

Article

You Better Repeat It: Complex CO₂ × Temperature Effects in Atlantic Silverside Offspring Revealed by Serial Experimentation

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Abstract: Concurrent ocean warming and acidification demand experimental approaches that assess biological sensitivities to combined effects of these potential stressors. Here, we summarize five CO₂ × temperature experiments on wild Atlantic silverside, *Menidia menidia*, offspring that were reared under factorial combinations of CO₂ (nominal: 400, 2200, 4000, and 6000 µatm) and temperature (17, 20, 24, and 28 °C) to quantify the temperature-dependence of CO₂ effects in early life growth and survival. Across experiments and temperature treatments, we found few significant CO₂ effects on response traits. Survival effects were limited to a single experiment, where elevated CO₂ exposure reduced embryo survival at 17 and 24 °C. Hatch length displayed CO₂ × temperature interactions due largely to reduced hatch size at 24 °C in one experiment but increased length at 28 °C in another. We found no overall influence of CO₂ on larval growth or survival to 9, 10, 15 and 13–22 days post-hatch, at 28, 24, 20, and 17 °C, respectively. Importantly, exposure to cooler (17 °C) and warmer (28 °C) than optimal rearing temperatures (24 °C) in this species did not appear to increase CO₂ sensitivity. Repeated experimentation documented substantial inter- and intra-experiment variability, highlighting the need for experimental replication to more robustly constrain inherently variable responses. Taken together, these results demonstrate that the early life stages of this ecologically important forage fish appear largely tolerate to even extreme levels of CO₂ across a broad thermal regime.

Keywords: early life history; factorial experiment; global warming; growth; *Menidia menidia*; ocean acidification; survival

1. Introduction

The current anthropogenic increase in atmospheric and therefore oceanic carbon dioxide (CO₂) concentrations has been unparalleled over the past 66 million years [1]. Resultant changes in ocean pH and carbon chemistry (ocean acidification, OA) are likely to have major impacts on marine ecosystems [2] by changing species abundances, interactions and trophic dynamics, all of which depend ultimately on the CO₂ sensitivities of individual organisms [3–5]. Laboratory experiments have played an important role in quantifying these CO₂ sensitivities, suggesting that they are greater in sessile, calcifying invertebrates than in active, non-calcifying vertebrates, and greater in early life stages than adults [6–8]. The latter has been particularly well documented for marine fish, where adults are largely tolerant of acute high-CO₂ levels far exceeding predicted OA conditions [9,10]. By contrast, fish early life-stages (embryos and early larvae) that are still developing effective acid-base regulation have exhibited reduced survival [11,12], reduced growth [13,14], defective development [14,15], otolith over-calcification [16,17], and behavioral abnormalities in response to high-CO₂ conditions

in the laboratory [18,19]. Experiments showing no discernible CO₂ effects are also common [20–24]. This complexity of empirical evidence remains challenging to reconcile [25], but is consistent with the emerging consensus of species- and population-specific CO₂ sensitivities, particularly for fish adapted to high CO₂ and pH variability in their habitats [26].

To date, experimental approaches have largely been guided by open-ocean predictions for administering CO₂ treatments (see [27]). It is now recognized, however, that many marine organisms experience considerable diel and seasonal pH/CO₂ fluctuations in their habitats [26,28–30]. Short-term pH/CO₂ variability can be attributed to ephemeral upwelling [31], river input [32], and metabolic processes that dominate CO₂ variability in coastal habitats [33] and in oxygen minimum zones [34]. The seasonal intensification of community respiration in highly productive coastal systems (e.g., saltmarshes and mangrove lagoons) can increase both average and extreme CO₂ levels to nearly double the open-ocean OA predictions for the next 300 years [35]. Given the thermal sensitivity of microbial respiration rates, metabolically driven acidification is generally most extreme during peak summer temperatures [36]. Hence, to better understand climate change effects on coastal species, experiments should implement CO₂ and temperature conditions that reflect the range of modern and predicted conditions of their source ecosystems, rather than relying on average global predictions.

While single-factor CO₂ experiments are a necessary initial step, it is now widely recognized that OA proceeds in concert with ocean warming and deoxygenation. Experiments are needed to address species sensitivities to multiple stressors of marine climate change [37,38]. Warming may be the primary driver of ecological disruption, as there is already evidence of shifting fish distributions and phenologies [39,40], which likely reflect the need for ectotherms to maintain environments within their scope of physiological optima [41]. The capacity of organisms to maintain performance at temperatures approaching or exceeding their thermal tolerance is a key metric in determining climate sensitivity [42]. Elevated environmental CO₂ may increase energetic costs associated with acid-base regulation [43] and could compromise the functional capacity of other vital processes [44] and therefore increase an organisms' sensitivity to thermal extremes [38]. Thus, CO₂ × temperature experiments are not only more realistic, they may also discover important stressor interactions that elude single-stressor approaches [45].

The majority of studies evaluating CO₂ × temperature effects in fish have focused on stenothermal taxa from polar [46–50] or tropical habitats [51–54]. These fish are presumably adapted to their relatively stable thermal environments and may thus show limited acclimation capacity to combined climate stressors [55,56]. By contrast, temperate species are often eurythermal, i.e., capable of acclimating to broad seasonal temperature fluctuations. However, they are still adapted to specific thermal regimes [41] and often show narrower thermal requirements during CO₂ sensitive early life stages [14,57].

Many fitness-relevant traits such as growth or survival are highly variable in nature during fish early life stages, thus producing variable outcomes even under most meticulously controlled experimental conditions [58,59]. Variations in offspring due to parentage, food quality and quantity, or water sources can introduce additional variability, hence underscoring the risks of generalizing results from single experiments to population or species characteristic such as CO₂ or temperature sensitivity. More robust depictions of CO₂ and temperature sensitivity are likely to emerge if experiments are replicated and analyzed together, but this approach is still underutilized in studies of climate change effects on marine organisms.

Here we report on five factorial CO₂ × temperature experiments conducted on offspring of wild Atlantic silversides, *Menidia menidia*, an ecologically important and abundant coastal forage fish with a broad distribution along the east coast of North America [60]. Wild silverside offspring are amenable to experimental manipulations and have thus become a widely used model in OA experiments [58,61–66]. Over the course of three years, we repeatedly reared Atlantic silverside offspring at different factorial combinations of CO₂ and temperature to quantify the temperature-dependence of CO₂ effects in growth and survival. We hypothesized that negative responses to high-CO₂ levels would largely occur

at the species lower and upper thermal limits, while predicting fewer or no CO₂ effects at optimal thermal conditions.

2. Methods

2.1. Field Sampling and Experimental Designs

Collections of wild, spawning ripe Atlantic silverside were made during high tide 1–3 days prior to full or new moons following the species' semi-lunar spawning periodicity during spring and early summer [61]. Adults were caught with a 30 × 2 m beach seine from local salt marshes and transported live to our laboratory facilities. For the 2014 experiment (experiment 1), adults were collected from Poquot Beach (40°56.85' N, 73°6.15' W), and the experiment took place at Stony Brook University's Flax Pond Marine Laboratory. During 2016 and 2017 (experiments 2–5), spawning adults were collected from Mumford Cove (41°19'25'' N 72°01'07'' W), and experiments were conducted in the Rankin Seawater Facility at University of Connecticut's Avery Point campus. Ripe adults were held overnight at 20 °C in aerated tanks at low densities with no food and strip-spawned the next day. Fertilization dates for each experiment are reported in Table 1.

