

A climate migrant escapes its parasites

Running head: Parasite escape in fiddler crabs

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ABSTRACT

When a species colonizes a new range, it can escape enemies found in its original range. Examples of enemy escape abound for invasive species, but are rare for climate migrants, which are populations of a species that colonize a new range due to climate-driven range shifts or expansions. The fiddler crab, *Minuca* (= *Uca*) *pugnax*, is found in the intertidal salt marshes of the east coast of the United States. It recently expanded its range north into the Gulf of Maine due to ocean warming. We tested the hypothesis that *M. pugnax* had escaped its parasite enemies. Parasite richness and trematode intensity were lower in populations in the expanded range than those in the historical range, but infection prevalence did not differ. Although *M. pugnax* escaped most of its historical parasites when it migrated northward, it was infected with black-gill lamellae (indicative of *Synophrya hypertrophica*), which was found in the historical range, and by the trematode *Odhneria* cf. *odhneri*, which was not found in the historical range. To our knowledge, this is the first time that *Odhneria* cf. *odhneri* has been reported in fiddler crabs. These results demonstrate that although *M. pugnax* escaped some of its historical parasites when it expanded its range, it appears to have gained a new parasite (*O. cf. odhneri*) in the expanded range. Overall, our results demonstrate that climate migrants can escape their enemies despite colonizing habitats adjacent to their enemy-filled historical range.

Keywords

Parasite escape, enemy-release hypothesis, climate change, global change, range expansions, range shift, climate migrant

1. INTRODUCTION

Species across the planet are shifting or expanding their ranges towards greater elevations, latitudes or depths in response to climate change (Parmesan & Yohe 2003, Sorte et al. 2010, Telwala et al. 2013, Johnson 2015, Pershing et al. 2015, Hale et al. 2017). Populations of a species that colonize a new habitat or range as a result of climate-driven range shifts or expansions are climate migrants (Johnson et al. 2019). The successful recruitment of a climate migrant to the expanded range can be attributed to a change in abiotic factors, such as warming temperatures (Pinsky et al. 2013, Pecl et al. 2017). This is logical from a biogeographical point of view as it follows Shelford (1931)'s law of tolerance and Grinnell (1917)'s niche concept which predict that a species' distribution is set by its tolerance for abiotic factors. While changes in abiotic factors such as temperature are the primary drivers of climate-migrant colonization, not all species expand or shift their range as temperatures rise (Sorte et al. 2010, Hale et al. 2017). This may be due, in part, to stochastic factors such as dispersal ability or deterministic factors such as interactions with other species (Pöyry et al. 2009).

Charles Elton hypothesized that some non-native species, specifically invasive species, fail to colonize a new habitat because they "meet resistance" and are rebuffed by enemies, such as predators, parasites, and competitors (Elton 1958). He called this "ecological resistance," though "biotic resistance" is the term more commonly used today (e.g., Kimbro et al. 2013). However, many non-native species successfully establish outside of their original range. For example, the European green crab, *Carcinus maenas*, is an invasive species that has established in coastal habitats worldwide, in part, because it has met little biotic resistance (Torchin et al. 1996, Carlton & Cohen 2003). The enemy-release hypothesis is commonly invoked to explain the

1 success of non-native species in new ranges (Elton 1958, Keane & Crawley 2002, Colautti et al.
2 2004). It predicts that a non-native species will successfully colonize a new range if two
3 conditions are met: it escapes enemies found in the original range and that escape benefits its
4 fitness (e.g., body size, fecundity) (Wolfe 2002, Keane & Crawley 2002). This hypothesis has
5 been well-tested for invasive species, but rarely for climate migrants (but see Menéndez et al.
6 2008, Hopper et al. 2014). Here, we focus on the first condition of the enemy-release hypothesis
7 and test the hypothesis that a climate migrant was able to escape its parasite enemies.

8
9 Climate migrants, like invasive species, are non-natives in their expanded ranges. Both climate
10 migrants and invasive species involve the movement of individuals from a source (sometimes
11 called ‘donor’) community into a recipient one. Following Hopper et al. (2014), we define
12 climate migrants as having an historical (source) range and an expanded (recipient) range, which
13 is analogous to the terminology of native (source) and introduced (recipient) range used for
14 invasive species (Sorte et al. 2010). Climate migrants differ from invasive species in at least two
15 ways. First, invasive species colonize their introduced ranges as a result of direct human
16 transport, whereas, climate migrants expand or shift their range as a result of indirect effects of
17 human activities (i.e., climate change) (Sorte et al. 2010, Hopper et al. 2014). Second, invasive
18 species often colonize areas geographically separated from their source range, whereas, climate
19 migrants colonize areas adjacent to their source range (Sorte et al. 2010, Hopper et al. 2014).
20 Because the expanded range of a climate migrant is adjacent to its historical range, but typically
21 separated by biogeographic barriers, one might predict little chance for enemy-escape as enemies
22 would presumably track climate as well. However, model predictions (Moorcroft et al. 2006) and

1 empirical evidence (Menéndez et al. 2008, Phillips et al. 2010, Hopper et al. 2014) suggest that
2 some species can escape their enemies as their ranges shift or expand.

3
4 Parasites are fundamental members of ecological communities. Yet, they are poorly studied in
5 terms of their ecology relative to their free-living counterparts (Blakeslee et al. 2012, Johnson &
6 Heard 2017). Nonetheless, host-parasite relationships are excellent for testing the enemy-escape
7 hypothesis because of the intimate symbiotic relationship between host and parasite.

8 Furthermore, aspects of their relationship can be used analyze patterns. For example, climate-
9 migrant hosts may escape parasites with indirect life cycles (i.e., have multiple hosts) if other
10 hosts are missing in the expanded range. Similarly, we can look to invasive species for evidence
11 of parasite escape. For example, in a wide-ranging review that included molluscan, fish,
12 crustacean, avian and mammalian hosts, Torchin et al. (2003) found that most invasive species
13 had fewer parasite species in their introduced ranges than conspecifics in their native ranges,
14 suggesting parasite escape.

15
16
17 Cape Cod, Massachusetts, United States is a well-known biogeographic boundary between the
18 Acadian and Virginian provinces and the northern limit for many warm-water marine species
19 (Briggs 1974). North of Cape Cod is the Gulf of Maine, which is much cooler than the ocean
20 south of Cape Cod (Briggs 1974). The Gulf of Maine is warming rapidly (Pershing et al. 2015),
21 and as a result warm-water species are expanding northward (Johnson 2014, Johnson 2015,
22 McDermott & Kraeuter 2015, Wilson & Pohle 2016). One such species is the Atlantic mud
23 fiddler crab *Minuca* (= *Uca*) *pugnax* (Smith 1870) (Figure 1). *Minuca pugnax* is a small crab (up

to 26 mm carapace width) that lives in salt marshes, which are intertidal grasslands, on the east coast of the United States. It is a burrowing crab with strong sexual dimorphism in which males have a single enlarged claw used in defense and courtship displays and a smaller claw used for foraging and burrowing. *Minuca pugnax* is an excellent species to test the parasite-escape hypothesis for climate migrants because it has a clearly defined historical range and expanded range. Historically, *M. pugnax* ranged from northern Florida to Cape Cod, Massachusetts (Williams 1984). It was first detected in the Gulf of Maine in 2003 (Sanford et al. 2006) and has expanded at least to New Hampshire (Johnson 2014). Based on the evidence for parasite escape in invasive hosts, we predicted that the climate migrant, *M. pugnax*, would have lower parasite diversity, prevalence, and intensity in the expanded range (i.e., the Gulf of Maine) than the historical range.

