone.0059687)
- 62. Tubby KV, Webber JF. 2010 Pests and diseases threatening urban trees under a changing climate. Forestry 83, 451–459. (doi:10.1093/forestry/cpq027)
- McBride JR, Laćan I. 2018 The impact of climatechange induced temperature increases on the suitability of street tree species in California (USA) cities. *Urban For. Urban Green.* 34, 348–356. (doi:10.1016/j.ufug.2018.07.020)
- Brandt L, Derby Lewis A, Fahey R, Scott L, Darling L, Swanston C. 2016 A framework for adapting urban forests to climate change. *Environ. Sci. Policy* 66, 393–402. (doi:10.1016/j.envsci.2016.06.005)

- Kendal D, Dobbs C, Lohr VI. 2014 Global patterns of diversity in the urban forest: is there evidence to support the 10/20/30 rule? *Urban For. Urban Green.* 411–417. (doi:10.1016/j.ufuq.2014.04.004)
- 66. Moore GM. 2014 Wind-thrown trees: storms or management? *Arboric. Urban For.* **40**, 53–69.
- Paerl HW, Paul VJ. 2012 Climate change: links to global expansion of harmful cyanobacteria. Water Res.
 46, 1349–1363. (doi:10.1016/j.watres.2011.08.002)
- 68. Wagner C, Adrian R. 2009 Cyanobacteria dominance: quantifying the effects of climate change. *Limnol. Oceanogr.* **54**, 2460–2468. (doi:10. 4319/lo.2009.54.6_part_2.2460)
- Ladwig R, Furusato E, Kirillin G, Hinkelmann R, Hupfer M. 2018 Climate change demands adaptive management of urban lakes: model-based assessment of management scenarios for Lake Tegel (Berlin, Germany). Water 10, 186. (doi:10.3390/ w10020186)
- Waajen GWAM, Faassen EJ, Lürling M. 2014
 Eutrophic urban ponds suffer from cyanobacterial blooms: Dutch examples. *Environ. Sci. Pollut. Res.*
 21, 9983–9994. (doi:10.1007/s11356-014-2948-y)
- 71. Elmore AJ, Kaushal SS. 2008 Disappearing headwaters: patterns of stream burial due to urbanization. *Front. Ecol. Environ.* **6**, 308–312. (doi:10.1890/070101)
- Kaushal SS, Likens GE, Jaworski NA, Pace ML, Sides AM, Seekell D, Belt KT, Secor DH, Wingate RL. 2010 Rising stream and river temperatures in the United States. Front. Ecol. Environ. 8, 461–466. (doi:10. 1890/090037)
- Nelson KC, Palmer MA, Pizzuto JE, Moglen GE, Angermeier PL, Hilderbrand RH, Dettinger M, Hayhoe K. 2009 Forecasting the combined effects of urbanization and climate change on stream ecosystems: from impacts to management options. J. Appl. Ecol. 46, 154–163. (doi:10.1111/j.1365-2664.2008.01599.x)
- 74. Chester MV, Allenby B. 2018 Toward adaptive infrastructure: flexibility and agility in a non-stationarity age. *Sustain. Resilient Infrastruct.* **4**, 173–191.
- Grimm NB, Schindler S. 2018 Nature of cities and nature in cities: prospects for conservation and design of urban nature in human habitat. In *Rethinking* environmentalism: linking justice, sustainability, and diversity. Strüngmann forum reports (eds S Lele, ES Brondizio, J Byrne, GM Mace, J Martinez-Alier), pp. 99–126. Cambridge, MA: MIT Press.
- ACSE. 2017 2017 report card for America's infrastructure. American Society of Civil Engineers. See https://www.infrastructurereportcard.org/.
- Kim Y, Chester DA, Eisenberg DA, Redman CL. 2019
 The infrastructure trolley problem: positioning safe-to-fail infrastructure for climate change adaptation. *Earth's Future* 7, 704–717 (doi:10.1029/2019EF001208).