Table 1. Summary of five CO₂ × temperature experiments on *M. menidia* offspring. Treatment levels for pCO₂ (µatm) and temperature (°C) represent target conditions, actual measured values are presented in Table 2. Trait are abbreviated as embryo survival (ES), hatch length (HL), larval survival (LS), and growth rate (GR).

Exp Num	Fertilization Date	Target Treatment Levels		Number of Replicates	Measured Traits
		pCO ₂	Temp		
1	5/5/2014	400, 2200, 6000	17, 24	5	ES, LS
2	4/22/2016	400, 2200	17, 24	5	ES, HL, LS, GR
3	5/19/2016	400, 2200, 4000	17, 20, 24	5	ES, HL, LS, GR
4	4/28/2017	400, 2200	24, 28	3	ES, HL, LS, GR
5	5/26/2017	400, 2200, 4000	24, 28	5	ES, HL, LS, GR

Strip-spawning protocols maximized fertilization success, while enabling random distribution of embryos across replicates [61,63]. For each experiment, eggs from 12+ running-ripe females were gently mixed into shallow plastic dishes lined with 1-mm plastic window screening. Milt from each of 20+ males was collected and pooled into 500-mL glass beakers, mixed with seawater, stirred, then gently poured into spawning dishes and mixed with eggs for ~15 min. The number of spawners used for each experiment and their length measurements are reported in Table S1. In this species, fertilized embryos uncoil chorionic filaments, which readily attach to screening. Window screens were cut into smaller sections where embryos were counted under low magnification with high accuracy.

Table 2. Mean ($\pm$ SD) pH and temperature ($^{\circ}$ C) from daily measurements. Mean ($\pm$ SD) salinity, total alkalinity (A_T ; $\mu\text{mol kg}^{-1}$), dissolved inorganic carbon (C_T ; $\mu\text{mol kg}^{-1}$), partial pressure of CO_2 ($p\text{CO}_2$; μatm), fugacity of CO_2 ($f\text{CO}_2$; μatm), and carbonate ion concentration (CO_3^{2-} ; $\mu\text{mol kg}^{-1}$) measured from replicated seawater samples of each treatment. Salinity was measured via refractometer and A_T from endpoint titrations. C_T and $p\text{CO}_2$, $f\text{CO}_2$ and CO_3^{2-} were calculated in CO_2SYS .

Exp Num	Target Temp	Measured Temp	Target $p\text{CO}_2$	Measured pH	Salinity	A_T	C_T	$p\text{CO}_2$	$f\text{CO}_2$	CO_3^{2-}
1	17	17.5 $\pm$ 0.1	400	8.24 $\pm$ 0.02	26	2514 $\pm$ 17	2302 $\pm$ 12	433 $\pm$ 29	431 $\pm$ 29	168 $\pm$ 8
		17.5 $\pm$ 0.1	2200	7.49 $\pm$ 0.05	26	2539 $\pm$ 22	2581 $\pm$ 5	2564 $\pm$ 94	2556 $\pm$ 94	38 $\pm$ 1
		17.5 $\pm$ 0.1	6000	7.14 $\pm$ 0.05	26	2492 $\pm$ 33	2680 $\pm$ 11	5753 $\pm$ 277	6733 $\pm$ 276	17 $\pm$ 1
1	24	24.0 $\pm$ 0.2	400	8.20 $\pm$ 0.06	26	2501 $\pm$ 7	2258 $\pm$ 11	474 $\pm$ 27	472 $\pm$ 27	191 $\pm$ 7
		24.0 $\pm$ 0.2	2200	7.47 $\pm$ 0.05	26	2474 $\pm$ 81	2504 $\pm$ 8	2881 $\pm$ 172	2872 $\pm$ 172	42 $\pm$ 2
		24.0 $\pm$ 0.2	6000	7.14 $\pm$ 0.05	26	2472 $\pm$ 49	2634 $\pm$ 13	6195 $\pm$ 378	6174 $\pm$ 378	20 $\pm$ 1
2	17	16.9 $\pm$ 0.3	400	8.17 $\pm$ 0.12	30	2038 $\pm$ 17	1851 $\pm$ 8	368 $\pm$ 18	367 $\pm$ 18	135 $\pm$ 6
		16.9 $\pm$ 0.3	2200	7.49 $\pm$ 0.13	30	2031 $\pm$ 12	2058 $\pm$ 21	2037 $\pm$ 188	2030 $\pm$ 188	32 $\pm$ 2
		23.5 $\pm$ 0.3	400	8.13 $\pm$ 0.09	30	204 $\pm$ 11	1838 $\pm$ 16	427 $\pm$ 29	426 $\pm$ 29	150 $\pm$ 7
2	24	23.6 $\pm$ 0.3	2200	7.49 $\pm$ 0.12	30	2041 $\pm$ 11	2048 $\pm$ 7	2190 $\pm$ 277	2183 $\pm$ 276	5 $\pm$ 5
		17.4 $\pm$ 0.2	400	8.22 $\pm$ 0.01	31	2054 $\pm$ 8	1838 $\pm$ 26	322 $\pm$ 12	321 $\pm$ 12	153 $\pm$ 2
		17.6 $\pm$ 0.3	2200	7.51 $\pm$ 0.01	31	2047 $\pm$ 20	2066 $\pm$ 21	1952 $\pm$ 39	1945 $\pm$ 39	35 $\pm$ 1
3	17	17.4 $\pm$ 0.2	4200	7.20 $\pm$ 0.02	31	2053 $\pm$ 24	2174 $\pm$ 16	4056 $\pm$ 204	4042 $\pm$ 203	18 $\pm$ 1
		19.7 $\pm$ 0.2	400	8.20 $\pm$ 0.02	31	2048 $\pm$ 29	1833 $\pm$ 3	345 $\pm$ 15	345 $\pm$ 15	160 $\pm$ 6
		19.6 $\pm$ 0.3	2200	7.51 $\pm$ 0.03	31	2031 $\pm$ 14	2039 $\pm$ 10	1964 $\pm$ 109	1957 $\pm$ 108	38 $\pm$ 2
3	20	19.7 $\pm$ 0.2	4200	7.21 $\pm$ 0.02	31	2058 $\pm$ 6	2153 $\pm$ 37	4066 $\pm$ 227	4063 $\pm$ 226	20 $\pm$ 1
		23.7 $\pm$ 0.2	400	8.22 $\pm$ 0.02	31	2044 $\pm$ 9	1798 $\pm$ 8	331 $\pm$ 14	330 $\pm$ 14	185 $\pm$ 5
		23.7 $\pm$ 0.3	2200	7.49 $\pm$ 0.02	31	2048 $\pm$ 22	2050 $\pm$ 25	2157 $\pm$ 92	2151 $\pm$ 92	42 $\pm$ 1
4	24	23.6 $\pm$ 0.2	4200	7.20 $\pm$ 0.02	31	2059 $\pm$ 1	2140 $\pm$ 8	4339 $\pm$ 169	4325 $\pm$ 169	22 $\pm$ 1
		23.6 $\pm$ 0.3	400	8.19 $\pm$ 0.03	31	2096 $\pm$ 63	1842 $\pm$ 64	368 $\pm$ 38	367 $\pm$ 38	180 $\pm$ 10
		23.7 $\pm$ 0.3	2200	7.51 $\pm$ 0.03	31	2124 $\pm$ 51	2122 $\pm$ 44	2155 $\pm$ 83	2148 $\pm$ 82	45 $\pm$ 4
4	28	28.1 $\pm$ 0.2	400	8.22 $\pm$ 0.03	32	2164 $\pm$ 88	1860 $\pm$ 85	356 $\pm$ 35	355 $\pm$ 34	216 $\pm$ 11
		27.9 $\pm$ 0.4	2200	7.52 $\pm$ 0.03	32	2164 $\pm$ 117	2146 $\pm$ 113	2217 $\pm$ 134	2210 $\pm$ 133	54 $\pm$ 6
		24.3 $\pm$ 0.4	400	8.19 $\pm$ 0.02	32	2137 $\pm$ 3	1897 $\pm$ 13	389 $\pm$ 23	388 $\pm$ 23	175 $\pm$ 8
5	24	24.1 $\pm$ 0.2	2200	7.50 $\pm$ 0.04	32	2151 $\pm$ 14	2156 $\pm$ 27	2265 $\pm$ 228	2258 $\pm$ 227	43 $\pm$ 4
		24.2 $\pm$ 0.3	4200	7.21 $\pm$ 0.02	32	2130 $\pm$ 27	2230 $\pm$ 25	4432 $\pm$ 180	4418 $\pm$ 179	23 $\pm$ 1
		28.2 $\pm$ 0.2	400	8.23 $\pm$ 0.02	32	2157 $\pm$ 24	1857 $\pm$ 29	350 $\pm$ 19	348 $\pm$ 19	215 $\pm$ 4
5	28	28.1 $\pm$ 0.2	2200	7.48 $\pm$ 0.02	32	2176 $\pm$ 50	2172 $\pm$ 48	2439 $\pm$ 84	2431 $\pm$ 83	49 $\pm$ 2
		28.2 $\pm$ 0.3	4200	7.20 $\pm$ 0.03	32	2155 $\pm$ 20	2244 $\pm$ 18	4720 $\pm$ 217	4714 $\pm$ 204	26 $\pm$ 1