2. METHODS

2.1 Field collections

From 15-25 August 2017, we collected crabs from ten marshes, spanning almost 12° latitude from Sapelo Island, Georgia to Portsmouth, New Hampshire, United States. (Figure 2, Table 1). We refer to their range from Cape Cod, Massachusetts to Florida as ‘historical’ and their range from Cape Cod, Massachusetts north (i.e., the Gulf of Maine) as ‘expanded.’ At each site, we haphazardly collected the first 50 adult crabs we encountered by hand on the surface or excavated from burrows they escape into. Crabs were collected within 3 m of the marsh edge in stands of smooth cordgrass, *Spartina alterniflora*. After collection, crabs were kept in 70-L plastic bins containing a thin layer of sediment and water from the collection site and transported

to the laboratory for processing. Crabs were dissected within 5 days of collection. All crabs were alive when dissected.

2.2 Laboratory processing

Prior to dissection, crabs were sexed and measured for size (carapace width) to the nearest 0.1 mm with Vernier calipers. For dissection, the carapace was physically removed, and the body placed in clean, filtered seawater for further dissection under a stereo-microscope. The following tissues were examined for the presence of parasites: branchial chamber, gills, hepatopancreas, thoracic nerve ganglia, brain and associated optic ganglia, sternal and apodemal musculature, epidermis of the dorsal carapace, gonads, antennal glands, midgut, hindgut, and foregut. Egg clutches of ovigerous crabs were also examined for parasites. The clutches were removed from the abdomen and examined with a stereo-microscope.

The number and location of the various parasites and any gross conditions were recorded. Parasites were counted *in situ* and removed to separate well plates for further identification. Metacercarial cysts of trematodes were excysted using gentle warming provided by an electric lamp placed ~15 cm overhead. Those that excysted were identified with a compound microscope by morphological comparisons with trematodes in existing literature (e.g., Hunter & Vernberg 1953, Heard 1966, Heard & Overstreet 1983, Schell 1985, Dunn et al.1990). Nematodes were identified based on their morphology and the tissues in which they were encysted as in Wong, Anderson & Bartlett (1989) and Wong & Anderson (1990). Non-metazoan parasites were identified from primary references (e.g., DeTurk 1940, Tuzet & Manier 1962, McCloskey & Caldwell 1965, Bourdon & Bowman 1970, Mattson 1988) and recorded as present or absent.

2.3 Variables

Except for rarefaction analysis of diversity, we quantified parasite infection at the site-level for each of the ten sites. Prevalence of infection (number of infected crabs/total number of crabs) was estimated for each parasite taxa. For statistical analysis, we used total prevalence (the prevalence of infection by any parasite). Mean intensity (mean number of parasites per infected host per site) was calculated for metazoan parasite (i.e., trematodes, nematodes, crustaceans) but not for non-metazoan parasites. We grouped all trematode infections into one variable because trematodes that use crabs as intermediate hosts can be considered as a guild of similar resource users. Parasite richness (number of parasite taxa) was calculated for all taxa present (metazoans plus non-metazoans). We also estimated diversity in each range with rarefaction analyses on all hosts using Coleman curves in EstimateS, 8.0, (Colwell 2006). Coleman rarefaction curves are a powerful technique to examine differences in diversity between different communities. Crabs were entered randomly into the rarefaction analysis and were grouped collectively by range, not site.

2.4 Statistics

To test the hypothesis that crabs in the expanded range had escaped their historical-range parasites we used generalized linear mixed effects models (GLMM) in the lme4 package (Bates et al. 2012) in R (version 3.6.1, “Action of the Toes,” R Core Team 2014). We set range (expanded versus historical) as the fixed factor and site nested within range. We used the following error distributions and their standard link functions for the following dependent variables: total parasite richness (Poisson), total parasite prevalence (binomial), and trematode

intensity (negative binomial). Because trematodes were absent from two sites in the historical range, they were excluded from the intensity analysis. P-values were obtained via stepwise model reduction using likelihood-ratio tests comparing full and reduced models (Crawley 2007).

3. RESULTS

3.1 Overall findings

We dissected 500 fiddler crabs, 50 from each of the 10 sites. We identified 13 different parasite taxa from *M. pugnax* from across the study. For metazoans these included six trematode species (*Odhneria* cf. *odhneri*, *Maritrema* sp., *Gynaecotyla adunca*, *Microphallus* cf. *basodactylophallus*, *Levinseniella* sp., *Maritrema* cf. *heardi*), two nematode species (*Ancyracanthopsis winegardi*, *Skrjabinoclava inornatae*), and a bopyrid isopod (*Leidya distorta*). *Microphallus* cf. *basodactylophallus* and *Levinseniella* sp. had morphologically similar metacercarial cysts that did not all excyst for identification; hence they were grouped in the analyses. Microbial symbionts included an unidentified bacterial infection, blackened gill lamellae indicative of *Synophyra hypertrophica*, the haplosporidian hyperparasite *Urosporidium crescens* infecting individuals of *Microphallus* cf. *basodactylophallus* and *Levinseniella* sp., commensal peritrich ciliates in the gills, and a species of *Enterobryus*, an endocommensal fungi in the order *Eccrinales* infesting the foregut and hindgut (Table 2). *Urosporidium crescens* is a microparasite that infects individual microphallid trematodes with high intensity infections that are difficult to quantify; therefore, we calculated its mean intensity as the mean intensity of infected trematodes in their crab hosts; its overall prevalence is the percentage of crabs with infected metacercariae in the population.

3.2 Parasite diversity

Parasite richness was significantly greater in the historical range (mean richness = 5.4) than in the expanded range (mean richness = 1.2) ($\chi^2 = 14.45$, $df = 1$, $p < 0.001$) (Figure 3A). In the expanded range, only one metazoan parasite was found, *Odhneria* cf. *odhneri*. This parasite was not found in crabs from the historical range (Table 2). Black-gill lamellae characteristic of *Synophrya hypertrophica* were found at low prevalence-levels in the expanded range and the historical range. When symbionts were grouped by geographic region in rarefaction analysis, parasite richness reached an asymptote of 13 in the historical range and 2 in the expanded range (Figure 4). With a sample size of 50 randomly selected crabs, richness was estimated at 9.8 species in the historical range and 1.5 species in the expanded range. The mean and maximum richness in the historical range calculated by rarefaction is higher than those calculated by site because it included all the parasites found in the crabs dissected from within the historical range.

3.3 Parasite prevalence

Despite significant differences in parasite richness, there was no difference in total parasite prevalence between the historical (mean = 0.58) and expanded ranges (mean = 0.49) ($\chi^2 = 0.40$, $df = 1$, $p = 0.53$) (Figure 3B). There were, however, differences in the prevalence patterns for individual symbionts (Appendix Table A1) such as the trematode *Odhneria* cf. *odhneri*, which was only found in crabs from the expanded range. There were also several differences in prevalence levels of individual species of parasites between sites within the historical range. For example, crabs from Stonington, Connecticut, in the historical range had higher prevalence levels of the parasitic isopod *Leidya distorta* compared to the other sites, as well as few other helminths. Most trematode infections occurred in crabs from the southern and central portion of

the historical range, and this was in sharp contrast to the other historical sites that had low levels of these parasites, with the notable exception of the aforementioned *Odhneria cf. odhneri* infections in the expanded range (Appendix Table A1).