- 78. Gilroy KL, McCuen RH. 2012 A nonstationary flood frequency analysis method to adjust for future climate change and urbanization. *J. Hydrol.* **414**, 40–48. (doi:10.1016/j.jhydrol.2011.10.009)
- 79. Royal Society. 2014 *Resilience to extreme weather*. London, UK: The Royal Society Science Policy Centre.

- 80. Larsen L. 2015 Urban climate and adaptation strategies. *Front. Ecol. Environ.* **13**, 486–492. (doi:10.1890/150103)
- 81. Askarizadeh A *et al.* 2015 From rain tanks to catchments: use of low-impact development to address hydrologic symptoms of the urban stream syndrome. *Environ. Sci. Technol.* **49**, 11 264–11 280. (doi:10.1021/acs.est.5b01635)
- Arkema KK, Guannel G, Verutes G, Wood SA, Guerry A, Ruckelshaus M, Kareiva P, Lacayo M, Silver JM. 2013 Coastal habitats shield people and property from sea-level rise and storms. *Nat. Clim. Change* 3, 913–918. (doi:10.1038/nclimate1944)
- 83. Spalding MD *et al.* 2014 Coastal ecosystems: a critical element of risk reduction. *Conserv. Lett.* **7**, 293–301. (doi:10.1111/conl.12074)
- 84. Sutton-Grier AE, Wowk K, Bamford H. 2015 Future of our coasts: the potential for natural and hybrid infrastructure to enhance the resilience of our coastal communities, economies and ecosystems. *Environ. Sci. Policy* **51**, 137–148. (doi:10.1016/j. envsci.2015.04.006)
- 85. Hino M, Field CB, Mach KJ. 2017 Managed retreat as a response to natural hazard risk. *Nat. Clim. Change* **7**, 364–370. (doi:10.1038/nclimate3252)
- Bowler DE, Buyung-Ali L, Knight TM, Pullin AS.
 2010 Urban greening to cool towns and cities: a systematic review of the evidence. *Landscape Urban Plan.* 97, 147–155. (doi:10.1016/j.landurbplan. 2010.05.006)
- 87. Gill SE, Handley JF, Ennos AR, Pauleit S. 2007 Adapting cities for climate change: the role of the green infrastructure. *Built Environ*. **33**, 115–133. (doi:10.2148/benv.33.1.115)
- Baró F, Gómez-Baggethun E. 2017 Assessing the potential of regulating ecosystem services as nature-based solutions in urban areas. In *Nature-based solutions to climate change adaptation in urban areas: linkages between science, policy and practice* (eds N Kabisch, H Korn, J Stadler, A Bonn), pp. 139–158. Berlin, Germany: Springer Open.
- Georgescu M, Morefield PE, Bierwagen BG, Weaver CP. 2014 Urban adaptation can roll back warming of emerging megapolitan regions. *Proc. Natl Acad. Sci. USA* 111, 2909–2914. (doi:10.1073/pnas. 1322280111)
- Stone B, Vargo J, Liu P, Habeeb D, DeLucia A, Trail M, Hu Y, Russell A. 2014 Avoided heat-related mortality through climate adaptation strategies in three US cities. PLoS ONE 9, e100852. (doi:10.1371/ journal.pone.0100852)
- Ziter CD, Pedersen EJ, Kucharik CJ, Turner MG. 2019 Scale-dependent interactions between tree canopy cover and impervious surfaces reduce daytime urban heat during summer. *Proc. Natl Acad. Sci. USA* 116, 7575–7580. (doi:10.1073/pnas.1817561116)
- 92. Demuzere M *et al.* 2014 Mitigating and adapting to climate change: multi-functional and multi-scale assessment of green urban infrastructure. *J. Environ. Manage.* **146**, 107–115. (doi:10.1016/j.jenvman. 2014.07.025)
- Pennino MJ, McDonald RI, Jaffe PR. 2016
 Watershed-scale impacts of stormwater green

- infrastructure on hydrology, nutrient fluxes, and combined sewer overflows in the mid-Atlantic region. *Sci. Total Environ.* **565**, 1044–1053. (doi:10. 1016/j.scitotenv.2016.05.101)
- Zellner M, Massey D, Minor E, Gonzales-Meler M. 2016 Exploring the effects of green infrastructure placement on neighborhood-level flooding via spatially explicit simulations. *Comput. Environ. Urban Syst.* 59, 116–118. (doi:10.1016/j. compenvurbsys.2016.04.008)
- 95. Emilsson T, Sang ÅO. 2017 Impacts of climate change on urban areas and nature-based solutions for adaptation. In *Nature-based solutions to climate change adaptation in urban areas: linkages between science, policy and practice* (eds N Kabisch, H Korn, J Stadler, A Bonn), pp. 15–27. Berlin, Germany: Springer Open.