Experiments were initiated within 2 h of fertilization when replicate rearing-containers (20-L cylindrical polyethylene buckets) received precisely 100 embryos. Rearing-containers were filled with clean seawater (filtered to 1 μm and UV sterilized). Optimal salinity (27–31) and light conditions (15 h light:9 h dark) for rearing *M. menidia* were maintained across experiments [60]. The number of $\text{CO}_2 \times$ temperature treatments and replicates varied between experiments (see Table 1). For actual $\text{CO}_2 \times$ temperature treatments administered see Table 2. Starting four days post-fertilization (dpf), each rearing-container was checked for hatched larvae. On the morning of first observed hatch, larvae were immediately provided with equal rations of powdered weaning diet (Otohime Marine Fish Diet, size A1, Reed Mariculture®, Campbell, CA, USA) to stimulate feeding and *ad libitum* levels of newly hatched brine shrimp nauplii (*Artemia salina*, San Francisco strain, brineshrimpdirect.com, Ogden, UT, USA). Larvae were fed daily *ad libitum* rations of newly hatched nauplii for the remainder of the experiment. To quantify survival to hatch, one day post-hatch larvae were counted by gently scooping small groups into replacement rearing-containers. For initial hatch standard length (SL, nearest 0.01 mm) measurements, larvae were randomly sub-sampled ($N = 10$) from each replicate were preserved in 5% formaldehyde/freshwater solution buffered with saturated sodium tetraborate. The timing of hatch sub-samples varied slightly between experiments and temperatures (see Table 3). Rearing-containers were siphoned of waste daily, and treatment water was partially exchanged with new seawater every other day. Levels of ammonia waste were monitored daily (Saltwater Ammonia Test Kit, API®, Chalfont, PA, USA) to maintain uncritical levels below 0.25 ppm. All experiments were terminated when larvae reached ~10 mm SL within temperature treatments (determined by visual estimates). Using body size rather than set time intervals allowed comparing CO_2 effects on offspring during the same developmental period (i.e., fertilization to ~10 mm SL) across temperature treatments. Experiment durations ranged from 14 to 36 days (Table 3). At termination, all survivors were counted and measured for SL (nearest 0.01 mm) via calibrated digital images (Image Pro Premier V9.0, Media Cybernetics®, Rockville, MD, USA).

2.2. CO_2 and Temperature Levels

We applied a target CO_2 level of 400 μatm (~8.15 pH) for control treatments, a level characteristic of the open ocean and of coastal systems at the onset of the silverside spawning season (spawning typically begins early April and extends through July) (Figure 1 [61]). The target level for high CO_2 was 2200 μatm (~7.50 pH), a level that is commonly experienced by silverside offspring in late spring and summer (Figure 1), but also represents the maximum prediction of average OA for the next 300 years [67] and therefore a common benchmark in many OA studies [61–63]. The target level for the extreme CO_2 treatment was 6000 μatm (~7.15 pH) during experiment 1, but was reduced to 4200 μatm (~7.20 pH) for experiments 3 and 5. These represent extreme CO_2 conditions rarely reached in contemporary coastal systems, but may become more common under future climate and eutrophication scenarios [33].

Table 3. Summary of all measured *M. meridia* response traits across five experiments; embryo survival (%), hatch length (mm), larval survival (%) and larval growth rate (mm d^{−1}) represented as treatments means ± SD. CO₂ levels are shown as control (C), high (H), and extreme (E) (see Table 2 for values). Sample times are given as days post fertilization (dpf).

Exp Num	Temp (°C)	Treatment CO ₂	Days to First Hatch	Age at Hatch Sample (dpf)	Embryo Survival (%)	Hatch Length (mm)	Age at Larval Sample (dpf)	Larval Survival (%)	Growth Rate (mm d ^{−1})
1	17	C	13	14	68 ± 4		26	34 ± 9	
		H	13	14	74 ± 3		26	50 ± 22	
		E	13	14	56 ± 6		26	43 ± 28	
	24	C	6	7	65 ± 4		16	44 ± 10	
		H	6	7	65 ± 3		16	53 ± 13	
		E	6	7	56 ± 8		16	37 ± 28	
2	17	C	13	15	92 ± 3	5.32 ± 0.05	30	21 ± 8	0.18 ± 0.03
		H	13	15	87 ± 11	5.29 ± 0.05	30	11 ± 7	0.16 ± 0.04
		E	13	15	88 ± 7	5.30 ± 0.14	16	32 ± 33	0.41 ± 0.05
	24	C	6	7	76 ± 6	5.35 ± 0.06	16	26 ± 7	0.40 ± 0.05
		H	6	7	93 ± 5	5.37 ± 0.05	36	32 ± 8	0.21 ± 0.02
		E	13	15	95 ± 5	5.42 ± 0.12	36	56 ± 21	0.20 ± 0.02
3	17	C	13	15	89 ± 6	5.42 ± 0.11	25	59 ± 14	0.19 ± 0.01
		H	13	15	96 ± 5	5.55 ± 0.11	25	82 ± 10	0.33 ± 0.02
		E	10	11	94 ± 7	5.42 ± 0.09	25	77 ± 14	0.32 ± 0.02
	20	C	10	11	95 ± 5	5.62 ± 0.09	25	75 ± 22	0.33 ± 0.03
		H	10	11	95 ± 5	5.51 ± 0.09	16	72 ± 8	0.37 ± 0.02
		E	6	7	95 ± 6	5.32 ± 0.05	16	69 ± 14	0.35 ± 0.03
4	24	C	6	6	62 ± 9	4.98 ± 0.07	16	33 ± 10	0.33 ± 0.04
		H	6	6	51 ± 7	4.90 ± 0.16	16	29 ± 20	0.36 ± 0.05
		E	6	6	83 ± 12	4.83 ± 0.10	16	14 ± 7	0.40 ± 0.04
	28	C	5	5	46 ± 5	4.76 ± 0.04	14	31 ± 35	0.48 ± 0.09
		H	5	5	49 ± 3	4.62 ± 0.09	14	40 ± 27	0.44 ± 0.08
		E	5	5	74 ± 13	4.78 ± 0.07	16	41 ± 27	0.33 ± 0.06
5	24	C	6	6	55 ± 3	4.83 ± 0.10	16	14 ± 9	0.42 ± 0.03
		H	6	6	80 ± 13	4.69 ± 0.10	14	9 ± 6	0.38 ± 0.08
		E	5	5	72 ± 13	4.70 ± 0.05	14		

