



Effects of Long-Term Cover Cropping on Weed Seedbanks

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Cool-season cover crops have been shown to reduce soil erosion and nutrient discharge from maize (*Zea mays* L.) and soybean [*Glycine max* (L.) Merr.] production systems. However, their effects on long-term weed dynamics are not well-understood. We utilized five long-term research trials in Iowa to quantify germinable weed seedbank densities and compositions after 10+ years of cover cropping treatments. All five trials consisted of zero-tillage maize-soybean rotations managed with and without the inclusion of a yearly winter rye (*Secale cereal* L.) cover crop. Seedbank sampling was conducted in the early spring before crop planting at all locations, with three of the five trials having grown a soybean crop the preceding year, and two a maize crop. Two of the trials (both previously soybean) showed significant and biologically relevant decreases (4,070 and 927 seeds m^{-2} , respectively) in seedbank densities in cover crop treatments compared to controls. In another two trials, one previously maize and one previously soybean, no difference was detected in seedbank densities. In the fifth trial (previously maize), there was a significant, but biologically unimportant increase of 349 seeds m^{-2} . All five trials' weed communities were dominated by common waterhemp [*Amaranthus tuberculatus* (Moq.)], and changes in seedbank composition from cover-cropping were driven by changes in this species. Although previous studies have shown that increases in cover crop biomass are strongly correlated with weed suppression, in our study we did not find a relationship between seedbank changes and the mean amount of cover crop biomass produced over a 10-years period (experiment means ranging from 0.5 to 2.0 Mg $ha^{-1} yr^{-1}$), the stability of the cover crop biomass production, nor the amount produced going into the previous crop's growing season. We conclude that long-term use of a winter rye cover crop in a maize-soybean system has the potential to meaningfully reduce the size of weed seedbanks compared to winter fallows. However, identifying the mechanisms by which this occurs requires further research into processes such as seed predation and seed decay in cover cropped systems.

Keywords: maize (*Zea mays* L.), cover crop, sustainable weed management, corn belt, waterhemp [*Amaranthus tuberculatus* (Moq.) J. D. Sauer], germinable seed bank

INTRODUCTION

One-third of the global maize (*Zea mays* L.) and soybean [*Glycine max* (L.) Merr.] production comes from the United States (US; Food and Agriculture Organization of the United Nations, 2020). The majority of US production occurs in the Midwest region (USDA, 2020a), and 80% of the agricultural land in the two top-producing states, Iowa and Illinois, is dedicated to a rotation

consisting solely of these two crops (USDA, 2020b). Maize-soybean cropping systems traditionally leave the soil fallow over the winter and early spring, resulting in high levels of nutrient and soil export that render the sustainability of the system questionable (O'Neal et al., 2005; Dold et al., 2017; Nearing et al., 2017; Jones et al., 2018). Incorporation of an over-wintering rye (*Secale cereal* L.) cover crop into these systems can significantly reduce soil erosion and nutrient leaching (Strock et al., 2004; Kaspar and Singer, 2011; Kaspar et al., 2012), and may offer additional long-term benefits to the soil (Moore et al., 2014; Basche et al., 2016a,b). Surveys indicate farmers consider cover crops to be a valuable component of an integrated approach to weed management (Arbuckle and Lasley, 2013). Moreover, ecologically-based approaches to weed management such as cover crops are becoming more critical as weeds develop herbicide resistance to multiple modes of action (Patzoldt et al., 2005; Price et al., 2011; Bunchek et al., 2020; MacLaren et al., 2020). However, the effects of over-wintering cover crops on weed dynamics in these systems is not well-understood.

