

Comparative Quantitative Toxicology and QSAR Modeling of the Haloacetonitriles: Forcing Agents of Water Disinfection Byproduct Toxicity

Xiao Wei, Mengting Yang, Qingyao Zhu, Elizabeth D. Wagner, and Michael J. Plewa*



Cite This: *Environ. Sci. Technol.* 2020, 54, 8909–8918



Read Online

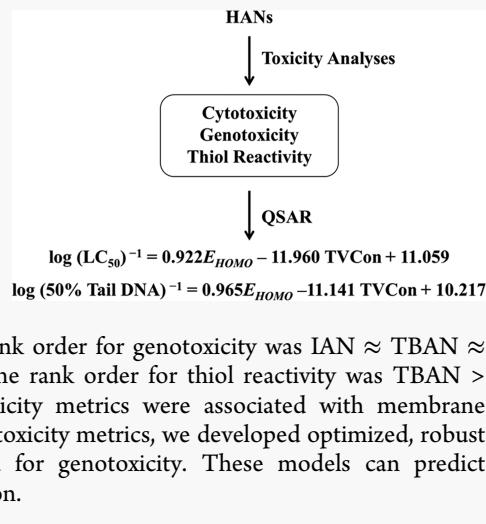
ACCESS |

Metrics & More

Article Recommendations

Supporting Information

ABSTRACT: The haloacetonitriles (HANs) is an emerging class of nitrogenous-disinfection byproducts (N-DBPs) present in disinfected drinking, recycled, processed wastewaters, and reuse waters. HANs were identified as primary forcing agents that accounted for DBP-associated toxicity. We evaluated the toxic characteristics of iodoacetonitrile (IAN), bromoacetonitrile (BAN), dibromoacetonitrile (DBAN), bromochloroacetonitrile (BCAN), tribromoacetonitrile (TBAN), chloroacetonitrile (CAN), dichloroacetonitrile (DCAN), trichloroacetonitrile (TCAN), bromodichloroacetonitrile (BDCAN), and chlorodibromoacetonitrile (CDBAN). This research generated the first quantitative, comparative analyses on the mammalian cell cytotoxicity, genotoxicity and thiol reactivity of these HANs. The descending rank order for HAN cytotoxicity was TBAN \approx DBAN $>$ BAN \approx IAN $>$ BCAN \approx CDBAN $>$ BDCAN \approx DCAN \approx CAN \approx TCAN. The rank order for genotoxicity was IAN \approx TBAN \approx DBAN $>$ BAN $>$ CDBAN \approx BCAN \approx CAN \approx TCAN \approx DCAN. The rank order for thiol reactivity was TBAN $>$ BDCAN \approx CDBAN $>$ DBAN $>$ BCAN $>$ BAN \approx IAN $>$ TCAN. These toxicity metrics were associated with membrane permeability and chemical reactivity. Based on their physiochemical parameters and toxicity metrics, we developed optimized, robust quantitative structure activity relationship (QSAR) models for cytotoxicity and for genotoxicity. These models can predict cytotoxicity and genotoxicity of novel HANs prior to analytical biological evaluation.



INTRODUCTION

Although the disinfection of drinking water was an outstanding public health achievement of the 20th century,¹ an unintended consequence was the generation of toxic disinfection byproducts (DBPs).^{2,3} Since their discovery^{4,5} over 600 DBPs were characterized,^{2,6} a fraction of the total organic halogen in disinfected water.⁷ Approximately 100 DBPs have undergone systematic, quantitative, comparative toxicological analyses.^{8–10} The U.S. EPA regulates 11 DBPs; none are nitrogen-containing (N-DBPs) or iodinated (I-DBPs).¹¹ China regulates 14 DBPs in drinking water, including one N-DBP, cyanogen chloride.¹² Yet unregulated N-DBPs and I-DBPs represent the most toxic classes in drinking waters.^{9,13–18}

Because of the concentration of haloacetonitriles (HANs) in drinking water and their toxicity, HANs are emerging as major forcing agents in the toxicity of disinfected waters.^{7,19–24} The HANs contain a cyano group attached to the α -carbon with halogen substitution; the α -carbon and cyano carbon are reactive centers.²⁵ The formation and degradation of HANs as metastable DBPs may affect the toxicity of water within a distribution network.^{26,27} The World Health Organization guidelines for dichloroacetonitrile (DCAN) and dibromoacetonitrile (DBAN) are 20 $\mu\text{g/L}$ and 70 $\mu\text{g/L}$, respectively.²⁸ From the U.S., EPA reports HANs were found up to 41 mg/L.²⁹ Other HANs, including brominated species, were defined

in the U.S. Nationwide Occurrence Study with total HANs at a maximum of 14 mg/L.⁷ In the present study the toxic characteristics of 10 HANs were evaluated: iodoacetonitrile (IAN), bromoacetonitrile (BAN), DBAN, bromochloroacetonitrile (BCAN), tribromoacetonitrile (TBAN), chloroacetonitrile (CAN), DCAN, trichloroacetonitrile (TCAN), bromodichloroacetonitrile (BDCAN), and chlorodibromoacetonitrile (CDBAN) (Table 1, Supporting Information (SI) Table S1).

Source waters^{30–32} and wastewaters²³ with high bromide and nitrogen concentrations enhanced the formation of N-DBPs. Natural organic matter, pharmaceutical, and personal care products in source waters may lead to N-DBP formation.^{33,34} Characteristics of source waters as well as disinfection methods can generate increased levels of N-DBPs.^{21,32,35–37} N-DBPs were more toxic than carbonaceous DBPs^{9,15} and HANs were forcing agents of measured or calculated toxicity.^{19,21–23,35,38–40} These facts indicate the importance of this DBP class in potential health risks.^{8,16,41}

Received: April 1, 2020

Revised: June 16, 2020

Accepted: June 18, 2020

Published: June 18, 2020



ACS Publications

© 2020 American Chemical Society

8909

<https://dx.doi.org/10.1021/acs.est.0c02035>
Environ. Sci. Technol. 2020, 54, 8909–8918

Table 1. Values of Physical–chemical, Quantum-Chemical and Topological Descriptors of 10 Haloacetonitriles^a

| haloacetonitrile, IAN, 624–75–9 | logP | R | E_{HOMO} (eV) | E_{LUMO} (eV) | μ | ClSC | ShpA | TIndx | Blndx | SDe | TVCon | TCon | Windx | S (Å ²) | |
|---|------|------|------------------------|------------------------|-------|------|------|-------|-------|-----|-------|-------|-------|---------------------|--------|
| iodoacetonitrile, IAN, 590–17–0 | 0.61 | 24.2 | −9.46 | −0.584 | 3.313 | 4 | 2.25 | 76 | 237 | 6 | 11.49 | 0.227 | 0.500 | 10 | 102.03 |
| bromoacetonitrile, DBAN, 590–17–0 | 0.20 | 19.0 | −10.81 | −0.676 | 3.290 | 4 | 2.25 | 76 | 237 | 6 | 12.00 | 0.158 | 0.500 | 10 | 96.51 |
| dibromoacetonitrile, BCAN, 3252–43–5 | 0.47 | 26.7 | −10.83 | −1.103 | 2.626 | 5 | 3.20 | 112 | 712 | 8 | 14.00 | 0.129 | 0.408 | 18 | 111.42 |
| bromochloroacetonitrile, BCAN, 83463–62–1 | 0.38 | 23.8 | −11.15 | −1.133 | 2.538 | 5 | 3.20 | 112 | 712 | 8 | 13.78 | 0.146 | 0.408 | 18 | 107.16 |
| tribromoacetonitrile, TBAN, 75519–19–6 | 1.48 | 34.4 | −10.96 | −1.531 | 1.963 | 6 | 4.17 | 152 | 1661 | 10 | 16.00 | 0.112 | 0.354 | 28 | 136.77 |
| chloroacetonitrile, CAN, 107–14–2 | 0.11 | 16.1 | −11.55 | −0.499 | 3.255 | 4 | 2.25 | 76 | 237 | 6 | 11.78 | 0.179 | 0.500 | 10 | 92.27 |
| dichloroacetonitrile, DCAN, 3018–12–0 | 0.29 | 20.9 | −11.83 | −0.967 | 2.421 | 5 | 3.20 | 112 | 712 | 8 | 13.56 | 0.166 | 0.408 | 18 | 102.94 |
| trichloroacetonitrile, TCAN, 545–06–2 | 1.21 | 25.7 | −12.12 | −1.258 | 1.342 | 6 | 4.17 | 152 | 1661 | 10 | 15.33 | 0.163 | 0.354 | 28 | 123.84 |
| bromodichloroacetonitrile, BDCAN, 60523–73–1 | 1.30 | 28.6 | −11.37 | −1.549 | 1.654 | 6 | 4.17 | 152 | 1661 | 10 | 15.56 | 0.144 | 0.354 | 28 | 128.12 |
| chlorodibromoacetonitrile, CDBAN, 144772–39–4 | 1.39 | 31.5 | −11.03 | −1.566 | 1.838 | 6 | 4.17 | 152 | 1661 | 10 | 15.78 | 0.127 | 0.354 | 28 | 132.44 |

^aLogP = octanol–water partition coefficient. R = molar refractivity. E_{HOMO} = energy of the highest occupied molecular orbital. E_{LUMO} = energy of the lowest unoccupied molecular orbital. μ = dipole moment. ClSC = cluster count. ShpA = shape attribute. TIndx = molecular topological index. Blndx = Balaban index. SDe = sum of degrees. SVDe = sum of valence degrees. TVCon = total valence connectivity. TCon = total connectivity. Windx = Wiener index. S = polar surface area.

The mechanisms of adverse biological effects induced by HANs fall into two general categories, acute direct impacts and delayed cell cycle impacts. Under *in vivo* exposure, N-DBPs induced adverse effects in zebrafish;⁴² these results agreed with Chinese hamster ovary (CHO) cell toxicity.⁹ DCAN induced developmental toxicity, reduced hatchability, and increased malformations.⁴³ A metabolomics study in mice found that HANs increased toxicity with increasing numbers of bromine substituents; these bromo-HANs induced oxidative stress-associated disruptions in amino acid, energy and lipid metabolic processes.⁴⁴ Oxidative stress was a significant mechanism of DCAN-induced hepatic mitochondrial injury in rats.⁴⁵ Toxicity pathway-based studies uncovered a potential to damage or inhibit proteins and enzymes. A soft electrophilic mechanism of action was suggested.^{10,46} When analyzed with antioxidant response element (ARE)-driven transcription of a reporter gene, the HANs were among the most potent in inducing oxidative stress in human cells. In addition BAN induced altered expression for genes related to inflammation and immune responses.⁴⁷ The HANs induced p53 activation which is used as a marker for mammalian genotoxicity and carcinogenicity.¹⁰ Quantitative structure–activity relationship (QSAR) modeling suggested that the genotoxicity of DCAN, DBAN, CAN, and IAN may be related to their molecular topological properties.⁴⁸

HANs expressed aberrant impacts on the cell cycle.^{49,50} Exposure of IAN, BAN, or CAN to CHO cells, at noncytotoxic concentrations, followed by the release from HAN treatment resulted in the accumulation of hyperploid (8N) cells over time. The potency for cell cycle alteration followed the rank order of IAN > BAN ≫ CAN. Proliferation of HAN-treated cells was suppressed for as long as 52 h. Enlarged cell size was observed without cytotoxicity with HAN treatment-induced mitosis override. This cell cycle M-phase blockage may involve the inhibition of nuclear topoisomerases.⁵¹ Cells with multiple genomes would result in aneuploidy since extra centrosomes could compromise the assembly of bipolar spindles.^{49,51} In yeast cells, DBAN delayed the transition from G1 to S phase in the cell cycle and blocked checkpoint kinase 1 (Chk1) at compromised DNA replication forks.⁵⁰ HANs may adversely impact a process or a protein that is necessary at the beginning of S phase which is also required at damaged DNA replication forks. Genomic DNA damage, the induction of aneuploidy and DNA replication stress are associated with cancer progression.^{52,53} Specific HANs are carcinogenic in rats and mice.^{54,55}

With increased concern of HANs in drinking water and as forcing agents for toxicity, we expanded the comparative CHO cell toxicity database. The objectives of this research included, (i) to generate the first quantitative, comparative analyses on the mammalian cell cytotoxicity, genotoxicity and thiol reactivity of 10 HANs, (ii) to determine the rank order of their cytotoxicity, genotoxicity, and thiol reactivity based on statistical analyses, and (iii) to determine an association among selected physicochemical characteristics based on correlation analyses and QSAR modeling of HANs and toxicity metrics.

MATERIALS AND METHODS

Haloacetonitriles. The sources, purity and physicochemical characteristics of the HANs are presented in Table 1 and SI Table S1.

Biological and Chemical Reagents, CHO Cells. For the *in vitro* cytotoxicity and genotoxicity experiments CHO K1 cell line (A52, clone 11–4–8) was employed. These CHO

Table 2. CHO Cell Cytotoxicity of the Haloacetonitrile DBPs

| HAN | lowest cytotoxic conc. (μM) ^a | mean LC_{50} value ($\mu\text{M} \pm \text{SE}$) ^b | r^2 ^c | ANOVA test statistic ^d | mean CTI value $\pm \text{SE}$ ^e |
|-------|---|--|--------------------|-------------------------------------|---|
| IAN | 0.1 | 3.27 \pm 0.05 | 0.98 | $F_{12, 163} = 148.4; P \leq 0.001$ | 307.28 \pm 4.41 |
| BAN | 1.0 | 3.10 \pm 0.06 | 0.98 | $F_{11, 228} = 98.3; P \leq 0.001$ | 325.83 \pm 7.05 |
| DBAN | 1.0 | 2.79 \pm 0.09 | 0.99 | $F_{11, 179} = 271.5; P \leq 0.001$ | 364.57 \pm 11.98 |
| BCAN | 7.0 | 8.20 \pm 0.51 | 0.96 | $F_{11, 171} = 36.2; P \leq 0.001$ | 130.84 \pm 8.24 |
| TBAN | 1.0 | 2.71 \pm 0.04 | 0.99 | $F_{11, 100} = 401.8; P \leq 0.001$ | 369.56 \pm 5.37 |
| CAN | 50.0 | 66.09 \pm 1.63 | 0.99 | $F_{13, 188} = 65.9; P \leq 0.001$ | 15.3 \pm 0.38 |
| DCAN | 10.0 | 55.03 \pm 3.23 | 0.99 | $F_{10, 171} = 63.4; P \leq 0.001$ | 19.48 \pm 1.29 |
| TCAN | 25.0 | 158.55 \pm 6.01 | 0.93 | $F_{17, 282} = 36.8; P \leq 0.001$ | 6.55 \pm 0.25 |
| BDCAN | 6.0 | 10.22 \pm 0.12 | 0.97 | $F_{15, 131} = 141.1; P \leq 0.001$ | 98.10 \pm 1.12 |
| CDBAN | 6.0 | 8.14 \pm 0.18 | 0.98 | $F_{11, 79} = 78.5; P \leq 0.001$ | 123.59 \pm 3.22 |

^aThe lowest HAN concentration that induced a statistically significant reduction as compared to the negative controls. ^b LC_{50} is the HAN concentration that induced a cell density of 50% of the negative controls. The mean and the standard error (SE) were derived using bootstrap statistics. ^cThe r^2 is the coefficient of determination for the regression analysis of the concentration–response data. ^dThe ANOVA degrees of freedom and the resulting probability value. ^e $\text{CTI} = \text{LC}_{50}^{-1} \times 10^3$.

cells are genetically stable, adherent, have normal morphology, express cell contact inhibition and grow as a monolayer without expression of neoplastic foci.^{56,57} A description of the growth conditions for the CHO cells is in the SI.

CHO Cell Chronic Cytotoxicity Analyses. CHO cell chronic cytotoxicity captures adverse biological impacts that result in the reduction in cell density after exposure to each HAN for 3 days.⁹ In this study the cytotoxicity of TBAN, BDCAN, and CDBAN was analyzed; data for the other HANs were previously published.²⁴ Detailed procedures for this assay were published.^{9,58}

Single Cell Gel Electrophoresis Genotoxicity Analyses. Single cell gel electrophoresis (SCGE or comet) analytically determines genomic DNA damage including DNA single- and double-strand breaks, incomplete excision repair sites, and alkali-labile sites in nuclei.^{59–61} The SCGE metric was the average %Tail DNA value and the 50%Tail DNA value was calculated after regression analyses. The details of SCGE analyses are presented in the SI and were published.⁹

N-Acetylcysteine Thiol Reactivity Analyses. The N-acetylcysteine (NAC) thiol reactivity screen was developed to identify potential adverse biological effects induced by toxic agents.^{46,62–64} HANs were reacted with NAC and the response was recorded spectroscopically. The details for these HAN thiol reactivity analyses are presented in the SI and the procedure was previously published.⁶²

Statistical Analyses for the Analytical Biology. Using a one-way analysis of variance (ANOVA) a test for significance was conducted. If a significant F value of $P \leq 0.05$ was obtained, a Holm-Sidak multiple comparison versus the control group analysis was conducted with the power $(1-\beta) \geq 0.8$ at $\alpha = 0.05$.^{65,66} LC_{50} values, 50%Tail DNA values and EC_{50} values were determined for cytotoxicity, genotoxicity, and NAC-thiol reactivity, respectively. The mean toxicity index values ($\pm \text{SE}$) were calculated employing bootstrap statistics.^{67,68} The definitions for the index values are defined in the SI and in Tables 2–4. The Pearson product-moment correlation test analyzed functional associations among HAN groups and biological and physicochemical metrics.^{14,69} HAN concentration–response curves for cytotoxicity (SI Figures S1–S10), genotoxicity (SI Figures S11–S20), and thiol reactivity (SI Figures S21–S30) with corresponding statistical analyses (SI Tables S2–S27) are presented in the SI.

Development of QSAR Model. By using stepwise multiple linear regression (MLR),⁷⁰ two QSAR models were

developed for the CHO assays based on the LC_{50} values for cytotoxicity and the 50%Tail DNA values for genotoxicity. For the descriptors, the $\log P$ values were estimated using the KOWWIN program (v. 1.69). The values of the energy of the highest occupied molecular orbital (E_{HOMO}), the energy of the lowest unoccupied molecular orbital (E_{LUMO}) and dipole moment (μ) were calculated with MOPAC2016 using the PM7 method. The values of molar refractivity (R) and molar surface area (S) were generated from Percepta Platform software (ACD Laboratories) and ChemAxon, respectively. Other descriptors (Table 1) were calculated using ChemOffice 2010 (Cambridge Soft). By performing the stepwise elimination of independent variables with the SPSS 22.0 software, appropriate variables were screened and selected from 15 candidate descriptors.

Statistical Metrics for Validation of the Developed QSAR Models. Several statistical parameters acquired from the regression equations including the significance level (P), the determination coefficient (R^2), variance ratio (F), the root-mean-square error (RMSE), and mean absolute error (MAE) were used to evaluate the prediction error. The established QSAR models were validated by using cross-validation through the leave-one-out (LOO) procedure to confirm their preferable prediction performance and practicability.⁷¹ Y-scrambling validation was applied to check the chance correlation of the QSAR models.⁷² For each developed QSAR equation, values of R^2 and Q^2_{LOO} were obtained from 50 randomly generated QSAR models, which should be lower than those of the developed model. The collinearity among the modeling variables needed to be reduced by evaluating the variance inflation factor (VIF).⁷³ The applicability domains (ADs) of the developed models were assessed using Williams plots of standardized residuals (δ) versus leverage (h) which expressed simple and straightforward graphical visualization of outliers. The leverage threshold (h^*) was calculated as $h^* = \frac{3k}{n}$, where k is the number of predictors plus one and n was the number of the tested compounds.⁷⁴

RESULTS AND DISCUSSION

The HANs and other N-DBPs accounted for the majority of toxicity associated with disinfected waters; HANs are more cytotoxic and genotoxic than regulated DBPs.^{9,15,18,22,24} With the increased demand for more quantitative comparative

toxicity data we expanded the number of HANs analyzed with additional toxicity end points plus QSAR modeling.

Comparative CHO Cell Chronic Cytotoxicity. TBAN, BDCAN, and CDBAN were compared with other HANs (Figure 1, Table 2). The individual cytotoxicity concen-

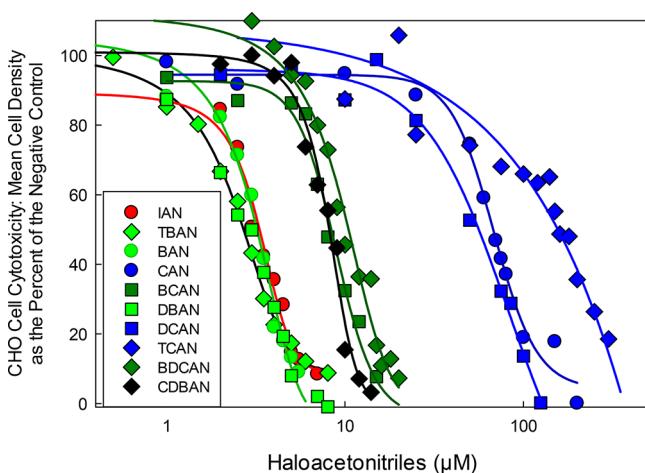


Figure 1. CHO cell chronic cytotoxicity concentration-response curves of the haloacetonitriles. SI Figures S1–S10 present each HAN cytotoxicity concentration-response curve with error bars. The ANOVA statistical analysis for TBAN, BDCAN, and CDBAN cytotoxicity are presented in SI Tables S2–S4.

tration-response curves for 10 HANs are presented in SI Figures S1–S10 and the ANOVA test of the cytotoxicity data of TBAN, BDCAN and CDBAN are presented in SI Tables S2–S4. These trihalo-HANs were highly toxic with mean LC_{50} values for TBAN, BDCAN, and CDBAN of 2.71, 10.22, and 8.14 μM , respectively (Table 2). Using cytotoxic index (CTI) values the statistical rank order from most toxic to least toxic was TBAN \approx DBAN $>$ BAN \approx IAN $>$ BCAN \approx CDBAN $>$ BDCAN $>$ DCAN \approx CAN \approx TCAN (Table 2). Applying an ANOVA test (Holm-Sidak all pairwise multiple comparison ($F_{9,175} = 744.8$; $P < 0.001$) SI Table S5) of the CTI values, the HANs separated by $>$ were significantly different while those separated by \approx were not.

Comparative CHO Cell Genotoxicity. We published the SCGE genomic DNA damage analyses of seven HANs with the tail moment as the biological metric.²⁴ However, the current preferred metric for SCGE is the %Tail DNA value.^{59,75} Using % Tail DNA values as the genotoxicity metric, a comparison of the SCGE concentration-response curves are presented in Figure 2 (individual concentration-response curves are presented in SI Figures S11–S20). Table 3 presents the statistical analyses of the genotoxicity of 10 HANs including the lowest concentration that induced a significant genotoxic response as well as the 50% Tail DNA values (SI Tables S6–S15). Using genotoxic index (GTI) values the statistical rank order from most genotoxic to least genotoxic was IAN \approx TBAN \approx DBAN $>$ BAN $>$ CDBAN \approx BDCAN \approx BCAN \approx CAN \approx TCAN \approx DCAN (where CDBAN > BCAN, CAN, TCAN, DCAN, and BDCAN > CAN, TCAN, DCAN) (Table 3). Applying an ANOVA test (Holm-Sidak all pairwise multiple comparison ($F_{9,104} = 279.9$; $P < 0.001$) SI Table S16 of the GTI values, those HANs separated by $>$ were significantly different while those separated by \approx were not.

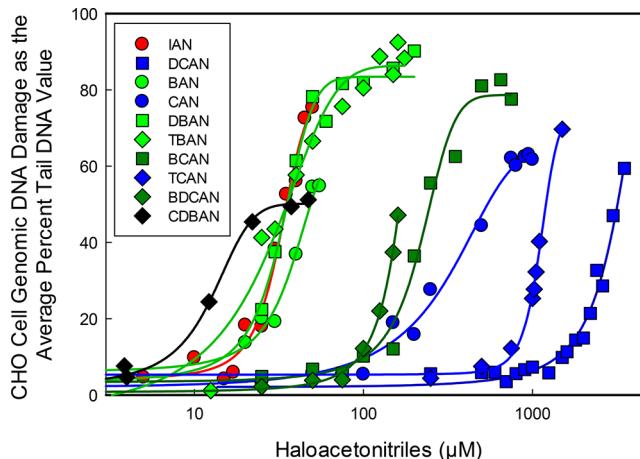


Figure 2. CHO cell genotoxicity, as the %Tail DNA, concentration-response curves of the haloacetonitriles. SI Figures S11–S20 present each HAN genotoxicity concentration-response curve with error bars, acute cytotoxicity and its ANOVA statistical analysis (SI Tables S6–S15) are presented in the SI.

Comparative NAC Thiol Reactivity. The NAC thiol reactivity was assessed; the comparative concentration-response curves are presented in Figure 3 (SI Figures S21–S30). A statistical analyses of the thiol reactivity including the lowest concentration that induced a significant response as well as their EC_{50} values is presented in Table 4 (SI Tables S17–S26). The rank order from most thiol reactive to least thiol reactive using the thiol reactivity index (TRI) values was TBAN $>$ BDCAN \approx CDBAN $>$ DBAN $>$ BCAN \approx BAN \approx IAN $>$ TCAN (Table 4). Applying an ANOVA test of the TRI values (Holm-Sidak all pairwise multiple comparison ($F_{7,53} = 809.6$; $P < 0.001$) (SI Table S27), those HANs separated by $>$ were significantly different while those separated by \approx were not. CAN and DCAN did not express NAC thiol reactivity.