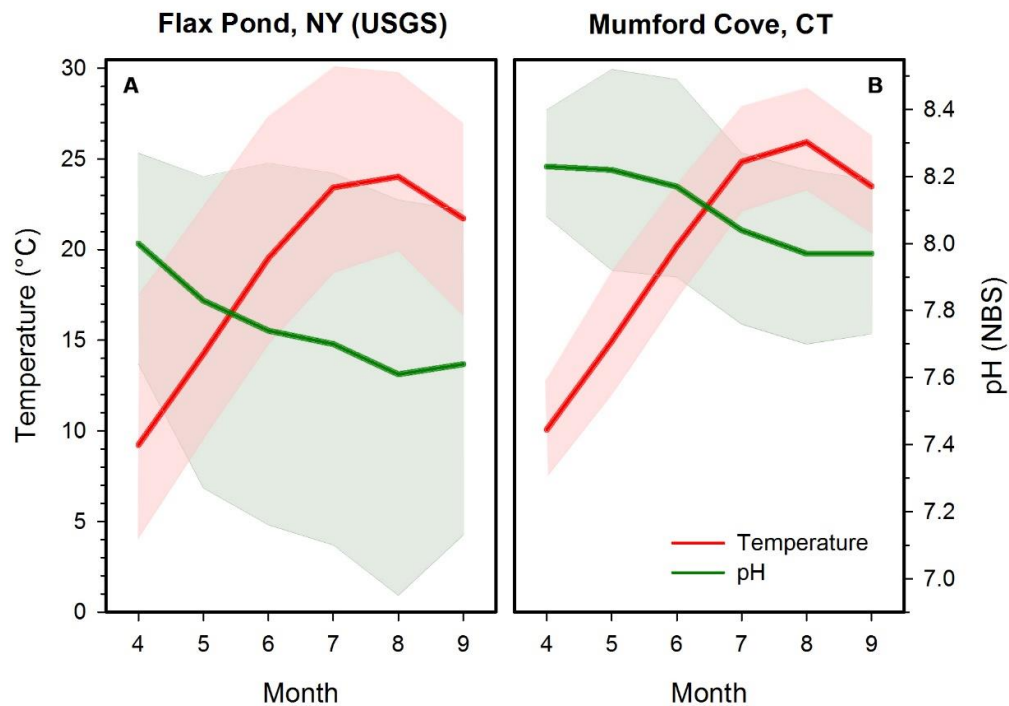


Figure 1. Average mean (+/− minimum/maximum) monthly temperature and pH conditions during the spawning and growing season of Atlantic silversides in (A) Flax Pond, Long Island, New York and (B) Mumford Cove, Connecticut. The sites provided wild spawners for experiment 1 (A) and experiments 2–5 (B). Long-term averages were derived from monitoring data collected in 15 min intervals by (A) USGS station #01304057 between 2008 and 2018 and (B) the Baumann lab in Mumford Cove between 2015–2018.

We administered four temperature treatments over the course of the five experiments; 17, 20, 24 and 28 °C. The first three temperatures represent local conditions found during the onset (late-April), peak (early-June), and end (July) of the silverside spawning season, respectively (Figure 1). At the latitude of our source populations (~41° N), silverside spawning habitats rarely reach temperatures of 28 °C, however, these conditions may become more common given projected increases of 2–3 °C in global mean ocean temperature [68]. The optimal culturing temperature for *M. menidia* from northern latitudes is ~24 °C; thus, 20 °C and 24 °C treatments were considered near-optimal temperatures, while 17 °C and 28 °C treatments represented sub-optimal thermal conditions [60].

2.3. CO₂ × Temperature Manipulations and Measurements

All experiments followed established best practices and guidelines for seawater acidification in OA research [27]. For 2 × 2 and 3 × 2 factorial designs (see Table 1 for overview of experiments and designs), replicate rearing-containers were placed into large temperature-controlled water baths. Elevated CO₂ levels were achieved via gas proportioners (ColeParmer®, Vernon Hills, IL, USA) mixing air with 100% CO₂ (bone dry grade) that was delivered continuously to the bottom of each replicate rearing-container via airstone. To counteract metabolic CO₂ accumulation, control CO₂ conditions were achieved by forcing compressed laboratory air through a series of CO₂-stripping units containing granular soda lime (AirGas®, Waterford, CT, USA), a particle filter (1 μm), and then to each rearing-container via airstone. Target pH levels were monitored daily using a handheld pH probe (Orion Ross Ultra pH/ATC Triode with Orion Star A121 pH Portable Meter (Thermo Fisher Scientific®, Waltham, MA, USA); Intellical PHC281 pH Electrode with HQ11D Handheld pH/ORP Meter (Hach®, Loveland, CO, USA) calibrated bi-weekly with National Institute of Standards and

Technology (NIST) certified 2-point pH references. Continuous bubbling maintained dissolved oxygen (DO) saturation (>8 mg/L) in rearing vessels. Target treatment temperatures were controlled by thermostats (Aqualogic[®], San Diego, CA, USA) which powered chillers (DeltaStar[®], Lynchburg, VA, USA) or glass submersible heaters to maintain water bath temperatures.