There is evidence cover crops can reduce weed biomass in many production contexts (Baraibar et al., 2018; MacLaren et al., 2019; Smith et al., 2020), and specifically in midwestern maize-soybean systems (Nichols et al., 2020a). In other production systems, there is also evidence cover crops can reduce weed seed densities in the soil (Moonen and Bàrberi, 2004; Mirsky et al., 2010; Alonso-Ayuso et al., 2018) and decrease the survival success of herbicide-resistant weeds (Cholette et al., 2018; Wallace et al., 2019). However, the majority of relevant studies have been conducted in plots where cover crop treatments were in place <3 years, so the long-term effect of cover cropping on weed dynamics in these systems is unclear. The density and species composition of emerged weeds can vary greatly from year to year based on weather conditions, rendering the more subtle effects of management practices difficult to discern in short-term studies (Teasdale et al., 2018). Additionally, weed seeds can persist in the soil for several years, creating legacy effects that can overwhelm short-term changes in management. Measurements taken in long-term, replicated settings may therefore more accurately reflect management-induced changes.

Aboveground measurements of weeds are useful, but the potential for annual weed species to interfere with crop growth and yield is ultimately an expression of the weed seedbank. In the midwestern US, management practices that target weed seedbanks are particularly relevant, as the majority of problematic weeds are annual species whose persistence depends on replenishing seedbanks (Davis, 2006). While seedbank sizes are of primary concern, the seedbank composition can provide insight into weed dynamics and differences in composition can be used to assess the relative strength of the filters defining the weed community (e.g., Ryan et al., 2010). Additionally, there is some evidence that crop yield loss and weed diversity are negatively correlated (Adeux et al., 2019) and more diverse assemblages of weed seeds in the soil may reflect the impacts of more sustainable management strategies (Storkey and Neve, 2018). Information about the size and composition of weed seedbanks after two or more full crop rotation sequences may

therefore provide a more complete picture of weed responses to cover cropping.

To address the lack of data concerning long-term effects of cover cropping on weed seedbanks in maize-soybean systems, we measured the size and composition of the germinable weed seedbank sampled from five trials in Iowa where rye cover crop treatments had been in place for at least 10 years. We hypothesized that long-term use of over-wintering rye cover crops in maize-soybean rotation systems would: (1) reduce the size and (2) increase the species diversity of the weed seedbank.

METHODS AND MATERIALS

Site Descriptions

Three research sites were used for this study (Table 1). The West and East sites were grain production fields on commercial farms, and only one phase of the maize-soybean rotation was present each year. The Central site had both a grain-based maize-soybean rotation and a silage-based rotation. In the silage rotation, the maize phase was harvested for silage at the milk stage (R3; Abendroth et al., 2011). The trials were part of a larger study (Kaspar et al., 2007, 2012) and had both phases of the rotations of both systems (grain- and silage-based) present each year, but each phase was located in a separate field.

All trials consisted of two treatments that had been in place for at least 10 years: (1) a maize-soybean rotation (either grain- or silage-based) with a winter rye cover crop planted in the fall following cash crop harvest and terminated in the spring, and (2) the same rotation without a cover crop. Every trial was arranged in a randomized complete block design with four (West and East) or five (Central) replicates. More detailed accounts of agronomic management at the Central site have been published elsewhere (Moore et al., 2014). None of the studies were originally set-up with the goal of assessing weed dynamics; as such there are unfortunately no baseline measurements of the weed seedbank available.

The plots within each trial were managed identically save for the planting of the cover crop in the fall. All sites applied herbicide 1–2 weeks before maize or soybean planting to all plots and in certain years an additional herbicide application shortly after cash crop planting (Table 2). The exact herbicide and nutrient programs varied by site, reflective of their particular managers and contexts (Supplementary Material). All sites had sub-surface tile drainage and were managed without tillage since initiation of the trials.

Weed Seedbank Sampling

Midwestern row crop production fields typically have early spring seedbank densities well-above 500 seed m⁻² in the top 10 cm of the soil profile (Forcella et al., 1992; Felix and Owen, 2001). For these expected values, 20 soil samples 5 cm in diameter are expected to provide a high level of precision when estimating seedbank densities (Dessaint et al., 1996; Forcella et al., 2003). We used these estimates to guide our sampling protocol.

A soil sampler was constructed using PVC pipe with an inner diameter of 5.25 cm and a line indicating 10 cm sampling depth to extract a total of 52.5 cm³ of soil per core. In no-till systems,

TABLE 1 | Summary of the four trials sampled.