QSAR Models. Table 1 lists the calculated values of the candidate descriptors. To avoid chance correlation, the ratio of compound number to variable number in models should be $>5:1$. Thus, only two descriptors were involved based on 10 HANs. Since CAN and DCAN did not express a thiol-reactivity response the corresponding QSAR equation was not developed. To select the most appropriate descriptors and to develop the QSAR models, stepwise multiple regression based on the Elimination Selection Stepwise Regression (ES-SWR) algorithm was performed using SPSS 22.0. The optimum QSAR models for cytotoxicity and genotoxicity are shown in eqs 1 and 2, respectively. To check model predictability and robustness, model validation was conducted and both models expressed high goodness-of-fit (Figures 4 and 5); detail criteria and explanation for validation tests are listed in the SI.

$$\log(LC_{50})^{-1} = 0.922E_{\text{HOMO}} - 11.960\text{TVCon} + 11.059 \quad (1)$$

$n = 10$, $R^2 = 0.933$, $Q^2_{\text{LOO}} = 0.617$, $R^2_{\text{YS}} = 0.230$, $Q^2_{\text{YS}} = -0.678$, RMSE = 0.190, MAE = 0.128, $F = 48.6$, $P < 0.0001$, VIF = 1.19

$$\begin{aligned} \log(50\%\text{Tail DNA})^{-1} \\ = 0.965E_{\text{HOMO}} - 11.141\text{TVCon} + 10.217 \end{aligned} \quad (2)$$

$n = 10$, $R^2 = 0.887$, $Q^2_{\text{LOO}} = 0.714$, $R^2_{\text{YS}} = 0.221$, $Q^2_{\text{YS}} = -0.465$, RMSE = 0.261, MAE = 0.209, $F = 27.6$, $P < 0.001$, VIF = 1.19 where n is the number of compounds in the data

Table 3. CHO Cell SCGE %Tail DNA Genotoxicity Analyses of the Haloacetonitrile DBPs

| HAN | lowest genotoxic conc. (μM) ^a | mean 50%Tail DNA ($\mu\text{M} \pm \text{SE}$) ^b | r^2 ^c | ANOVA test statistic ^d | mean GTI value $\pm \text{SE}$ ^e |
|-------|---|---|--------------------|-----------------------------------|---|
| IAN | 25 | 34.24 \pm 0.76 | 0.98 | $F_{11,56} = 57.72; P \leq 0.001$ | 29.35 \pm 0.63 |
| BAN | 20 | 48.19 \pm 1.06 | 0.98 | $F_{6,36} = 68.42; P \leq 0.001$ | 20.9 \pm 0.47 |
| DBAN | 25 | 35.50 \pm 0.37 | 0.98 | $F_{9,40} = 182.3; P \leq 0.001$ | 28.21 \pm 0.29 |
| BCAN | 200 | 250.1 \pm 12.3 | 0.98 | $F_{11,44} = 53.46; P \leq 0.001$ | 4.11 \pm 0.23 |
| TBAN | 25 | 37.76 \pm 3.23 | 0.97 | $F_{11,51} = 19.60; P \leq 0.001$ | 28.37 \pm 2.28 |
| CAN | 250 | 517.9 \pm 26.2 | 0.98 | $F_{13,44} = 46.49; P \leq 0.001$ | 2.00 \pm 0.13 |
| DCAN | 2200 | 3083 \pm 85.4 | 0.98 | $F_{17,62} = 19.20; P \leq 0.001$ | 0.33 \pm 0.01 |
| TCAN | 750 | 1187 \pm 15.2 | 0.98 | $F_{8,33} = 160.4; P \leq 0.001$ | 0.84 \pm 0.01 |
| BDCAN | 125 | 163.29 \pm 1.45 | 0.99 | $F_{7,29} = 16.25; P \leq 0.001$ | 6.13 \pm 0.05 |
| CDBAN | 100 | 139.73 \pm 2.89 | 0.83 | $F_{8,41} = 26.81; P \leq 0.001$ | 7.19 \pm 0.15 |

^aThe lowest HAN concentration that induced a statistically significant increase in the electrophoretic migration of genomic DNA from the nucleus as compared to the negative control. ^bThe HAN concentration that induced a DNA migration from the nuclei of 50%. The mean and the standard error (SE) were derived using bootstrap statistics. ^cThe r^2 is the coefficient of determination. ^dThe ANOVA degrees of freedom and the resulting probability value. ^eGTI = 50% Tail DNA⁻¹ $\times 10^3$.

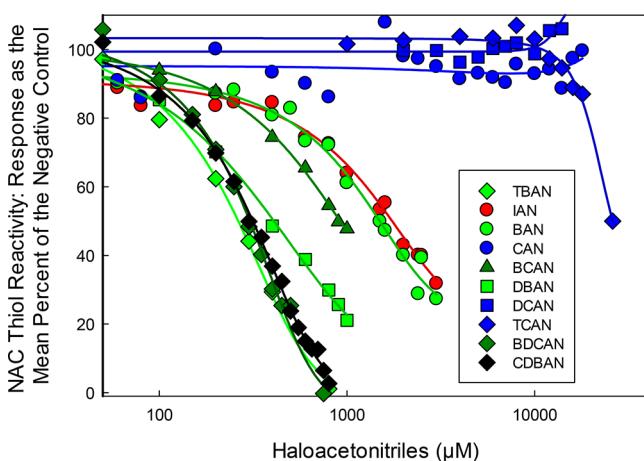


Figure 3. NAC thiol reactivity concentration-response curves of the haloacetonitriles. SI Figures S21–S30 present each HAN NAC thiol reactivity concentration-response curve with error bars and its ANOVA statistical analysis (SI Tables S17–S26) are presented in the SI.

set; R^2 is the determination coefficient; Q^2_{LOO} is the leave-one-out cross-validation coefficient; R^2_{YS} and Q^2_{YS} are Y-scrambling technique parameters; RMSE and MAE are the root-mean-

square error and the mean absolute error for the data set, respectively; P is the significance level.

Two descriptors E_{HOMO} and TVCon are involved in the optimum QSAR models. These two descriptors may reveal the toxicity mechanisms of HANs to some extent. Many chemical reactions are inextricably linked to the frontier molecular orbitals of reacting species.^{76,77} E_{HOMO} is related to the relative nucleophilicity; a higher value indicates the corresponding compound may possess higher electron donating ability, thus having a higher potential to react with electrophiles.^{78,79} Various adverse biological outcomes can result from the electrophilic-nucleophilic interactions via different mechanisms (such as Michael addition, Schiff's base formation, and nucleophilic substitution); however, the reactions are not specific.⁸⁰ It was reported that the metabolism of diethyl-stilbestrol (a carcinogenic synthetic estrogen) generates peroxides, which may react with the electrophilic sites in nucleic acids and lipids.⁸¹ HANs may attack electrophilic sites in biomacromolecules within cells via electrophilic-nucleophilic interactions, leading to cytotoxicity and genotoxicity. Iodine/bromine-containing HANs have higher toxicity than their chlorinated analogues and express higher E_{HOMO} values. Previous studies reported that cytotoxicity and developmental toxicity of aromatic DBPs correlated well with E_{HOMO} .^{69,77} Cytotoxicity induced by halobenzenes were highly related with E_{HOMO} and oxidation may a toxicity mechanism.⁸² The

Table 4. NAC Thiol Reactivity Analyses of the Haloacetonitrile DBPs

| HAN | lowest NAC response (mM) ^a | EC_{50} value (mM \pm SE) ^b | r^2 ^c | ANOVA test statistic ^d | mean TRI value \pm SE ^e |
|-------|---------------------------------------|--|--------------------|-----------------------------------|--------------------------------------|
| IAN | 0.020 | 1.71 \pm 0.07 | 0.97 | $F_{18,76} = 59.68; P \leq 0.001$ | 0.589 \pm 0.026 |
| BAN | 0.060 | 1.50 \pm 0.02 | 0.97 | $F_{18,34} = 67.71; P \leq 0.001$ | 0.666 \pm 0.008 |
| DBAN | 0.100 | 0.404 \pm 0.004 | 0.99 | $F_{7,17} = 505.4; P \leq 0.001$ | 2.48 \pm 0.02 |
| BCAN | 0.100 | 0.913 \pm 0.01 | 0.99 | $F_{7,17} = 435.1; P \leq 0.001$ | 1.10 \pm 0.01 |
| TBAN | 0.100 | 0.263 \pm 0.007 | 0.99 | $F_{8,39} = 433.5; P \leq 0.001$ | 3.82 \pm 0.09 |
| CAN | NA ^f | NA | | NS ^g | NS |
| DCAN | NA | NA | | NS | NS |
| TCAN | 18.0 | 26.15 \pm 0.01* | 0.86 | $F_{10,22} = 3.698; P \leq 0.001$ | 0.0382 \pm 0.0001 |
| BDCAN | 0.100 | 0.302 \pm 0.002 | 0.98 | $F_{12,37} = 239.3; P \leq 0.001$ | 3.31 \pm 0.02 |
| CDBAN | 0.100 | 0.314 \pm 0.006 | 0.99 | $F_{17,63} = 270.0; P \leq 0.001$ | 3.19 \pm 0.055 |

^aThe lowest HAN concentration that induced a statistically significant reduction as compared to the negative control. ^b EC_{50} value is the HAN concentration that induced a reduction in NAC thiol concentration by 50%. The * denotes that the EC_{50} value was generated by extrapolation. The mean and the standard error (SE) were derived using bootstrap statistics. ^cThe r^2 is the coefficient of determination for the regression analysis of the concentration-response data. ^dANOVA degrees of freedom and the resulting probability value. ^eTRI = $EC_{50}^{-1} \times 10^3$. ^fNA = not applicable. ^gNS = not significant.

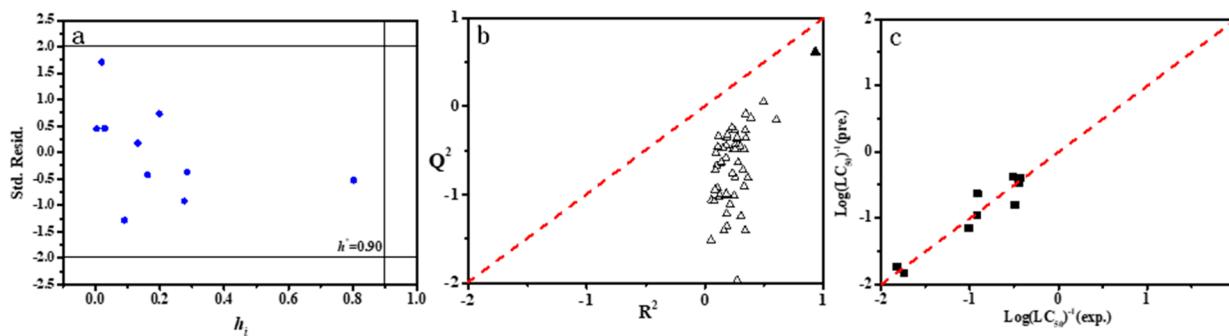


Figure 4. (a) Applicability domain of the developed QSAR model for cytotoxicity, (b) Scatter plot of the recorded Q^2_{YS} and R^2_{YS} for the haloacetonitriles (the filled triangle and empty triangles correspond to the developed QSAR model and random models, respectively), (c) Relationship between the experimental and predicted cytotoxicity data (the red dash straight line is the 45-degree benchmark line).

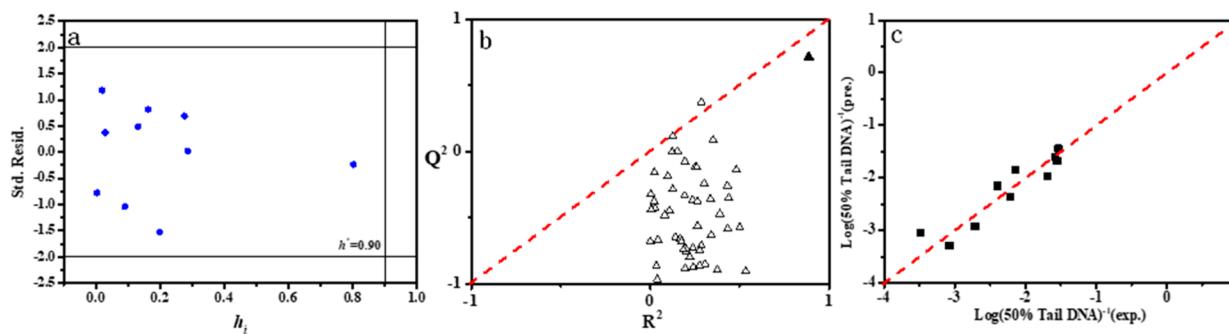


Figure 5. (a) Applicability domain of the developed QSAR model for genotoxicity, (b) Scatter plot of the recorded Q^2_{YS} and R^2_{YS} for the haloacetonitriles (the filled triangle and empty triangles correspond to the developed QSAR model and random models, respectively), (c) Relationship between the experimental and predicted genotoxicity data (the red dash straight line is the 45-degree benchmark line).

genotoxicity of DCAN, DBAN, CAN, and IAN may be related to their molecular topological properties.⁴⁸ The two HAN QSAR models in our study confirm the importance of topological properties, not only for genotoxicity, but also for cytotoxicity. Among seven topological parameters, TVCon correlated with cytotoxicity and genotoxicity. The index encodes structural characteristics, e.g., molecular size, degree of branching, shape, unsaturation, heteroatom content, and cyclicity.⁸³

These QSAR models may predict cytotoxicity and genotoxicity of novel HANs, especially iodinated HANs. If new iodinated HANs are detected in disinfected waters, toxicity data predicted by these two models would prioritize those HANs to be synthesized for quantitative toxicity. The predicted LC_{50} and 50% Tail DNA values of the recently discovered chloroiodoacetonitrile and other HAN that have not been identified in water are presented in Table 5.⁸⁴

Toxicity Correlations. For the first time a quantitative biological analyses of 10 HANs were compared using the same biological platform. The importance of the halogen atom(s) bound to the α -carbon on toxicity was calculated using index values; the iodinated and brominated HANs were approximately 18 \times more cytotoxic and 16 \times more genotoxic than their chlorinated analogues (Tables 2 and 3). The CTI and GTI values were highly and significantly correlated ($r = 0.97$; $P \leq 0.001$).

We conducted multiple correlation analyses using the CTI, GTI, and TRI values with the physicochemical parameters (Table 1). In terms of biological activity, the parameters that demonstrated a correlation with toxicity end points may be broadly divided into membrane permeability ($\log P$, S, R) and

Table 5. List of Haloacetonitriles and Their QSAR Model-Predicted Cytotoxicity (LC_{50}) and Genotoxicity (50% Tail DNA) Values

| haloacetonitrile | E_{HOMO} (eV) | TVCon | predicted LC_{50} (μM) | predicted 50% TDNA (μM) |
|--------------------------|-----------------|-------|---------------------------------------|--------------------------------------|
| diidoacetonitrile | -9.447 | 0.266 | 6.76 | 72.53 |
| triiodoacetonitrile | -9.461 | 0.330 | 41.13 | 391.30 |
| bromoiodoacetonitrile | -9.632 | 0.185 | 1.09 | 13.86 |
| chloroiodoacetonitrile | -9.708 | 0.210 | 2.53 | 30.99 |
| bromodiidoacetonitrile | -9.538 | 0.230 | 3.07 | 35.61 |
| chlorodiidoacetonitrile | -9.588 | 0.261 | 7.99 | 87.74 |
| dibromoiodoacetonitrile | -9.741 | 0.160 | 0.69 | 9.34 |
| dichloroiodoacetonitrile | -9.892 | 0.206 | 3.37 | 42.27 |

chemical reactivity (E_{HOMO} , E_{LUMO} , TVCon, SVDe) (Table 1). CTI and GTI values were significantly correlated with E_{HOMO} ($r = 0.73$; $P \leq 0.02$ and $r = 0.79$; $P \leq 0.007$, respectively). NAC thiol reactivity (TRI values) was weakly correlated with $\log P$ ($r = 0.60$) and highly correlated with R ($r = 0.83$; $P \leq 0.01$). TRI was significantly correlated with SVDe ($r = 0.66$; $P \leq 0.05$), TVCon ($r = -0.71$; $P \leq 0.05$), S ($r = 0.72$; $P \leq 0.05$), and E_{LUMO} ($r = -0.75$; $P \leq 0.04$).

Cytotoxicity, genotoxicity, and thiol reactivity were all highly and significantly correlated with the relative alkylation potency of the HANs ($r = 0.99$; $P \leq 0.002$, $r = 0.97$; $P \leq 0.01$, and $r = 0.99$; $P \leq 0.06$, respectively). The associations with toxicity, alkylation potential⁸⁵ and thiol reactivity suggest that the interaction of HANs with biological molecules, especially proteins, may play a part in their overall toxicity.

The HANs are associated with the induction of reactive oxygen species and oxidative stress. In a study on DNA damage pathway analyses, HAN-mediated interference at the DNA replication fork was reported. Based on ARE activation, the HANs were among the most potent DBPs tested for the induction of oxidative stress in human cells.^{10,47} HANs induced p53 activation which is an indicator for mammalian genotoxicity and carcinogenicity.¹⁰ The activation of DNA damage and repair pathways suggested that the genotoxicity of the HANs were structure-dependent, reflected oxidative damage to DNA and were related to their topological properties.⁴⁸ Cytotoxicity and genotoxicity are relatively immediate toxic responses. In addition HANs exhibit a delayed toxic response in that they interfere with transcription elements and/or enzymes involved in cell division. The capacity of HANs to react with biological thiols (Table 4) may not only reduce cellular defenses against oxidative stress; direct damage to proteins may be involved. HAN-mediated cell cycle M-phase blockage, the possible inhibition of associated nuclear topoisomerases, HAN effects on the G1 to S phase in the cell cycle, and blocked Chk1 checkpoint kinase strongly imply a direct adverse impact on cellular proteins involved with genomic stability.^{49–51}

When disinfected waters were analyzed for DBP levels and for calculated or measured cytotoxicity or genotoxicity, the N-DBPs, and not the regulated DBPs, were found to be the primary forcing agents driving toxicity.^{19,21–23,40} In 9 of 11 European drinking waters the primary forcing agents for both cytotoxicity and genotoxicity were N-DBPs; of these the HANs accounted for approximately 70%.^{86,87} Attention is being focused on the HANs because of their impact on the toxicity and possible chronic health effects of disinfected water. Information continues to accrue that challenges the primacy of regulated DBPs as health risks as compared to the N-DBPs.^{8,22,88} The HANs will play a central role in future evaluations of the risks to the environment and the public health posed by disinfection byproducts.

ASSOCIATED CONTENT

Supporting Information

The Supporting Information is available free of charge at <https://pubs.acs.org/doi/10.1021/acs.est.0c02035>.

QSAR validation, methods for statistical analyses for the analytical biology, biological and chemical reagents, CHO cells, CHO cell chronic cytotoxicity analyses, detailed concentration-response curves for CHO cell chronic cytotoxicity, acute SCGE genotoxicity, and response curves for the NAC thiol reactivity analyses, statistical analyses (PDF)

AUTHOR INFORMATION

Corresponding Author

Michael J. Plewa — Department of Crop Sciences and Safe Global Water Institute, University of Illinois at Urbana-Champaign, Urbana, Illinois 61801, United States; orcid.org/0000-0001-8307-1629; Email: mplewa@illinois.edu

Authors

Xiao Wei — Department of Occupational and Environmental Health, School of Public Health, Guangxi Medical University, Nanning, Guangxi 530021, China; Department of Crop

Sciences, University of Illinois at Urbana-Champaign, Urbana, Illinois 61801, United States; orcid.org/0000-0001-8907-8602

Mengting Yang — College of Chemistry and Environmental Engineering, Shenzhen University, Shenzhen, Guangdong 518000, China; orcid.org/0000-0003-4352-558X

Qingyao Zhu — College of Chemistry and Environmental Engineering, Shenzhen University, Shenzhen, Guangdong 518000, China

Elizabeth D. Wagner — Department of Crop Sciences and Safe Global Water Institute, University of Illinois at Urbana-Champaign, Urbana, Illinois 61801, United States; orcid.org/0000-0002-3198-2727

Complete contact information is available at: <https://pubs.acs.org/10.1021/acs.est.0c02035>

Notes

The authors declare no competing financial interest.

ACKNOWLEDGMENTS

We acknowledge funding from the Guangxi Medical University Training Program for Distinguished Young Scholars Grant 2017 (XW), and support from the U.S. National Science Foundation grants CBET 17-06862 and CBET 17-06575 (MP). We thank Dr. S. D. Richardson for providing the TBAN sample.

REFERENCES

- Calderon, R. L. The epidemiology of chemical contaminants of drinking water. *Food Chem. Toxicol.* **2000**, *38* (1 Suppl), S13–20.
- Richardson, S. D.; Postigo, C., Formation of DBPs: state of the science. In *Recent Advances in Disinfection By-Products*; Karanfil, T.; Mitch, W. A.; Westerhoff, P.; Xie, Y., Eds. American Chemical Society: Washington, D.C., 2015; pp 189–214.
- Richardson, S. D.; Plewa, M. J.; Wagner, E. D.; Schoeny, R.; DeMarini, D. M. Occurrence, genotoxicity, and carcinogenicity of regulated and emerging disinfection by-products in drinking water: A review and roadmap for research. *Mutat. Res., Rev. Mutat. Res.* **2007**, *636*, 178–242.
- Rook, J. J. Formation of haloforms during chlorination of natural waters. *Water Treat. Exam.* **1974**, *23*, 234–243.
- Bellar, T. A.; Lichtenberg, J. J.; Kroner, R. C. The occurrence of organohalides in chlorinated drinking waters. *J. - Am. Water Works Assoc.* **1974**, *66* (12), 703–706.
- Kimura, S. Y.; Cuthbertson, A. A.; Byer, J. D.; Richardson, S. D. The DBP exposome: Development of a new method to simultaneously quantify priority disinfection by-products and comprehensively identify unknowns. *Water Res.* **2019**, *148*, 324–333.
- Krasner, S. W.; Weinberg, H. S.; Richardson, S. D.; Pastor, S. J.; Chinn, R.; Scimenti, M. J.; Onstad, G. D.; Thruston, A. D., Jr. The occurrence of a new generation of disinfection by-products. *Environ. Sci. Technol.* **2006**, *40* (23), 7175–7185.
- Plewa, M. J.; Wagner, E. D., Charting a new path to resolve the adverse health effects of DBPs. In *Occurrence, Formation, Health Effects, and Control of Disinfection By-Products*, Karanfil, T.; Mitch, W.; Westerhoff, P.; Xie, Y., Eds.; American Chemical Society: Washington, D.C., 2015; Vol. 1190, pp 3–23.
- Wagner, E. D.; Plewa, M. J. CHO cell cytotoxicity and genotoxicity analyses of disinfection by-products: an updated review. *J. Environ. Sci.* **2017**, *58*, 64–76.
- Stalter, D.; O'Malley, E.; von Gunten, U.; Escher, B. I. Fingerprinting the reactive toxicity pathways of 50 drinking water disinfection by-products. *Water Res.* **2016**, *91*, 19–30.
- U. S. Environmental Protection Agency. National primary drinking water regulations: Stage 2 disinfectants and disinfection byproducts rule. *Fed. Regist.* **2006**, *71* (2), 387–493.

(12) Wang, X.; Mao, Y.; Tang, S.; Yang, H.; Xie, Y. F. Disinfection byproducts in drinking water and regulatory compliance: A critical review. *Front. Environ. Sci. Eng.* **2015**, *9* (1), 3–15.

(13) Plewa, M. J.; Wagner, E. D.; Jazwierska, P.; Richardson, S. D.; Chen, P. H.; McKague, A. B. Halonitromethane drinking water disinfection byproducts: chemical characterization and mammalian cell cytotoxicity and genotoxicity. *Environ. Sci. Technol.* **2004**, *38* (1), 62–68.

(14) Plewa, M. J.; Wagner, E. D.; Richardson, S. D.; Thruston, A. D., Jr.; Woo, Y. T.; McKague, A. B. Chemical and biological characterization of newly discovered iodoacid drinking water disinfection byproducts. *Environ. Sci. Technol.* **2004**, *38* (18), 4713–4722.

(15) Plewa, M. J.; Wagner, E. D.; Muellner, M. G.; Hsu, K. M.; Richardson, S. D., Comparative mammalian cell toxicity of N-DBPs and C-DBPs. In *Occurrence, Formation, Health Effects and Control of Disinfection By-Products in Drinking Water*; Karanfil, T.; Krasner, S. W.; Westerhoff, P.; Xie, Y., Eds.; American Chemical Society: Washington, D.C., 2008; Vol. 995, pp 36–50.

(16) Mian, H. R.; Hu, G.; Hewage, K.; Rodriguez, M. J.; Sadiq, R. Prioritization of unregulated disinfection by-products in drinking water distribution systems for human health risk mitigation: A critical review. *Water Res.* **2018**, *147*, 112–131.

(17) Vu, T. N.; Kimura, S. Y.; Plewa, M. J.; Richardson, S. D.; Mariñas, B. J. Predominant N-haloacetamide and haloacetonitrile formation in drinking water via the aldehyde reaction pathway. *Environ. Sci. Technol.* **2019**, *53* (2), 850–859.

(18) Richardson, S. D.; Plewa, M. J. To regulate or not to regulate? What to do with more toxic DBPs. *J. Environ. Chem. Eng.* **2020**, *8*, 103939.

(19) Plewa, M. J.; Wagner, E. D.; Richardson, S. D. TIC-Tox: A preliminary discussion on identifying the forcing agents of DBP-mediated toxicity of disinfected water. *J. Environ. Sci.* **2017**, *58*, 208–216.

(20) Cuthbertson, A. A.; Kimura, S. Y.; Liberatore, H. K.; Summers, R. S.; Knappe, D. R. U.; Stanford, B. D.; Maness, J. C.; Mulhern, R. E.; Selbes, M.; Richardson, S. D. Does granular activated carbon with chlorination produce safer drinking water? From disinfection byproducts and total organic halogen to calculated toxicity. *Environ. Sci. Technol.* **2019**, *53* (10), 5987–5999.

(21) Krasner, S. W.; Lee, T. C.; Westerhoff, P.; Fischer, N.; Hanigan, D.; Karanfil, T.; Beita-Sandi, W.; Taylor-Edmonds, L.; Andrews, R. C. Granular activated carbon treatment may result in higher predicted genotoxicity in the presence of bromide. *Environ. Sci. Technol.* **2016**, *50*, 9583–9591.

(22) Lau, S. S.; Wei, X.; Bokenkamp, K.; Wagner, E. D.; Plewa, M. J.; Mitch, W. A. Assessing additivity of cytotoxicity associated with disinfection byproducts in potable reuse and conventional drinking waters. *Environ. Sci. Technol.* **2020**, *54*, 5729–5736.

(23) Le Roux, J.; Plewa, M. J.; Wagner, E. D.; Nihemaiti, M.; Dad, A.; Croue, J. P. Chloramination of wastewater effluent: Toxicity and formation of disinfection byproducts. *J. Environ. Sci.* **2017**, *58*, 135–145.

(24) Muellner, M. G.; Wagner, E. D.; McCalla, K.; Richardson, S. D.; Woo, Y. T.; Plewa, M. J. Haloacetonitriles vs. regulated haloacetic acids: Are nitrogen containing DBPs more toxic? *Environ. Sci. Technol.* **2007**, *41* (2), 645–651.

(25) Lin, E. L.; Guion, C. W. Interaction of haloacetonitriles with glutathione and glutathione-S-transferase. *Biochem. Pharmacol.* **1989**, *38* (4), 685–688.

(26) Reckhow, D. A.; Platt, T. L.; MacNeill, A. L.; McClellan, J. N. Formation and degradation of dichloroacetonitrile in drinking waters. *J. Water. Supply Res. Technol.-Aqua* **2001**, *50* (1), 1–13.

(27) Yu, Y.; Reckhow, D. A. Kinetic analysis of haloacetonitrile stability in drinking waters. *Environ. Sci. Technol.* **2015**, *49* (18), 11028–11036.

(28) World Health Organization. *Guidelines for Drinking Water Quality*, 4th ed.; United Nations: Geneva, Switzerland, 2011; Vol. 1, p 541.

(29) McGuire, M. J.; McLain, J. L.; Obolensky, A. *Information Collection Rule Data Analysis*; American Water Works Association Research Foundation and AWWA: Denver, CO, 2002.

(30) Kristiana, I.; Liew, D.; Henderson, R. K.; Joll, C. A.; Linge, K. L. Formation and control of nitrogenous DBPs from Western Australian source waters: Investigating the impacts of high nitrogen and bromide concentrations. *J. Environ. Sci.* **2017**, *58*, 102–115.

(31) Regli, S.; Chen, J.; Messner, M.; Elovitz, M. S.; Letkiewicz, F. J.; Pegram, R. A.; Pepping, T. J.; Richardson, S. D.; Wright, J. M. Estimating potential increased bladder cancer risk due to increased bromide concentrations in sources of disinfected drinking waters. *Environ. Sci. Technol.* **2015**, *49* (22), 13094–13102.

(32) Postigo, C.; Emiliano, P.; Barcelo, D.; Valero, F. Chemical characterization and relative toxicity assessment of disinfection byproduct mixtures in a large drinking water supply network. *J. Hazard. Mater.* **2018**, *359*, 166–173.

(33) Ding, S.; Chu, W.; Bond, T.; Wang, Q.; Gao, N.; Xu, B.; Du, E. Formation and estimated toxicity of trihalomethanes, haloacetonitriles, and haloacetamides from the chlor(am)ination of acetaminophen. *J. Hazard. Mater.* **2018**, *341*, 112–119.

(34) Li, X. F.; Mitch, W. A. Drinking water disinfection byproducts (DBPs) and human health effects: Multidisciplinary challenges and opportunities. *Environ. Sci. Technol.* **2018**, *52*, 1681–1689.