For 3×3 factorial experiments, we constructed an automated acidification system composed of nine discrete recirculation-units designed for larval fish rearing. Each recirculating-unit consists of a sump (90 L), a header tank (40 L) and a main tank (240 L) that holds up to five replicate rearing-containers (20 L) fitted with screened overflow holes (100 μ m). In these units, seawater continuously circulates from the sump through a UV sterilizer into the header tank, where it is gravity fed to the bottom of each rearing-container, from which it overflows in the main tank and back into the sump. We designed a LabView (National Instruments[®], Austin, TX, USA) based program to fully automate the control of seawater chemistry. The software interfaces with the recirculating-units via a data-acquisition module (NI cDAQ-9184, National Instruments[®]), which controls nine sampling-pumps (one per tank) and a series of gas and water solenoid valves, while receiving input from a central pH electrode (Hach pH^D[®] digital electrode calibrated weekly using NIST 2-point pH references) and DO probe (Hach LDO[®] Model 2). The software sequentially assesses the pH conditions in each recirculating-unit (once per hour) by pumping water for ~ 7.5 min through the housing of the central pH probe, comparing measured pH levels to set-points and then adjusting levels by bubbling standardized amounts of 100% CO₂ (bone dry grade, AirGas[®]) or CO₂-stripped air into the sump of each tank. The software also maintains DO saturation (>8 mg/L) by bubbling in CO₂-stripped air. LabView logs current pH, temperature, and DO conditions before cycling to the next unit. Temperatures were controlled by thermostats (Aqualogic[®]) that powered submersible heaters or in-line chillers (DeltaStar[®]).

Actual treatment CO₂ levels were determined based on measurements of pH, temperature, salinity, and total alkalinity (A_T). Treatment tanks were sampled three times per experiment for measurements of A_T (μ mol kg⁻¹). Seawater was siphoned and filtered (to 10 μ m) into 300-mL borosilicate bottles. Salinity was measured at the time of sampling using a refractometer. Bottles were stored at 3 °C and measured for A_T within two weeks of sampling using an endpoint titration (G20 Potentiometric Titrator, Mettler Toledo[®], Columbus, OH, USA). Methodological accuracy (within $\pm 1\%$) of alkalinity titrations were verified and calibrated using Dr. Andrew Dickson's (University of California San Diego, Scripps Institution of Oceanography, https://www.nodc.noaa.gov/ocads/oceans/Dickson_CRM/batches.html) certified reference material for A_T in seawater. The partial pressure and fugacity of CO₂ ($p\text{CO}_2$, $f\text{CO}_2$; μ atm) as well as dissolved inorganic carbon (C_T ; μ mol kg⁻¹) and carbonate ion concentration (CO_3^{2-} ; μ mol kg⁻¹) were calculated in CO2SYS (V2.1, <http://cdiac.ornl.gov/ftp/co2sys>) based on measured A_T , pH, temperature, and salinity using K1 and K2 constants from Mehrbach et al. [69] refitted by Dickson and Millero [70] and Dickson [71] for KHSO₄. An overview of pH and carbonate chemistry measurements for each experiment is given in Table 2.

2.4. Response Traits and Statistical Analysis

For all replicates in each experiment we quantified time (d) to first-hatch (day of fertilization to day of first-hatch), the % of embryo survival (fertilization to hatch), the % of larval survival (hatch to end of experiment), SL at hatch, and post-hatch growth rate ((mean final SL – mean hatch SL)/number days reared post-hatch). For experiment 1, only survival traits were quantified. Time to first-hatch was invariant between CO₂ levels and thus was not analyzed statistically. Proportional survival data were logit transformed [$=\log_{10}(\text{survival}/(1 - \text{survival}))$] prior to analysis [72]. Grubb's test [73] was used to identify potential outlying replicates, resulting in the removal of three replicates throughout the dataset for low embryo survival (Grubb's test, $p < 0.05$).

Statistical analyses were computed using SPSS (V20, IBM). As a first step, we used linear mixed effects models incorporating data from all experiments to test for significant effects ($\alpha < 0.05$) of CO₂, temperature, their interaction (fixed factors) and experiment (random factor) for each response trait:

$$\text{Response trait} = \text{CO}_2 + \text{temperature} + \text{CO}_2 \times \text{temperature} + \text{experiment} + \text{error}.$$

Response trait data were checked for variance homogeneity and assumption of normality using Levene's and Shapiro-Wilk tests ($\alpha < 0.05$), respectively. If linear mixed effects models identified traits with significant CO₂ or CO₂ $\times$ temperature effects, we used two-way analysis of variance (ANOVA) to test for significant effects of CO₂, temperature, and their interaction within each experiment:

$$\text{Response trait} = \text{CO}_2 + \text{temperature} + \text{CO}_2 \times \text{temperature} + \text{error}.$$

This approach was implemented to characterize how CO₂ effects differed between experiments. If two-way ANOVAs detected significant ($\alpha < 0.05$) CO₂ or CO₂ $\times$ temperature interactive effects, we used one-way ANOVAs to test for significant CO₂ effects within temperature treatments. Where necessary, least-significant-difference (LSD) post-hoc tests were used for multiple comparisons. We conducted two- and one-way ANOVAs on experiment 1 separately because the extreme CO₂ level implemented there (~6000 μatm) was higher than in experiments 3 and 5 (~4200 μatm). ANOVA groups were checked for variance homogeneity and assumption of normality using Levene's and Shapiro-Wilk tests ($\alpha < 0.05$), respectively.

Linear mixed effects models indicated that response traits were highly variable between experiments. Hence, we implemented an additional approach to better describe CO₂ effects across temperature treatments. We quantified the temperature-specific CO₂ effect sizes for each trait (T) for each experiment by calculating the log-transformed CO₂ response ratio (lnRR) to high and extreme levels of CO₂ exposure. LnRRs evaluate the average proportional change in a trait relative to control treatments, with negative LnRRs indicating negative CO₂ effects. LnRRs have become a common metric for evaluating CO₂ effects in meta-analyses when comparing variable responses across studies [6,25,74]. LnRRs were calculated as:

$$\ln\text{RR}(T) = \ln(T_{\text{high or extreme CO}_2}) - \ln(T_{\text{control CO}_2}).$$

Overall temperature-specific CO₂ responses were calculated as the mean lnRR(T) across experiments.

3. Results

3.1. Embryo Survival

Among-replicate time to first hatch was 5, 6, 10, and 13 dpf at 28, 24, 20, and 17 °C, respectively. CO₂ level had no influence on time to hatch (Table 3). Across experiments, mean embryo survival ranged from 46–96% (mean 76%, Figure 2A–D, Table 3). A linear mixed effect model found no influence of CO₂, temperature, or their interaction on embryo survival (Table 4). However, there was a significant effect of experiment ($F(4, 118) = 33.581$, $p < 0.001$), because in experiment 1, high CO₂ reduced embryo survival (two-way ANOVA, $F(2, 28) = 18.965$, $p < 0.001$) by 11% at 17 °C (LSD, $p = 0.003$) and 24 °C (LSD, $p = 0.027$) relative to control treatments. This effect was not observed in subsequent experiments with less extreme CO₂ treatments.