Trial	Latitude, longitude	Year started	Number of replicates	Plot size	30-years annual mean		Mean cover crop biomass (Mg ha ⁻¹)		2018 crop	2019 sampling date
					Air temperature (°C)	Precipitation (mm)	5-years	10-years		
WEST										
1	42°03'N 94°20'W	2008	4	25 × 250 m	9.5	880	0.24	0.45	Soybean	April 17
CENTRAL SILAGE										
2	42°00'N 94°12'W	2002	5	3.8 × 55 m	9.8	907	2.38	1.98	Soybean	April 16
CENTRAL GRAIN										
3	42°00'N	2009	5	3.8 × 55 m	9.8	907	1.53	0.88	Soybean	April 8
4	94°12'W						1.93	1.34	Maize	April 9
EAST										
5	41°19'N 92°17'W	2009	4	25 × 275 m	10.2	947	1.73	1.32	Maize	April 6

TABLE 2 | Summary of herbicide active ingredients applied at each site during 2017–2019 growing seasons.

Site	Maize year			Soybean year		
	Pre-plant	At planting/post-emergence	Herbicide groups	Pre-plant	At planting/post-emergence	Herbicide groups
West	Glyphosate	Metolachlor Atrazine Mesotrione	5, 9, 15, 27	Glyphosate	Glyphosate Fluthiacet-methyl	9, 14
Central	Glyphosate	Metolachlor Atrazine Mesotrione	5, 9, 15, 27	Glyphosate	Glyphosate Hand weeding in late July	9
East	Glyphosate Acetochlor	Atrazine Acetochlor Glyphosate	5, 9, 15	Glyphosate Chlorimuron-ethyl Flumioxazin Pyroxasulfone	Dicamba Acetochlor	2, 4, 9, 14, 15

The same herbicide treatments were applied to the cover-crop and no-cover plots.

this represents a generous depth from which most midwestern US weed seedlings can emerge (Mohler, 1993), so we assumed our sampling efforts accurately recovered seeds with the potential to contribute to weed infestations in maize and soybean crops.

Sampling was done in April 2019 at all locations. Each plot was divided longitudinally into five sampling areas. Within each sampling area, four cores were taken. The East and West locations' plots were wide, so the cores were taken randomly within each of the five sampling areas. For the Central sites, which had narrower plots, the cores were taken from the middle of the sampling area to minimize edge effects. Within each sampling area, four cores were taken and the soil was emptied into a bucket, thoroughly mixed, then placed in a sealed polyethylene bag and stored for a maximum of 5 h in a cooler for transportation. Each plot had a total of 1,050 cm³ of soil sampled (20 cores, each 52.5 cm³). Sampling occurred before maize (West, Central-grain)

or soybean (East, Central-grain, Central-silage) planting at each site. At the Central site, both phases of the grain rotation were sampled, while only one phase of the silage rotation was sampled due to time constraints (Table 1).

Germinable Seedbank Measurements

The germinable seedbank method was chosen over the extractable seedbank method based on practicality, and its applicability for assessing treatment differences (Reinhardt and Leon, 2018). Soil processing, as described below, occurred on the same day as collection.

The field-wet soil was weighed to ensure each plot had approximately the same mass of soil sampled (Supplementary Material). The soil from each plot's five

sampling points was then combined and sieved through a 5 mm wire mesh screen into a bucket and transported to a greenhouse.

Plastic 25 × 50 cm trays with drainage holes were filled with vermiculite to a depth of 1 cm (Greenhouse Megastore, Danville, Illinois, US). The bulked soil from each plot was evenly distributed into three trays, creating a 1 cm soil layer covering the vermiculite. The soil was saturated using a three-hole fine-mist brass nozzle (Greenhouse Megastore). The greenhouse area had no artificial lighting and was maintained near 28°C. Germination from soil samples occurred between April and July, during a period with 13–15 h of daylight.

Trays were checked 1–3 times per day to ensure proper germination conditions. Weed seedlings were identified, counted, and pulled daily, after which the trays were randomly relocated within the greenhouse to avoid the effects of micro-environments on germination. When no new seedlings appeared for at least 3 days, the tray was allowed to dry in order to avoid conditions that would promote decay of un-germinated seeds. Once all trays were dried (~2 months after sampling), each tray's soil was recollected, re-sieved, redistributed into the same tray, and again saturated. This process was repeated twice, and after the second soil re-sieving no seedlings emerged. The total number of emerged seedlings was reported as the seedbank density.