(35) Zhang, Y.; Chu, W.; Yao, D.; Yin, D. Control of aliphatic halogenated DBP precursors with multiple drinking water treatment processes: Formation potential and integrated toxicity. *J. Environ. Sci.* **2017**, *58*, 322–330.

(36) Yang, Y.; Komaki, Y.; Kimura, S.; Hu, H. Y.; Wagner, E. D.; Marinas, B. J.; Plewa, M. J. Toxic impact of bromide and iodide on drinking water disinfected with chlorine or chloramines. *Environ. Sci. Technol.* **2014**, *48* (20), 12362–12369.

(37) Nihemaiti, M.; Le Roux, J.; Hoppe-Jones, C.; Reckhow, D. A.; Croue, J. P. Formation of haloacetonitriles, haloacetamides, and nitrogenous heterocyclic byproducts by chloramination of phenolic compounds. *Environ. Sci. Technol.* **2017**, *51* (1), 655–663.

(38) Chuang, Y. H.; Mitch, W. A. Effect of ozonation and biological activated carbon treatment of wastewater effluents on formation of N-nitrosamines and halogenated disinfection byproducts. *Environ. Sci. Technol.* **2017**, *51* (4), 2329–2338.

(39) Szczuksa, A.; Parker, K. M.; Harvey, C.; Hayes, E.; Vengosh, A.; Mitch, W. A. Regulated and unregulated halogenated disinfection byproduct formation from chlorination of saline groundwater. *Water Res.* **2017**, *122*, 633–644.

(40) Zeng, T.; Plewa, M. J.; Mitch, W. A. N-Nitrosamines and halogenated disinfection byproducts in U.S. Full Advanced Treatment trains for potable reuse. *Water Res.* **2016**, *101*, 176–186.

(41) Lu, G.; Qin, D.; Wang, Y.; Liu, J.; Chen, W. Single and combined effects of selected haloacetonitriles in a human-derived hepatoma line. *Ecotoxicol. Environ. Saf.* **2018**, *163*, 417–426.

(42) Hanigan, D.; Truong, L.; Simonich, M.; Tanguay, R.; Westerhoff, P. Zebrafish embryo toxicity of 15 chlorinated, brominated, and iodinated disinfection by-products. *J. Environ. Sci.* **2017**, *58*, 302–310.

(43) Lin, T.; Zhou, D.; Dong, J.; Jiang, F.; Chen, W. Acute toxicity of dichloroacetonitrile (DCAN), a typical nitrogenous disinfection by-product (N-DBP), on zebrafish (*Danio rerio*). *Ecotoxicol. Environ. Saf.* **2016**, *133*, 97–104.

(44) Deng, Y.; Zhang, Y.; Lu, Y.; Lu, K.; Bai, H.; Ren, H. Metabolomics evaluation of the *in vivo* toxicity of bromoacetonitriles: One class of high-risk nitrogenous disinfection byproducts. *Sci. Total Environ.* **2017**, *579*, 107–114.

(45) Dong, Y.; Li, F.; Shen, H.; Lu, R.; Yin, S.; Yang, Q.; Li, Z.; Wang, S. Evaluation of the water disinfection by-product dichloroacetonitrile-induced biochemical, oxidative, histopathological, and mitochondrial functional alterations: Subacute oral toxicity in rats. *Toxicol. Ind. Health* **2018**, *34* (3), 158–168.

(46) Pals, J. A.; Wagner, E. D.; Plewa, M. J. Energy of the lowest unoccupied molecular orbital, thiol reactivity, and toxicity of three

monobrominated water disinfection byproducts. *Environ. Sci. Technol.* 2016, 50 (6), 3215–3221.

(47) Procházka, E.; Melvin, S. D.; Escher, B. I.; Plewa, M. J.; Leusch, D. L. Global transcriptional changes in non-transformed human intestinal epithelial cells (FHs 74 Int) after exposure to selected drinking water disinfection by-products. *Environ. Health Perspect.* 2019, 127 (11), 117006–1–117006–11.

(48) Lan, J.; Rahman, M. B.; Gou, N.; Jiang, T.; Plewa, M. J.; Gu, A. Z. Genotoxicity assessment of drinking water disinfection by-products by DNA damage pathway profiling analysis. *Environ. Sci. Technol.* 2018, 52 (11), 6565–6575.

(49) Komaki, Y.; Marinas, B. J.; Plewa, M. J. Toxicity of drinking water disinfection by-products: cell cycle alterations induced by monohaloacetonitriles. *Environ. Sci. Technol.* 2014, 48 (19), 11662–11669.

(50) Caspari, T.; Dyer, J.; Fenner, N.; Dunn, C.; Freeman, C. The drinking water contaminant dibromoacetonitrile delays G1-S transition and suppresses Chk1 activation at broken replication forks. *Sci. Rep.* 2017, 7 (1), 12730.

(51) Komaki, Y.; Plewa, M. J. Investigation of nuclear enzyme topoisomerase as a putative molecular target of monohaloacetonitrile disinfection by-products. *J. Environ. Sci.* 2017, 58, 231–238.

(52) Davoli, T.; de Lange, T. The causes and consequences of polyploidy in normal development and cancer. *Annu. Rev. Cell Dev. Biol.* 2011, 27 (1), 585–610.

(53) Gaillard, H.; Garcia-Muse, T.; Aguilera, A. Replication stress and cancer. *Nat. Rev. Cancer* 2015, 15 (5), 276–89.

(54) Bull, R. J.; Meier, J. R.; Robinson, M.; Ringhand, H. P.; Laurie, R. D.; Stober, J. A. Evaluation of mutagenic and carcinogenic properties of brominated and chlorinated acetonitriles: by-products of chlorination. *Fundam. Appl. Toxicol.* 1985, 5 (6 Pt 1), 1065.

(55) National Toxicology Program, Toxicology and carcinogenesis studies of dibromoacetonitrile (CAS No. 3252–43–5) in F344/N rats and B6C3F1 mice (drinking water studies). *National Toxicology Program Technical Report Series* 2010, (544) 194.

(56) Wagner, E. D.; Rayburn, A. L.; Anderson, D.; Plewa, M. J. Analysis of mutagens with single cell gel electrophoresis, flow cytometry, and forward mutation assays in an isolated clone of Chinese hamster ovary cells. *Environ. Mol. Mutagen.* 1998, 32 (4), 360–368.

(57) Wagner, E. D.; Rayburn, A. L.; Anderson, D.; Plewa, M. J. Calibration of the single cell gel electrophoresis assay, flow cytometry analysis and forward mutation in Chinese hamster ovary cells. *Mutagenesis* 1998, 13 (1), 81–84.

(58) Plewa, M. J.; Wagner, E. D. *Mammalian Cell Cytotoxicity and Genotoxicity of Disinfection By-Products*; Water Research Foundation: Denver, CO, 2009; p 134.

(59) Tice, R. R.; Agurell, E.; Anderson, D.; Burlinson, B.; Hartmann, A.; Kobayashi, H.; Miyamae, Y.; Rojas, E.; Ryu, J. C.; Sasaki, Y. F. Single cell gel/comet assay: guidelines for in vitro and in vivo genetic toxicology testing. *Environ. Mol. Mutagen.* 2000, 35 (3), 206–221.

(60) Rundell, M. S.; Wagner, E. D.; Plewa, M. J. The comet assay: genotoxic damage or nuclear fragmentation? *Environ. Mol. Mutagen.* 2003, 42 (2), 61–67.

(61) Wagner, E. D.; Plewa, M. J. Microplate-based comet assay. In *The Comet Assay in Toxicology*; Dhawan, A.; Anderson, D., Eds.; Royal Society of Chemistry: London, 2009; pp 79–97.

(62) Dong, S.; Page, M. A.; Wagner, E. D.; Plewa, M. J. Thiol reactivity analyses to predict mammalian cell cytotoxicity of water samples. *Environ. Sci. Technol.* 2018, 52, 8822–8829.

(63) Pals, J. A.; Wagner, E. D.; Plewa, M. J.; Xia, M.; Attene-Ramos, M. S. Monohalogenated acetamide-induced cellular stress and genotoxicity are related to electrophilic softness and thiol/thiolate reactivity. *J. Environ. Sci.* 2017, 58, 224–230.

(64) Townsend, D. M.; Tew, K. D.; Tapiero, H. The importance of glutathione in human disease. *Biomed. Pharmacother.* 2003, 57 (3–4), 145–155.

(65) Box, G. E. P.; Hunter, W. G.; Hunter, J. S. *Statistics for Experimenters: An Introduction to Design, Data Analysis, and Model Building*; Wiley & Sons Inc.: New York, NY., 1978.

(66) Lovell, D. P.; Omori, T. Statistical issues in the use of the comet assay. *Mutagenesis* 2008, 23 (3), 171–182.

(67) Efron, B. Better bootstrap confidence intervals. *J. Am. Stat. Assoc.* 1987, 82 (397), 171–185.

(68) Singh, K.; Xie, M. *Bootstrap: A Statistical Method*; Rutgers University: New Brunswick, NJ, 2008; p 14.

(69) Li, J.; Moe, B.; Vemula, S.; Wang, W.; Li, X.-F. Emerging disinfection byproducts, halobenzoquinones: effects of isomeric structure and halogen substitution on cytotoxicity, formation of reactive oxygen species, and genotoxicity. *Environ. Sci. Technol.* 2016, 50 (13), 6744–6752.

(70) Herrig, I. M.; Böer, S. I.; Brennholt, N.; Manz, W. Development of multiple linear regression models as predictive tools for fecal indicator concentrations in a stretch of the lower Lahn River, Germany. *Water Res.* 2015, 85, 148–157.

(71) Acharya, K.; Werner, D.; Dolfig, J.; Barycki, M.; Meynet, P.; Mrozik, W.; Komolafe, O.; Puzyn, T.; Davenport, R. J. A quantitative structure-biodegradation relationship (QSBR) approach to predict biodegradation rates of aromatic chemicals. *Water Res.* 2019, 157, 181–190.

(72) Borhani, T. N. G.; Saniedanesh, M.; Bagheri, M.; Lim, J. S. QSPR prediction of the hydroxyl radical rate constant of water contaminants. *Water Res.* 2016, 98, 344–353.

(73) Yang, X.; Liu, H.; Yang, Q.; Liu, J.; Chen, J.; Shi, L. Predicting anti-androgenic activity of bisphenols using molecular docking and quantitative structure-activity relationships. *Chemosphere* 2016, 163, 373–381.

(74) Zou, X.; Zhou, X.; Lin, Z.; Deng, Z.; Yin, D. A docking-based receptor library of antibiotics and its novel application in predicting chronic mixture toxicity for environmental risk assessment. *Environ. Monit. Assess.* 2013, 185 (6), 4513–4527.

(75) Kumaravel, T. S.; Jha, A. N. Reliable Comet assay measurements for detecting DNA damage induced by ionising radiation and chemicals. *Mutat. Res., Genet. Toxicol. Environ. Mutagen.* 2006, 605 (1–2), 7–16.

(76) Karelson, M.; Lobanov, V. S.; Katritzky, A. R. Quantum-chemical descriptors in QSAR/QSPR studies. *Chem. Rev.* 1996, 96 (3), 1027–1044.

(77) Yang, M.; Zhang, X. Comparative developmental toxicity of new aromatic halogenated DBPs in a chlorinated saline sewage effluent to the marine polychaete *Platynereis dumerilii*. *Environ. Sci. Technol.* 2013, 47 (19), 10868–10876.

(78) Morell, C.; Grand, A.; Toro-Labbe, A. New dual descriptor for chemical reactivity. *J. Phys. Chem. A* 2005, 109 (1), 205–212.

(79) Zhang, Z.; Zhu, Q.; Huang, C.; Yang, M.; Li, J.; Chen, Y.; Yang, B.; Zhao, X. Comparative cytotoxicity of halogenated aromatic DBPs and implications of the corresponding developed QSAR model to toxicity mechanisms of those DBPs: Binding interactions between aromatic DBPs and catalase play an important role. *Water Res.* 2020, 170, 115283.

(80) Netzeva, T. I.; Pavan, M.; Worth, A. P. Review of (quantitative) structure–activity relationships for acute aquatic toxicity. *QSAR Comb. Sci.* 2008, 27 (1), 77–90.

(81) McCaskill, M. L.; Rogan, E.; Thomas, R. D. Diallyl sulfide inhibits diethylstilbestrol induced DNA damage in human breast epithelial cells (MCF-10A). *Steroids* 2014, 92, 96–100.

(82) Chan, K.; Jensen, N. S.; Silber, P. M.; O'Brien, P. J. Structure–activity relationships for halobenzene induced cytotoxicity in rat and human hepatocytes. *Chem.-Biol. Interact.* 2007, 165 (3), 165–174.

(83) Rastija, V.; Medić-Šarić, M. QSAR study of antioxidant activity of wine polyphenols. *Eur. J. Med. Chem.* 2009, 44 (1), 400–408.

(84) Postigo, C.; DeMarini, D. M.; Armstrong, M. D.; Liberatore, H. K.; Lamann, K.; Kimura, S. Y.; Cuthbertson, A. A.; Warren, S. H.; Richardson, S. D.; McDonald, T.; Sey, Y. M.; Ackerson, N. O. B.; Duirk, S. E.; Simmons, J. E. Chlorination of source water containing iodinated X-ray contrast media: mutagenicity and identification of

new iodinated disinfection byproducts. *Environ. Sci. Technol.* **2018**, *52* (22), 13047–13056.

(85) Lin, E. L.; Daniel, F. B.; Herren-Freund, S. L.; Pereira, M. A. Haloacetonitriles: metabolism, genotoxicity, and tumor-initiating activity. *Environ. Health Perspect.* **1986**, *69*, 67–71.

(86) Jeong, C. H.; Wagner, E. D.; Siebert, V. R.; Anduri, S.; Richardson, S. D.; Daiber, E. J.; McKague, A. B.; Kogevinas, M.; Villanueva, C. M.; Goslan, E. H.; Luo, W.; Isabelle, L. M.; Pankow, J. F.; Grazuleviciene, R.; Cordier, S.; Edwards, S. C.; Righi, E.; Nieuwenhuijsen, M. J.; Plewa, M. J. The occurrence and toxicity of disinfection byproducts in European drinking waters in relation with the HIWATE epidemiology study. *Environ. Sci. Technol.* **2012**, *46* (21), 12120–12128.

(87) Plewa, M. J., Forcing Agents of DBP-Mediated Toxicity of Disinfected Water: Are Regulated DBPs Important? In *International Water Association Second DBP Conference*, Beijing, China, 2018.

(88) Li, Y.; Jiang, J.; Li, W.; Zhu, X.; Zhang, X.; Jiang, F. Volatile DBPs contributed marginally to the developmental toxicity of drinking water DBP mixtures against *Platynenreis dumerilii*. *Chemosphere* **2020**, *252*, 126611.

Comparative quantitative toxicology and QSAR modeling of the haloacetonitriles: forcing agents of water disinfection by-product toxicity

Supporting Information

Xiao Wei ^{†,‡}, Mengting Yang [#], Qingyao Zhu [#], Elizabeth D. Wagner ^{‡,§}, Michael J. Plewa ^{‡,§,*}

[†] Department of Occupational and Environmental Health, School of Public Health, Guangxi Medical University, Nanning, Guangxi 530021, China.

[‡] Department of Crop Sciences, University of Illinois at Urbana-Champaign, Urbana, IL, 61801, USA

[§] Safe Global Water Institute, University of Illinois at Urbana-Champaign, Urbana, IL, 61801, USA

[#] College of Chemistry and Environmental Engineering, Shenzhen University, Shenzhen, Guangdong 518000 China

* Corresponding author. Department of Crop Sciences and the Safe Global Water Institute, University of Illinois at Urbana-Champaign, 1101 W. Peabody Dr., Urbana, IL, 61801, USA

E-mail address: mplewa@illinois.edu (M.J. Plewa). Co-corresponding author. X. Wei.

Contains 47 pages, 27 tables, and 30 figures

QSAR model validation

According to the acceptable criteria reported previously ($R^2 > 0.700$, $Q^2_{\text{LOO}} > 0.600$, $p < 0.05$),¹ the obtained R^2 , Q^2_{LOO} and p values indicate the two models with high goodness-of-fit and robustness. The $RMSE$ and MAE values for both models were relatively small. The F values were relatively greater in certain degrees of freedom. The applicability domains which indicated the areas of reliable predictions of the models were characterized using the Williams plot.¹ The HANs in the two data sets are in the corresponding domains, indicating that both the cytotoxicity and genotoxicity data sets have great representativeness (Figures 4a and 5a). According to the Y-scrambling test criteria, the obtained random models have significantly lower prediction accuracies than the two developed models based on experimental data, indicating no accidental correlation in the QSAR models (Figures 4b and 5b).² The variance inflation factor (VIF) of the two variables is lower than 10, indicating that there is no serious multi-collinearity among the variables and the established models are stable and acceptable. The plot of observed versus predicted $\log(\text{LC}_{50})^{-1}$ values is shown in Figure 4c, further demonstrating that the LC_{50} values predicted from the developed cytotoxicity model are generally coincident with the observed values. Also, the predicted 50% Tail DNA values generally coincide with the observed values (Figure 5c).

Table S1. Haloacetonitrile source and purity

| Haloacetonitrile & Abbreviation | Formula & CAS | Molecular Weight | Source & Purity |
|------------------------------------|--------------------------------------|------------------|---------------------------|
| Iodoacetonitrile IAN | <chem>C2H2IN</chem> 624-75-9 | 166.95 | Sigma Aldrich 98% |
| Bromoacetonitrile BAN | <chem>C2H2BrN</chem> 590-17-0 | 119.95 | Chem Service 97% |
| Dibromoacetonitrile DBAN | <chem>C2HBr2N</chem> 3252-43-5 | 198.84 | Chem Service 97% |
| Bromochloroacetonitrile BCAN | <chem>C2HBrClN</chem> 83463-62-1 | 154.39 | Chem Service Tech |
| Tribromoacetonitrile TBAN | <chem>C2Br3N</chem> 75519-19-6 | 277.74 | Cansyn Chem Corp. >90% |
| Chloroacetonitrile CAN | <chem>C2H2ClN</chem> 107-14-2 | 75.497 | Chem Service 99.5% |
| Dichloroacetonitrile DCAN | <chem>C2HCl2N</chem> 3018-12-0 | 109.94 | Chem Service 99.5% |
| Trichloroacetonitrile TCAN | <chem>C2Cl3N</chem> 545-06-2 | 144.39 | Sigma Aldrich 98% |
| Bromodichloroacetonitrile BDCAN | <chem>C2BrCl2N</chem> 60523-73-1 | 188.84 | Toronto Res. Chem. 98% |
| Chlorodibromoacetonitrile CDBAN | <chem>C2Br2ClN</chem> 144772-39-4 | 233.29 | Cansyn Chem Corp. >88% |

Methods for statistical analyses for the analytical biology

Statistical analyses were conducted for each toxicological assay. After a concentration-response curve from combined replicate experiments (>3) was generated, a test for significance using a one-way analysis of variance (ANOVA) test was conducted. If a significant F value of $P \leq 0.05$ was obtained, a Holm-Sidak multiple comparison versus the control group analysis was conducted with the power $(1-\beta) \geq 0.8$ at $\alpha = 0.05$ to identify the lowest concentration that was

significantly different from the negative control.^{3, 4} After regression analyses, LC₅₀ values were determined for CHO cell cytotoxicity, 50%Tail DNA values for CHO cell genotoxicity, and EC₅₀ values for NAC-thiol reactivity. Bootstrap statistics were conducted for each assay dataset^{5, 6} and mean toxicity index values (\pm SE) were calculated. We used index values (expressed as μ M) such that the larger the value, the more toxic or reactive the sample. The cytotoxicity index (CTI) value is the LC₅₀⁻¹ \times 10³; the genotoxicity index (GTI) value is the 50%Tail DNA⁻¹ \times 10³; the thiol reactivity index (TRI) value is the EC₅₀⁻¹ \times 10³. Using these index values, an ANOVA test was conducted to identify significant differences among specific groups. The Pearson product-moment correlation test analyzed functional associations amongst HAN groups and biological and physicochemical metrics.^{7, 8}

Biological and chemical reagents, CHO cells

For the in vitro cytotoxicity and genotoxicity experiments CHO K1 cell line (AS52, clone 11-4-8) was employed.^{9, 10} Cells were grown in Hams F12 medium containing 5% fetal bovine serum (FBS), 1% L-glutamine, and 1% antibiotics (0.25 μ g/mL amphotericin B, 100 μ g/mL streptomycin sulfate, and 100 units/mL sodium penicillin G in 0.85% saline) at 37 °C in a mammalian cell incubator with a humidified atmosphere of 5% CO₂.

CHO cell chronic cytotoxicity analyses

CHO cell cytotoxicity was measured as the reduction in cell density after exposure of CHO cells to each HAN for 72 h compared to untreated concurrent negative controls.¹¹ Cytotoxicity uncovers a wide array of toxic insults and adverse biological impacts. In this study the

cytotoxicity of TBAN, BDCAN and CDBAN was analyzed; data for the other HANs were previously published.¹² Detailed procedures for this assay were published.^{11, 13}

The concentration-response graphs illustrating the CHO cell chronic cytotoxicity of the haloacetonitriles (HANs) are presented in Figures S1 to S10.

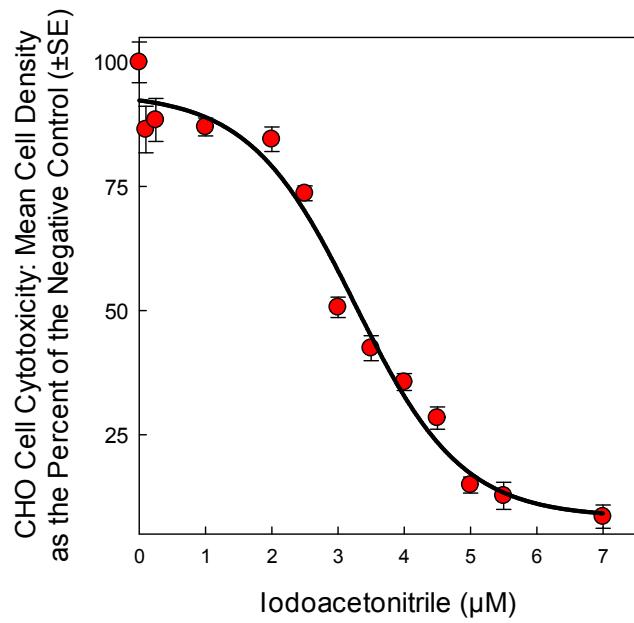


Figure S1. CHO cell cytotoxicity concentration-response curve for IAN. Mean (\pm SE) LC₅₀ value was $3.27 \pm 0.05 \mu$ M.

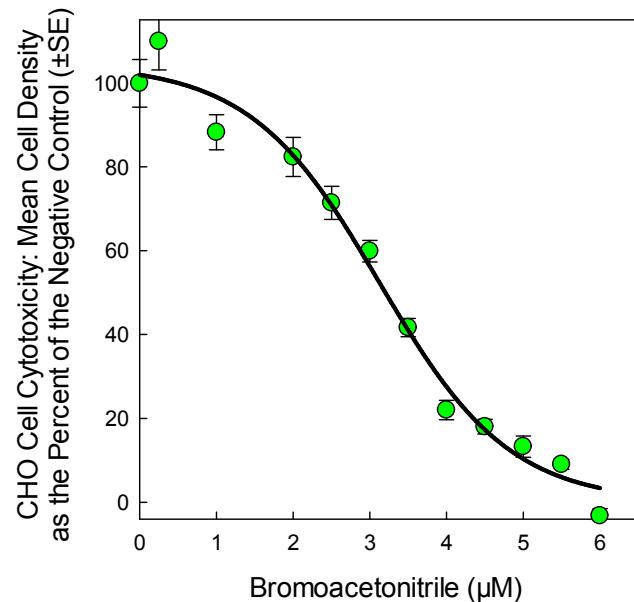


Figure S2. CHO cell cytotoxicity concentration-response curve for BAN. Mean (\pm SE) LC₅₀ value was $3.10 \pm 0.06 \mu$ M.

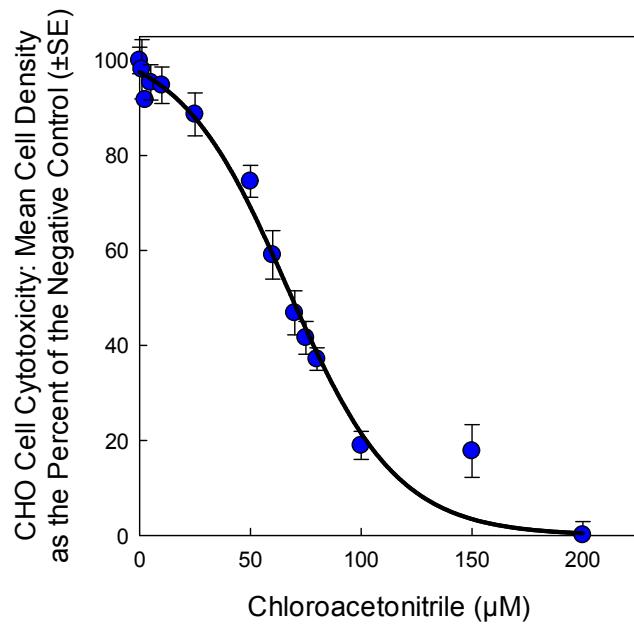


Figure S3. CHO cell cytotoxicity concentration-response curve for CAN. Mean (\pm SE) LC₅₀ value was $66.09 \pm 1.63 \mu$ M.

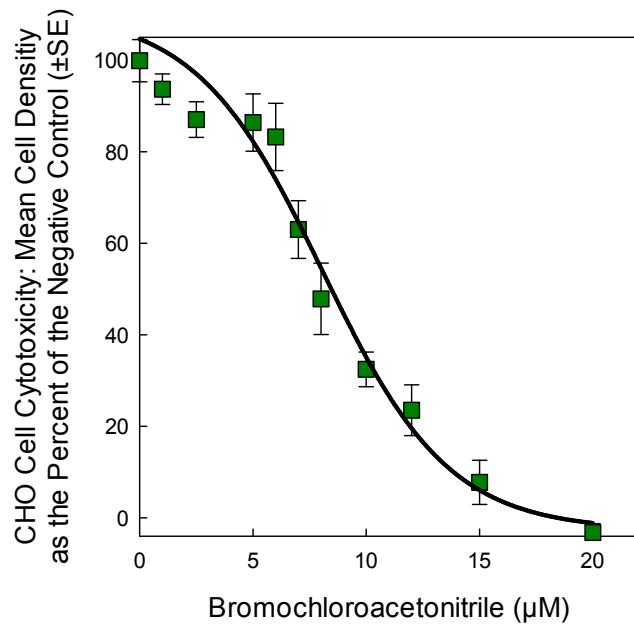


Figure S4. CHO cell cytotoxicity concentration-response curve for BCAN. Mean (\pm SE) LC₅₀ value was $8.20 \pm 0.51 \mu$ M.

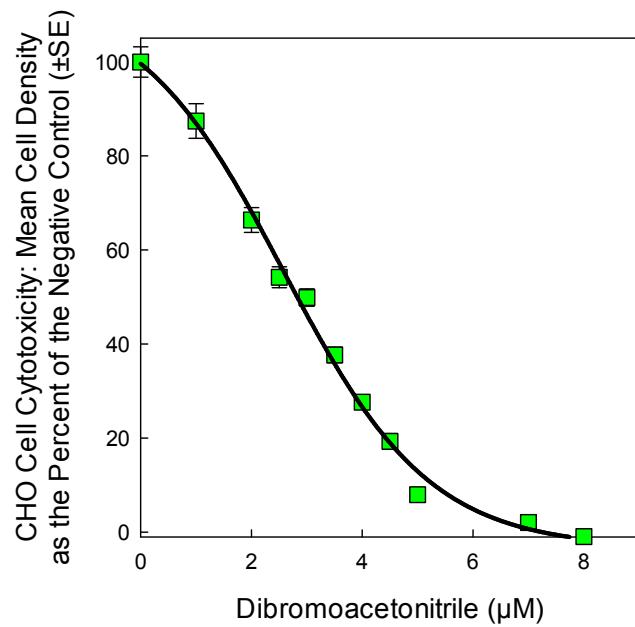


Figure S5. CHO cell cytotoxicity concentration-response curve for DBAN. Mean (\pm SE) LC₅₀ value was $2.79 \pm 0.09 \mu\text{M}$.