3.4 Trematode intensity

The mean intensity of trematode infections was more than 4x higher in the historical range (mean = 14) than in the expanded range (mean = 3) (Figure 3C). This difference was significant ($\chi^2 = 4.63$, $df = 1$, $p = 0.03$).

4. DISCUSSION

Our study demonstrates that the climate migrant, *Minuca pugnax*, escaped all but one of its historical parasites when it expanded its range. We know of only two other documented examples of parasite escape in climate migrants: Kellet's whelk, *Kelletia kelletii*, on the West Coast of the United States (Hopper et al. 2014) and the brown Argus butterfly, *Aricia agestis*, in Britain (Menéndez et al. 2008). Taken together, climate migrants appear to follow the same parasite-escape pattern found in invasive species: they have fewer parasites in their expanded ranges versus their historical ones (Torchin et al. 2003, Blakeslee et al. 2012, Blakeslee et al. 2013). Thus, climate migrants can escape their enemies despite colonizing habitats that are adjacent to their enemy-filled historical range.

4.1 Mechanisms of parasite escape

At least four mechanisms can be invoked to explain why a climate migrant may escape its parasites. (1) A climate migrant may arrive in the expanded range while in a life stage that is not

1 infected. In marine systems, non-native species often arrive as planktonic larvae and are likely
2 free of parasites, which typically infect juvenile and adult stages or kill their larval host before
3 metamorphosis (Shields 2012, Shields et al. 2015). This mechanism may apply to *Minuca*
4 *pugnax* because it disperses as planktonic larvae and has no known parasites that infect its larval
5 and post-larval stages. (2) A climate migrant may expand ahead of intermediate or definitive
6 hosts required to complete the multi-host, indirect life cycles of some parasites (Hopper et al.
7 2014). It is unlikely that this mechanism explains the parasite escape we see for *M. pugnax* in the
8 expanded range because many of intermediate and definitive hosts used by parasites that infect
9 *M. pugnax* are found in both ranges. For instance, migratory birds, such as Virginia rails (*Rallus*
10 *limicola*) and saltmarsh sparrows (*Ammodramus caudacutus*), are definitive hosts and move
11 between the expanded (Gulf of Maine) and historical ranges seasonally. Many parasites of *M.*
12 *pugnax* are trematodes, which use specific gastropods as first intermediate hosts. Several
13 gastropod hosts including the coffee-bean snail, *Melampus bidentatus*, the Atlantic mudsnail,
14 *Tritia* (= *Ilyanassa*) *obsoleta*, and hydrobiid snails are found in marshes in the expanded and
15 historical ranges (Gosner 1978, Johnson et al. 2009, Johnson 2011, Johnson & Short 2013,
16 Johnson & Heard 2017). (3) Phenological mismatch between hosts and parasites may limit *M.*
17 *pugnax* infection in the expanded range. Crabs and their parasites in the expanded range are at
18 higher latitudes and thus have a shorter growing season. Migratory birds, as definitive hosts, may
19 arrive too late in northern latitudes for certain parasites to develop and successfully overwinter in
20 their snail or fiddler-crab intermediate hosts (Paull & Johnson 2014). For instance, trematodes
21 have well-known seasonal cycles that match the migration patterns of their hosts, which are
22 being disrupted by climate change, leading to phenological mismatch (Galaktionov et al. 2006).
23 (4) *Minuca pugnax* may have escaped parasites due to its small population size in the Gulf of

1 Maine, which can limit density-dependent transmission (Blakeslee et al. 2012). Crab densities at
2 the leading edge of the expanded range (i.e., northern Massachusetts) are $<6\text{ m}^{-2}$ on average (K.
3 Martinez-Soto and D.S. Johnson unpublished data), which is much lower than those found in the
4 historical range ($60\text{-}120\text{ m}^{-2}$, Culbertson et al. 2007, Luk & Zajac 2013).

5 6 4.2 Acquisition of a novel parasite?

7 Although *M. pugnax* in the expanded range escaped all metazoan parasites found in the historical
8 range, they may have acquired a new one, the trematode *Odhneria* cf. *odhneri*. To our
9 knowledge, this is the first time *O. cf. odhneri* has been reported in *M. pugnax*. This parasite was
10 not found in crabs in the historical range. Moreover, although it was initially reported in
11 *Palaemonetes vulgaris* from the Woods Hole area (Stunkard, 1979), it has not been observed in a
12 large collection of this host from Georgia (Pung et al. 2002), nor has it been reported in
13 crustaceans from locations south of Cape Cod, Massachusetts. Notably, *O. cf. odhneri* was not
14 found in our collections in the northernmost portion of the historical range, which is less than 8
15 km away from the area where Stunkard (1979) found it in grass shrimp. Thus, although *O. cf.*
16 *odhneri* infects *M. pugnax* in the expanded range, it does not appear to infect *M. pugnax* in the
17 historical range even though they are found in the same locations.

18 Invasive species can acquire new parasites in their invaded ranges (Torchin et al. 2003,
19 Kelly et al. 2009). For example, the invasive shrimp, *Palaemon macrodactylus*, was infected by
20 a species of *Odhneria* and a balanid barnacle when it invaded Argentina (Martorelli et al. 2012).
21 In another example, Torchin et al. (2002) found that the invasive European green crab, *Carcinus*
22 *maenas*, gained novel parasites in its invaded ranges worldwide. Because *O. cf. odhneria* appears
23 to be a novel parasite for *M. pugnax*, then, as with invasive species, the crab host was able to

1 escape its historical parasites in terms of parasite diversity, but in the expanded range it did not
2 escape its vulnerability to host generalists (Torchin et al. 2002, Selechnik et al. 2017).

4 4.3 Parasite escape or parasite reduction?

5 Although *M. pugnax* had an average of one parasite per site in the expanded range and an
6 average of five in the historical range, total parasite prevalence was similar between the ranges.
7 These results indicate that *M. pugnax* is equally susceptible to parasite infections in their
8 historical and expanded ranges. Parasite intensity (number of parasites per infected host) can
9 strongly determine a parasite's impact on host fitness (Lafferty & Morris 1996, Shields & Wood
10 1993, Colautti et al. 2004), though sometimes only one parasite is required to influence host
11 fitness (Lafferty & Kuris 2009). For instance, the trematode *Levinsiniella byrdi* turns its second
12 intermediate host, the amphipod *Orchestia grillus*, orange and negates the photatic response with
13 an average infection intensity of 1 metaceariae per host, presumably making it more susceptible
14 to predators (Johnson & Heard 2017). We found that trematode intensity was significantly lower
15 in crabs in the expanded than in the historical range, further evidence of parasite escape in *M.*
16 *pugnax*. The lower parasite richness and intensity may benefit *M. pugnax* fitness in the expanded
17 range. In a previous study, we found that male *M. pugnax* are larger and females more fecund in
18 the expanded range relative to the historical range (Johnson et al. 2019). Similarly, the invasive
19 green crab, *Carcinus maenas* is larger in its invaded range than in its native Europe, which
20 Torchin et al. (2001) attribute to parasite escape. These results support the second prediction of
21 the enemy-release hypothesis: a release from parasite enemies will benefit the host's fitness
22 (Keane and Crawley 2002). More work is required to explicitly test this prediction.