Cover Crop Biomass Sampling

Cover crop biomass was sampled in each trial since initiation. For the East and West experiments, four 76 × 48 cm quadrats were collected per plot before cover crop termination. For the Central experiments, cover crop biomass was sampled before cover crop termination using an 81 × 30 cm quadrat, with two quadrat samples per plot. Only two quadrats were used at this site because the plots were small and removing more biomass could affect the long-term plots' integrity. Biomass from all sites was dried at 60°C for at least 48 h and then weighed. Carbon-to-nitrogen ratios of the biomass were collected in select years, but due to the inconsistency of data collection those results are not presented here. No other cover crop metrics were collected (e.g., height, stand count, stage). Mean values at each trial for each year are available in **Supplementary Material** and in the published dataset (Nichols et al., 2020b).

Data Analysis

The raw dataset is available on Iowa State University's DataShare platform (Nichols et al., 2020b) and as an R package available on github (<https://github.com/vanichols/PFIweeds2020>). All data management, visualization, and statistical analyses were conducted using R version 3.6.1 (R Core Team, 2016); all code is publicly available (https://github.com/vanichols/PFIweeds2020_analysis). The *tidyverse* meta package (Wickham et al., 2019) was used for data manipulation and visualization, in addition to several other packages (Becker et al., 2018; Wickham and Bryan, 2018; Wilke, 2019). All packages used for statistical analyses are cited below.

Seedbank Size

The number of emerged seedlings was assumed to represent the seedbank density. The distribution of measured seedbank densities exhibited a high right-skewness and over-dispersion typical of count data. Several candidate statistical models were evaluated, and the detailed exploration process can be found in an online format (<https://lydiae.com/2020/04/22/many-models/>). We chose to use a generalized linear mixed-effect model (McCulloch and Neuhaus, 2005) using a log-linked Poisson distribution and observation-level random effects to account for overdispersion (Harrison, 2014), fit using the *glmer* function from the *lme4* package (Bates et al., 2015). We used the trial as a fixed effect, which had five levels (West, Central-grain/soybean, Central-grain/maize, Central-silage/soybean, East). Additionally, the cover crop treatment (cover, no-cover) and its interaction with the trial were included as fixed effects. In addition to the random intercept for each observation to address overdispersion, we included a random intercept term for the blocks nested within the trial. All pair-wise comparisons were conducted using the *emmeans* package (Lenth et al., 2018), which calculates the least-squares means and computes contrasts. Raw seedling counts were converted to seeds m⁻² based on the PVC sampling tube diameter.

We ran a leave-one-out sensitivity analysis wherein the statistical model was run on datasets with one observation removed to explore the sensitivity of our results to any single experimental unit. One cover-cropped plot in the West location had a waterhemp [*Amaranthus tuberculatus* (Moq.)] seed count of more than 16,000 seeds m⁻², while the plot with the next highest observed waterhemp density at that site (a no-cover plot) was <10,000 seeds m⁻². We ran all models both with and without the outlier (**Supplementary Material**), and found it did affect the magnitude of the cover crop treatment effect in that experiment, but not the direction of the effect. We felt this large value may not be a realistic representation of the actual seed density in the plot, as the producer did not recall that plot having twice the weed biomass of other plots (*personal comm*). Due to the mixing of individual soil cores that was done in the field, it is not possible to isolate whether this large value was caused by a single core. Waterhemp plants grown in highly competitive environments can still produce 10,000 seeds (Schwartz et al., 2016), so it is conceivable we captured the seed rain from a single plant. We chose to present the results with the outlier removed as we felt it was more representative, but note the effect of the outlier when relevant throughout the results.