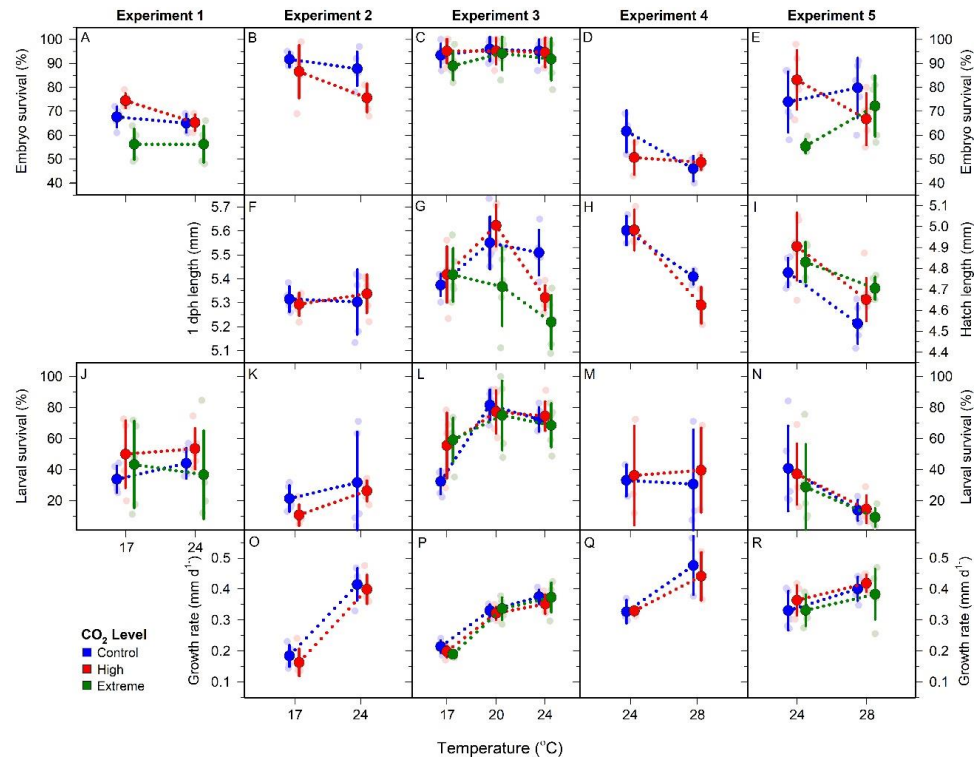


Figure 2. *M. menidia*. Offspring responses to control (blue), high (red), and extreme (green) CO₂ conditions at four temperatures across five CO₂ × temperature factorial experiments. Traits include embryo survival (A–E), hatch length (F–I), larval survival (J–N) and larval growth rate (O–R). Individual replicates are represented by small faded circles. Treatment means (±SD) are depicted by large, bold circles and connected by dotted lines. Note: different scales used for hatch length measurements due to differences in sample timing; panels F and G use 1dph length Y axis (left) while panels H and I use hatch length Y axis (right).

Table 4. Summary statistics for linear mixed models testing the effects of CO₂, temperature and their interaction (fixed factors) and experiment (random factor) on four response traits; embryo survival (ES), hatch length (HL), larval survival (LS), and growth rate (GR) of *M. menidia* offspring. Significant ($\alpha < 0.05$) factors are denoted by *p*-values in bold.

Trait	Factor	<i>F</i>	df	<i>p</i>
ES	CO ₂	2.992	2	0.058
	Temp	1.140	3	0.336
	CO ₂ × Temp	0.677	6	0.669
	Experiment	33.581	4	<0.001
HL	CO ₂	1.895	2	0.156
	Temp	19.518	3	<0.001
	CO ₂ × Temp	3.021	6	0.010
	Experiment	75.361	3	<0.001
LS	CO ₂	0.296	2	0.756
	Temp	9.429	3	<0.001
	CO ₂ × Temp	0.759	6	0.614
	Experiment	12.385	4	<0.001
GR	CO ₂	0.457	2	0.595
	Temp	77.964	3	<0.001
	CO ₂ × Temp	0.515	6	0.838
	Experiment	3.330	3	0.012

3.2. Hatch Length

Across experiments mean hatch SL ranged from 4.54–5.62 mm, but within experiment variation was small (± 0.4 mm) (Figure 2F–I, Table 3). A linear mixed effects model indicated significant temperature ($F(3, 90) = 19.518, p < 0.001$), $\text{CO}_2 \times \text{temperature}$ ($F(6, 90) = 3.021, p = 0.010$), and experiment effects ($F(3, 90) = 75.361, p < 0.001$) (Table 4). The $\text{CO}_2 \times \text{temperature}$ interaction was driven by divergent $\text{CO}_2 \times \text{temperature}$ responses between experiments 3 and 5. During experiment 3, a two-way ANOVA found hatch lengths were significantly affected by CO_2 ($F(2, 44) = 7.600, p = 0.002$), temperature ($F(2, 44) = 14.857, p < 0.001$), and their interaction ($F(4, 44) = 5.522, p = 0.001$). Hatch lengths were similar at 17 °C, but exposure to extreme CO_2 significantly reduced offspring size at 20 °C relative to the control treatment (one-way ANOVA, $F(2, 14) = 5.947, p = 0.016$). The CO_2 -induced reduction in hatch length was largest at 24 °C (one-way ANOVA, $F(2, 14) = 13.342, p = 0.001$), with exposure to high and extreme CO_2 reducing larval size by 3% (LSD, $p = 0.006$) and 5% (LSD, $p < 0.001$), respectively, compared to controls. During experiment 5, a two-way ANOVA found hatch lengths were significantly affected by CO_2 ($F(2, 28) = 5.222, p = 0.013$) and temperature ($F(1, 28) = 25.544, p < 0.001$). Contrary to experiment 3, there was no influence of CO_2 at 24 °C. However, at 28 °C, high and extreme CO_2 significantly increased hatch length (one-way ANOVA, $F(2, 14) = 5.942, p = 0.016$) by 3% (LSD, $p = 0.014$) and 4% (LSD, $p = 0.010$), respectively, compared to the control treatment.

3.3. Larval Survival

Larval survival was highly variable both within and between experiments, with treatment means ranging from 9–82% (mean 42%, Figure 2J–N, Table 3). Across all experiments, a linear mixed effects model found significant effects of temperature ($F(31, 21) = 9.918, p < 0.001$) and experiment ($F(41, 21) = 12.798, p < 0.001$), but no influence of CO_2 or $\text{CO}_2 \times \text{temperature}$ (Table 4). Within temperature mean ($\pm \text{SD}$) larval survival was highest at 20 °C ($78 \pm 14\%$), with reduced survival observed at 24 °C ($46 \pm 25\%$), 17 °C ($38 \pm 22\%$), and 28 °C ($19 \pm 19\%$).

3.4. Larval Growth Rate

Growth rates ranged from 0.16–0.48 mm d^{-1} (mean 0.33 mm d^{-1} , Figure 2O–R, Table 3). A linear mixed effects model identified significant temperature ($F(3, 92) = 80.189, p < 0.001$) and experiment ($F(3, 92) = 3.838, p = 0.012$) effects, but no influence of CO_2 or $\text{CO}_2 \times \text{temperature}$ (Table 4). Across experiments, growth rates increased similarly across CO_2 levels with increasing temperatures. Within-temperature means ($\pm \text{SD}$) ranged from 0.19 ± 0.03 , 0.33 ± 0.02 , 0.36 ± 0.05 , and 0.42 ± 0.06 mm d^{-1} at 17, 20, 24, and 28 °C, respectively.

3.5. Overall CO_2 Effect Size (LnRR)

The overall CO_2 effect on embryo survival was small in response to high CO_2 conditions (within ± 0.06) and similar across temperature treatments (Figure 3A). For offspring exposed to extreme CO_2 , all responses were negative (-0.04 to -0.30), but there was no apparent trend with temperature (Figure 3B). For hatch size, overall effects were small both at high and extreme CO_2 treatments (± 0.03), again with no apparent temperature dependency (Figure 3C,D). The overall effect of high CO_2 conditions on larval survival was highly variable (-0.64 to 0.55) and overall neutral across temperatures (Figure 3E). Interestingly, the effect of extreme CO_2 conditions on larval survival was positive at 17 °C (0.42) but became increasingly negative with increasing temperatures (-0.18 at 24 °C and -0.44 at 28 °C, Figure 3F). Average CO_2 effects for growth rate were small (within ± 0.10), but exhibited a dome-shaped response across temperatures at both CO_2 levels, with negative growth responses at sub-optimal rearing temperatures (Figure 3G,H).