4.4 The honeymoon phase

A species that escapes its enemies today may not escape them tomorrow. This may be particularly true for climate migrants whose expanded range is adjacent to an enemy-filled historical range. The honeymoon hypothesis predicts that when a climate migrant or invasive species escapes its enemies those enemies will eventually migrate into the expanded/invaded range from nearby sources (Phillips et al. 2010) or native parasites will colonize the novel host and accumulate over time (Kołodziej-Sobocińska et al. 2018). Again, we can look to invasive species for precedence. For example, in Australia, the invasive cane toad, *Rhinella marina* (formerly, *Bufo marinus*), is expanding westward. Toads at the leading edge of the expansion escape the lungworm parasite, *Rhabdias pseudosphaerocephala*. After 1-3 years, however, lungworms also expand their range and catch-up to infect previously uninfected toads (Phillips et al. 2010). In a marine example, Blakeslee et al. (2012) found that the periwinkle, *Littorina saxatilis*, escaped more parasites than the mud snail, *Tritia obsoleta*, when introduced into San Francisco Bay, United States. They suggest that the differences in parasite escape are driven, in part, to the fact the *L. saxatilis* was introduced more recently than *T. obsoleta*. *Minuca pugnax* was first observed in its expanded range in 2003 (Sanford et al. 2006). Because we found no historical-range metazoan parasites in expanded-range crabs, our results indicate that *M. pugnax* is currently enjoying at least a 14-year honeymoon from their historical metazoan parasites.

4.5 Conclusions

We show that in terms of parasite richness and intensity, *Minuca pugnax* escaped parasites from its historical range as it migrated into its expanded range. If this result is typical of other climate migrants, then it suggests that they follow the same parasite-escape pattern seen in invasive

1 species: they have fewer parasites in their expanded ranges versus their historical ones (Torchin
2 et al. 2003, Blakeslee et al. 2012, Blakeslee et al. 2013). Thus, our results suggest that climate
3 migrants can escape their parasites despite colonizing habitats that are adjacent to their enemy-
4 filled historical range. However, that escape may only be temporary as parasites catch-up (i.e.,
5 honeymoon phase, Phillips et al. 2010, Kołodziej-Sobocińska et al. 2018) or as host-generalist
6 parasites start using expanding populations of a new host (Kelly et al. 2009). Despite the
7 ubiquity of parasites and the growing number of documented climate-migrant hosts (Sorte et al.
8 2010, Pecl et al. 2017), our study is one of only three that we are aware of to investigate the
9 parasites of climate migrants (Menéndez et al. 2008, Hopper et al. 2014). We advocate exploring
10 host-parasite relationships in the context of climate change to answer fundamental ecological
11 questions.

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LITERATURE CITED

- Bates D, Maechler M, Bolker, B (2012) lme4 version 1.1-5: linear mixed-effects models using Eigen and S4.
- Blakeslee AMH, Altman I, Miller AW, Byers JE, Hamer CE, Ruiz GM (2012) Parasites and invasions: a biogeographic examination of parasites and hosts in native and introduced ranges: Invasion history influences parasite biogeographic patterns. *Journal of Biogeography* 39:609–622.
- Blakeslee AMH, Fowler AE, Keogh CL (2013) Marine invasions and parasite escape: updates and new perspectives: In *Advances in Marine Biology*. Academic Press, p 87–169
- Bourdon R, Bowman TE (1970) Western Atlantic species of the parasitic genus *Leidyia* (Epicaridea: Bopyridae). *Proceedings of the Biological Society of Washington*, 83:409-424.
- Briggs, JC (1974) *Marine zoogeography*. McGraw-Hill, New York.
- Carlton JT, Cohen AN (2003) Episodic global dispersal in shallow water marine organisms: the case history of the European shore crabs *Carcinus maenas* and *C. aestuarii*. *Journal of Biogeography* 30:1809–1820.
- Colautti RI, Ricciardi A, Grigorovich IA, MacIsaac HJ (2004) Is invasion success explained by the enemy release hypothesis? *Ecology Letters* 7:721–733.
- Colwell RK (2006) Estimates, Version 8.0: statistical estimation of species richness and shared species from samples (Software and User's Guide). Freeware for Windows and Mac OS.
- Crawley MJ (2007) *The R Book*. John Wiley & Sons Ltd., Chichester, England.
- Culbertson JB, Valiela I, Peacock EE, Reddy CM, Carter A, VanderKruik R (2007) Long-term biological effects of petroleum residues on fiddler crabs in salt marshes. *Marine Pollution Bulletin* 54:955–962.
- Dery DW (1958) A description of *Maritreminoides raminellae*, n. sp. (Trematoda: Microphallidae). *Proceedings of the Helminthological Society of Washington* 25: 40-43.
- DeTurk WE (1940) The occurrence and development of a hyper-parasite, *Urosporidium crescens* sp. nov. (Sporozoa, Haplosporidia), which infests the metacercariae of *Spelotrema nicolli*, parasitic in *Callinectes sapidus*. *Journal of the Elisha Mitchell Scientific Society* 56:231-232.
- Dunn TS, Ownbey TC, Vannarath T (1990) In vitro excystment of the metacercariae of *Gynaecotyla adunca* and *Probolocoryphe uca* (Microphallidae) from the fiddler crab, *Uca pugilator*. *Canadian Journal of Zoology* 68:2376–2384.
- Elton CS (1958) *The ecology of invasions by animals and plants*—Methuen & Co. Ltd., London, 181.
- Galaktionov KV, Irwin SWB, Prokofiev VV, Saville DH, Nikolaev KE, Levakin IA (2006) Trematode transmission in coastal communities - temperature dependence and climate change perspectives. In *International Congress of Parasitology XI. MEDIMOND sRL*, p 85-90.
- Gosner K (1978) *A Field Guide to the Atlantic Seashore*. Houghton Mifflin Company, New York, New York.
- Grinnell J (1917) The Niche-Relationships of the California thrasher. *The Auk* 34:427–433.
- Hale SS, Buffum HW, Kiddon JA, Hughes MM (2017) Subtidal benthic invertebrates shifting northward along the US Atlantic Coast. *Estuaries and Coasts* 40:1744–1756.