We used a first- and second-order stochastic dominance analysis to compare the cumulative distribution curves of seedbank size for no-cover and cover-cropped production systems (Levy, 1992). Stochastic dominance is a tool commonly used in risk-assessments to identify scenarios with a higher probability of favorable outcomes (e.g., Goplen et al., 2018). We assumed producers want to minimize the size of the weed seedbank, and therefore used the inverse of the cumulative probability distributions to assess outcomes from using a cover crop compared to no cover crop. Comparing the cumulative distributions at a given value of weed seedbank densities provides information concerning outcomes of a practice (first-order),

while comparing the area under the cumulative probability curves provides information about the risk associated with a particular practice (second-order).

To quantify cover crop biomass production for each trial, we calculated different metrics to capture varying functional aspects of the cover crop that might affect weed seedbanks. Using cover crop biomass data from 2009 through 2019, we calculated the following metrics using both the previous 10 years of data and only the previous 5 years: (1) mean biomass production, (2) variance in biomass production, (3) maximum biomass production, (4) number of years with $>1 \text{ Mg ha}^{-1}$ production, (5) number of years with $>2 \text{ Mg ha}^{-1}$ production, (6) mean-to-standard-deviation ratio of biomass production (stability), and (7) biomass production the year of sampling, as well as the year prior.

We used non-parametric Spearman rank correlations to assess the association between the metrics listed above and cover crop effect on seedbank densities (relative and absolute).

Seedbank community composition

Changes in the weed seedbank community were assessed using both uni- and multivariate approaches. For the univariate approach, linear mixed-effect models with trial, cover crop treatment, and their interaction as fixed effects and random intercepts for nested blocks were used to assess the impact of cover cropping on seedbank diversity metrics. Our diversity metrics included species richness, Shannon Hill diversity and evenness (Jost, 2006) for each experimental unit (a plot) using the raw seedling counts and the following equations:

$$\text{Shannon Hill diversity} = \exp(H')$$

$$\text{Evenness} = \frac{H'}{\log(S)}$$

Where

$$S = \text{species richness}$$

$$H' = - \sum_i^N p_i \log(p_i)$$

Evenness describes how a given species richness is distributed and ranges from 0 to 1, with 1 signifying all species are equally present. Shannon Hill diversity can be interpreted as the “effective” number of species; when evenness is 1, Shannon Hill diversity is equal to species richness.

For the multivariate approach, species composition was compared across trials and cover crop treatments using non-metric multidimensional scaling (NMDS) implemented through the *vegan* package (Oksanen et al., 2019). NMDS assists in visualizing and analyzing similarities between groups of individuals (Prentice, 1977). The removal of rare species from multi-variate analyses can impact interpretations (Poos and Jackson, 2012), so we performed all analyses on both the full dataset and on a dataset containing only the species comprising

more than 5% of the observations and found the results did not change. Bray-Curtis dissimilarities were calculated on raw seed counts. Variation in distance matrices were partitioned into trial and cover crop contributions using permutations implemented through the *adonis* function of the *vegan* package. The *adonis* function works by creating permutations of the raw data wherein data is randomly assigned to a group. It finds the centroids and squared-deviations for each permutation, and by comparing those values to the raw data, the significance tests reflect the probability of observing the true data assuming no group structure exists.

RESULTS

Weed Seedbank Size

The West (previous crop of soybean) trial had the largest estimated mean seedbank size ($5,647 \text{ seeds m}^{-2}$), followed by the Central-silage (previously soybean; 935 seeds m^{-2}), with the Central-grain (previously maize, soybean) and East (previously maize) locations having similarly low mean densities ($382\text{--}482 \text{ seeds m}^{-2}$; **Figure 1**). Due to a significant interaction between trial and cover crop treatment, all results are reported on a per-trial basis.