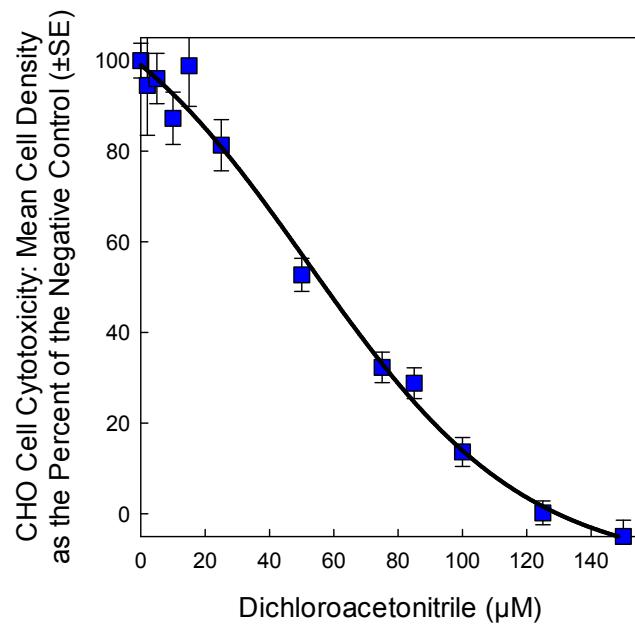


Figure S6. CHO cell cytotoxicity concentration-response curve for DCAN. Mean (\pm SE) LC₅₀ value was $55.03 \pm 3.23 \mu\text{M}$.

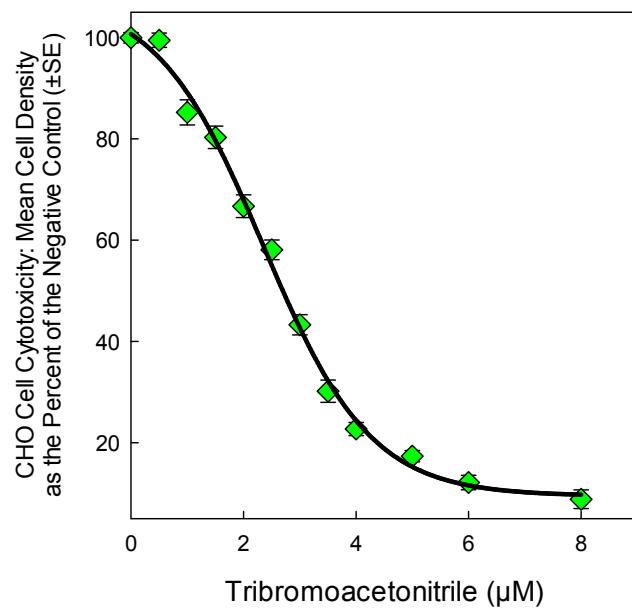


Figure S7. CHO cell cytotoxicity concentration-response curve for TBAN. Mean (\pm SE) LC₅₀ value was $2.71 \pm 0.04 \mu$ M.

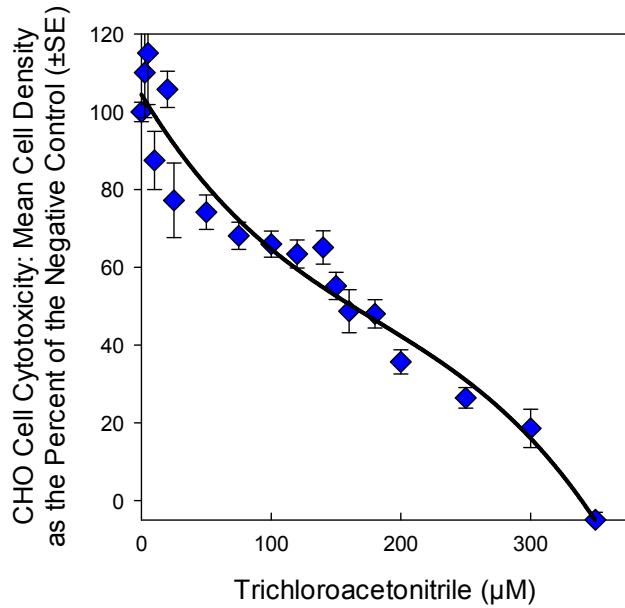


Figure S8. CHO cell cytotoxicity concentration-response curve for TCAN. Mean (\pm SE) LC₅₀ value was $158.55 \pm 6.01 \mu$ M.

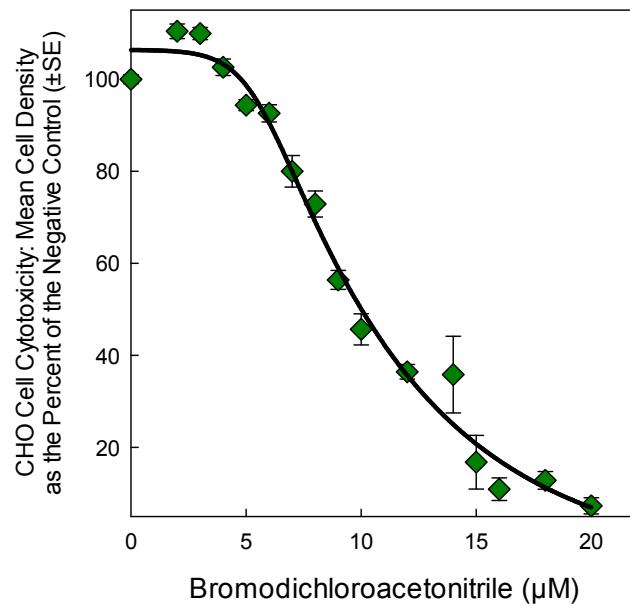


Figure S9. CHO cell cytotoxicity concentration-response curve for BDCAN. Mean (\pm SE) LC₅₀ value was $10.22 \pm 0.12 \mu$ M.

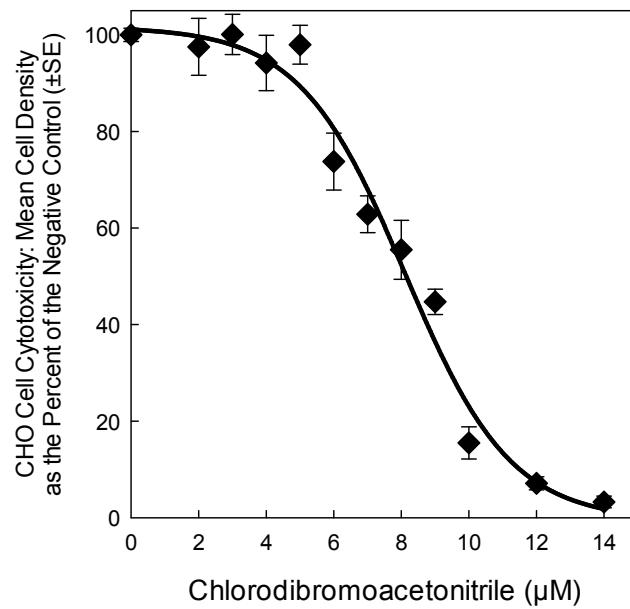


Figure S10. CHO cell cytotoxicity concentration-response curve for CDBAN. Mean (\pm SE) LC₅₀ value was $8.14 \pm 0.18 \mu$ M.

The ANOVA test statistic to determine significant decreases in cell viability for the three newly evaluated HANs are presented in Table S2 (TBAN), Table S3 (BDCAN) and Table S4 (CDBAN).

Table S2. One Way Analysis of Variance: Tribromoacetonitrile (TBAN) CHO cell cytotoxicity. Percent of the negative control.

| Group Name | N | Missing | Mean | Std Dev | SEM |
|----------------------------|-----------|----------------|-------------|----------------|------------|
| 0 TBAN μ M | 24 | 0 | 99.974 | 5.116 | 1.044 |
| 0.5 TBAN μ M | 4 | 0 | 99.478 | 2.826 | 1.413 |
| 1 TBAN μ M | 8 | 0 | 85.243 | 7.056 | 2.495 |
| 1.5 TBAN μ M | 8 | 0 | 80.312 | 6.252 | 2.210 |
| 2 TBAN μ M | 10 | 0 | 66.713 | 7.067 | 2.235 |
| 2.5 TBAN μ M | 8 | 0 | 58.093 | 5.508 | 1.948 |
| 3 TBAN μ M | 8 | 0 | 43.266 | 5.667 | 2.003 |
| 3.5 TBAN μ M | 8 | 0 | 30.159 | 6.233 | 2.204 |
| 4 TBAN μ M | 10 | 0 | 22.686 | 4.133 | 1.307 |
| 5 TBAN μ M | 8 | 0 | 17.336 | 3.137 | 1.109 |
| 6 TBAN μ M | 10 | 0 | 12.131 | 4.548 | 1.438 |
| 8 TBAN μ M | 6 | 0 | 8.824 | 4.506 | 1.840 |
| Source of Variation | DF | SS | MS | F | P |
| Between Groups | 11 | 128912.588 | 11719.326 | 401.784 | <0.001 |
| Residual | 100 | 2916.820 | 29.168 | | |
| Total | 111 | 131829.409 | | | |

The differences in the mean values among the treatment groups are greater than would be expected by chance; there is a statistically significant difference P = <0.001).

Power of performed test with alpha = 0.050: 1.000

Multiple Comparisons versus Control Group (Holm-Sidak method):
Overall significance level = 0.05

Comparisons for factor:

| Comparison | Diff of Means | t | P | P<0.050 |
|-------------------------------------|----------------------|----------|----------|-------------------|
| 0 TBAN μ M vs. 6 TBAN μ M | 87.843 | 43.213 | <0.001 | Yes |
| 0 TBAN μ M vs. 4 TBAN μ M | 77.288 | 38.021 | <0.001 | Yes |
| 0 TBAN μ M vs. 5 TBAN μ M | 82.639 | 37.480 | <0.001 | Yes |
| 0 TBAN μ M vs. 8 TBAN μ M | 91.150 | 36.976 | <0.001 | Yes |
| 0 TBAN μ M vs. 3.5 TBAN μ M | 69.815 | 31.664 | <0.001 | Yes |
| 0 TBAN μ M vs. 3 TBAN μ M | 56.709 | 25.720 | <0.001 | Yes |
| 0 TBAN μ M vs. 2.5 TBAN μ M | 41.881 | 18.995 | <0.001 | Yes |
| 0 TBAN μ M vs. 2 TBAN μ M | 33.261 | 16.363 | <0.001 | Yes |
| 0 TBAN μ M vs. 1.5 TBAN μ M | 19.662 | 8.918 | <0.001 | Yes |
| 0 TBAN μ M vs. 1 TBAN μ M | 14.731 | 6.681 | <0.001 | Yes |
| 0 TBAN μ M vs. 0.5 TBAN μ M | 0.497 | 0.170 | 0.865 | No |

Table S3. One Way Analysis of Variance: Bromodichloroacetonitrile (BDCAN) CHO cell cytotoxicity. Percent of the negative control.

| Group Name | N | Missing | Mean | Std Dev | SEM |
|-------------------|----------|----------------|-------------|----------------|------------|
| 0 BDCAN | 32 | 0 | 100.007 | 5.657 | 1.000 |
| 2 BDCAN | 9 | 0 | 110.412 | 4.775 | 1.592 |
| 3 BDCAN | 4 | 0 | 109.911 | 2.635 | 1.318 |
| 4 BDCAN | 13 | 0 | 102.589 | 6.452 | 1.790 |
| 5 BDCAN | 8 | 0 | 94.370 | 3.412 | 1.206 |
| 6 BDCAN | 17 | 0 | 92.592 | 7.649 | 1.855 |
| 7 BDCAN | 4 | 0 | 79.994 | 6.856 | 3.428 |
| 8 BDCAN | 13 | 0 | 72.903 | 10.203 | 2.830 |
| 9 BDCAN | 4 | 0 | 56.405 | 4.114 | 2.057 |
| 10 BDCAN | 9 | 0 | 45.637 | 10.176 | 3.392 |
| 12 BDCAN | 8 | 0 | 36.424 | 4.518 | 1.597 |
| 14 BDCAN | 8 | 0 | 35.808 | 23.578 | 8.336 |
| 15 BDCAN | 5 | 0 | 16.787 | 13.050 | 5.836 |
| 16 BDCAN | 4 | 0 | 10.924 | 4.901 | 2.451 |
| 18 BDCAN | 4 | 0 | 12.848 | 3.847 | 1.924 |
| 20 BDCAN | 5 | 0 | 7.322 | 3.932 | 1.758 |

| Source of Variation | DF | SS | MS | F | P |
|----------------------------|-----------|------------|-----------|----------|----------|
| Between Groups | 15 | 159535.937 | 10635.729 | 141.080 | <0.001 |
| Residual | 131 | 9875.827 | 75.388 | | |
| Total | 146 | 169411.764 | | | |

The differences in the mean values among the treatment groups are greater than would be expected by chance; there is a statistically significant difference (P = <0.001).

Power of performed test with alpha = 0.050: 1.000

Multiple Comparisons versus Control Group (Holm-Sidak method):

Overall significance level = 0.05

Comparisons for factor:

| Comparison | Diff of Means | t | P | P<0.050 |
|----------------------|----------------------|----------|----------|-------------------|
| 0 BDCAN vs. 20 BDCAN | 92.685 | 22.198 | <0.001 | Yes |
| 0 BDCAN vs. 15 BDCAN | 83.220 | 19.931 | <0.001 | Yes |
| 0 BDCAN vs. 16 BDCAN | 89.083 | 19.346 | <0.001 | Yes |
| 0 BDCAN vs. 18 BDCAN | 87.159 | 18.928 | <0.001 | Yes |
| 0 BDCAN vs. 14 BDCAN | 64.200 | 18.706 | <0.001 | Yes |
| 0 BDCAN vs. 12 BDCAN | 63.583 | 18.526 | <0.001 | Yes |
| 0 BDCAN vs. 10 BDCAN | 54.370 | 16.596 | <0.001 | Yes |
| 0 BDCAN vs. 8 BDCAN | 27.104 | 9.491 | <0.001 | Yes |
| 0 BDCAN vs. 9 BDCAN | 43.602 | 9.469 | <0.001 | Yes |
| 0 BDCAN vs. 7 BDCAN | 20.014 | 4.346 | <0.001 | Yes |
| 0 BDCAN vs. 2 BDCAN | 10.405 | 3.176 | 0.009 | Yes |
| 0 BDCAN vs. 6 BDCAN | 7.415 | 2.846 | 0.020 | Yes |
| 0 BDCAN vs. 3 BDCAN | 9.904 | 2.151 | 0.097 | No |
| 0 BDCAN vs. 5 BDCAN | 5.637 | 1.642 | 0.195 | No |
| 0 BDCAN vs. 4 BDCAN | 2.582 | 0.904 | 0.368 | No |

Table S4. One Way Analysis of Variance: Chlorodibromoacetonitrile (CDBAN) CHO cell cytotoxicity. Percent of the negative control.

| Group Name | N | Missing | Mean | Std Dev | SEM |
|----------------------------|-----------|----------------|-------------|----------------|------------|
| 0uM CDBAN | 16 | 0 | 100.007 | 5.512 | 1.378 |
| 2 CDBAN | 5 | 0 | 97.517 | 13.185 | 5.897 |
| 3 CDBAN | 4 | 0 | 100.095 | 8.360 | 4.180 |
| 4 CDBAN | 9 | 0 | 94.191 | 17.199 | 5.733 |
| 5 CDBAN | 8 | 0 | 97.966 | 11.399 | 4.030 |
| 6 CDBAN | 9 | 0 | 73.771 | 17.656 | 5.885 |
| 7 CDBAN | 4 | 0 | 62.866 | 7.620 | 3.810 |
| 8 CDBAN | 9 | 0 | 55.506 | 18.335 | 6.112 |
| 9 CDBAN | 4 | 0 | 44.707 | 5.272 | 2.636 |
| 10 CDBAN | 9 | 0 | 15.487 | 9.994 | 3.331 |
| 12 CDBAN | 9 | 0 | 7.124 | 4.014 | 1.338 |
| 14 CDBAN | 5 | 0 | 3.261 | 2.716 | 1.215 |
| Source of Variation | DF | SS | MS | F | P |
| Between Groups | 11 | 120619.108 | 10965.373 | 78.499 | <0.001 |
| Residual | 79 | 11035.320 | 139.688 | | |
| Total | 90 | 131654.428 | | | |

The differences in the mean values among the treatment groups are greater than would be expected by chance; there is a statistically significant difference (P = <0.001).

Power of performed test with alpha = 0.050: 1.000

Multiple Comparisons versus Control Group (Holm-Sidak method):
Overall significance level = 0.05

Comparisons for factor:

| Comparison | Diff of Means | t | P | P<0.050 |
|----------------------|----------------------|----------|----------|-------------------|
| 0 CDBAN vs. 12 CDBAN | 92.883 | 18.861 | <0.001 | Yes |
| 0 CDBAN vs. 10 CDBAN | 84.521 | 17.163 | <0.001 | Yes |
| 0 CDBAN vs. 14 CDBAN | 96.746 | 15.977 | <0.001 | Yes |
| 0 CDBAN vs. 8 CDBAN | 44.502 | 9.037 | <0.001 | Yes |
| 0 CDBAN vs. 9 CDBAN | 55.301 | 8.370 | <0.001 | Yes |
| 0 CDBAN vs. 7 CDBAN | 37.141 | 5.621 | <0.001 | Yes |
| 0 CDBAN vs. 6 CDBAN | 26.236 | 5.328 | <0.001 | Yes |
| 0 CDBAN vs. 4 CDBAN | 5.817 | 1.181 | 0.668 | No |
| 0 CDBAN vs. 2 CDBAN | 2.490 | 0.411 | 0.968 | No |
| 0 CDBAN vs. 5 CDBAN | 2.042 | 0.399 | 0.905 | No |
| 0 CDBAN vs. 3 CDBAN | 0.0874 | 0.0132 | 0.989 | No |

Table S5. One Way Analysis of Variance: 10 HAN CTI Comparisons.

| Group Name | N | Missing | Mean | Std Dev | SEM |
|---------------------|-----|-------------|------------|---------|--------|
| IAN CTI | 17 | 0 | 307.280 | 18.180 | 4.409 |
| BAN CTI | 25 | 0 | 325.828 | 35.268 | 7.054 |
| DBAN CTI | 16 | 0 | 364.565 | 47.938 | 11.985 |
| BCAN CTI | 19 | 0 | 130.839 | 35.932 | 8.243 |
| TBAN CTI | 14 | 0 | 369.561 | 20.086 | 5.368 |
| CAN CTI | 19 | 0 | 15.298 | 1.659 | 0.381 |
| DCAN CTI | 19 | 0 | 19.483 | 5.614 | 1.288 |
| TCAN CTI | 27 | 0 | 6.546 | 1.293 | 0.249 |
| BDCAN CTI | 18 | 0 | 98.097 | 4.737 | 1.117 |
| CDBAN CTI | 11 | 0 | 123.593 | 10.668 | 3.217 |
| Source of Variation | DF | SS | MS | F | P |
| Between Groups | 9 | 3840984.828 | 426776.092 | 744.800 | <0.001 |
| Residual | 175 | 100276.348 | 573.008 | | |
| Total | 184 | 3941261.176 | | | |

The differences in the mean values among the treatment groups are greater than would be expected by chance; there is a statistically significant difference (P = <0.001).

Power of performed test with alpha = 0.050: 1.000

All Pairwise Multiple Comparison Procedures (Holm-Sidak method):
Overall significance level = 0.05

Comparisons for factor:

| Comparison | Diff of Means | t | P | P<0.050 |
|------------------------|---------------|--------|--------|---------|
| BAN CTI vs. TCAN CTI | 319.282 | 48.056 | <0.001 | Yes |
| DBAN CTI vs. TCAN CTI | 358.019 | 47.406 | <0.001 | Yes |
| TBAN CTI vs. TCAN CTI | 363.014 | 46.047 | <0.001 | Yes |
| DBAN CTI vs. CAN CTI | 349.267 | 43.001 | <0.001 | Yes |
| BAN CTI vs. CAN CTI | 310.530 | 42.623 | <0.001 | Yes |
| DBAN CTI vs. DCAN CTI | 345.082 | 42.486 | <0.001 | Yes |
| BAN CTI vs. DCAN CTI | 306.345 | 42.048 | <0.001 | Yes |
| TBAN CTI vs. CAN CTI | 354.262 | 42.017 | <0.001 | Yes |
| TBAN CTI vs. DCAN CTI | 350.077 | 41.521 | <0.001 | Yes |
| IAN CTI vs. TCAN CTI | 300.734 | 40.577 | <0.001 | Yes |
| IAN CTI vs. CAN CTI | 291.982 | 36.536 | <0.001 | Yes |
| IAN CTI vs. DCAN CTI | 287.797 | 36.013 | <0.001 | Yes |
| DBAN CTI vs. BDCAN CTI | 266.469 | 32.398 | <0.001 | Yes |
| TBAN CTI vs. BDCAN CTI | 271.464 | 31.824 | <0.001 | Yes |
| BAN CTI vs. BDCAN CTI | 227.731 | 30.776 | <0.001 | Yes |
| DBAN CTI vs. BCAN CTI | 233.727 | 28.776 | <0.001 | Yes |
| TBAN CTI vs. BCAN CTI | 238.722 | 28.314 | <0.001 | Yes |
| BAN CTI vs. BCAN CTI | 194.989 | 26.764 | <0.001 | Yes |
| IAN CTI vs. BDCAN CTI | 209.184 | 25.839 | <0.001 | Yes |
| DBAN CTI vs. CDBAN CTI | 240.973 | 25.702 | <0.001 | Yes |
| TBAN CTI vs. CDBAN CTI | 245.968 | 25.503 | <0.001 | Yes |
| BAN CTI vs. CDBAN CTI | 202.235 | 23.350 | <0.001 | Yes |
| IAN CTI vs. BCAN CTI | 176.441 | 22.079 | <0.001 | Yes |
| IAN CTI vs. CDBAN CTI | 183.687 | 19.831 | <0.001 | Yes |
| BCAN CTI vs. TCAN CTI | 124.293 | 17.340 | <0.001 | Yes |
| BCAN CTI vs. CAN CTI | 115.541 | 14.877 | <0.001 | Yes |

| | | | | |
|-------------------------|---------|--------|--------|-----|
| BCAN CTI vs. DCAN CTI | 111.355 | 14.338 | <0.001 | Yes |
| CDBAN CTI vs. TCAN CTI | 117.047 | 13.670 | <0.001 | Yes |
| BDCAN CTI vs. TCAN CTI | 91.550 | 12.569 | <0.001 | Yes |
| CDBAN CTI vs. CAN CTI | 108.295 | 11.941 | <0.001 | Yes |
| CDBAN CTI vs. DCAN CTI | 104.109 | 11.479 | <0.001 | Yes |
| BDCAN CTI vs. CAN CTI | 82.798 | 10.516 | <0.001 | Yes |
| BDCAN CTI vs. DCAN CTI | 78.613 | 9.985 | <0.001 | Yes |
| TBAN CTI vs. IAN CTI | 62.280 | 7.209 | <0.001 | Yes |
| DBAN CTI vs. IAN CTI | 57.285 | 6.871 | <0.001 | Yes |
| TBAN CTI vs. BAN CTI | 43.733 | 5.473 | <0.001 | Yes |
| DBAN CTI vs. BAN CTI | 38.738 | 5.055 | <0.001 | Yes |
| BCAN CTI vs. BDCAN CTI | 32.742 | 4.159 | <0.001 | Yes |
| CDBAN CTI vs. BDCAN CTI | 25.496 | 2.783 | 0.041 | Yes |
| BAN CTI vs. IAN CTI | 18.548 | 2.465 | 0.085 | No |
| DCAN CTI vs. TCAN CTI | 12.937 | 1.805 | 0.315 | No |
| CAN CTI vs. TCAN CTI | 8.752 | 1.221 | 0.637 | No |
| BCAN CTI vs. CDBAN CTI | 7.246 | 0.799 | 0.810 | No |
| TBAN CTI vs. DBAN CTI | 4.995 | 0.570 | 0.814 | No |
| DCAN CTI vs. CAN CTI | 4.185 | 0.539 | 0.591 | No |

Single cell gel electrophoresis genotoxicity analyses. Single cell gel electrophoresis (SCGE or comet) quantitatively measures genomic DNA damage including DNA strand breaks, alkali-labile sites, incomplete excision repair sites, and interstrand crosslinks in the nuclei of cells.¹⁴⁻¹⁶ CHO cells were treated for 4 h with a minimum of 10 concentrations; a range finding experiment plus a minimum of two repeated experiments were conducted. The biological metric was the average %Tail DNA value; a regression analysis of the SCGE concentration-response curve was conducted to obtain the concentration that induced a 50%Tail DNA value. The details of SCGE analyses were published.¹¹

The concentration-response graphs illustrating the CHO cell SCGE genomic DNA damage of the HANs are presented in Figures S11 to S20.

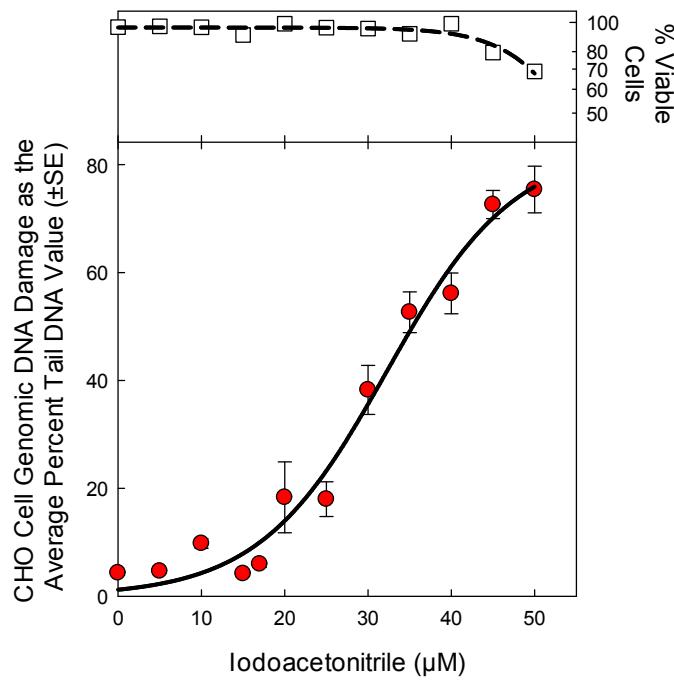


Figure S11. CHO cell genotoxicity concentration-response curve for IAN. The top panel illustrates the acute cytotoxicity and the bottom panel presents the genotoxicity as the Mean ($\pm\text{SE}$) 50% Tail DNA value that was $34.24 \pm 0.76 \mu\text{M}$.