- 1 Heard, R.W. (1970). Parasites of the Clapper Rail, *Rallus longirostris* Boddaert. II. Some
2 trematodes and cestodes from *Spartina* marshes of the Eastern United States. Proceedings
3 of the Helminthological Society of Washington 37:147-153.
- 4 Heard RW, Overstreet RM (1983) Taxonomy and Life Histories of Two North American Species
5 of “*Carneophallus*” (= *Microphallus*) (Digenea: Microphallidae). Proceedings of the
6 Helminthological Society of Washington 50: 170-174.
- 7 Hopper JV, Kuris AM, Lorda J, Simmonds SE, White C, Hechinger RF (2014) Reduced parasite
8 diversity and abundance in a marine whelk in its expanded geographical range. Journal of
9 Biogeography 41:1674–1684.
- 10 Hunter WS, Vernberg WB (1953) Early stages in the life cycle of the trematode, *Gynaecotyla*
11 *adunca* (Linton, 1905). Transactions of the American Microscopical Society 72:163-170.
- 12 Johnson DS, Fleeger J, Deegan L (2009) Large-scale manipulations reveal that top-down and
13 bottom-up controls interact to alter habitat utilization by saltmarsh fauna. Marine Ecology
14 Progress Series 377:33–41.
- 15 Johnson DS (2011) High-marsh invertebrates are susceptible to eutrophication. Marine Ecology
16 Progress Series 438:143–152.
- 17 Johnson DS, Short MI (2013) Chronic nutrient enrichment increases the density and biomass of
18 the mudsnail, *Nassarius obsoletus*. Estuaries and coasts 36:28–35.
- 19 Johnson DS (2014) Fiddler on the roof: a northern range extension for the marsh fiddler crab
20 *Uca pugnax*. Journal of Crustacean Biology 34:671–673.
- 21 Johnson DS (2015) The savory swimmer swims north: a northern range extension of the blue
22 crab *Callinectes sapidus*? Journal of Crustacean Biology 35:105–110.
- 23 Johnson DS, Heard RW (2017) Bottom-up control of parasites. Ecosphere 8.
- 24 Johnson MW (1957) The copepod *Choniosphaera cancrorum* parasitizing a new host, the green
25 crab *Carcinides maenas*. The Journal of Parasitology 43:470-473.
- 26 Keane RM, Crawley MJ (2002) Exotic plant invasions and the enemy release hypothesis. Trends
27 in Ecology & Evolution 17:164–170.
- 28 Kelly DW, Paterson RA, Townsend CR, Poulin R, Tompkins DM (2009) Parasite spillback: a
29 neglected concept in invasion ecology? Ecology 90:2047–2056.
- 30 Kimbro DL, Cheng BS, Grosholz ED (2013) Biotic resistance in marine environments. Ecology
31 Letters 16:821–833.
- 32 Kołodziej-Sobocińska M, Brzeziński M, Niemczynowicz A, Zalewski A (2018) High parasite
33 infection level in non-native invasive species: it is just a matter of time. Ecography
34 41:1283–1294.
- 35 Kuris AM, Gurney R (1997) Survey of Tasmanian crabs for parasites: a progress report.
36 Proceedings of the first international workshop on the demography, impacts and
37 management of the introduced populations of the European crab, *Carcinus maenas*.
38 Centre for Research on Introduced Marine Pests 11: 92-94. Technical Report.
- 39 Lafferty KD, Kuris AM (2009) Parasitic castration: the evolution and ecology of body snatchers.
40 Trends in Parasitology 25:564–572.
- 41 Lafferty KD, Morris AK (1996) Altered behavior of parasitized killifish increases susceptibility
42 to predation by bird final hosts. Ecology 77:1390–1397.

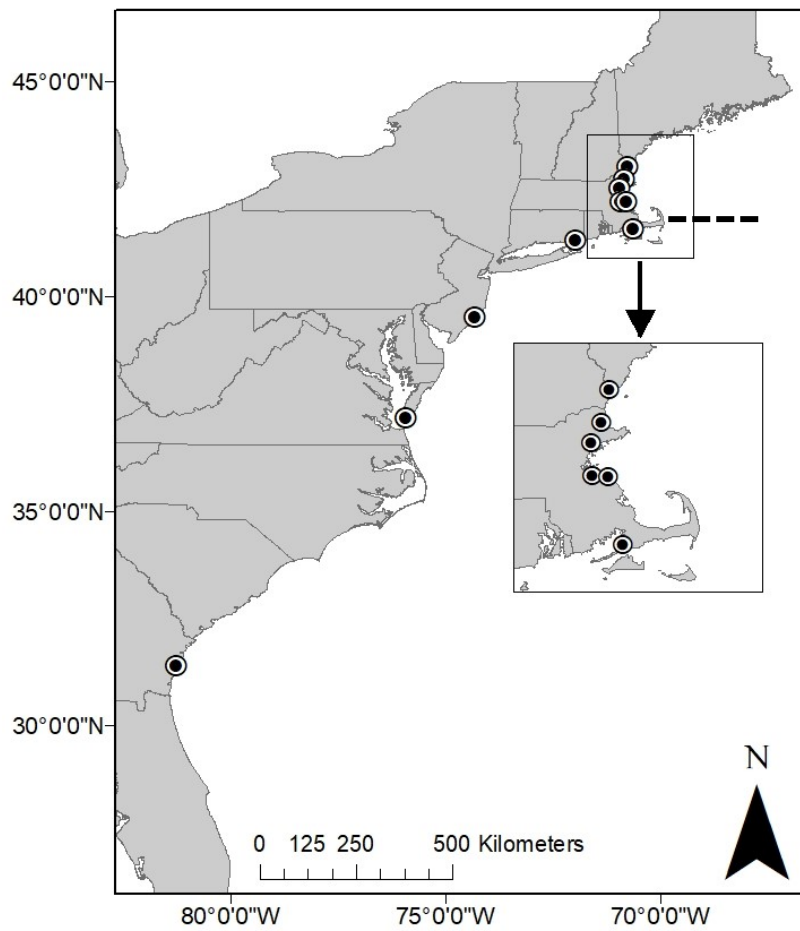
- 1 Luk YC, Zajac RN (2013) Spatial ecology of fiddler crabs, *Uca pugnax*, in southern New
2 England salt marsh landscapes: potential habitat expansion in relation to salt marsh
3 change. *Northeastern Naturalist* 20:255–274.
- 4 Martorelli SR, Alda P, Marcotegui P, Montes M, La Sala L (2012) New locations and
5 parasitological findings for the invasive shrimp *Palaemon macrodactylus* in temperate
6 southwestern Atlantic coastal waters. *Aquatic Biology* 15:153–157.
- 7 Mattson RA (1988) Occurrence and abundance of eccrinaceous fungi (Trichomycetes) in
8 brachyuran crabs from Tampa Bay, Florida. *Journal of Crustacean Biology* 8:20–30.
- 9 McCloskey LR, Caldwell SP (1965) *Enteromyces callianassae* Lichtwardt (Trichomycetes,
10 Eccrinales) in the mud shrimp *Upogebia affinis* (Say). *Journal of the Elisha Mitchell*
11 *Scientific Society* 114:117.
- 12 McDermott JJ, Kraeuter JN (2015) Occurrence of first crab instars of the Atlantic ghost crab
13 *Ocypode quadrata* (Decapoda: Brachyura: Ocypodidae) along the coast of Maine, U.S.A.
14 *Proceedings of the Biological Society of Washington* 128:98–102.
- 15 Menéndez R, González-Megías A, Lewis OT, Shaw MR, Thomas CD (2008) Escape from
16 natural enemies during climate-driven range expansion: a case study. *Ecological*
17 *Entomology* 33:413–421.
- 18 Moorcroft PR, Pacala SW, Lewis MA (2006) Potential role of natural enemies during tree range
19 expansions following climate change. *Journal of Theoretical Biology* 241:601–616.
- 20 Parmesan C, Yohe G (2003) A globally coherent fingerprint of climate change impacts across
21 natural systems. *Nature* 421:37–42.
- 22 Paull SH, Johnson PTJ (2014) Experimental warming drives a seasonal shift in the timing of
23 host-parasite dynamics with consequences for disease risk. *Ecology Letters* 17:445–453.
- 24 Pecl GT, Araújo MB, Bell JD, Blanchard J, Bonebrake TC, Chen I-C, Clark TD, Colwell RK,
25 Danielsen F, Evengård B, Falconi L, Ferrier S, Frusher S, Garcia RA, Griffis RB,
26 Hobday AJ, Janion-Scheepers C, Jarzyna MA, Jennings S, Lenoir J, Linnetved HI,
27 Martin VY, McCormack PC, McDonald J, Mitchell NJ, Mustonen T, Pandolfi JM,
28 Pettorelli N, Popova E, Robinson SA, Scheffers BR, Shaw JD, Sorte CJB, Strugnell JM,
29 Sunday JM, Tuanmu M-N, Vergés A, Villanueva C, Wernberg T, Wapstra E, Williams
30 SE (2017) Biodiversity redistribution under climate change: Impacts on ecosystems and
31 human well-being. *Science* 355:eaai9214.
- 32 Pershing AJ, Alexander MA, Hernandez CM, Kerr LA, Le Bris A, Mills KE, Nye JA, Record
33 NR, Scannell HA, Scott JD, Sherwood GD, Thomas AC (2015) Slow adaptation in the
34 face of rapid warming leads to collapse of the Gulf of Maine cod fishery. *Science*
35 350:809–812.
- 36 Phillips BL, Kelehear C, Pizzatto L, Brown GP, Barton D, Shine R (2010) Parasites and
37 pathogens lag behind their host during periods of host range advance. *Ecology* 91:872–
38 881.
- 39 Pinsky ML, Worm B, Fogarty MJ, Sarmiento JL, Levin SA (2013) Marine taxa track local
40 climate velocities. *Science* 341:1239–1242.
- 41 Pöyry J, Luoto M, Heikkinen RK, Kuussaari M, Saarinen K (2009) Species traits explain recent
42 range shifts of Finnish butterflies. *Global Change Biology* 15:732–743.