- McPhearson PT, Feller M, Felson A, Karty R, Lu JW, Palmer MI, Wenskus T. 2011 Assessing the effects of the urban forest restoration effort of MillionTreesNYC on the structure and functioning of New York City ecosystems. Cities Environment (CATE) 3. 1–21.
- McPherson EG, Simpson JR, Xiao Q, Wu C. 2011
 Million trees Los Angeles canopy cover and benefit
 assessment. Landscape Urban Plan. 99, 40–50.
 (doi:10.1016/j.landurbplan.2010.08.011)
- 98. Kuehler E, Hathaway J, Tirpak A. 2017 Quantifying the benefits of urban forest systems as a component of the green infrastructure stormwater treatment network. *Ecohydrology* **10**, e1813. (doi:10.1002/eco. 1813)
- De Sousa MRC, Montalto FA, Spatari S. 2012 Using life cycle assessment to evaluate green and grey combined sewer overflow control strategies. *J. Ind. Ecol.* 16, 901–913. (doi:10.1111/j.1530-9290.2012. 00534.x)
- Walsh CJ et al. 2016 Principles for urban stormwater management to protect stream ecosystems.
 Freshwater Sci. 35, 398–411. (doi:10.1086/685284)
- 101. Flynn C, Davidson CI, Mahoney J. 2014 Transformational changes associated with sustainable stormwater management practices in Onondaga Country, New York. ICSJ 2014. Reston, VA: American Society of Civil Engineers.
- 102. Tao W, Bays JS, Meyer D, Smardon RC, Levy ZF. 2014 Constructed wetlands for treatment of combined sewer overflow in the US: a review of design challenges and application status. Water 6, 3362–3385. (doi:10.3390/w6113362)
- 103. Roach WJ, Grimm NB. 2011 Denitrification mitigates N flux through the stream-floodplain complex of a desert city. *Ecol. Appl.* 21, 2618–2636. (doi:10. 1890/10-1613.1)
- 104. Bettez ND, Groffman PM. 2012 Denitrification potential in stormwater control structures and natural riparian zones in an urban landscape. *Environ. Sci. Technol.* 46, 10 909–10 917. (doi:10. 1021/es301409z)
- 105. Reisinger AJ, Woytowitz E, Majcher E, Rosi EJ, Belt KT, Duncan JM , Kaushal SS, Groffman PM. 2019 Changes in long-term water quality of Baltimore streams are associated with both gray and green

- infrastructure. *Limnol. Oceanogr.* **64**, S60–S76. (doi:10.1002/lno.10947)
- 106. Janke B, Finlay JC, Hobbie SE. 2017 Trees and streets as drivers of urban stormwater nutrient pollution. *Environ. Sci. Technol.* **51**, 9569–9579.