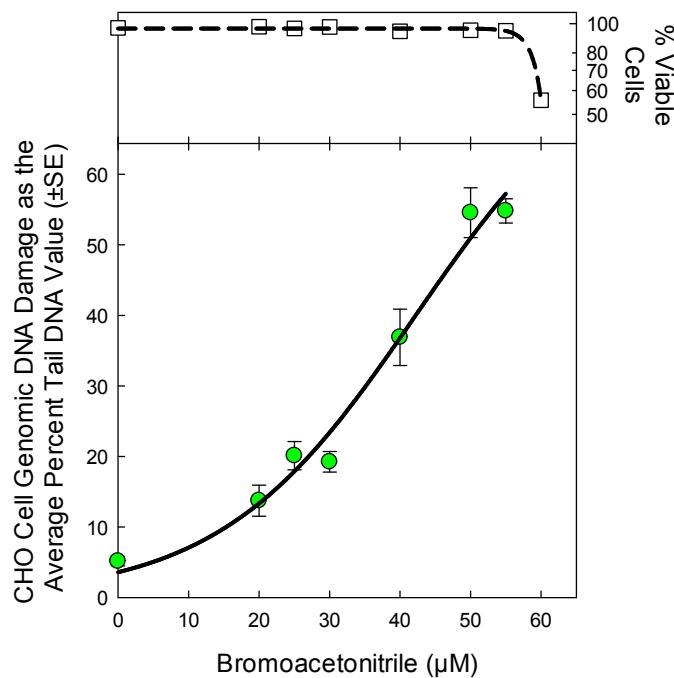


Figure S12. CHO cell genotoxicity concentration-response curve for BAN. The top panel illustrates the acute cytotoxicity and the bottom panel presents the genotoxicity as the Mean ($\pm\text{SE}$) 50% Tail DNA value that was $48.19 \pm 1.06 \mu\text{M}$.

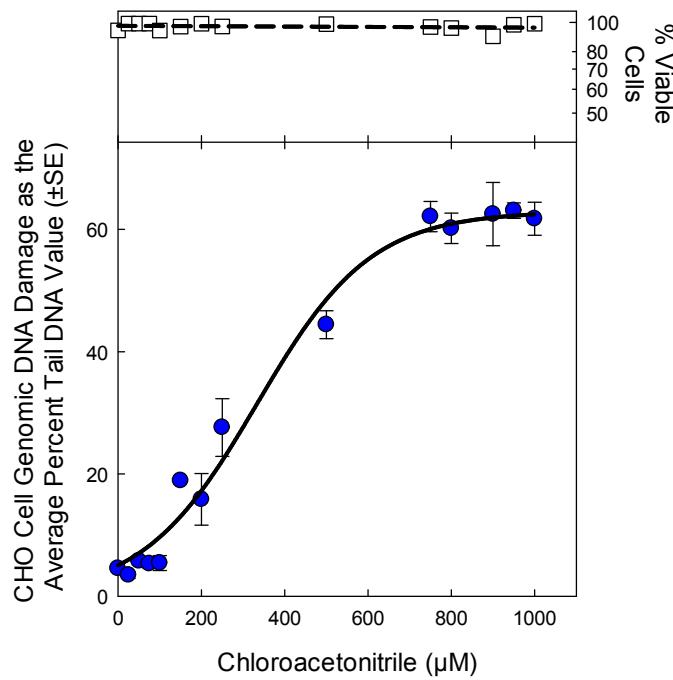


Figure S13. CHO cell genotoxicity concentration-response curve for CAN. The top panel illustrates the acute cytotoxicity and the bottom panel presents the genotoxicity as the Mean (\pm SE) 50% Tail DNA value that was $517.9 \pm 26.2 \mu\text{M}$.

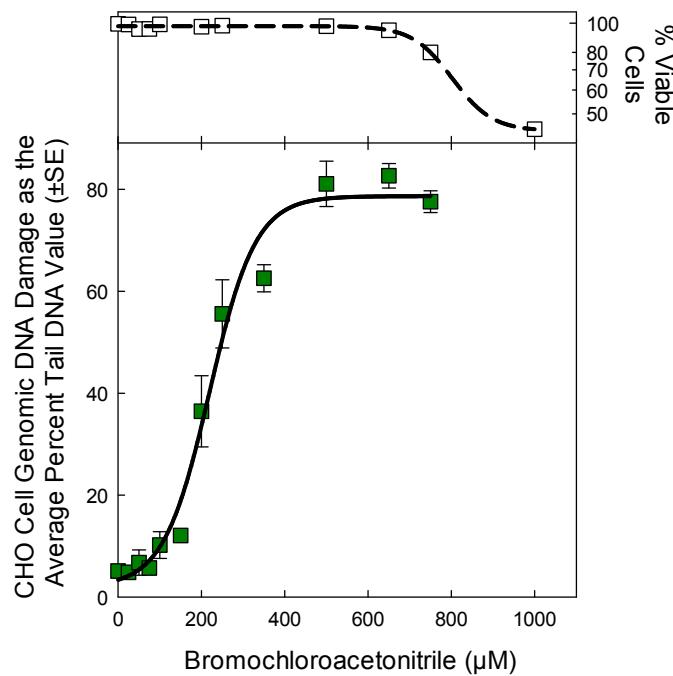


Figure S14. CHO cell genotoxicity concentration-response curve for BCAN. The top panel illustrates the acute cytotoxicity and the bottom panel presents the genotoxicity as the Mean (\pm SE) 50% Tail DNA value that was $250.1 \pm 12.3 \mu\text{M}$.

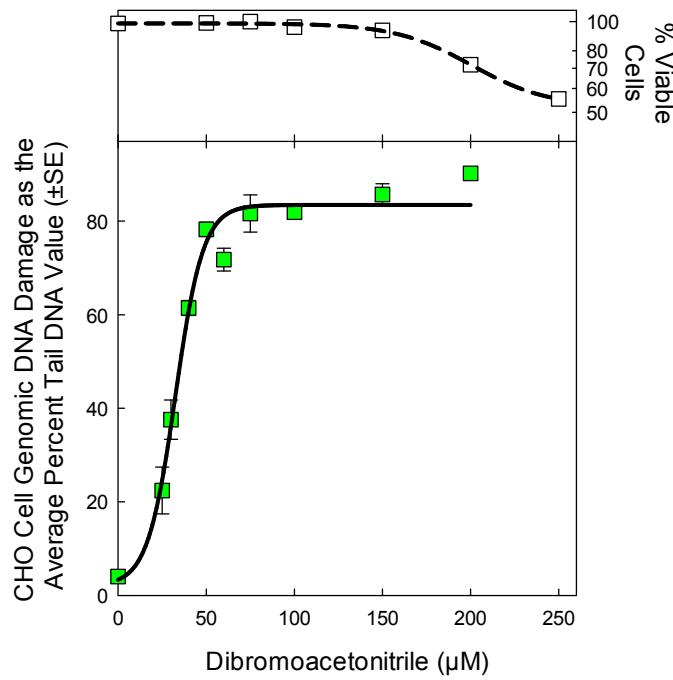


Figure S15. CHO cell genotoxicity concentration-response curve for DBAN. The top panel illustrates the acute cytotoxicity and the bottom panel presents the genotoxicity as the Mean (\pm SE) 50% Tail DNA value that was $35.50 \pm 0.37 \mu\text{M}$.

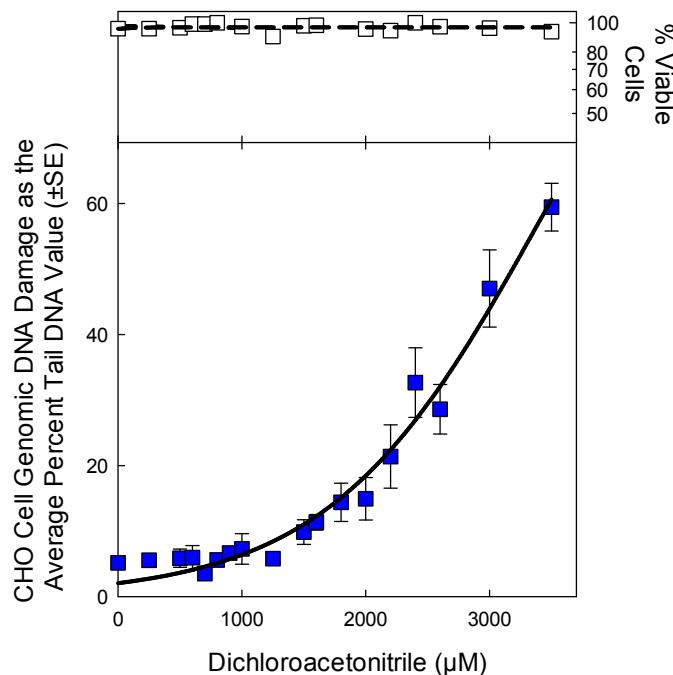


Figure S16. CHO cell genotoxicity concentration-response curve for DCAN. The top panel illustrates the acute cytotoxicity and the bottom panel presents the genotoxicity as the Mean (\pm SE) 50% Tail DNA value that was $3083 \pm 85.4 \mu\text{M}$.

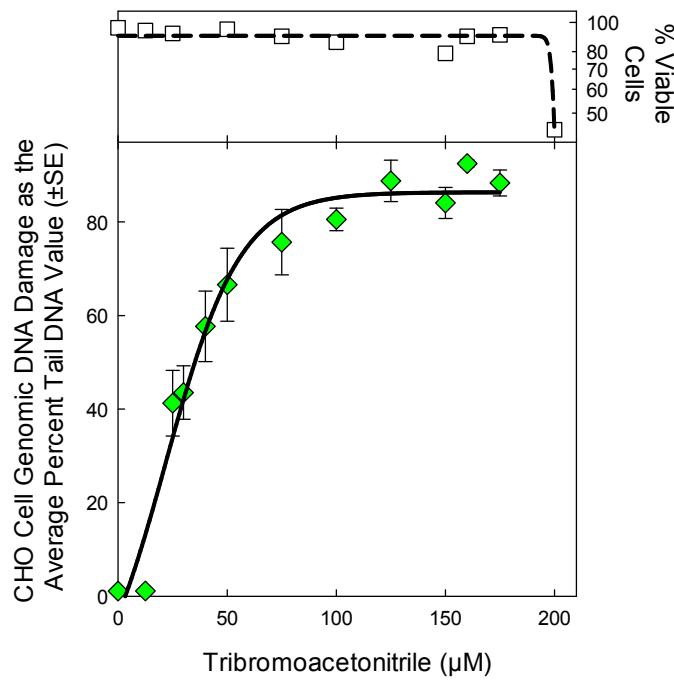


Figure S17. CHO cell genotoxicity concentration-response curve for TBAN. The top panel illustrates the acute cytotoxicity and the bottom panel presents the genotoxicity as the Mean ($\pm\text{SE}$) 50% Tail DNA value that was $37.76 \pm 3.23 \mu\text{M}$.

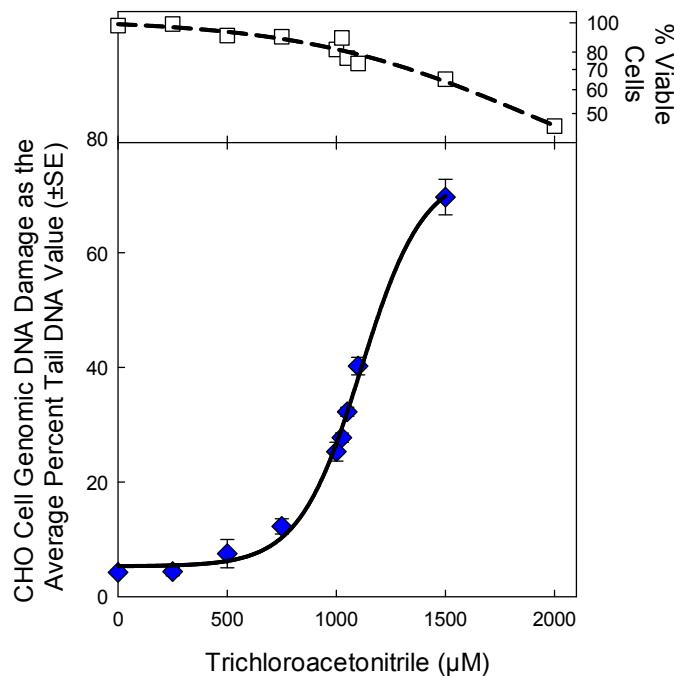


Figure S18. CHO cell genotoxicity concentration-response curve for TCAN. The top panel illustrates the acute cytotoxicity and the bottom panel presents the genotoxicity as the Mean ($\pm\text{SE}$) 50% Tail DNA value that was $1187 \pm 15.2 \mu\text{M}$.

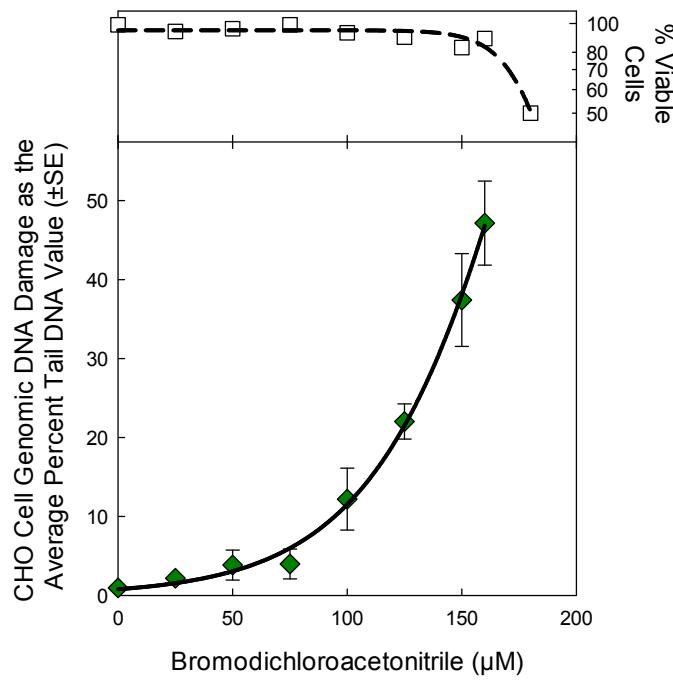


Figure S19. CHO cell genotoxicity concentration-response curve for BDCAN. The top panel illustrates the acute cytotoxicity and the bottom panel presents the genotoxicity as the Mean (\pm SE) 50% Tail DNA value that was $163.29 \pm 1.45 \mu$ M.

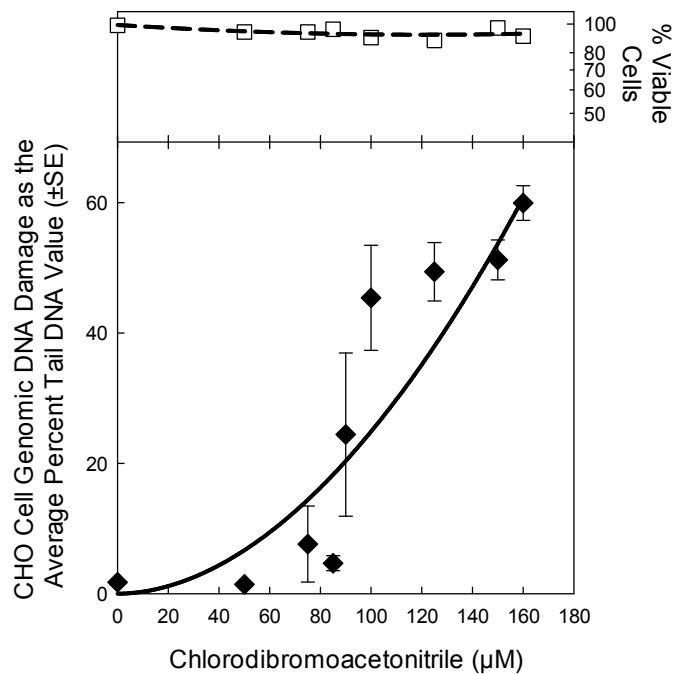


Figure S20. CHO cell genotoxicity concentration-response curve for CDBAN. The top panel illustrates the acute cytotoxicity and the bottom panel presents the genotoxicity as the Mean (\pm SE) 50% Tail DNA value that was $139.73 \pm 2.89 \mu$ M.

The ANOVA test statistic to determine if a significant increase in genomic DNA damage (%TDNA) over their concurrent negative control for the HANs are presented in Tables S6 – Table S15.

Table S6. One Way Analysis of Variance: Iodoacetonitrile (IAN) CHO cell genotoxicity SCGE %Tail DNA.

| Group Name | N | Missing | Mean | Std Dev | SEM |
|----------------------------|-----------|----------------|-------------|----------------|------------|
| IAN 0 µM | 8 | 0 | 4.325 | 1.105 | 0.391 |
| IAN 5 | 2 | 0 | 4.656 | 0.507 | 0.358 |
| IAN 10 | 2 | 0 | 9.779 | 1.234 | 0.872 |
| IAN 15 | 4 | 0 | 4.194 | 0.862 | 0.431 |
| IAN 17 | 6 | 0 | 5.968 | 1.444 | 0.590 |
| IAN 20 | 6 | 0 | 18.331 | 16.081 | 6.565 |
| IAN 25 | 6 | 0 | 17.993 | 7.881 | 3.217 |
| IAN 30 | 6 | 0 | 38.251 | 11.131 | 4.544 |
| IAN 35 | 8 | 0 | 52.653 | 10.681 | 3.776 |
| IAN 40 | 6 | 0 | 56.148 | 9.226 | 3.766 |
| IAN 45 | 10 | 0 | 72.612 | 8.229 | 2.602 |
| IAN 50 | 4 | 0 | 75.403 | 8.648 | 4.324 |
| Source of Variation | DF | SS | MS | F | P |
| Between Groups | 11 | 48799.515 | 4436.320 | 57.723 | <0.001 |
| Residual | 56 | 4303.882 | 76.855 | | |
| Total | 67 | 53103.396 | | | |

The differences in the mean values among the treatment groups are greater than would be expected by chance; there is a statistically significant difference (P = <0.001).

Power of performed test with alpha = 0.050: 1.000

Multiple Comparisons versus Control Group (Holm-Sidak method):
Overall significance level = 0.05

Comparisons for factor:

| Comparison | Diff of Means | t | P | P<0.050 |
|---------------------|----------------------|----------|----------|-------------------|
| IAN 0 µM vs. IAN 45 | 68.287 | 16.421 | <0.001 | Yes |
| IAN 0 µM vs. IAN 50 | 71.079 | 13.240 | <0.001 | Yes |
| IAN 0 µM vs. IAN 35 | 48.328 | 11.025 | <0.001 | Yes |
| IAN 0 µM vs. IAN 40 | 51.824 | 10.946 | <0.001 | Yes |
| IAN 0 µM vs. IAN 30 | 33.926 | 7.166 | <0.001 | Yes |
| IAN 0 µM vs. IAN 20 | 14.006 | 2.958 | 0.027 | Yes |
| IAN 0 µM vs. IAN 25 | 13.668 | 2.887 | 0.027 | Yes |
| IAN 0 µM vs. IAN 10 | 5.454 | 0.787 | 0.898 | No |
| IAN 0 µM vs. IAN 17 | 1.643 | 0.347 | 0.980 | No |
| IAN 0 µM vs. IAN 5 | 0.332 | 0.0479 | 0.999 | No |
| IAN 0 µM vs. IAN 15 | 0.130 | 0.0243 | 0.981 | No |

Table S7. One Way Analysis of Variance: Bromoacetonitrile (BAN) CHO cell genotoxicity SCGE %Tail DNA.

| Group Name | N | Missing | Mean | Std Dev | SEM |
|----------------------------|-----------|----------------|-------------|----------------|------------|
| BAN 0 μ M | 8 | 0 | 5.151 | 1.878 | 0.664 |
| BAN 20 | 4 | 0 | 13.727 | 4.414 | 2.207 |
| BAN 25 | 6 | 0 | 20.110 | 4.909 | 2.004 |
| BAN 30 | 6 | 0 | 19.237 | 3.575 | 1.459 |
| BAN 40 | 6 | 0 | 36.887 | 9.771 | 3.989 |
| BAN 50 | 7 | 0 | 54.552 | 9.332 | 3.527 |
| BAN 55 | 6 | 0 | 54.809 | 4.210 | 1.719 |
| Source of Variation | DF | SS | MS | F | P |
| Between Groups | 6 | 15462.675 | 2577.113 | 68.419 | <0.001 |
| Residual | 36 | 1356.006 | 37.667 | | |
| Total | 42 | 16818.681 | | | |

The differences in the mean values among the treatment groups are greater than would be expected by chance; there is a statistically significant difference (P = <0.001).

Power of performed test with alpha = 0.050: 1.000

Multiple Comparisons versus Control Group (Holm-Sidak method):
Overall significance level = 0.05

Comparisons for factor:

| Comparison | Diff of Means | t | P | P<0.050 |
|--------------------------|----------------------|----------|----------|-------------------|
| BAN 0 μ M vs. BAN 50 | 49.401 | 15.553 | <0.001 | Yes |
| BAN 0 μ M vs. BAN 55 | 49.658 | 14.982 | <0.001 | Yes |
| BAN 0 μ M vs. BAN 40 | 31.736 | 9.575 | <0.001 | Yes |
| BAN 0 μ M vs. BAN 25 | 14.959 | 4.513 | <0.001 | Yes |
| BAN 0 μ M vs. BAN 30 | 14.086 | 4.250 | <0.001 | Yes |
| BAN 0 μ M vs. BAN 20 | 8.576 | 2.282 | 0.029 | Yes |

Table S8. One Way Analysis of Variance: Chloroacetonitrile (CAN) CHO cell genotoxicity SCGE %Tail DNA.

| Group Name | N | Missing | Mean | Std Dev | SEM |
|----------------------------|-----------|----------------|-------------|----------------|------------|
| CAN 0 μ M | 5 | 0 | 4.561 | 0.700 | 0.313 |
| CAN 25 | 2 | 0 | 3.464 | 0.770 | 0.544 |
| CAN 50 | 2 | 0 | 5.774 | 0.000 | 0.000 |
| CAN 75 | 2 | 0 | 5.330 | 1.051 | 0.743 |
| CAN 100 | 3 | 0 | 5.429 | 2.112 | 1.219 |
| CAN 150 | 2 | 0 | 18.919 | 0.000 | 0.000 |
| CAN 200 | 4 | 0 | 15.841 | 8.438 | 4.219 |
| CAN 250 | 6 | 0 | 27.600 | 11.586 | 4.730 |
| CAN 500 | 6 | 0 | 44.428 | 5.585 | 2.280 |
| CAN 750 | 6 | 0 | 62.103 | 6.059 | 2.474 |
| CAN 800 | 6 | 0 | 60.184 | 6.105 | 2.492 |
| CAN 900 | 6 | 0 | 62.503 | 12.683 | 5.178 |
| CAN 950 | 2 | 0 | 63.072 | 1.779 | 1.258 |
| CAN 1000 | 6 | 0 | 61.750 | 6.611 | 2.699 |
| Source of Variation | DF | SS | MS | F | P |
| Between Groups | 13 | 33642.338 | 2587.872 | 46.490 | <0.001 |
| Residual | 44 | 2449.285 | 55.666 | | |
| Total | 57 | 36091.623 | | | |

The differences in the mean values among the treatment groups are greater than would be expected by chance; there is a statistically significant difference (P = <0.001).

Power of performed test with alpha = 0.050: 1.000

Multiple Comparisons versus Control Group (Holm-Sidak method):
Overall significance level = 0.05

Comparisons for factor:

| Comparison | Diff of Means | t | P | P<0.050 |
|----------------------------|----------------------|----------|----------|-------------------|
| CAN 0 μ M vs. CAN 900 | 57.943 | 12.825 | <0.001 | Yes |
| CAN 0 μ M vs. CAN 750 | 57.542 | 12.737 | <0.001 | Yes |
| CAN 0 μ M vs. CAN 1000 | 57.190 | 12.659 | <0.001 | Yes |
| CAN 0 μ M vs. CAN 800 | 55.623 | 12.312 | <0.001 | Yes |
| CAN 0 μ M vs. CAN 950 | 58.511 | 9.373 | <0.001 | Yes |
| CAN 0 μ M vs. CAN 500 | 39.867 | 8.824 | <0.001 | Yes |
| CAN 0 μ M vs. CAN 250 | 23.039 | 5.100 | <0.001 | Yes |
| CAN 0 μ M vs. CAN 150 | 14.359 | 2.300 | 0.147 | No |
| CAN 0 μ M vs. CAN 200 | 11.280 | 2.254 | 0.138 | No |
| CAN 0 μ M vs. CAN 50 | 1.213 | 0.194 | 0.999 | No |
| CAN 0 μ M vs. CAN 25 | 1.097 | 0.176 | 0.997 | No |
| CAN 0 μ M vs. CAN 100 | 0.868 | 0.159 | 0.984 | No |
| CAN 0 μ M vs. CAN 75 | 0.769 | 0.123 | 0.903 | No |

Table S9. One Way Analysis of Variance: Bromochloroacetonitrile (BCAN) CHO cell genotoxicity SCGE %Tail DNA.

| Group Name | N | Missing | Mean | Std Dev | SEM |
|----------------------------|-----------|----------------|-------------|----------------|------------|
| BCAN 0 μ M | 6 | 0 | 5.093 | 1.420 | 0.580 |
| BCAN 25 | 2 | 0 | 4.835 | 0.897 | 0.635 |
| BCAN 50 | 2 | 0 | 6.761 | 3.501 | 2.476 |
| BCAN 75 | 2 | 0 | 5.700 | 0.731 | 0.517 |
| BCAN 100 | 4 | 0 | 10.190 | 5.251 | 2.625 |
| BCAN 150 | 6 | 0 | 12.061 | 3.209 | 1.310 |
| BCAN 200 | 6 | 0 | 36.449 | 17.118 | 6.988 |
| BCAN 250 | 6 | 0 | 55.574 | 16.365 | 6.681 |
| BCAN 350 | 6 | 0 | 62.565 | 6.521 | 2.662 |
| BCAN 500 | 6 | 0 | 81.086 | 10.892 | 4.447 |
| BCAN 650 | 6 | 0 | 82.674 | 5.914 | 2.415 |
| BCAN 750 | 4 | 0 | 77.587 | 4.281 | 2.141 |
| Source of Variation | DF | SS | MS | F | P |
| Between Groups | 11 | 53426.540 | 4856.958 | 53.455 | <0.001 |
| Residual | 44 | 3997.866 | 90.861 | | |
| Total | 55 | 57424.406 | | | |

The differences in the mean values among the treatment groups are greater than would be expected by chance; there is a statistically significant difference (P = <0.001).

Power of performed test with alpha = 0.050: 1.000

Multiple Comparisons versus Control Group (Holm-Sidak method):
Overall significance level = 0.05

Comparisons for factor:

| Comparison | Diff of Means | t | P | P<0.050 |
|-----------------------------|----------------------|----------|----------|-------------------|
| BCAN 0 μ M vs. BCAN 650 | 77.581 | 14.097 | <0.001 | Yes |
| BCAN 0 μ M vs. BCAN 500 | 75.992 | 13.808 | <0.001 | Yes |
| BCAN 0 μ M vs. BCAN 750 | 72.493 | 11.782 | <0.001 | Yes |
| BCAN 0 μ M vs. BCAN 350 | 57.471 | 10.443 | <0.001 | Yes |
| BCAN 0 μ M vs. BCAN 250 | 50.481 | 9.173 | <0.001 | Yes |
| BCAN 0 μ M vs. BCAN 200 | 31.356 | 5.698 | <0.001 | Yes |
| BCAN 0 μ M vs. BCAN 150 | 6.967 | 1.266 | 0.696 | No |
| BCAN 0 μ M vs. BCAN 100 | 5.096 | 0.828 | 0.880 | No |
| BCAN 0 μ M vs. BCAN 50 | 1.668 | 0.214 | 0.995 | No |
| BCAN 0 μ M vs. BCAN 75 | 0.606 | 0.0779 | 0.996 | No |
| BCAN 0 μ M vs. BCAN 25 | 0.258 | 0.0332 | 0.974 | No |

Table S10. One Way Analysis of Variance: Dibromoacetonitrile (DBAN) CHO cell genotoxicity SCGE %Tail DNA.

| Group Name | N | Missing | Mean | Std Dev | SEM |
|----------------------------|-----------|----------------|-------------|----------------|------------|
| DBAN 0 μ M | 6 | 0 | 4.056 | 1.114 | 0.455 |
| DBAN 25 | 2 | 0 | 22.427 | 7.081 | 5.007 |
| DBAN 30 | 2 | 0 | 37.559 | 5.935 | 4.197 |
| DBAN 40 | 2 | 0 | 61.450 | 2.251 | 1.592 |
| DBAN 50 | 8 | 0 | 78.244 | 4.248 | 1.502 |
| DBAN 60 | 2 | 0 | 69.927 | 1.935 | 1.368 |
| DBAN 75 | 6 | 0 | 81.632 | 9.744 | 3.978 |
| DBAN 100 | 12 | 0 | 82.427 | 2.748 | 0.793 |
| DBAN 150 | 6 | 0 | 85.688 | 5.653 | 2.308 |
| DBAN 200 | 4 | 0 | 90.201 | 1.262 | 0.631 |
| Source of Variation | DF | SS | MS | F | P |
| Between Groups | 9 | 38930.391 | 4325.599 | 182.313 | <0.001 |
| Residual | 40 | 949.047 | 23.726 | | |
| Total | 49 | 39879.438 | | | |

The differences in the mean values among the treatment groups are greater than would be expected by chance; there is a statistically significant difference (P = <0.001).