- Pung OJ, Khan RN, Vives SP, Walker CB (2002) Prevalence, geographic distribution, and fitness effects of *Microphallus turgidus* (Trematoda: Microphallidae) in grass shrimp (*Palaemonetes* spp.) from coastal Georgia. *Journal of Parasitology* 88:89–92.
- R Core Team (2019) R: A language and environment for statistical computing. R Foundation for Statistical Computing, Vienna, Austria. URL <https://www.R-project.org/>.
- Sanford E, Holzman SB, Haney RA, Rand DM, Bertness MD (2006) Larval tolerance, gene flow, and the northern geographic range limit of fiddler crabs. *Ecology* 87:2882–2894.
- Schell SC (1985) Handbook of trematodes of North America north of Mexico. University Press of Idaho.
- Selechnik D, Rollins LA, Brown GP, Kelehear C, Shine R (2017) The things they carried: the pathogenic effects of old and new parasites following the intercontinental invasion of the Australian cane toad (*Rhinella marina*). *International Journal for Parasitology: Parasites and Wildlife* 6:375–385.
- Shelford VE (1931) Some concepts of bioecology. *Ecology* 12:455–467.
- Shields JD, Wood FE (1993) Impact of parasites on the reproduction and fecundity of the blue sand crab *Portunus pelagicus* from Moreton Bay, Australia. *Marine Ecology Progress Series* 92:159–170.
- Shields JD (2012) The impact of pathogens on exploited populations of decapod crustaceans. *Journal of Invertebrate Pathology* 110:211–224.
- Shields JD, Williams JD, Boyko CB (2015) Parasites and diseases of Brachyura. In: The Crustacea. Treatise on Zoology-Anatomy, Taxonomy, Biology, Castro P, Davie PJF, Guinot D, Schram F, von Vaupel Klein J (eds) Brill, Leiden, p 639–774
- Sinclair NR (1971) A review of *Odhneria odhneri* Travassos, 1921 (Trematoda: Microphallidae). *The Journal of Parasitology* 57:980.
- Smith, S. I. 1870. Notes on American Crustacea. No. I. Ocypodoidea. *Transactions of the Connecticut Academy of Arts and Sciences* 2: 113-176.
- Sorte CJB, Williams SL, Carlton JT (2010) Marine range shifts and species introductions: comparative spread rates and community impacts: Range shifts and non-native species introductions. *Global Ecology and Biogeography* 19:303–316.
- Stunkard HW (1979) The morphology, life-history, and taxonomic relations of *Odhneria odhneri* Travassos, 1921 (Digenea: Microphallidae). *The Biological Bulletin* 156:234–245.
- Telwala Y, Brook BW, Manish K, Pandit MK (2013) Climate-induced elevational range shifts and increase in plant species richness in a Himalayan biodiversity epicentre. *PLoS ONE* 8:e57103.
- Torchin ME, Lafferty KD, Dobson AP, McKenzie VJ, Kuris AM (2003) Introduced species and their missing parasites. *Nature* 421:628–630.
- Torchin ME, Lafferty KD, Kuris AM (1996) Infestation of an introduced host, the European green crab, *Carcinus maenas*, by a symbiotic nemertean egg predator, *Carcinonemertes epialti*. *The Journal of Parasitology* 82:449.
- Torchin ME, Lafferty KD, Kuris AM (2002) Parasites and marine invasions. *Parasitology* 124:137–151.
- Tuzet O, Manier JF (1962) *Enteromyces callianassae* Licht-wardt Trichomycète Eccrinale commensal de l'estomac de *Uca pugilator* Latreille. *Ann Sci Nat Bot, Paris Série* 12:615-617.

- 1 Wang Q, An S, Ma Z, Zhao B, Chen J, Li B (2006) Invasive *Spartina alterniflora*: biology,
2 ecology and management. *Acta Phytotaxonomica Sinica* 44:559-588.
- 3 Williams AB (1984) Shrimps, lobsters, and crabs of the Atlantic coast of the eastern United
4 States, Maine to Florida. Smithsonian Institution Press, Washington, D.C., USA.
- 5 Wilson BM Pohle, GW (2016) Northern range expansion of the American talon crab,
6 *Euchirograpsus americanus* A. Milne-Edwards, 1880 (Decapoda, Grapsoidea,
7 Plagusiidae), to the Bay of Fundy, Canada. *Crustaceana* 89:163–173.
- 8 Wolfe LM (2002) Why alien invaders succeed: support for the escape-from-enemy hypothesis.
9 *The American Naturalist* 160:705-711.
- 10 Wong PL, Anderson RC, Bartlett CM (1989) Development of *Skrjabinoclava inornatae*
11 (Nematoda: Acuarioidea) in fiddler crabs (*Uca* spp.) (Crustacea) and western willets
12 (*Catoptrophorus semipalmatus inornatus*) (Aves: Scolopacidae). *Canadian Journal of*
13 *Zoology* 67:2893–2901.
- 14 Wong PL, Anderson RC (1990) *Ancyracanthopsis winegardi* n.sp. (Nematoda: Acuarioidea)
15 from *Pluvialis squatarola* (Aves: Charadriidae) and *Ancyracanthus heardi* n.sp. from
16 *Rallus longirostris* (Aves: Rallidae), and a review of the genus. *Canadian Journal of*
17 *Zoology* 68:1297–1306.