- 107. Selbig WR. 2016 Evaluation of leaf removal as a means to reduce nutrient concentrations and loads in urban stormwater. *Sci. Total Environ.* 571, 124–133. (doi:10.1016/j.scitotenv.2016.07.003)
- 108. Buffam I, Mitchell ME, Durtsche RD. 2016 Environmental drivers of seasonal variation in green roof runoff water quality. *Ecol. Eng.* 91, 506–514. (doi:10.1016/j.ecoleng.2016.02.044)
- 109. Birkmann J, Buckle P, Jaeger J, Pelling M, Setiadi N, Garschagen M, Fernando N, Kropp J. 2010 Extreme events and disasters: a windown of opportunity for change? Analysis of organizational, institutional and political changes, formal and informal responses after mega-disasters. *Nat. Hazards* 55, 637–655. (doi:10.1007/s11069-008-9319-2)
- 110. Solecki W et al. 2019 Extreme events and climate adaptation-mitigation linkages: understanding low-carbon transitions in the era of global urbanization. Wiley Interdiscip. Rev. Climate Change 10, e616. (doi:10.1002/wcc.616)
- 111. Puppim de Oliviera JA. 2013 Learning how to align climate, environmental and development objectives in cities: lessons from the implementation of climate co-benefits initiatives in urban Asia. *J. Clean. Prod.* 58, 7–14. (doi:10.1016/j.jdepro. 2013.08.009)
- 112. Mguni P, Herslund L, Jensen MB. 2016 Sustainable urban drainage systems: examining the potential for green infrastructure-based stormwater management for sub-Saharan cities. *Nat. Hazards* **82**, 241–257. (doi:10.1007/s11069-016-2309-x)
- 113. Kondo M, Fluehr J, McKeon T, Branas C. 2018 Urban green space and its impact on human health. *Int. J. Environ. Res. Public Health* **15**, 445. (doi:10. 3390/ijerph15030445)
- 114. Hartig T, Mitchell R, De Vries S, Frumkin H. 2014 Nature and health. *Annu. Rev. Public Health* **35**, 207–228. (doi:10.1146/annurev-publhealth-032013-182443)
- 115. Bogar S, Beyer KM. 2016 Green space, violence, and crime: a systematic review. *Trauma Violence Abuse* **17**, 160–171. (doi:10.1177/1524838015576412)
- 116. Andersson E, Borgström S, McPhearson T. 2017 Double insurance in dealing with extremes: ecological and social factors for making nature-based solutions last. In Nature-based solutions to climate change adaptation in urban areas: linkages

- between science, policy and practice (eds N Kabisch, H Korn, J Stadler, A Bonn), pp. 51–64. Berlin, Germany: Springer Open.
- 117. Aronson MFJ, Lepczyk CA, Evans KL, Goddard MA, Lerman SB, MacIvor JS, Nilon CH, Vargo T. 2017 Biodiversity in the city: key challenges for urban green space management. Front. Ecol. Environ. 15, 189–196. (doi:10.1002/fee.1480)
- 118. Connop S, Vandergert P, Eisenberg B, Collier MJ, Nash C, Cough J, Newport D. 2016 Renaturing cities using a regionally-focused biodiversity-led multifunctional benefits approach to urban green infrastructure. *Environ. Sci. Policy* 62, 99–111. (doi:10.1016/j.envsci.2016.01.013)
- Garmendia E, Apostolopoulou E, Adams WM, Bormpoudakis D. 2016 Biodiversity and green infrastructure in Europe: boundary object or ecological trap? *Land Use Policy* 56, 315–319. (doi:10.1016/j.landusepol.2016.04.003)
- 120. Green TL, Kronenberg J, Andersson E, Elmqvist T, Gomez-Baggethun E. 2016 Insurance value of green infrastructure in and around cities. *Ecosystems* 19, 1051–1063. (doi:10.1007/s10021-016-9986-x)
- 121. Grabowski Z, Matsler A, Thiel C, McPhillips L, Hum R, Bradshaw A, Miller T, Redman C. 2017 Infrastructures as socio-eco-technical systems: five considerations for interdisciplinary dialogue. J. Infrastruct. Syst. 23, 02517002. (doi:10.1061/(ASCE)IS.1943-555X.0000383)
- 122. Grimm NB, Cook EM, Hale RL, Iwaniec DM. 2016 A broader framing of ecosystem services in cities: benefits and challenges of built, natural, or hybrid system function. In *The Routledge handbook on urbanization and global environmental change* (eds K Seto, WD Solecki, CA Griffith). London, UK: Routledge.