Power of performed test with alpha = 0.050: 1.000

Multiple Comparisons versus Control Group (Holm-Sidak method):
Overall significance level = 0.05

Comparisons for factor:

| Comparison | Diff of Means | t | P | P<0.050 |
|-----------------------------|----------------------|----------|----------|-------------------|
| DBAN 0 μ M vs. DBAN 100 | 78.371 | 32.179 | <0.001 | Yes |
| DBAN 0 μ M vs. DBAN 150 | 81.632 | 29.027 | <0.001 | Yes |
| DBAN 0 μ M vs. DBAN 50 | 74.187 | 28.202 | <0.001 | Yes |
| DBAN 0 μ M vs. DBAN 75 | 77.576 | 27.585 | <0.001 | Yes |
| DBAN 0 μ M vs. DBAN 200 | 86.145 | 27.398 | <0.001 | Yes |
| DBAN 0 μ M vs. DBAN 60 | 65.871 | 16.562 | <0.001 | Yes |
| DBAN 0 μ M vs. DBAN 40 | 57.394 | 14.431 | <0.001 | Yes |
| DBAN 0 μ M vs. DBAN 30 | 33.503 | 8.424 | <0.001 | Yes |
| DBAN 0 μ M vs. DBAN 25 | 18.371 | 4.619 | <0.001 | Yes |

Table S11. One Way Analysis of Variance: Dichloroacetonitrile (DCAN) CHO cell genotoxicity SCGE %Tail DNA.

| Group Name | N | Missing | Mean | Std Dev | SEM |
|-------------------|----------|----------------|-------------|----------------|------------|
| DCAN 0 µM | 8 | 0 | 5.150 | 1.150 | 0.407 |
| DCAN 250 | 2 | 0 | 5.550 | 0.955 | 0.675 |
| DCAN 500 | 2 | 0 | 5.859 | 1.964 | 1.389 |
| DCAN 600 | 2 | 0 | 5.958 | 2.601 | 1.839 |
| DCAN 700 | 2 | 0 | 3.503 | 0.713 | 0.504 |
| DCAN 800 | 2 | 0 | 5.582 | 0.906 | 0.640 |
| DCAN 900 | 2 | 0 | 6.651 | 0.308 | 0.218 |
| DCAN 1000 | 2 | 0 | 7.287 | 3.298 | 2.332 |
| DCAN 1250 | 4 | 0 | 5.795 | 1.286 | 0.643 |
| DCAN 1500 | 6 | 0 | 9.848 | 4.613 | 1.883 |
| DCAN 1600 | 6 | 0 | 11.341 | 2.903 | 1.185 |
| DCAN 1800 | 6 | 0 | 14.381 | 7.152 | 2.920 |
| DCAN 2000 | 6 | 0 | 14.925 | 7.947 | 3.244 |
| DCAN 2200 | 6 | 0 | 21.373 | 11.836 | 4.832 |
| DCAN 2400 | 6 | 0 | 32.670 | 13.014 | 5.313 |
| DCAN 2600 | 6 | 0 | 28.597 | 9.284 | 3.790 |
| DCAN 3000 | 6 | 0 | 47.039 | 14.420 | 5.887 |
| DCAN 3500 | 6 | 0 | 59.444 | 8.916 | 3.640 |

| Source of Variation | DF | SS | MS | F | P |
|----------------------------|-----------|-----------|-----------|----------|----------|
| Between Groups | 17 | 21970.765 | 1292.398 | 19.199 | <0.001 |
| Residual | 62 | 4173.536 | 67.315 | | |
| Total | 79 | 26144.301 | | | |

The differences in the mean values among the treatment groups are greater than would be expected by chance; there is a statistically significant difference (P = <0.001).

Power of performed test with alpha = 0.050: 1.000

Multiple Comparisons versus Control Group (Holm-Sidak method):

Overall significance level = 0.05

Comparisons for factor:

| Comparison | Diff of Means | t | P | P<0.050 |
|-------------------------|----------------------|----------|----------|-------------------|
| DCAN 0 µM vs. DCAN 3500 | 54.294 | 12.253 | <0.001 | Yes |
| DCAN 0 µM vs. DCAN 3000 | 41.889 | 9.454 | <0.001 | Yes |
| DCAN 0 µM vs. DCAN 2400 | 27.520 | 6.211 | <0.001 | Yes |
| DCAN 0 µM vs. DCAN 2600 | 23.447 | 5.292 | <0.001 | Yes |
| DCAN 0 µM vs. DCAN 2200 | 16.223 | 3.661 | 0.007 | Yes |
| DCAN 0 µM vs. DCAN 2000 | 9.774 | 2.206 | 0.316 | No |
| DCAN 0 µM vs. DCAN 1800 | 9.231 | 2.083 | 0.372 | No |
| DCAN 0 µM vs. DCAN 1600 | 6.191 | 1.397 | 0.840 | No |
| DCAN 0 µM vs. DCAN 1500 | 4.698 | 1.060 | 0.956 | No |
| DCAN 0 µM vs. DCAN 1000 | 2.137 | 0.329 | 1.000 | No |
| DCAN 0 µM vs. DCAN 700 | 1.647 | 0.254 | 1.000 | No |
| DCAN 0 µM vs. DCAN 900 | 1.501 | 0.231 | 1.000 | No |
| DCAN 0 µM vs. DCAN 1250 | 0.645 | 0.128 | 1.000 | No |
| DCAN 0 µM vs. DCAN 600 | 0.808 | 0.125 | 1.000 | No |
| DCAN 0 µM vs. DCAN 500 | 0.709 | 0.109 | 0.999 | No |
| DCAN 0 µM vs. DCAN 800 | 0.432 | 0.0666 | 0.997 | No |
| DCAN 0 µM vs. DCAN 250 | 0.400 | 0.0617 | 0.951 | No |

Table S12. One Way Analysis of Variance: Tribromoacetonitrile (TBAN) CHO cell genotoxicity SCGE %Tail DNA.

| Group Name | N | Missing | Mean | Std Dev | SEM |
|----------------------------|-----------|----------------|-------------|----------------|------------|
| TBAN 0 μ M | 8 | 0 | 1.097 | 0.866 | 0.306 |
| TBAN 12.5 | 2 | 0 | 1.134 | 0.785 | 0.555 |
| TBAN 25 | 9 | 0 | 41.263 | 21.067 | 7.022 |
| TBAN 30 | 9 | 0 | 43.533 | 17.197 | 5.732 |
| TBAN 40 | 7 | 0 | 57.707 | 19.929 | 7.532 |
| TBAN 50 | 7 | 0 | 66.584 | 20.603 | 7.787 |
| TBAN 75 | 5 | 0 | 75.683 | 15.594 | 6.974 |
| TBAN 100 | 4 | 0 | 80.542 | 4.796 | 2.398 |
| TBAN 125 | 2 | 0 | 88.778 | 6.257 | 4.425 |
| TBAN 150 | 4 | 0 | 84.061 | 6.629 | 3.315 |
| TBAN 160 | 4 | 0 | 92.461 | 1.588 | 0.794 |
| TBAN 175 | 2 | 0 | 88.321 | 3.930 | 2.779 |
| Source of Variation | DF | SS | MS | F | P |
| Between Groups | 11 | 51105.081 | 4645.916 | 19.602 | <0.001 |
| Residual | 51 | 12087.810 | 237.016 | | |
| Total | 62 | 63192.891 | | | |

The differences in the mean values among the treatment groups are greater than would be expected by chance; there is a statistically significant difference (P = <0.001).

Power of performed test with alpha = 0.050: 1.000

Multiple Comparisons versus Control Group (Holm-Sidak method):

Overall significance level = 0.05

Comparisons for factor:

| Comparison | Diff of Means | t | P | P<0.050 |
|------------------------------|----------------------|----------|----------|-------------------|
| TBAN 0 μ M vs. TBAN 160 | 91.364 | 9.691 | <0.001 | Yes |
| TBAN 0 μ M vs. TBAN 150 | 82.964 | 8.800 | <0.001 | Yes |
| TBAN 0 μ M vs. TBAN 75 | 74.586 | 8.498 | <0.001 | Yes |
| TBAN 0 μ M vs. TBAN 100 | 79.445 | 8.427 | <0.001 | Yes |
| TBAN 0 μ M vs. TBAN 50 | 65.487 | 8.219 | <0.001 | Yes |
| TBAN 0 μ M vs. TBAN 125 | 87.681 | 7.204 | <0.001 | Yes |
| TBAN 0 μ M vs. TBAN 175 | 87.224 | 7.166 | <0.001 | Yes |
| TBAN 0 μ M vs. TBAN 40 | 56.609 | 7.105 | <0.001 | Yes |
| TBAN 0 μ M vs. TBAN 30 | 42.436 | 5.673 | <0.001 | Yes |
| TBAN 0 μ M vs. TBAN 25 | 40.166 | 5.369 | <0.001 | Yes |
| TBAN 0 μ M vs. TBAN 12.5 | 0.0370 | 0.00304 | 0.998 | No |

Table S13. One Way Analysis of Variance: Trichloroacetonitrile (TCAN) CHO cell genotoxicity SCGE %Tail DNA.

| Group Name | N | Missing | Mean | Std Dev | SEM |
|----------------------------|-----------|----------------|-------------|----------------|------------|
| TCAN 0 μ M | 6 | 0 | 4.247 | 0.454 | 0.185 |
| TCAN 250 | 2 | 0 | 4.379 | 1.034 | 0.731 |
| TCAN 500 | 2 | 0 | 7.509 | 3.486 | 2.465 |
| TCAN 750 | 6 | 0 | 12.253 | 3.246 | 1.325 |
| TCAN 1000 | 6 | 0 | 25.298 | 4.041 | 1.650 |
| TCAN 1025 | 6 | 0 | 27.733 | 1.993 | 0.814 |
| TCAN 1050 | 6 | 0 | 32.247 | 1.920 | 0.784 |
| TCAN 1100 | 6 | 0 | 40.234 | 3.747 | 1.530 |
| TCAN 1500 | 2 | 0 | 69.713 | 4.380 | 3.097 |
| Source of Variation | DF | SS | MS | F | P |
| Between Groups | 8 | 10744.004 | 1343.001 | 160.401 | <0.001 |
| Residual | 33 | 276.302 | 8.373 | | |
| Total | 41 | 11020.306 | | | |

The differences in the mean values among the treatment groups are greater than would be expected by chance; there is a statistically significant difference (P = <0.001).

Power of performed test with alpha = 0.050: 1.000

Multiple Comparisons versus Control Group (Holm-Sidak method):

Overall significance level = 0.05

Comparisons for factor:

| Comparison | Diff of Means | t | P | P<0.050 |
|------------------------------|----------------------|----------|----------|-------------------|
| TCAN 0 μ M vs. TCAN 1500 | 65.466 | 27.709 | <0.001 | Yes |
| TCAN 0 μ M vs. TCAN 1100 | 35.987 | 21.541 | <0.001 | Yes |
| TCAN 0 μ M vs. TCAN 1050 | 28.000 | 16.760 | <0.001 | Yes |
| TCAN 0 μ M vs. TCAN 1025 | 23.486 | 14.058 | <0.001 | Yes |
| TCAN 0 μ M vs. TCAN 1000 | 21.051 | 12.601 | <0.001 | Yes |
| TCAN 0 μ M vs. TCAN 750 | 8.006 | 4.792 | <0.001 | Yes |
| TCAN 0 μ M vs. TCAN 500 | 3.261 | 1.380 | 0.322 | No |
| TCAN 0 μ M vs. TCAN 250 | 0.132 | 0.0557 | 0.956 | No |

Table S14. One Way Analysis of Variance: Bromodichloroacetonitrile (BDCAN) CHO cell genotoxicity SCGE %Tail DNA.

| Group Name | N | Missing | Mean | Std Dev | SEM |
|----------------------------|-----------|----------------|-------------|----------------|------------|
| 0 BDCAN μ M | 6 | 0 | 0.945 | 0.513 | 0.209 |
| 25 BDCAN | 2 | 0 | 2.175 | 0.0288 | 0.0204 |
| 50 BDCAN | 2 | 0 | 3.850 | 2.681 | 1.895 |
| 75 BDCAN | 2 | 0 | 3.979 | 2.673 | 1.890 |
| 100 BDCAN | 6 | 0 | 12.192 | 9.598 | 3.919 |
| 125 BDCAN | 6 | 0 | 22.021 | 5.456 | 2.227 |
| 150 BDCAN | 6 | 0 | 37.414 | 14.384 | 5.872 |
| 160 BDCAN | 7 | 0 | 47.150 | 14.066 | 5.316 |
| Source of Variation | DF | SS | MS | F | P |
| Between Groups | 7 | 11162.753 | 1594.679 | 16.245 | <0.001 |
| Residual | 29 | 2846.767 | 98.164 | | |
| Total | 36 | 14009.520 | | | |

The differences in the mean values among the treatment groups are greater than would be expected by chance; there is a statistically significant difference ($P = <0.001$).

Power of performed test with alpha = 0.050: 1.000

Multiple Comparisons versus Control Group (Holm-Sidak method):
Overall significance level = 0.05

Comparisons for factor:

| Comparison | Diff of Means | t | P | P<0.050 |
|-------------------------------|----------------------|----------|----------|-------------------|
| 0 BDCAN μ M vs. 160 BDCAN | 46.205 | 8.382 | <0.001 | Yes |
| 0 BDCAN μ M vs. 150 BDCAN | 36.469 | 6.375 | <0.001 | Yes |
| 0 BDCAN μ M vs. 125 BDCAN | 21.076 | 3.684 | 0.005 | Yes |
| 0 BDCAN μ M vs. 100 BDCAN | 11.247 | 1.966 | 0.216 | No |
| 0 BDCAN μ M vs. 75 BDCAN | 3.035 | 0.375 | 0.976 | No |
| 0 BDCAN μ M vs. 50 BDCAN | 2.905 | 0.359 | 0.923 | No |
| 0 BDCAN μ M vs. 25 BDCAN | 1.230 | 0.152 | 0.880 | No |

Table S15. One Way Analysis of Variance: Chlorodibromoacetonitrile (CDBAN) CHO cell genotoxicity SCGE %Tail DNA.

| Group Name | N | Missing | Mean | Std Dev | SEM |
|----------------------------|-----------|----------------|-------------|----------------|------------|
| 0 μ M CDBAN | 7 | 0 | 1.756 | 0.674 | 0.255 |
| 50 CDBAN | 3 | 0 | 1.427 | 0.496 | 0.286 |
| 75 CDBAN | 3 | 0 | 7.617 | 10.102 | 5.832 |
| 85 CDBAN | 4 | 0 | 4.677 | 2.303 | 1.152 |
| 90 CDBAN | 2 | 0 | 24.425 | 17.715 | 12.526 |
| 100 CDBAN | 6 | 0 | 45.406 | 19.735 | 8.057 |
| 125 CDBAN | 9 | 0 | 49.414 | 13.436 | 4.479 |
| 150 CDBAN | 10 | 0 | 51.228 | 9.701 | 3.068 |
| 160 CDBAN | 6 | 0 | 59.954 | 6.529 | 2.665 |
| Source of Variation | DF | SS | MS | F | P |
| Between Groups | 8 | 26100.281 | 3262.535 | 26.812 | <0.001 |
| Residual | 41 | 4988.874 | 121.680 | | |
| Total | 49 | 31089.156 | | | |

The differences in the mean values among the treatment groups are greater than would be expected by chance; there is a statistically significant difference (P = <0.001).

Power of performed test with alpha = 0.050: 1.000

Multiple Comparisons versus Control Group (Holm-Sidak method):
Overall significance level = 0.05

Comparisons for factor:

| Comparison | Diff of Means | t | P | P<0.050 |
|-------------------------------|----------------------|----------|----------|-------------------|
| 0 μ M CDBAN vs. 160 CDBAN | 58.198 | 9.483 | <0.001 | Yes |
| 0 μ M CDBAN vs. 150 CDBAN | 49.472 | 9.101 | <0.001 | Yes |
| 0 μ M CDBAN vs. 125 CDBAN | 47.658 | 8.573 | <0.001 | Yes |
| 0 μ M CDBAN vs. 100 CDBAN | 43.650 | 7.113 | <0.001 | Yes |
| 0 μ M CDBAN vs. 90 CDBAN | 22.669 | 2.563 | 0.055 | No |
| 0 μ M CDBAN vs. 75 CDBAN | 5.861 | 0.770 | 0.830 | No |
| 0 μ M CDBAN vs. 85 CDBAN | 2.921 | 0.422 | 0.894 | No |
| 0 μ M CDBAN vs. 50 CDBAN | 0.329 | 0.0432 | 0.966 | No |

Table S16. One Way Analysis of Variance: 10 HAN GTI Comparisons.

| Group Name | N | Missing | Mean | Std Dev | SEM |
|----------------------------|-----------|----------------|-------------|----------------|------------|
| IAN GTI | 11 | 0 | 29.347 | 2.099 | 0.633 |
| BAN GTI | 11 | 0 | 20.854 | 1.572 | 0.474 |
| DBAN GTI | 15 | 0 | 28.210 | 1.141 | 0.295 |
| BCAN GTI | 11 | 0 | 4.110 | 0.752 | 0.227 |
| TBAN GTI | 11 | 0 | 28.372 | 7.550 | 2.276 |
| CAN GTI | 11 | 0 | 1.997 | 0.444 | 0.134 |
| DCAN GTI | 11 | 0 | 0.327 | 0.0292 | 0.00881 |
| TCAN GTI | 11 | 0 | 0.844 | 0.0342 | 0.0103 |
| BDCAN GTI | 11 | 0 | 6.129 | 0.180 | 0.0544 |
| CDBAN GTI | 11 | 0 | 7.187 | 0.484 | 0.146 |
| Source of Variation | DF | SS | MS | F | P |
| Between Groups | 9 | 16162.572 | 1795.841 | 279.887 | <0.001 |
| Residual | 104 | 667.295 | 6.416 | | |
| Total | 113 | 16829.867 | | | |

The differences in the mean values among the treatment groups are greater than would be expected by chance; there is a statistically significant difference (P = <0.001).. Power of performed test with alpha = 0.050: 1.000

All Pairwise Multiple Comparison Procedures (Holm-Sidak method):
Overall significance level = 0.05

Comparisons for factor:

| Comparison | Diff of Means | t | P | P<0.050 |
|------------------------|----------------------|----------|----------|-------------------|
| DBAN GTI vs. DCAN GTI | 27.883 | 27.730 | <0.001 | Yes |
| DBAN GTI vs. TCAN GTI | 27.366 | 27.216 | <0.001 | Yes |
| IAN GTI vs. DCAN GTI | 29.020 | 26.868 | <0.001 | Yes |
| IAN GTI vs. TCAN GTI | 28.503 | 26.390 | <0.001 | Yes |
| DBAN GTI vs. CAN GTI | 26.213 | 26.069 | <0.001 | Yes |
| TBAN GTI vs. DCAN GTI | 28.045 | 25.965 | <0.001 | Yes |
| TBAN GTI vs. TCAN GTI | 27.528 | 25.487 | <0.001 | Yes |
| IAN GTI vs. CAN GTI | 27.350 | 25.322 | <0.001 | Yes |
| TBAN GTI vs. CAN GTI | 26.375 | 24.419 | <0.001 | Yes |
| DBAN GTI vs. BCAN GTI | 24.100 | 23.967 | <0.001 | Yes |
| IAN GTI vs. BCAN GTI | 25.237 | 23.365 | <0.001 | Yes |
| TBAN GTI vs. BCAN GTI | 24.262 | 22.463 | <0.001 | Yes |
| DBAN GTI vs. BDCAN GTI | 22.081 | 21.960 | <0.001 | Yes |
| IAN GTI vs. BDCAN GTI | 23.218 | 21.496 | <0.001 | Yes |
| DBAN GTI vs. CDBAN GTI | 21.023 | 20.908 | <0.001 | Yes |
| TBAN GTI vs. BDCAN GTI | 22.243 | 20.593 | <0.001 | Yes |
| IAN GTI vs. CDBAN GTI | 22.160 | 20.517 | <0.001 | Yes |
| TBAN GTI vs. CDBAN GTI | 21.185 | 19.614 | <0.001 | Yes |
| BAN GTI vs. DCAN GTI | 20.527 | 19.005 | <0.001 | Yes |
| BAN GTI vs. TCAN GTI | 20.010 | 18.526 | <0.001 | Yes |
| BAN GTI vs. CAN GTI | 18.857 | 17.458 | <0.001 | Yes |
| BAN GTI vs. BCAN GTI | 16.744 | 15.502 | <0.001 | Yes |
| BAN GTI vs. BDCAN GTI | 14.725 | 13.633 | <0.001 | Yes |
| BAN GTI vs. CDBAN GTI | 13.667 | 12.654 | <0.001 | Yes |
| IAN GTI vs. BAN GTI | 8.493 | 7.863 | <0.001 | Yes |
| DBAN GTI vs. BAN GTI | 7.356 | 7.316 | <0.001 | Yes |
| TBAN GTI vs. BAN GTI | 7.518 | 6.961 | <0.001 | Yes |
| CDBAN GTI vs. DCAN GTI | 6.860 | 6.351 | <0.001 | Yes |

| | | | | |
|-------------------------|-------|-------|--------|-----|
| CDBAN GTI vs. TCAN GTI | 6.343 | 5.873 | <0.001 | Yes |
| BDCAN GTI vs. DCAN GTI | 5.802 | 5.372 | <0.001 | Yes |
| BDCAN GTI vs. TCAN GTI | 5.285 | 4.893 | <0.001 | Yes |
| CDBAN GTI vs. CAN GTI | 5.190 | 4.805 | <0.001 | Yes |
| BDCAN GTI vs. CAN GTI | 4.132 | 3.826 | 0.003 | Yes |
| BCAN GTI vs. DCAN GTI | 3.783 | 3.503 | 0.008 | Yes |
| BCAN GTI vs. TCAN GTI | 3.266 | 3.024 | 0.034 | Yes |
| CDBAN GTI vs. BCAN GTI | 3.076 | 2.848 | 0.050 | Yes |
| BCAN GTI vs. CAN GTI | 2.113 | 1.957 | 0.388 | No |
| BDCAN GTI vs. BCAN GTI | 2.019 | 1.869 | 0.413 | No |
| CAN GTI vs. DCAN GTI | 1.670 | 1.546 | 0.608 | No |
| IAN GTI vs. DBAN GTI | 1.137 | 1.131 | 0.837 | No |
| CAN GTI vs. TCAN GTI | 1.153 | 1.068 | 0.817 | No |
| CDBAN GTI vs. BDCAN GTI | 1.058 | 0.979 | 0.798 | No |
| IAN GTI vs. TBAN GTI | 0.975 | 0.903 | 0.748 | No |
| TCAN GTI vs. DCAN GTI | 0.517 | 0.479 | 0.865 | No |
| TBAN GTI vs. DBAN GTI | 0.162 | 0.161 | 0.872 | No |

N-Acetylcysteine thiol reactivity analyses. The *N*-acetylcysteine (NAC) thiol reactivity high throughput assay is a screen to identify potential adverse biological effects.¹⁷⁻¹⁹ The cysteine thiol is a reductant against reactive toxicants.^{20, 21} HANs were reacted with NAC for 20 min in a volume of 50 μ L, followed by the addition of 50 μ L of 5,5-dithiobis (2-nitrobenzoic acid) (DTNB, 1 mM). Each experiment included concurrent negative controls (Tris buffer and NAC), positive controls (Tris buffer, NAC and maleimide), and corresponding blanks to correct for the background A_{412} . After incubation with NAC (20 min, 30 °C shaking, dark conditions), DTNB was added to quantify the available thiol groups. The microplate was analyzed at 412 nm using a Molecular Devices Spectramax Paradigm multimode microplate reader after linear shaking of 10 s. The data were saved in an Excel spreadsheet. The A_{412} values for each well were blank-corrected. The blank-corrected negative controls were averaged. This value was divided into the individual A_{412} values for each treatment group $\times 100$; the data were expressed as the percent of the concurrent negative controls. Using these normalized data, we generated concentration-response curves. Regression analyses were used to calculate the EC₅₀ values, the effective HAN concentration that induced a reduction in the NAC thiol response by 50% compared to the concurrent negative controls.

The concentration-response graphs illustrating the NAC thiol reactivity of the HANs are presented in Figures S21 to S30.

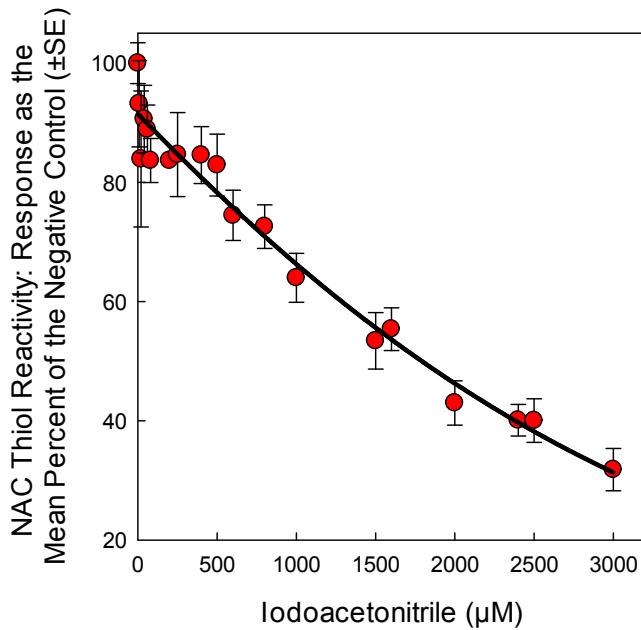


Figure S21. NAC thiol reactivity concentration-response curve for IAN. Mean (\pm SE) EC₅₀ value was 1714.47 \pm 70.15 μ M.