Figure 1: A male fiddler crab, *Minuca pugnax*, in Rowley, Massachusetts, which is in the expanded range. Note the blue face of the crab, a characteristic field marking of this species. Photo credit: DS Johnson.



1
2 Figure 2: Location of marshes sampled (circles). Dashed line indicates northern limit of
3 historical range.

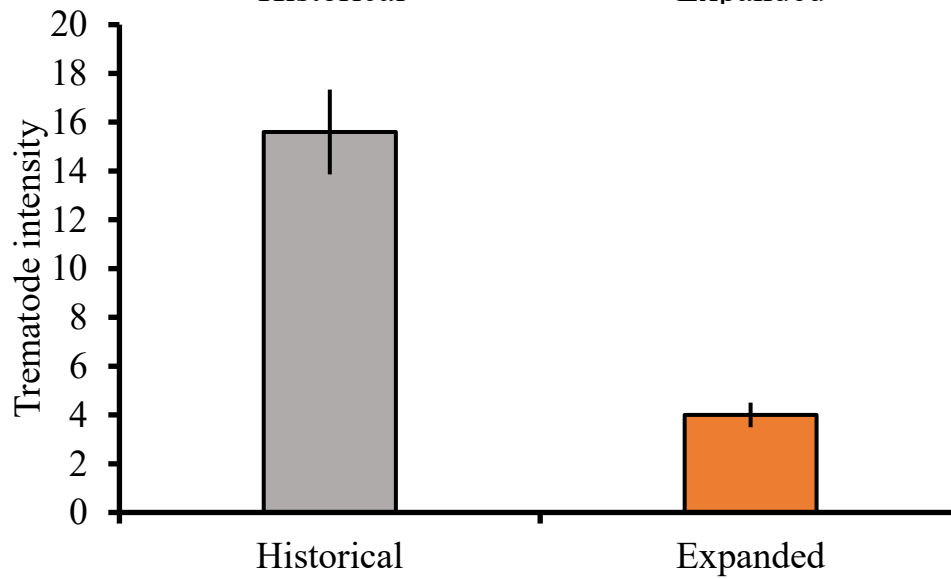
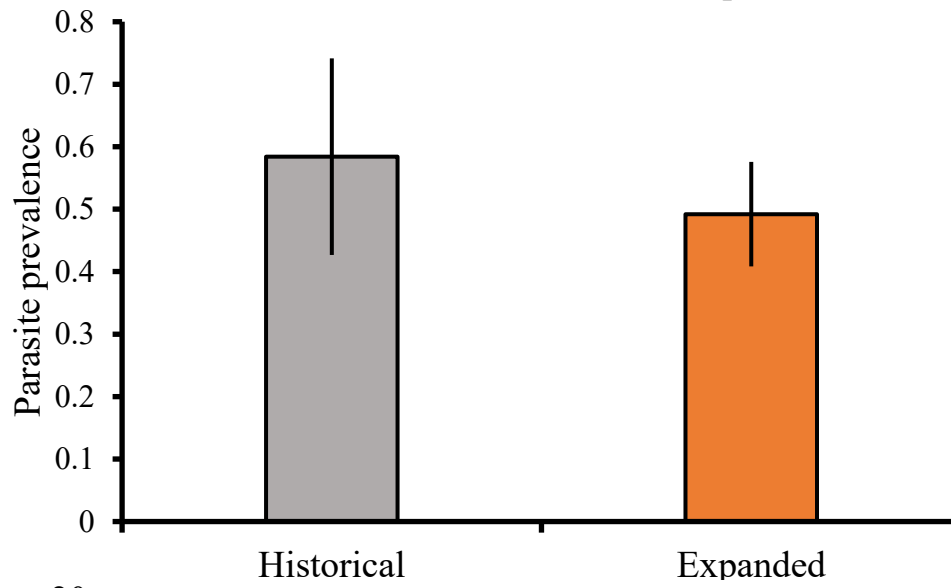
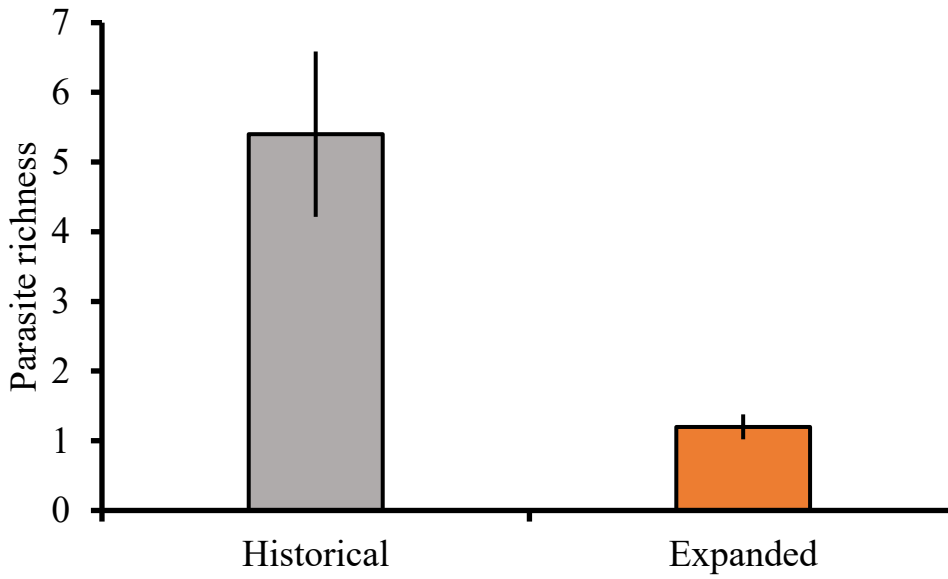


Figure 3: Mean (± 1 SE) A) taxonomic richness of parasites, B) total parasite prevalence (proportion of the population at each site infected with a parasite, regardless of species) and C) trematode intensity (number of parasites per infected crab) found in *Minuca pugnax* in its ranges.

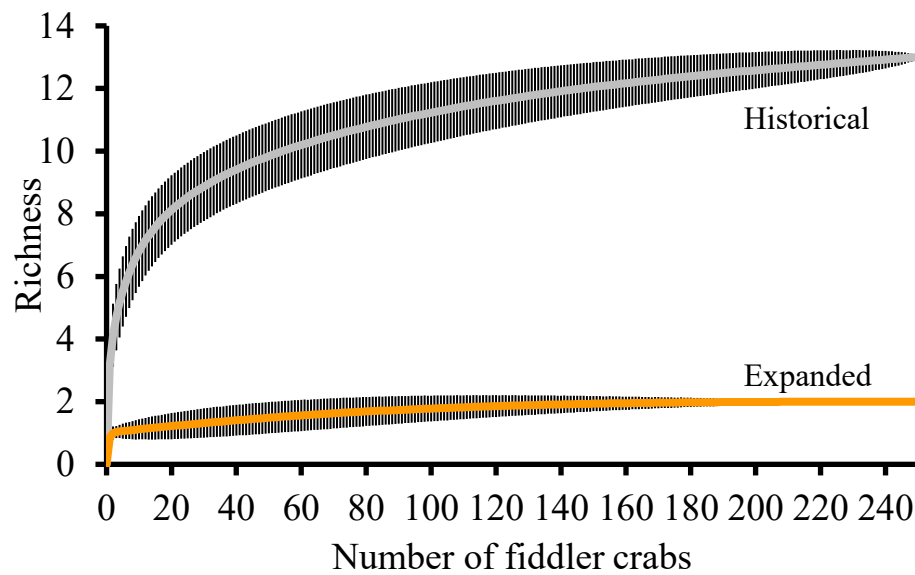


Figure 4: Parasite richness as a function of sample size for fiddler crabs sampled in their historical vs. expanded ranges (Coleman rarefaction curves). Solid line indicates richness estimate, whereas bars are standard deviations.