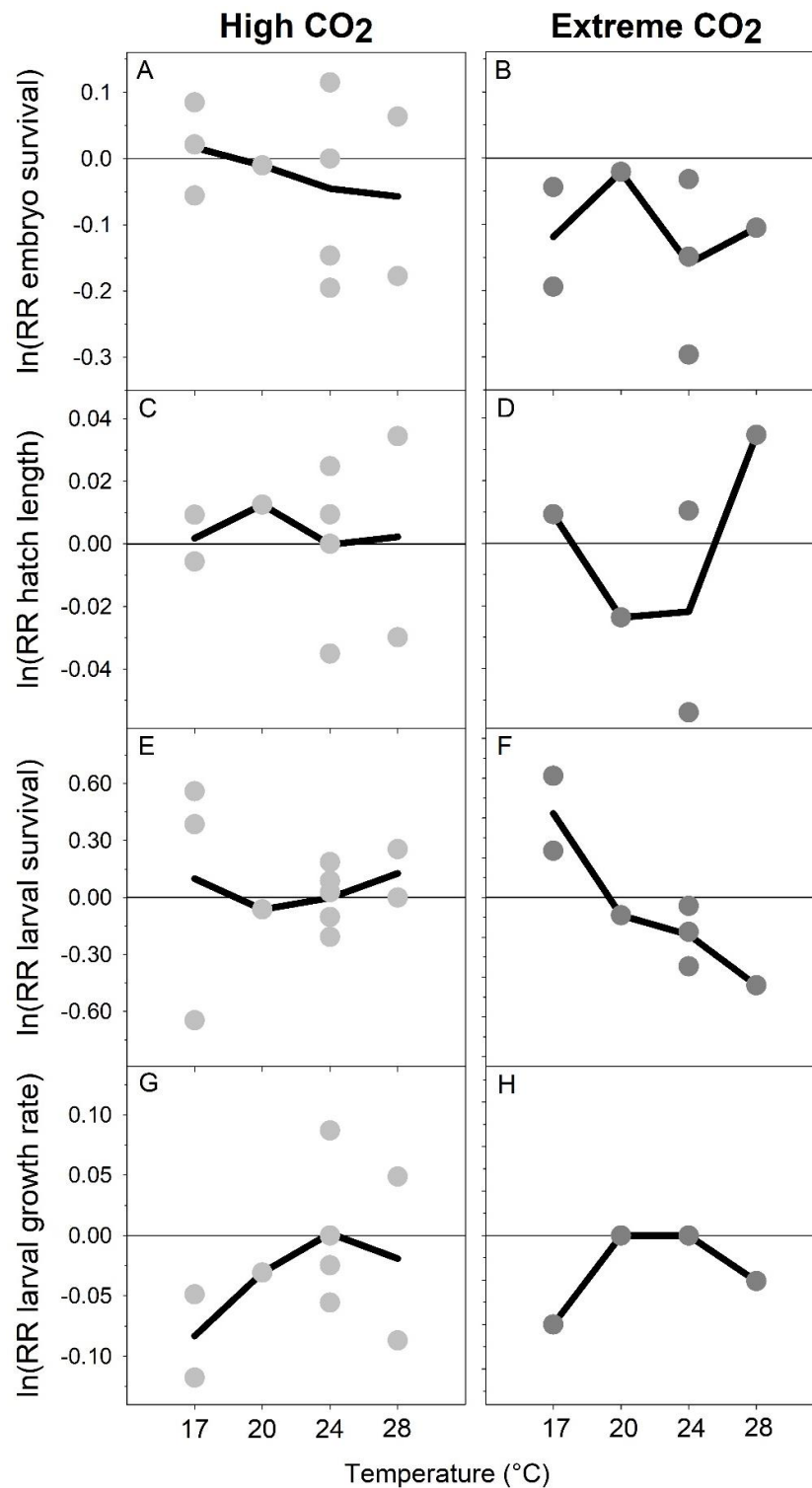


Figure 3. *M. menidia*. CO₂ effect sizes using log-transformed response ratios (lnRR) of high (light grey) and extreme (dark grey) CO₂ exposure across four rearing temperatures. Response traits include embryo survival (A,B), hatch length (C,D), larval survival (E,F), and growth rate (G,H). Circles represent lnRRs of each experiment, while black lines represent lnRRs averaged across experiments at each rearing temperature. Negative (positive) values indicate a trait decrease (increase) at elevated CO₂ levels compared to control CO₂ conditions.

4. Discussion

We conducted five factorial experiments to evaluate the sensitivity of *M. menidia* early life traits to high (2000–2800 μatm) and extreme CO_2 conditions (4000–6200 μatm) across four temperatures (17, 20, 24, and 28 °C) that encompassed contemporary and potential future conditions in nearshore silverside spawning habitats. The experiments showed few significant CO_2 effects on response traits. Significant reductions in embryo survival occurred at 17 and 24 °C in a single experiment and at the most extreme CO_2 treatment (~6000 μatm). Effects on hatch length showed evidence for $\text{CO}_2 \times$ temperature interactions, given that elevated CO_2 reduced hatch length at 24 °C during one experiment, while increasing hatch length at 28 °C during another. There were no significant effects of CO_2 on larval survival or growth rate. Together, these findings suggest that *M. menidia* offspring can tolerate high to extreme CO_2 levels across most of the species' thermal range.

The apparent CO_2 resilience of *M. menidia* offspring may reflect the pH/ CO_2 variability typical of their nursery habitat. Atlantic silversides spawn in shallow subtropical to temperate estuaries [75] where seasonal acidification elicits increasingly large diel pH fluctuations while progressively reducing daily mean and minimum pH levels [35]. Such patterns of seasonal pH/ CO_2 variation appear to be common in shallow nearshore habitats [36]. As a batch-spawning fish, silversides spawn fortnightly from late April to early July (at ~41° N) which coincides with the period of most rapid habitat acidification [61]. Thus, a single female will deposit subsequent batches of embryos into a progressively more pH variable and acidic environment. In a previous study, we found offspring CO_2 tolerance closely tracked temporal trends in habitat acidification [61]. Transgenerational plasticity is a possible explanation for this rapid shift, by which adults experiencing a progressively more acidic environment augment offspring phenotypes to better match current environmental conditions [76]. An additional source of CO_2 tolerance may arise from local adaptation. Despite being an annual fish with high population connectivity, Atlantic silversides exhibit local adaptation for traits involved in growth and environmental sex determination [77], which are likely maintained through the continuous selection of locally suited genotypes [78]. As previously demonstrated, early life survival under high CO_2 conditions is a heritable trait in this species, suggesting that CO_2 tolerance could evolve [63]. Local adaptation to acidified habitats through the selection and maintenance of CO_2 -tolerant traits has been demonstrated in other taxa [29]. For Atlantic silversides that spawn in habitats prone to acidification, adaptations that enable high- CO_2 tolerance are likely well represented in wild populations and would explain the observed CO_2 tolerance. Importantly, we found that exposure to ~6000 μatm p CO_2 did reduce embryo survival during experiment 1, while offspring were largely tolerant to ~4200 μatm in subsequent experiments. While 6000 μatm is an extreme, likely unrealistic CO_2 level for silverside spawning habitats, it may represent a tolerance threshold for *M. menidia*. Identifying such thresholds are necessary to accurately assign an organisms' sensitivity to future climate change [79].