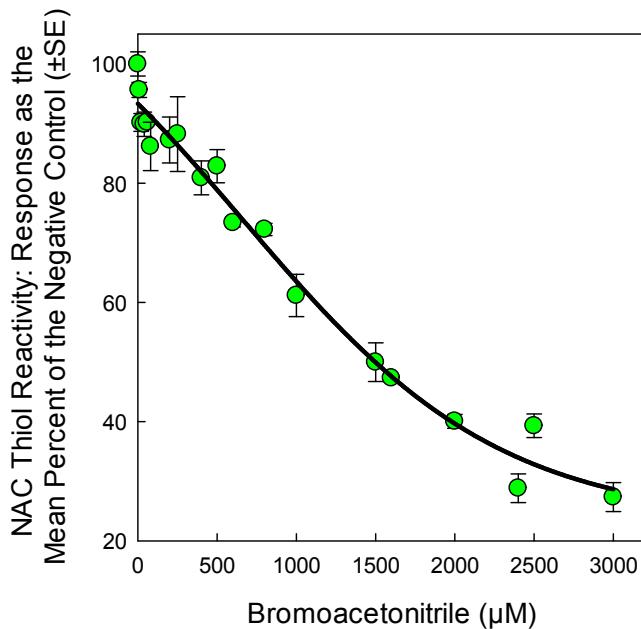


Figure S22. NAC thiol reactivity concentration-response curve for BAN. Mean (\pm SE) EC₅₀ value was 1502.96 \pm 16.88 μ M.

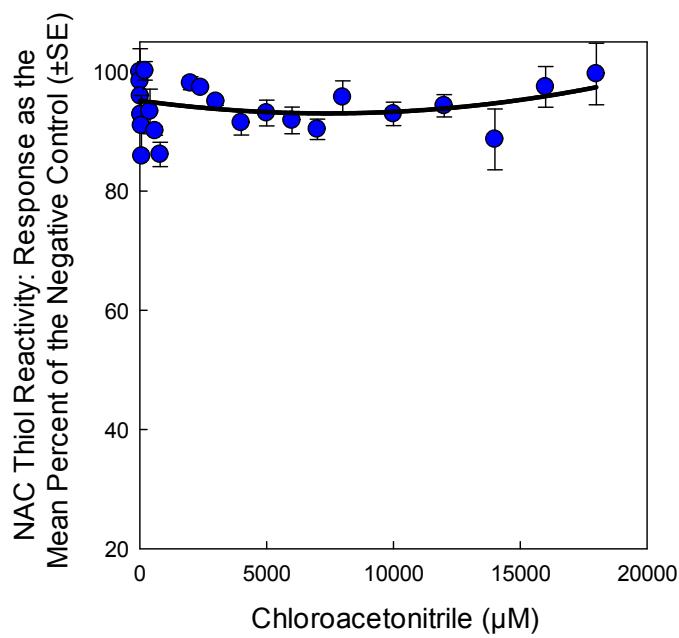


Figure S23. NAC thiol reactivity concentration-response curve for CAN.

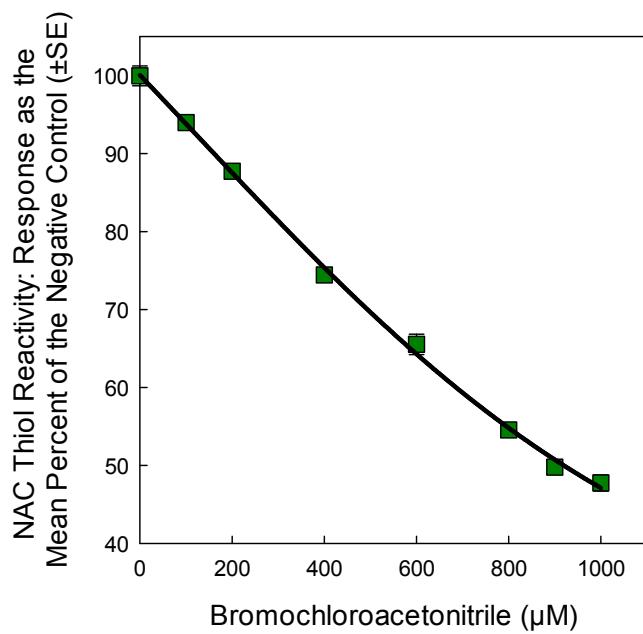


Figure S24. NAC thiol reactivity concentration-response curve for BCAN. Mean (\pm SE) EC₅₀ value was $912.84 \pm 11.11 \mu$ M.

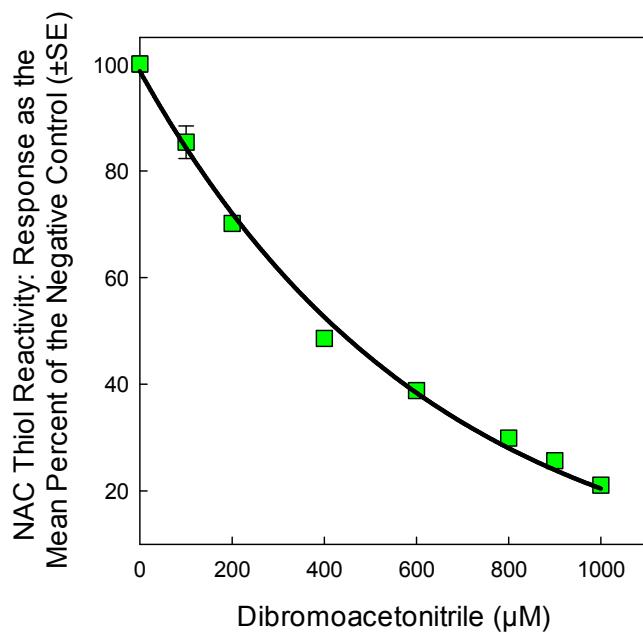


Figure S25. NAC thiol reactivity concentration-response curve for DBAN. Mean (\pm SE) EC₅₀ value was $403.78 \pm 3.92 \mu$ M.

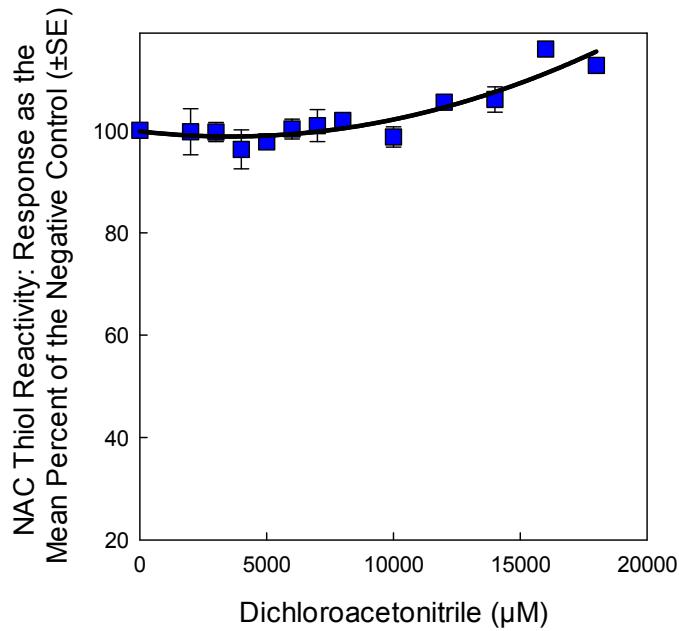


Figure S26. NAC thiol reactivity concentration-response curve for DCAN.

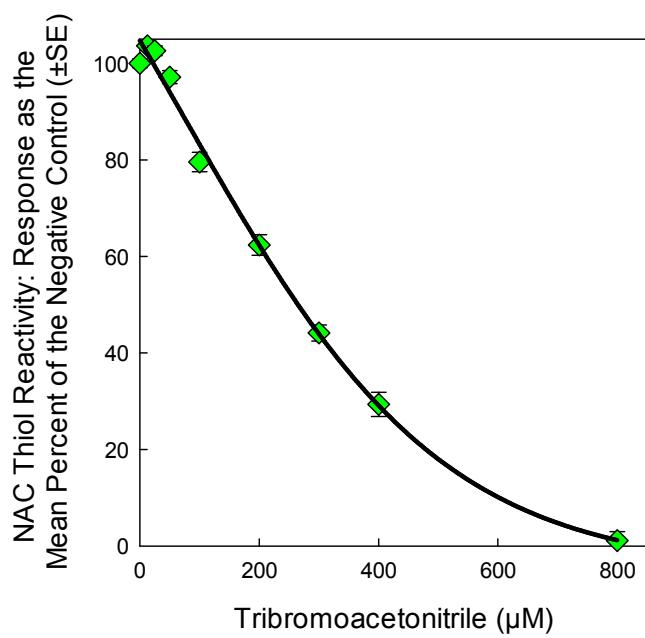


Figure S27. NAC thiol reactivity concentration-response curve for TBAN. Mean (\pm SE) EC₅₀ value was $263.06 \pm 6.73 \mu\text{M}$.

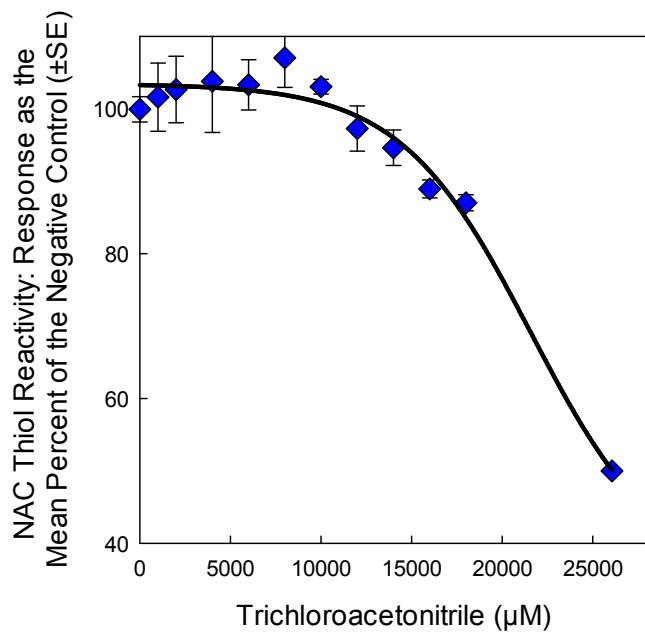


Figure S28. NAC thiol reactivity concentration-response curve for TCAN. The extrapolated mean (\pm SE) EC₅₀ value was $26148.67 \pm 9.83 \mu\text{M}$.

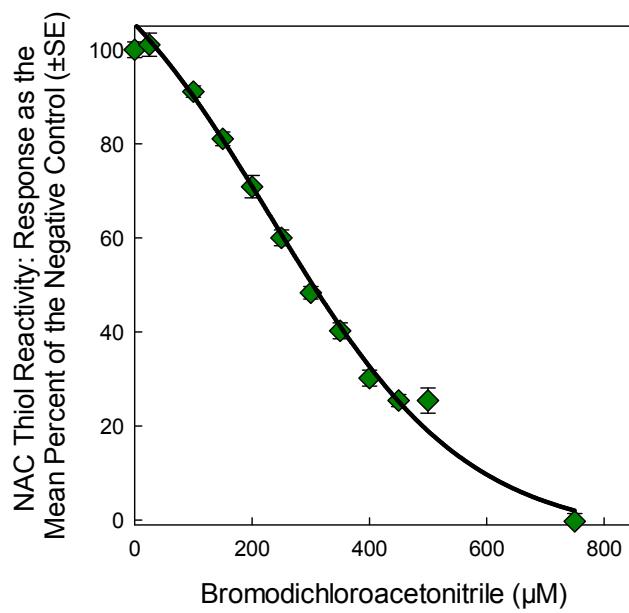


Figure S29. NAC thiol reactivity concentration-response curve for BDCAN. The extrapolated mean (\pm SE) EC₅₀ value was 302.31 \pm 2.04 μ M.

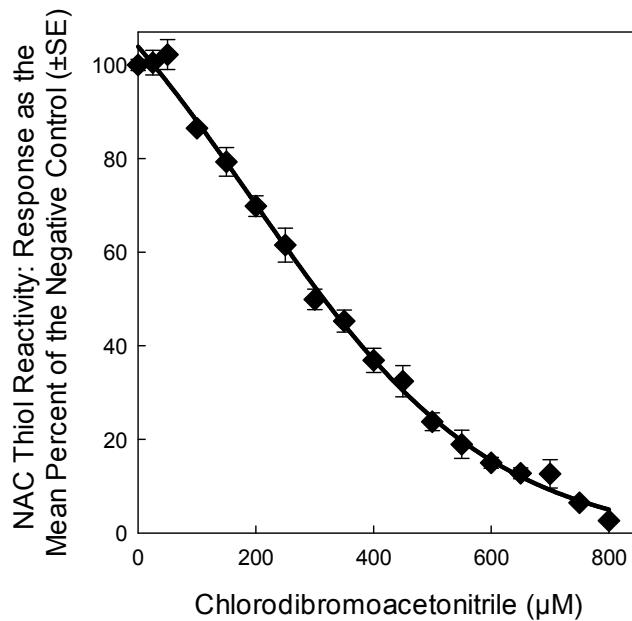


Figure S30. NAC thiol reactivity concentration-response curve for CDBAN. The extrapolated mean (\pm SE) EC₅₀ value was 314.43 \pm 5.74 μ M.

Table S17. One Way Analysis of Variance: Iodoacetonitrile (IAN) thiol reactivity response as the percent of the negative control.

| Group Name | N | Missing | Mean | Std Dev | SEM |
|-------------------|----------|----------------|-------------|----------------|------------|
| 0 IAN | 5 | 0 | 100.000 | 0.000 | 0.000 |
| 8 IAN | 5 | 0 | 94.640 | 7.924 | 3.544 |
| 20 IAN | 5 | 0 | 81.632 | 12.495 | 5.588 |
| 40 IAN | 5 | 0 | 89.466 | 6.206 | 2.775 |
| 60 IAN | 5 | 0 | 88.077 | 4.471 | 1.999 |
| 80 IAN | 5 | 0 | 82.926 | 4.048 | 1.810 |
| 200 IAN | 5 | 0 | 83.655 | 0.000 | 0.000 |
| 250 IAN | 5 | 0 | 89.427 | 10.813 | 4.836 |
| 400 IAN | 5 | 0 | 85.529 | 5.233 | 2.340 |
| 500 IAN | 5 | 0 | 84.905 | 7.103 | 3.176 |
| 600 IAN | 5 | 0 | 75.314 | 4.610 | 2.062 |
| 800 IAN | 5 | 0 | 73.314 | 3.998 | 1.788 |
| 1000 IAN | 5 | 0 | 60.759 | 6.692 | 2.993 |
| 1500 IAN | 5 | 0 | 55.306 | 6.387 | 2.856 |
| 1600 IAN | 5 | 0 | 54.652 | 3.937 | 1.761 |
| 2000 IAN | 5 | 0 | 41.919 | 5.755 | 2.574 |
| 2400 IAN | 5 | 0 | 40.621 | 2.903 | 1.298 |
| 2500 IAN | 5 | 0 | 37.108 | 6.007 | 2.686 |
| 3000 IAN | 5 | 0 | 32.876 | 4.591 | 2.053 |

| Source of Variation | DF | SS | MS | F | P |
|----------------------------|-----------|-----------|-----------|----------|----------|
| Between Groups | 18 | 41021.732 | 2278.985 | 59.679 | <0.001 |
| Residual | 76 | 2902.250 | 38.187 | | |
| Total | 94 | 43923.982 | | | |

The differences in the mean values among the treatment groups are greater than would be expected by chance; there is a statistically significant difference (P = <0.001). Power of performed test with alpha = 0.050: 1.000

Multiple Comparisons versus Control Group (Holm-Sidak method):

Overall significance level = 0.05

Comparisons for factor:

| Comparison | Diff of Means | t | P | P<0.050 |
|--------------------|----------------------|----------|----------|-------------------|
| 0 IAN vs. 3000 IAN | 67.124 | 17.175 | <0.001 | Yes |
| 0 IAN vs. 2500 IAN | 62.892 | 16.092 | <0.001 | Yes |
| 0 IAN vs. 2400 IAN | 59.379 | 15.193 | <0.001 | Yes |
| 0 IAN vs. 2000 IAN | 58.081 | 14.861 | <0.001 | Yes |
| 0 IAN vs. 1600 IAN | 45.348 | 11.603 | <0.001 | Yes |
| 0 IAN vs. 1500 IAN | 44.694 | 11.436 | <0.001 | Yes |
| 0 IAN vs. 1000 IAN | 39.241 | 10.040 | <0.001 | Yes |
| 0 IAN vs. 800 IAN | 26.686 | 6.828 | <0.001 | Yes |
| 0 IAN vs. 600 IAN | 24.686 | 6.316 | <0.001 | Yes |
| 0 IAN vs. 20 IAN | 18.368 | 4.700 | <0.001 | Yes |
| 0 IAN vs. 80 IAN | 17.074 | 4.369 | <0.001 | Yes |
| 0 IAN vs. 200 IAN | 16.345 | 4.182 | <0.001 | Yes |
| 0 IAN vs. 500 IAN | 15.095 | 3.862 | 0.001 | Yes |
| 0 IAN vs. 400 IAN | 14.471 | 3.703 | 0.002 | Yes |
| 0 IAN vs. 60 IAN | 11.923 | 3.051 | 0.013 | Yes |
| 0 IAN vs. 250 IAN | 10.573 | 2.705 | 0.025 | Yes |
| 0 IAN vs. 40 IAN | 10.534 | 2.695 | 0.017 | Yes |
| 0 IAN vs. 8 IAN | 5.360 | 1.372 | 0.174 | No |

Table S18. One Way Analysis of Variance: Bromoacetonitrile (BAN) thiol reactivity response as the percent of the negative control.

| Group Name | N | Missing | Mean | Std Dev | SEM |
|------------|---|---------|--------|---------|-------|
| 0 BAN | 6 | 0 | 99.997 | 4.971 | 2.029 |
| 8 BAN | 2 | 0 | 95.643 | 1.776 | 1.256 |
| 20 BAN | 2 | 0 | 90.184 | 2.104 | 1.488 |
| 40 BAN | 2 | 0 | 89.884 | 2.883 | 2.039 |
| 60 BAN | 2 | 0 | 90.217 | 0.608 | 0.430 |
| 80 BAN | 2 | 0 | 86.169 | 5.746 | 4.063 |
| 200 BAN | 2 | 0 | 87.246 | 5.438 | 3.845 |
| 250 BAN | 3 | 0 | 88.225 | 10.845 | 6.261 |
| 400 BAN | 2 | 0 | 80.894 | 4.024 | 2.845 |
| 500 BAN | 4 | 0 | 82.856 | 5.575 | 2.787 |
| 600 BAN | 2 | 0 | 73.367 | 1.004 | 0.710 |
| 800 BAN | 2 | 0 | 72.242 | 1.448 | 1.024 |
| 1000 BAN | 4 | 0 | 61.187 | 7.089 | 3.544 |
| 1500 BAN | 3 | 0 | 49.959 | 5.613 | 3.241 |
| 1600 BAN | 2 | 0 | 47.309 | 0.717 | 0.507 |
| 2000 BAN | 4 | 0 | 40.048 | 2.345 | 1.173 |
| 2400 BAN | 2 | 0 | 28.826 | 3.395 | 2.401 |
| 2500 BAN | 3 | 0 | 39.303 | 3.419 | 1.974 |
| 3000 BAN | 4 | 0 | 27.345 | 4.842 | 2.421 |

| Source of Variation | DF | SS | MS | F | P |
|---------------------|----|-----------|----------|--------|--------|
| Between Groups | 18 | 31768.426 | 1764.913 | 67.714 | <0.001 |
| Residual | 34 | 886.188 | 26.064 | | |
| Total | 52 | 32654.614 | | | |

The differences in the mean values among the treatment groups are greater than would be expected by chance; there is a statistically significant difference (P = <0.001). Power of performed test with alpha = 0.050: 1.000

Multiple Comparisons versus Control Group (Holm-Sidak method):

Overall significance level = 0.05

Comparisons for factor:

| Comparison | Diff of Means | t | P | P<0.050 |
|--------------------|---------------|--------|--------|---------|
| 0 BAN vs. 3000 BAN | 72.652 | 22.046 | <0.001 | Yes |
| 0 BAN vs. 2000 BAN | 59.950 | 18.191 | <0.001 | Yes |
| 0 BAN vs. 2400 BAN | 71.171 | 17.074 | <0.001 | Yes |
| 0 BAN vs. 2500 BAN | 60.695 | 16.813 | <0.001 | Yes |
| 0 BAN vs. 1500 BAN | 50.038 | 13.861 | <0.001 | Yes |
| 0 BAN vs. 1600 BAN | 52.688 | 12.640 | <0.001 | Yes |
| 0 BAN vs. 1000 BAN | 38.810 | 11.777 | <0.001 | Yes |
| 0 BAN vs. 800 BAN | 27.756 | 6.658 | <0.001 | Yes |
| 0 BAN vs. 600 BAN | 26.630 | 6.388 | <0.001 | Yes |
| 0 BAN vs. 500 BAN | 17.141 | 5.201 | <0.001 | Yes |
| 0 BAN vs. 400 BAN | 19.103 | 4.583 | <0.001 | Yes |
| 0 BAN vs. 80 BAN | 13.828 | 3.317 | 0.015 | Yes |
| 0 BAN vs. 250 BAN | 11.772 | 3.261 | 0.015 | Yes |
| 0 BAN vs. 200 BAN | 12.751 | 3.059 | 0.021 | Yes |
| 0 BAN vs. 40 BAN | 10.113 | 2.426 | 0.080 | No |
| 0 BAN vs. 20 BAN | 9.814 | 2.354 | 0.072 | No |
| 0 BAN vs. 60 BAN | 9.780 | 2.346 | 0.049 | Yes |
| 0 BAN vs. 8 BAN | 4.355 | 1.045 | 0.304 | No |

Table S19. One Way Analysis of Variance: Chloroacetonitrile (CAN) thiol reactivity response as the percent of the negative control.

| Group Name | N | Missing | Mean | Std Dev | SEM |
|----------------------------|-----------|----------------|-------------|----------------|------------|
| 0 CAN | 8 | 0 | 99.984 | 2.240 | 0.792 |
| 8 CAN | 2 | 0 | 98.432 | 7.681 | 5.431 |
| 20 CAN | 2 | 0 | 95.940 | 0.529 | 0.374 |
| 40 CAN | 2 | 0 | 92.835 | 1.087 | 0.768 |
| 60 CAN | 2 | 0 | 90.987 | 7.137 | 5.047 |
| 80 CAN | 2 | 0 | 85.841 | 0.374 | 0.265 |
| 200 CAN | 2 | 0 | 100.156 | 2.188 | 1.547 |
| 400 CAN | 2 | 0 | 93.359 | 5.294 | 3.744 |
| 600 CAN | 2 | 0 | 90.099 | 1.050 | 0.742 |
| 800 CAN | 2 | 0 | 86.153 | 2.886 | 2.040 |
| 1600 CAN | 2 | 0 | 107.954 | 0.000 | 0.000 |
| 2000 CAN | 3 | 0 | 98.094 | 1.896 | 1.095 |
| 2400 CAN | 2 | 0 | 97.362 | 0.000 | 0.000 |
| 3000 CAN | 3 | 0 | 95.029 | 1.138 | 0.657 |
| 4000 CAN | 3 | 0 | 91.461 | 3.579 | 2.066 |
| 5000 CAN | 3 | 0 | 93.095 | 3.739 | 2.159 |
| 6000 CAN | 3 | 0 | 91.853 | 3.864 | 2.231 |
| 7000 CAN | 3 | 0 | 90.364 | 2.971 | 1.715 |
| 8000 CAN | 6 | 0 | 95.747 | 6.696 | 2.734 |
| 10000 CAN | 3 | 0 | 92.945 | 3.402 | 1.964 |
| 12000 CAN | 3 | 0 | 94.295 | 3.245 | 1.874 |
| 14000 CAN | 3 | 0 | 88.672 | 8.823 | 5.094 |
| 16000 CAN | 3 | 0 | 97.471 | 5.899 | 3.406 |
| 18000 CAN | 3 | 0 | 99.639 | 8.949 | 5.167 |
| Source of Variation | DF | SS | MS | F | P |
| Between Groups | 23 | 1434.989 | 62.391 | 2.944 | <0.001 |
| Residual | 45 | 953.614 | 21.191 | | |
| Total | 68 | 2388.602 | | | |

The differences in the mean values among the treatment groups are greater than would be expected by chance; there is a statistically significant difference (P = <0.001).

NOTE No concentration response observed. Interpretation, no significant effect on NAC thiol reactivity observed.

Power of performed test with alpha = 0.050: 0.947

Multiple Comparisons versus Control Group (Holm-Sidak method):
Overall significance level = 0.05

Comparisons for factor:

| Comparison | Diff of Means | t | P | P<0.050 |
|---------------------|----------------------|----------|----------|-------------------|
| 0 CAN vs. 80 CAN | 14.142 | 3.886 | 0.008 | Yes |
| 0 CAN vs. 800 CAN | 13.831 | 3.800 | 0.009 | Yes |
| 0 CAN vs. 14000 CAN | 11.311 | 3.630 | 0.015 | Yes |
| 0 CAN vs. 7000 CAN | 9.620 | 3.087 | 0.067 | No |
| 0 CAN vs. 4000 CAN | 8.523 | 2.735 | 0.156 | No |
| 0 CAN vs. 600 CAN | 9.885 | 2.716 | 0.155 | No |
| 0 CAN vs. 6000 CAN | 8.131 | 2.609 | 0.190 | No |

| | | | | |
|---------------------|-------|--------|-------|----|
| 0 CAN vs. 60 CAN | 8.997 | 2.472 | 0.243 | No |
| 0 CAN vs. 10000 CAN | 7.039 | 2.258 | 0.355 | No |
| 0 CAN vs. 5000 CAN | 6.888 | 2.210 | 0.368 | No |
| 0 CAN vs. 1600 CAN | 7.971 | 2.190 | 0.360 | No |
| 0 CAN vs. 40 CAN | 7.149 | 1.964 | 0.497 | No |
| 0 CAN vs. 12000 CAN | 5.688 | 1.825 | 0.574 | No |
| 0 CAN vs. 400 CAN | 6.624 | 1.820 | 0.543 | No |
| 0 CAN vs. 8000 CAN | 4.237 | 1.704 | 0.594 | No |
| 0 CAN vs. 3000 CAN | 4.954 | 1.590 | 0.637 | No |
| 0 CAN vs. 20 CAN | 4.044 | 1.111 | 0.892 | No |
| 0 CAN vs. 16000 CAN | 2.513 | 0.806 | 0.964 | No |
| 0 CAN vs. 2400 CAN | 2.621 | 0.720 | 0.960 | No |
| 0 CAN vs. 2000 CAN | 1.890 | 0.606 | 0.958 | No |
| 0 CAN vs. 8 CAN | 1.552 | 0.426 | 0.965 | No |
| 0 CAN vs. 18000 CAN | 0.345 | 0.111 | 0.992 | No |
| 0 CAN vs. 200 CAN | 0.172 | 0.0473 | 0.962 | No |

Table S20. One Way Analysis of Variance: Bromochloroacetonitrile (BCAN) thiol reactivity response as the percent of the negative control.

| Group Name | N | Missing | Mean | Std Dev | SEM |
|----------------------------|-----------|----------------|-------------|----------------|------------|
| 0 BCAN | 4 | 0 | 99.971 | 2.548 | 1.274 |
| 100 BCAN | 3 | 0 | 93.972 | 0.298 | 0.172 |
| 200 BCAN | 3 | 0 | 87.731 | 1.383 | 0.798 |
| 400 BCAN | 3 | 0 | 74.443 | 1.368 | 0.790 |
| 600 BCAN | 3 | 0 | 65.519 | 2.260 | 1.305 |
| 800 BCAN | 3 | 0 | 54.553 | 1.575 | 0.909 |
| 900 BCAN | 3 | 0 | 49.780 | 1.559 | 0.900 |
| 1000 BCAN | 3 | 0 | 47.750 | 1.782 | 1.029 |
| Source of Variation | DF | SS | MS | F | P |
| Between Groups | 7 | 9604.957 | 1372.137 | 435.145 | <0.001 |
| Residual | 17 | 53.606 | 3.153 | | |
| Total | 24 | 9658.562 | | | |

The differences in the mean values among the treatment groups are greater than would be expected by chance; there is a statistically significant difference (P = <0.001).