10 Table 1: Location of marshes sampled.

Range	Marsh or Estuary	Location	Latitude (N)	Longitude (W)
Expanded	Sagamore Creek	Portsmouth, New Hampshire	43° 3' 3.58"	70° 46' 13.08"
Expanded	Plum Island Estuary	Rowley, Massachusetts	42° 44' 19.63"	70° 50' 23.23"
Expanded	Waters River	Danvers, Massachusetts	42° 32' 47.84"	70° 56' 19.95"
Expanded	Bare Cove	Weymouth, Massachusetts	42° 13' 47.27"	70° 55' 37.26"
Expanded	Musquashcut Brook	Scituate, Massachusetts	42° 13' 34.25"	70° 46' 35.92"
Historical	Great Sippewissett	Falmouth, Massachusetts	41° 35' 0.55"	70° 38' 14.22"
Historical	Mason's Island	Stonington, Connecticut	41° 19' 58.37"	71° 58' 16.98"
Historical	Great Bay Estuary	Little Egg Harbor Township, New Jersey	39° 31' 18.28"	74° 19' 6.32"
Historical	Cushman's Landing	Cape Charles, Virginia	37° 10' 30.94"	75° 56' 32.28"
Historical	Sapelo Island	McIntosh County, Georgia	31° 25' 20.74"	81° 17' 24.25"

32 Table 2. Prevalence and mean intensity of the parasites and symbionts in *Minuca pugnax* from its native and expanded range.

Parasite	Historical range		Expanded range	
	Prevalence (%)	Mean intensity $\pm$ sd	Prevalence (%)	Mean intensity $\pm$ sd
Microbial parasites and symbionts				
Bacterial infection	0.4	nd	0.0	
<i>Eccrinales</i> (Fungi)	4.0	nd	0.0	
Peritrich ciliates ^a	$\pm$	nd	$\pm$	nd
Black gill lamellae cf. <i>Synophrya hypertrophica</i>	0.8	nd	0.8	nd
<i>Urosporidium crescens</i>	8.8	2.82 ± 3.07^b	0.0	
Trematoda				
<i>Odhneria</i> cf. <i>odhneri</i>	0.0	0.0	48.4	4.00 ± 5.54
<i>Maritrema</i> sp.	21.6	5.17 ± 5.02	0.0	
<i>Gynaecotyla adunca</i>	18.4	22.24 ± 19.71	0.0	
<i>Microphallus</i> cf. <i>basodactylophallus</i> & <i>Levinesenilla</i> sp.	26.0	7.69 ± 8.89	0.0	
<i>Maritrema</i> cf. <i>heardi</i>	2.4	3.17 ± 2.99	0.0	
Nematoda				
<i>Ancyracanthopsis winegardi</i>	16.4	1.32 ± 0.72	0.0	
<i>Skrjabinoclava inornatae</i>	1.6	1.00 ± 0.00	0.0	
Isopoda				
<i>Leidya distorta</i> ^c	7.6	1.37 ± 0.50	0.0	

33 ^aThe peritrich ciliates were only noted when they had high population densities on individuals. They were not included in the analyses.

34 ^bBecause they are hyperparasites and microparasitic on their trematode hosts, the mean intensity for *U. crescens* is the mean number
35 of infected metacercariae in infected crabs. Prevalence is given as proportion of infected crab hosts with trematodes infected by *U.*
36 *crescens*.

37 ^cThe isopod is a parasitic castrator and the only metazoan parasite using the crab as a definitive host. Mature parasites typically occur
38 in pairs.

39 nd=no data.

42 Appendix: Table A1. Parasite prevalence (mean intensity $\pm$ sd) of *Minuca pugnax* by location from north to south. Only prevalence is
 43 given for bacterial, black gill and *Eccrinales* infections.

Range	Site	Bacterial infection	Black gill	<i>Eccrinales</i>	<i>Urosporidium crescens</i> ^a	<i>Odhneria</i> cf. <i>odhneri</i>	<i>Maritrema</i> sp.
Expanded	Portsmouth, NH	0	0	0	0	66 (3.85 $\pm$ 3.42)	0
Expanded	Rowley, MA	0	0	0	0	26 (1.46 $\pm$ 0.66)	0
Expanded	Danvers, MA	0	4	0	0	72 (7.00 $\pm$ 8.65)	0
Expanded	Weymouth, MA	0	0	0	0	40 (2.35 $\pm$ 1.53)	0
Expanded	Scituate, MA	0	0	0	0	38 (2.26 $\pm$ 2.68)	0
Historical	Falmouth, MA	2	0	0	0	0	0
Historical	Stonington, CT	0	0	8	0	0	0
Historical	Little Egg Harbor, NJ	0	0	12	0	0	54 (6.44 $\pm$ 6.22)
Historical	Cape Charles, VA	0	4	0	28 (2.07 $\pm$ 2.06)	0	20 (4.30 $\pm$ 3.56)
Historical	McIntosh County, GA	0	0	0	16 (4.13 $\pm$ 4.16)	0	34 (3.65 $\pm$ 2.76)

44
 45 Table A1, continued

Range	Site	<i>Microphallus</i> sp. A & <i>Levinseniella</i>	<i>Microphallus</i> sp. B	<i>Ancyracanthopsi s winegardi</i>	<i>Skrjabinonem a inornatae</i>	<i>Leidya distorta</i>	Parasite Richness
Expanded	Portsmouth, NH	0	0	0	0	0	1
Expanded	Rowley, MA	0	0	0	0	0	1
Expanded	Danvers, MA	0	0	0	0	0	2
Expanded	Weymouth, MA	0	0	0	0	0	1
Expanded	Scituate, MA	0	0	0	0	0	1
Historical	Falmouth, MA	0	0	0	0	0	1
Historical	Stonington, CT	0	0	2 (3.00 $\pm$ 0.00)	0	20 (1.50 $\pm$ 0.53)	3
Historical	Little Egg Harbor, NJ	0	0	28 (1.14 $\pm$ 0.36)	6 (1.00 $\pm$ 0.00)	0	4
Historical	Cape Charles, VA	58 (7.28 $\pm$ 6.55)	4 (1.00)	22 (1.64 $\pm$ 1.03)	2 (1.00 $\pm$ 0.00)	2 (2.00 $\pm$ 0.00)	9
Historical	McIntosh County, GA	72 (8.03 $\pm$ 10.48)	8 (4.25 $\pm$ 3.20)	30 (1.13 $\pm$ 0.52)	0	16 (1.13 $\pm$ 0.35)	7

46 ^aMean intensity of *U. crescens* is the mean intensity of infected metacercariae in the crab population. Prevalence is the percentage of
 47 crabs with infected worms.