- 123. Depietri Y, McPhearson T. 2017 Integrating the grey, green, and blue in cities: nature-based solutions for climate change adaptation and risk reduction. In Nature-based solutions to climate change adaptation in urban areas: linkages between science, policy and practice (eds N Kabisch, H Korn, J Stadler, A Bonn). Berlin, Germany: Springer Open.
- 124. Sun T, Grimmond CSB, Ni G-H. 2016 How do green roofs mitigate urban thermal stress under heat waves? *J. Geophys. Res. Atmos.* **121**, 5320–5335. (doi:10.1002/2016JD024873)
- 125. The Nature Conservancy. 2013 The case for green infrastructure. Joint-industry white paper. See https://www.nature.org/content/dam/tnc/nature/en/documents/the-case-for-green-infrastructure.pdf.

- 126. William R, Goodwell A, Richardson M, Le PVV, Kumar P, Stillwell AS. 2016 An environmental costbenefit analysis of alternative green roofing strategies. *Ecol. Eng.* 95, 1–9. (doi:10.1016/j. ecoleng.2016.06.091)
- Clark C, Adriaens P, Talbot FB. 2008 Green roof valuation: a probabilistic economic analysis of environmental benefits. *Environ. Sci. Technol.* 42, 2155–2161. (doi:10.1021/es0706652)
- 128. Narayan S *et al.* 2016 The effectiveness, costs and coastal protection benefits of natural and nature-based defences. *PLoS ONE* **11**, e0154735. (doi:10. 1371/journal.pone.0154735)
- 129. Wolch JR, Byrne J, Newell JP. 2014 Urban green space, public health, and environmental justice: the challenge of making cities 'just green enough'. Landscape Urban Plan. 125, 234–244. (doi:10.1016/j.landurbplan.2014.01.017)
- 130. Schwarz K *et al.* 2015 Trees grow on money: urban tree canopy cover and environmental justice. *PLoS ONE* **10**, e0122051. (doi:10.1371/journal.pone. 0122051)
- 131. Baviskar A. 2018 City limits: looking for ecology and equity in the urban context. In *Rethinking* environmentalism: linking justice, sustainability, and diversity (eds S Lele, ES Brondizio, J Byrne, GM Mace, J Martinez-Alier), pp. 85–97. Cambridge, MA: MIT Press.
- 132. Schäffler A, Swilling M. 2013 Valuing green infrastructure in an urban environment under pressure—the Johannesburg case. *Ecol. Econ.* 86, 246–257. (doi:10.1016/j.ecolecon. 2012.05.008)
- 133. Locke DH, Grove JM. 2016 Doing the hard work where it's easiest? Examining the relationships between urban greening programs and social and ecological characteristics. *Appl. Spatial Anal. Policy* **9**, 77–96. (doi:10.1007/s12061-014-9131-1)
- 134. Anguelovski I, Connolly JJ, Garcia-Lamarca M, Cole H, Pearsall H. 2018 New scholarly pathways on green gentrification: what does the urban 'green turn' mean and where is it going? *Prog. Hum. Geogr.* 43, 1064–1086. (doi:10.1177/ 0309132518803799)
- 135. Kabisch N, Haase D. 2014 Green justice or just green? Provision of urban green spaces in Berlin, Germany. *Landscape Urban Plan.* **122**, 129–139. (doi:10.1016/j.landurbplan.2013.11.016)
- 136. United Nations. 2015 United Nations sustainable development goals. See https://www.un.org/ sustainabledevelopment/.