Maternal provisioning of eggs through modifications of energy content or fatty acid composition may further influence offspring CO_2 sensitivity [58]. Such differences may have contributed to CO_2 effects on hatch length documented during experiments 3 and 5. In Atlantic cod, CO_2 induced reductions in hatch size were not accompanied by increased utilization of yolk reserves during embryogenesis [80]. The authors suggest that yolk utilization was already maximized, and increased demands on acid/base regulation resulted in a shift of endogenous energy use away from somatic growth. Conceivably, CO_2 reductions in hatch size during experiment 3 were the result of a similar mechanism. Conversely, differences in maternal provisioning of embryos from experiment 5 may have stimulated yolk utilization under elevated CO_2 , leading to increased embryonic growth and hatch size [13]. Fish embryos passively experience their environment with fixed energy reserves [81] but are likely most sensitive to elevated CO_2 [11]. Further investigations are needed into how $\text{CO}_2 \times$ temperature combinations influence embryo energetics.

The apparent tolerance of *M. menidia* offspring to combined climate stressors contrasts with the growing evidence for compounding effects of near future OA and warming in the early life stages

of other fish species. For example, combined treatments synergistically decreased embryo survival in Antarctic dragon fish (*Gymnodraco acuticeps*) [48] and compromised temperature acclimation and aerobic performance in emerald rockcod (*Trematomus bema*) [49]. As extreme stenotherms, polar species appear particularly vulnerable to combined climate effects [82], but eurythermal temperate species have demonstrated similar sensitivities. For example, exposure to acidification and warming reduced hatch size and larval survival in the Senegalese sole (*Solea senegalensis*) [14] and Atlantic cod (*Gadus morhua*) [80]. In the congeneric *M. beryllina*, a large reduction in survival was found when simultaneously exposed to high- CO_2 and 29 °C [59]. The $\text{CO}_2 \times$ temperature tolerance demonstrated by *M. menidia* offspring is likely a manifestation of conditions widely experienced by wild silverside early life stages. The acidification of their near shore nursery habitat is largely driven by seasonal changes in community respiration that generally peak with seasonally maximum water temperatures [36]. Thus, simultaneous occurrence of potentially stressful temperature and CO_2 levels are a regular feature of *M. menidia* spawning habitat. Furthermore, because seasonal habitat changes are of the same direction and similar magnitude to climate projections, existing phenotypic or genetic variation already present in silverside populations may confer some degree of tolerance to future marine climate change [83].

While *M. menidia* early life stages appear resistant to elevated CO_2 across a broad thermal regime, the addition of other stressors could potentially be detrimental. For example, temperature-dependent metabolic processes that drive coastal acidification simultaneously consume oxygen; hence, warming, acidification, and hypoxia co-occur in *M. menidia* nursery habitats [36,84]. Given their co-occurrence in nature, physiological responses to elevated CO_2 and low DO are likely connected. Intermediate CO_2 exposure can elicit important adaptive responses which may mediate sublethal effects of low DO [84], yet more extreme exposures may act synergistically to elevate stressor sensitivity [85]. Thus, factorial $\text{CO}_2 \times \text{DO} \times$ temperature experiments would be insightful for more robust characterizations of coastal climate effects on fish early life stages.

Whole lifecycle effects of elevated CO_2 exposure remain critically understudied in fish [25]. While acclimation to chronic hypercapnia likely has small metabolic costs [86], over longer timescales tradeoffs associated with increased acid/base regulation could compromise other physiological processes [43]. In a previous study, we documented small but significant size reductions in *M. menidia* reared under ~2200 μatm CO_2 and 17 °C for approximately a third of their lifespan [62]. Importantly, differences in length were only detected after two months of continuous high- CO_2 exposure. In the present study, CO_2 effect sizes calculated for growth rates displayed dome-shaped response curves, with more negative responses at sub-optimal rearing temperatures. For offspring reared under 17 °C and high CO_2 , the average growth effect size was -0.08 (i.e., -8%), a response of similar magnitude to previously documented growth reductions under the same conditions after four months [62]. Importantly, that study used large sample populations (>2000 individuals) providing the necessary power to statistically confirm a CO_2 effect. Arguably, many early-life experiments with smaller sample sizes lack the power to robustly detect small effects [87]. Thus, it is possible that small or undetectable CO_2 reductions in early-life growth accrue and become detectable during long-term exposures. Even minor changes to early life development may have important carry-over effects to later life stages and ultimately impact fitness [88]. As an annual fish, juvenile growth during summer is critically important for *M. menidia*, as larger individuals have higher overwintering survival [89]. How warming temperatures may interact with CO_2 over longer time-scales is presently unknown and represents a serious gap in our understanding of how combined climate stressors will impact fish [90].

Across experiments, CO_2 responses were highly complex, consistent with previous OA studies on silverside offspring [58,59,61]. Experiments produced functionally different outcomes within equivalent treatment conditions despite meticulously controlled experimental conditions. For all traits but growth rate, inter-experiment variation was more substantial than variability driven by CO_2 or temperature level. A portion of this variability could be elicited by small differences in food quantity or quality, water source, or realized CO_2 levels, but parentage likely constitutes the largest source of

variation in offspring, mediated through genetic or phenotypic inheritance and maternal provisioning. A limitation of the present study was the use of embryos from a single group of spawning adults for each experiment. This precluded the incorporation of parentage into the statistical analysis, thereby potentially underestimating this source of variation as previously reported in *M. menidia* [58]. Plastic offspring responses to CO₂ × temperatures conditions are likely adaptive in species like *M. menidia* that spawn in highly dynamic systems [91]. Thus, this inherent plasticity precludes broad generalizations based on single, short-term experiments, but understanding such plasticity is fundamental in assigning potential risks to ongoing climate change [83,92]. Highly variable CO₂ × temperature responses are common across taxonomic groups [90,93], thus experimental replication and inter-experiment statistical comparisons are necessary for robust evaluations of climate sensitivities in marine organisms.

In summary, we analyzed five CO₂ × temperature experiments together to robustly characterize CO₂ × temperature responses of important fitness related traits in *M. menidia* offspring. While individual experiments demonstrated some negative CO₂ effects, overall responses were largely neutral. Importantly, we found sub-optimal rearing temperatures did not increase sensitivity to even extreme CO₂ levels. Repeated experimentation documented substantial inter- and intra-experiment variability, highlighting the need for experimental replication to accurately describe inherently variable response traits.

Supplementary Materials: The following are available online at <http://www.mdpi.com/1424-2818/10/3/69/s1>, Table S1: Adult spawner counts and length measurements.

Author Contributions: C.S.M. and H.B. designed the experiments, C.S.M. carried out the experiments, C.S.M. and H.B. wrote the manuscript.

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