Power of performed test with alpha = 0.050: 1.000

Multiple Comparisons versus Control Group (Holm-Sidak method):
Overall significance level = 0.05

Comparisons for factor:

| Comparison | Diff of Means | t | P | P<0.050 |
|----------------------|----------------------|----------|----------|-------------------|
| 0 BCAN vs. 1000 BCAN | 52.222 | 38.504 | <0.001 | Yes |
| 0 BCAN vs. 900 BCAN | 50.192 | 37.008 | <0.001 | Yes |
| 0 BCAN vs. 800 BCAN | 45.418 | 33.488 | <0.001 | Yes |
| 0 BCAN vs. 600 BCAN | 34.453 | 25.403 | <0.001 | Yes |
| 0 BCAN vs. 400 BCAN | 25.528 | 18.823 | <0.001 | Yes |
| 0 BCAN vs. 200 BCAN | 12.241 | 9.025 | <0.001 | Yes |
| 0 BCAN vs. 100 BCAN | 6.000 | 4.424 | <0.001 | Yes |

Table S21. One Way Analysis of Variance: Dibromoacetonitrile (DBAN) thiol reactivity response as the percent of the negative control.

| Group Name | N | Missing | Mean | Std Dev | SEM |
|----------------------------|-----------|----------------|-------------|----------------|------------|
| 0 DBAN | 4 | 0 | 100.040 | 2.949 | 1.475 |
| 100 DBAN | 3 | 0 | 85.378 | 5.280 | 3.048 |
| 200 DBAN | 3 | 0 | 70.173 | 1.418 | 0.819 |
| 400 DBAN | 3 | 0 | 48.588 | 1.382 | 0.798 |
| 600 DBAN | 3 | 0 | 38.819 | 0.293 | 0.169 |
| 800 DBAN | 3 | 0 | 29.912 | 1.548 | 0.894 |
| 900 DBAN | 3 | 0 | 25.677 | 1.000 | 0.578 |
| 1000 DBAN | 3 | 0 | 21.082 | 0.818 | 0.472 |
| Source of Variation | DF | SS | MS | F | P |
| Between Groups | 7 | 20391.347 | 2913.050 | 505.400 | <0.001 |
| Residual | 17 | 97.986 | 5.764 | | |
| Total | 24 | 20489.333 | | | |

The differences in the mean values among the treatment groups are greater than would be expected by chance; there is a statistically significant difference ($P = <0.001$). Power of performed test with alpha = 0.050: 1.000

Multiple Comparisons versus Control Group (Holm-Sidak method):

Overall significance level = 0.05

Comparisons for factor:

| Comparison | Diff of Means | t | P | P<0.050 |
|----------------------|----------------------|----------|----------|-------------------|
| 0 DBAN vs. 1000 DBAN | 78.959 | 43.061 | <0.001 | Yes |
| 0 DBAN vs. 900 DBAN | 74.363 | 40.555 | <0.001 | Yes |
| 0 DBAN vs. 800 DBAN | 70.129 | 38.245 | <0.001 | Yes |
| 0 DBAN vs. 600 DBAN | 61.222 | 33.388 | <0.001 | Yes |
| 0 DBAN vs. 400 DBAN | 51.452 | 28.060 | <0.001 | Yes |
| 0 DBAN vs. 200 DBAN | 29.868 | 16.289 | <0.001 | Yes |
| 0 DBAN vs. 100 DBAN | 14.662 | 7.996 | <0.001 | Yes |

Table S22. One Way Analysis of Variance: Dichloroacetonitrile (DCAN) thiol reactivity response as the percent of the negative control.

| Group Name | N | Missing | Mean | Std Dev | SEM |
|----------------------------|-----------|----------------|-------------|----------------|------------|
| 0 DCAN | 9 | 0 | 100.024 | 10.540 | 3.513 |
| 2000 DCAN | 3 | 0 | 99.751 | 7.770 | 4.486 |
| 3000 DCAN | 3 | 0 | 99.700 | 3.220 | 1.859 |
| 4000 DCAN | 3 | 0 | 96.311 | 6.572 | 3.794 |
| 5000 DCAN | 3 | 0 | 97.769 | 2.245 | 1.296 |
| 6000 DCAN | 3 | 0 | 100.286 | 3.357 | 1.938 |
| 7000 DCAN | 3 | 0 | 100.949 | 5.423 | 3.131 |
| 8000 DCAN | 6 | 0 | 101.971 | 2.518 | 1.028 |
| 10000 DCAN | 3 | 0 | 98.737 | 3.435 | 1.983 |
| 12000 DCAN | 3 | 0 | 105.525 | 2.006 | 1.158 |
| 14000 DCAN | 3 | 0 | 106.053 | 4.293 | 2.479 |
| 16000 DCAN | 4 | 0 | 110.178 | 11.588 | 5.794 |
| 18000 DCAN | 4 | 0 | 105.787 | 13.892 | 6.946 |
| Source of Variation | DF | SS | MS | F | P |
| Between Groups | 12 | 670.924 | 55.910 | 0.903 | 0.552 |
| Residual | 37 | 2290.002 | 61.892 | | |
| Total | 49 | 2960.926 | | | |

The differences in the mean values among the treatment groups are not great enough to exclude the possibility that the difference is due to random sampling variability; there is not a statistically significant difference (P = 0.552).

Table S23. One Way Analysis of Variance: Tribromoacetonitrile (TBAN) thiol reactivity response as the percent of the negative control.

| Group Name | N | Missing | Mean | Std Dev | SEM |
|----------------------------|-----------|----------------|-------------|----------------|------------|
| 0 TBAN | 6 | 0 | 100.018 | 2.250 | 0.918 |
| 12.5 TBAN | 3 | 0 | 103.624 | 1.432 | 0.827 |
| 25 TBAN | 6 | 0 | 102.615 | 2.502 | 1.021 |
| 50 TBAN | 6 | 0 | 97.162 | 3.320 | 1.355 |
| 100 TBAN | 6 | 0 | 79.562 | 4.924 | 2.010 |
| 200 TBAN | 6 | 0 | 62.387 | 5.214 | 2.129 |
| 300 TBAN | 3 | 0 | 44.157 | 2.911 | 1.680 |
| 400 TBAN | 6 | 0 | 29.352 | 6.130 | 2.502 |
| 800 TBAN | 6 | 0 | 1.121 | 4.588 | 1.873 |
| Source of Variation | DF | SS | MS | F | P |
| Between Groups | 8 | 60741.593 | 7592.699 | 433.527 | <0.001 |
| Residual | 39 | 683.038 | 17.514 | | |
| Total | 47 | 61424.631 | | | |

The differences in the mean values among the treatment groups are greater than would be expected by chance; there is a statistically significant difference (P = <0.001).

Power of performed test with alpha = 0.050: 1.000

Multiple Comparisons versus Control Group (Holm-Sidak method):

Overall significance level = 0.05

Comparisons for factor:

| Comparison | Diff of Means | t | P | P<0.050 |
|----------------------|----------------------|----------|----------|-------------------|
| 0 TBAN vs. 800 TBAN | 98.897 | 40.931 | <0.001 | Yes |
| 0 TBAN vs. 400 TBAN | 70.666 | 29.247 | <0.001 | Yes |
| 0 TBAN vs. 300 TBAN | 55.860 | 18.877 | <0.001 | Yes |
| 0 TBAN vs. 200 TBAN | 37.631 | 15.574 | <0.001 | Yes |
| 0 TBAN vs. 100 TBAN | 20.456 | 8.466 | <0.001 | Yes |
| 0 TBAN vs. 12.5 TBAN | 3.606 | 1.219 | 0.544 | No |
| 0 TBAN vs. 50 TBAN | 2.856 | 1.182 | 0.429 | No |
| 0 TBAN vs. 25 TBAN | 2.597 | 1.075 | 0.289 | No |

Table S24. One Way Analysis of Variance: Trichloroacetonitrile (TCAN) thiol reactivity response as the percent of the negative control.

| Group Name | N | Missing | Mean | Std Dev | SEM |
|----------------------------|-----------|----------------|-------------|----------------|------------|
| 0 TCAN | 3 | 0 | 99.944 | 3.038 | 1.754 |
| 1000 TCAN | 3 | 0 | 101.627 | 8.160 | 4.711 |
| 2000 TCAN | 3 | 0 | 102.675 | 7.949 | 4.589 |
| 4000 TCAN | 3 | 0 | 103.806 | 12.212 | 7.051 |
| 6000 TCAN | 3 | 0 | 103.325 | 6.008 | 3.469 |
| 8000 TCAN | 3 | 0 | 107.048 | 7.076 | 4.086 |
| 10000 TCAN | 3 | 0 | 103.071 | 1.775 | 1.025 |
| 12000 TCAN | 3 | 0 | 97.290 | 5.395 | 3.115 |
| 14000 TCAN | 3 | 0 | 94.639 | 4.244 | 2.450 |
| 16000 TCAN | 3 | 0 | 88.948 | 2.162 | 1.248 |
| 18000 TCAN | 3 | 0 | 84.348 | 0.706 | 0.408 |
| Source of Variation | DF | SS | MS | F | P |
| Between Groups | 10 | 1444.913 | 144.491 | 3.698 | 0.005 |
| Residual | 22 | 859.490 | 39.068 | | |
| Total | 32 | 2304.403 | | | |

The differences in the mean values among the treatment groups are greater than would be expected by chance; there is a statistically significant difference (P = 0.005).

Power of performed test with alpha = 0.050: 0.860

Multiple Comparisons versus Control Group (Holm-Sidak method):
Overall significance level = 0.05

Comparisons for factor:

| Comparison | Diff of Means | t | P | P<0.050 |
|-----------------------|----------------------|----------|----------|-------------------|
| 0 TCAN vs. 18000 TCAN | 15.597 | 3.056 | 0.050 | Yes |
| 0 TCAN vs. 16000 TCAN | 10.996 | 2.155 | 0.323 | No |
| 0 TCAN vs. 8000 TCAN | 7.103 | 1.392 | 0.791 | No |
| 0 TCAN vs. 14000 TCAN | 5.306 | 1.040 | 0.925 | No |
| 0 TCAN vs. 4000 TCAN | 3.861 | 0.757 | 0.974 | No |
| 0 TCAN vs. 6000 TCAN | 3.381 | 0.662 | 0.973 | No |
| 0 TCAN vs. 10000 TCAN | 3.127 | 0.613 | 0.958 | No |
| 0 TCAN vs. 2000 TCAN | 2.730 | 0.535 | 0.935 | No |
| 0 TCAN vs. 12000 TCAN | 2.655 | 0.520 | 0.846 | No |
| 0 TCAN vs. 1000 TCAN | 1.683 | 0.330 | 0.745 | No |

Table S25. One Way Analysis of Variance: Bromodichloroacetonitrile (BDCAN) thiol reactivity response as the percent of the negative control.

| Group Name | N | Missing | Mean | Std Dev | SEM |
|----------------------------|-----------|----------------|-------------|----------------|------------|
| BDCAN 0 | 14 | 0 | 99.989 | 6.243 | 1.668 |
| 25 BDCAN | 3 | 0 | 101.056 | 4.277 | 2.469 |
| 50 BDCAN | 3 | 0 | 105.776 | 1.052 | 0.607 |
| 100 BDCAN | 3 | 0 | 91.096 | 2.062 | 1.190 |
| 150 BDCAN | 3 | 0 | 81.060 | 2.504 | 1.446 |
| 200 BDCAN | 3 | 0 | 70.891 | 4.122 | 2.380 |
| 250 BDCAN | 3 | 0 | 60.022 | 2.913 | 1.682 |
| 300 BDCAN | 3 | 0 | 48.350 | 2.309 | 1.333 |
| 350 BDCAN | 3 | 0 | 40.242 | 2.941 | 1.698 |
| 400 BDCAN | 3 | 0 | 30.183 | 2.925 | 1.689 |
| 450 BDCAN | 3 | 0 | 25.416 | 2.218 | 1.280 |
| 500 BDCAN | 3 | 0 | 25.409 | 4.644 | 2.681 |
| 750 BDCAN | 3 | 0 | -0.266 | 2.882 | 1.664 |
| Source of Variation | DF | SS | MS | F | P |
| Between Groups | 12 | 56808.208 | 4734.017 | 239.250 | <0.001 |
| Residual | 37 | 732.117 | 19.787 | | |
| Total | 49 | 57540.325 | | | |

The differences in the mean values among the treatment groups are greater than would be expected by chance; there is a statistically significant difference (P = <0.001).

Power of performed test with alpha = 0.050: 1.000

Multiple Comparisons versus Control Group (Holm-Sidak method):
Overall significance level = 0.05

Comparisons for factor:

| Comparison | Diff of Means | t | P | P<0.050 |
|-----------------------|----------------------|----------|----------|-------------------|
| BDCAN 0 vs. 750 BDCAN | 100.255 | 35.426 | <0.001 | Yes |
| BDCAN 0 vs. 500 BDCAN | 74.580 | 26.353 | <0.001 | Yes |
| BDCAN 0 vs. 450 BDCAN | 74.573 | 26.351 | <0.001 | Yes |
| BDCAN 0 vs. 400 BDCAN | 69.806 | 24.666 | <0.001 | Yes |
| BDCAN 0 vs. 350 BDCAN | 59.747 | 21.112 | <0.001 | Yes |
| BDCAN 0 vs. 300 BDCAN | 51.639 | 18.247 | <0.001 | Yes |
| BDCAN 0 vs. 250 BDCAN | 39.967 | 14.123 | <0.001 | Yes |
| BDCAN 0 vs. 200 BDCAN | 29.098 | 10.282 | <0.001 | Yes |
| BDCAN 0 vs. 150 BDCAN | 18.929 | 6.689 | <0.001 | Yes |
| BDCAN 0 vs. 100 BDCAN | 8.893 | 3.142 | 0.010 | Yes |
| BDCAN 0 vs. 50 BDCAN | 5.787 | 2.045 | 0.094 | No |
| BDCAN 0 vs. 25 BDCAN | 1.067 | 0.377 | 0.708 | No |

Table S26. One Way Analysis of Variance: Chlorodibromoacetonitrile (CDBAN) thiol reactivity response as the percent of the negative control.

| Group Name | N | Missing | Mean | Std Dev | SEM |
|---------------------|----|------------|----------|---------|--------|
| 0 CDBAN | 14 | 0 | 99.980 | 4.465 | 1.193 |
| 25 CDBAN | 4 | 0 | 100.473 | 5.246 | 2.623 |
| 50 CDBAN | 4 | 0 | 102.187 | 6.387 | 3.193 |
| 100 CDBAN | 4 | 0 | 86.473 | 1.304 | 0.652 |
| 150 CDBAN | 4 | 0 | 79.284 | 6.125 | 3.063 |
| 200 CDBAN | 4 | 0 | 69.843 | 4.396 | 2.198 |
| 250 CDBAN | 4 | 0 | 61.516 | 7.247 | 3.623 |
| 300 CDBAN | 4 | 0 | 49.923 | 4.354 | 2.177 |
| 350 CDBAN | 4 | 0 | 45.266 | 4.734 | 2.367 |
| 400 CDBAN | 4 | 0 | 36.870 | 5.150 | 2.575 |
| 450 CDBAN | 4 | 0 | 32.426 | 6.679 | 3.339 |
| 500 CDBAN | 4 | 0 | 23.777 | 3.786 | 1.893 |
| 550 CDBAN | 4 | 0 | 18.951 | 6.014 | 3.007 |
| 600 CDBAN | 4 | 0 | 14.995 | 2.339 | 1.170 |
| 650 CDBAN | 4 | 0 | 12.761 | 2.335 | 1.167 |
| 700 CDBAN | 4 | 0 | 12.632 | 6.065 | 3.032 |
| 750 CDBAN | 4 | 0 | 6.454 | 0.188 | 0.0939 |
| 800 CDBAN | 3 | 0 | 2.652 | 0.107 | 0.0618 |
| Source of Variation | DF | SS | MS | F | P |
| Between Groups | 17 | 104161.331 | 6127.137 | 270.028 | <0.001 |
| Residual | 63 | 1429.517 | 22.691 | | |
| Total | 80 | 105590.847 | | | |

The differences in the mean values among the treatment groups are greater than would be expected by chance; there is a statistically significant difference ($P = <0.001$). Power of performed test with alpha = 0.050: 1.000

Multiple Comparisons versus Control Group (Holm-Sidak method):

Overall significance level = 0.05

Comparisons for factor:

| Comparison | Diff of Means | t | P | P<0.050 |
|-----------------------|---------------|--------|--------|---------|
| 0 CDBAN vs. 750 CDBAN | 93.526 | 34.631 | <0.001 | Yes |
| 0 CDBAN vs. 700 CDBAN | 87.348 | 32.344 | <0.001 | Yes |
| 0 CDBAN vs. 650 CDBAN | 87.220 | 32.296 | <0.001 | Yes |
| 0 CDBAN vs. 800 CDBAN | 97.328 | 32.115 | <0.001 | Yes |
| 0 CDBAN vs. 600 CDBAN | 84.986 | 31.469 | <0.001 | Yes |
| 0 CDBAN vs. 550 CDBAN | 81.029 | 30.004 | <0.001 | Yes |
| 0 CDBAN vs. 500 CDBAN | 76.203 | 28.217 | <0.001 | Yes |
| 0 CDBAN vs. 450 CDBAN | 67.554 | 25.014 | <0.001 | Yes |
| 0 CDBAN vs. 400 CDBAN | 63.110 | 23.369 | <0.001 | Yes |
| 0 CDBAN vs. 350 CDBAN | 54.714 | 20.260 | <0.001 | Yes |
| 0 CDBAN vs. 300 CDBAN | 50.058 | 18.535 | <0.001 | Yes |
| 0 CDBAN vs. 250 CDBAN | 38.464 | 14.243 | <0.001 | Yes |
| 0 CDBAN vs. 200 CDBAN | 30.138 | 11.159 | <0.001 | Yes |
| 0 CDBAN vs. 150 CDBAN | 20.696 | 7.663 | <0.001 | Yes |
| 0 CDBAN vs. 100 CDBAN | 13.507 | 5.002 | <0.001 | Yes |
| 0 CDBAN vs. 50 CDBAN | 2.207 | 0.817 | 0.660 | No |
| 0 CDBAN vs. 25 CDBAN | 0.493 | 0.182 | 0.856 | No |

Table S27. One Way Analysis of Variance: HAN TRI Comparisons.

| Group Name | N | Missing | Mean | Std Dev | SEM |
|---------------------|----|---------|--------|----------|-----------|
| IAN TRI | 6 | 0 | 0.589 | 0.0631 | 0.0258 |
| BAN TRI | 6 | 0 | 0.666 | 0.0186 | 0.00761 |
| DBAN TRI | 6 | 0 | 2.478 | 0.0587 | 0.0240 |
| BCAN TRI | 6 | 0 | 1.096 | 0.0326 | 0.0133 |
| TBAN TRI | 9 | 0 | 3.820 | 0.279 | 0.0931 |
| TCAN TRI | 6 | 0 | 0.0382 | 0.000214 | 0.0000875 |
| BDCAN TRI | 11 | 0 | 3.309 | 0.0745 | 0.0225 |
| CDBAN TRI | 11 | 0 | 3.190 | 0.183 | 0.0553 |
| Source of Variation | DF | SS | MS | F | P |
| Between Groups | 7 | 113.463 | 16.209 | 809.632 | <0.001 |
| Residual | 53 | 1.061 | 0.0200 | | |
| Total | 60 | 114.524 | | | |

The differences in the mean values among the treatment groups are greater than would be expected by chance; there is a statistically significant difference (P = <0.001).

Power of performed test with alpha = 0.050: 1.000

All Pairwise Multiple Comparison Procedures (Holm-Sidak method):

Overall significance level = 0.05

Comparisons for factor:

| Comparison | Diff of Means | t | P | P<0.050 |
|-------------------------|---------------|--------|--------|---------|
| TBAN TRI vs. TCAN TRI | 3.782 | 50.717 | <0.001 | Yes |
| BDCAN TRI vs. TCAN TRI | 3.271 | 45.553 | <0.001 | Yes |
| CDBAN TRI vs. TCAN TRI | 3.152 | 43.896 | <0.001 | Yes |
| TBAN TRI vs. IAN TRI | 3.232 | 43.338 | <0.001 | Yes |
| TBAN TRI vs. BAN TRI | 3.155 | 42.302 | <0.001 | Yes |
| BDCAN TRI vs. IAN TRI | 2.721 | 37.890 | <0.001 | Yes |
| BDCAN TRI vs. BAN TRI | 2.644 | 36.814 | <0.001 | Yes |
| TBAN TRI vs. BCAN TRI | 2.724 | 36.529 | <0.001 | Yes |
| CDBAN TRI vs. IAN TRI | 2.602 | 36.233 | <0.001 | Yes |
| CDBAN TRI vs. BAN TRI | 2.525 | 35.157 | <0.001 | Yes |
| BDCAN TRI vs. BCAN TRI | 2.213 | 30.819 | <0.001 | Yes |
| DBAN TRI vs. TCAN TRI | 2.440 | 29.863 | <0.001 | Yes |
| CDBAN TRI vs. BCAN TRI | 2.094 | 29.162 | <0.001 | Yes |
| DBAN TRI vs. IAN TRI | 1.889 | 23.126 | <0.001 | Yes |
| DBAN TRI vs. BAN TRI | 1.812 | 22.181 | <0.001 | Yes |
| TBAN TRI vs. DBAN TRI | 1.343 | 18.005 | <0.001 | Yes |
| DBAN TRI vs. BCAN TRI | 1.381 | 16.911 | <0.001 | Yes |
| BCAN TRI vs. TCAN TRI | 1.058 | 12.952 | <0.001 | Yes |
| BDCAN TRI vs. DBAN TRI | 0.832 | 11.581 | <0.001 | Yes |
| CDBAN TRI vs. DBAN TRI | 0.713 | 9.925 | <0.001 | Yes |
| TBAN TRI vs. CDBAN TRI | 0.630 | 9.906 | <0.001 | Yes |
| TBAN TRI vs. BDCAN TRI | 0.511 | 8.035 | <0.001 | Yes |
| BAN TRI vs. TCAN TRI | 0.628 | 7.682 | <0.001 | Yes |
| IAN TRI vs. TCAN TRI | 0.550 | 6.736 | <0.001 | Yes |
| BCAN TRI vs. IAN TRI | 0.508 | 6.216 | <0.001 | Yes |
| BCAN TRI vs. BAN TRI | 0.431 | 5.270 | <0.001 | Yes |
| BDCAN TRI vs. CDBAN TRI | 0.119 | 1.972 | 0.105 | No |
| BAN TRI vs. IAN TRI | 0.0773 | 0.946 | 0.349 | No |

References

1. Chen, B.; Zhang, T.; Bond, T.; Gan, Y., Development of quantitative structure activity relationship (QSAR) model for disinfection byproduct (DBP) research: A review of methods and resources. *J. Hazard. Mater.* **2015**, *299*, 260-279.
2. Gramatica, P.; Cassani, S.; Sangion, A., Aquatic ecotoxicity of personal care products: QSAR models and ranking for prioritization and safer alternatives' design. *Green Chemistry* **2016**, *18*, (16), 4393-4406.
3. Box, G. E. P.; Hunter, W. G.; Hunter, J. S., *Statistics for Experimenters: An Introduction to Design, Data Analysis, and Model Building*. Wiley & Sons Inc.: New York, NY., 1978.
4. Lovell, D. P.; Omori, T., Statistical issues in the use of the comet assay. *Mutagenesis* **2008**, *23*, (3), 171-182.
5. Efron, B., Better bootstrap confidence intervals. *J. Am. Statis. Assoc.* **1987**, *82*, (397), 171-185.
6. Singh, K.; Xie, M. *Bootstrap: A Statistical Method*; Rutgers University: New Brunswick, NJ, 2008; p 14.
7. Plewa, M. J.; Wagner, E. D.; Richardson, S. D.; Thruston, A. D., Jr.; Woo, Y. T.; McKague, A. B., Chemical and biological characterization of newly discovered iodoacid drinking water disinfection byproducts. *Environ. Sci. Technol.* **2004**, *38*, (18), 4713-4722.
8. Li, J.; Moe, B.; Vemula, S.; Wang, W.; Li, X.-F., Emerging disinfection byproducts, halobenzoquinones: effects of isomeric structure and halogen substitution on cytotoxicity, formation of reactive oxygen species, and genotoxicity. *Environ. Sci. Technol.* **2016**, *50*, (13), 6744-6752.
9. Wagner, E. D.; Rayburn, A. L.; Anderson, D.; Plewa, M. J., Analysis of mutagens with single cell gel electrophoresis, flow cytometry, and forward mutation assays in an isolated clone of Chinese hamster ovary cells. *Environ. Mol. Mutagen.* **1998**, *32*, (4), 360-368.
10. Wagner, E. D.; Rayburn, A. L.; Anderson, D.; Plewa, M. J., Calibration of the single cell gel electrophoresis assay, flow cytometry analysis and forward mutation in Chinese hamster ovary cells. *Mutagenesis* **1998**, *13*, (1), 81-84.
11. Wagner, E. D.; Plewa, M. J., CHO cell cytotoxicity and genotoxicity analyses of disinfection by-products: an updated review. *J. Environ. Sci.* **2017**, *58*, 64-76.
12. Muellner, M. G.; Wagner, E. D.; McCalla, K.; Richardson, S. D.; Woo, Y. T.; Plewa, M. J., Haloacetonitriles vs. regulated haloacetic acids: Are nitrogen containing DBPs more toxic? *Environ. Sci. Technol.* **2007**, *41*, (2), 645-651.
13. Plewa, M. J.; Wagner, E. D., *Mammalian Cell Cytotoxicity and Genotoxicity of Disinfection By-Products*. Water Research Foundation: Denver, CO, 2009; p 134.
14. Tice, R. R.; Agurell, E.; Anderson, D.; Burlinson, B.; Hartmann, A.; Kobayashi, H.; Miyamae, Y.; Rojas, E.; Ryu, J. C.; Sasaki, Y. F., Single cell gel/comet assay: guidelines for in vitro and in vivo genetic toxicology testing. *Environ. Mol. Mutagen.* **2000**, *35*, (3), 206-221.
15. Rundell, M. S.; Wagner, E. D.; Plewa, M. J., The comet assay: genotoxic damage or nuclear fragmentation? *Environ. Mol. Mutagen.* **2003**, *42*, (2), 61-67.
16. Wagner, E. D.; Plewa, M. J., Microplate-based comet assay. In *The Comet Assay in Toxicology*, Dhawan, A.; Anderson, D., Eds. Royal Society of Chemistry: London, 2009; pp 79-97.

17. Dong, S.; Page, M. A.; Wagner, E. D.; Plewa, M. J., Thiol reactivity analyses to predict mammalian cell cytotoxicity of water samples. *Environ. Sci. Technol.* **2018**, *52*, 8822–8829.
18. Pals, J. A.; Wagner, E. D.; Plewa, M. J.; Xia, M.; Attene-Ramos, M. S., Monohalogenated acetamide-induced cellular stress and genotoxicity are related to electrophilic softness and thiol/thiolate reactivity. *J. Environ. Sci.* **2017**, *58*, 224-230.
19. Pals, J. A.; Wagner, E. D.; Plewa, M. J., Energy of the lowest unoccupied molecular orbital, thiol reactivity, and toxicity of three monobrominated water disinfection byproducts. *Environ. Sci. Technol.* **2016**, *50*, (6), 3215-3221.
20. Townsend, D. M.; Tew, K. D.; Tapiero, H., The importance of glutathione in human disease. *Biomed. Pharmacother.* **2003**, *57*, (3-4), 145-155.
21. Meister, A.; Anderson, M. E., Glutathione. *Annu. Rev. Biochem.* **1983**, *52*, 711-760.