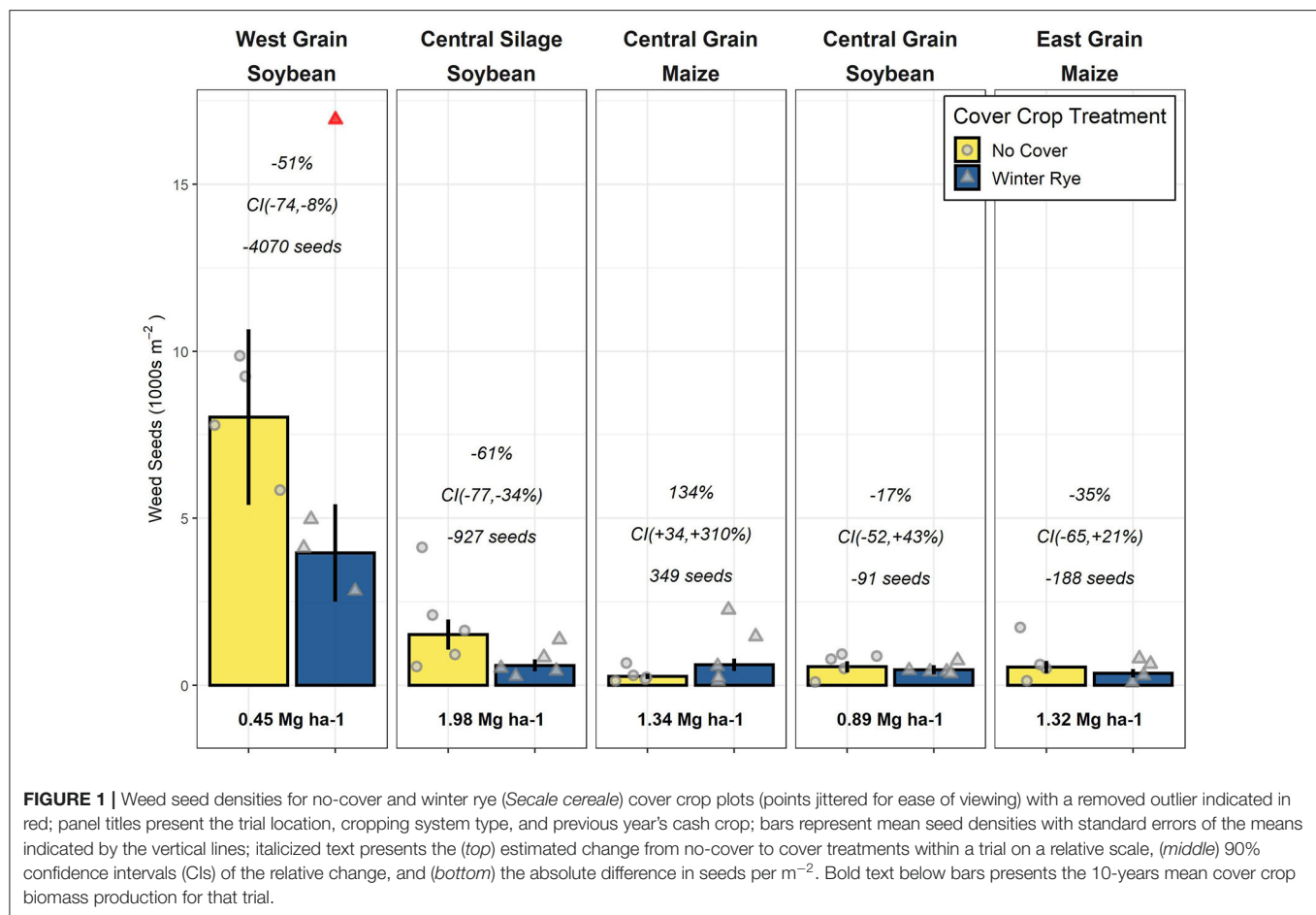
In the three trials with a soybean crop the preceding year, the seedbank density was lower in the cover crop treatment compared to the no-cover treatment by 91 (Central-grain), 927 (Central-silage), and 4,070 (West) seeds m^{-2} , respectively, corresponding to a 17, 61, and 51% reduction. The magnitude of the West results were sensitive to the inclusion of the outlier (**Supplementary Material**), but the direction of the effect was not. In the trials previously planted to maize, seedbank densities in cover cropped plots were lower in one trial (East; reduced by 188 seeds m^{-2} , 35%) and increased in another (Central-grain; increased by 349 seeds m^{-2} , 134%).

Neither the absolute nor relative differences between the cover crop and control treatments were meaningfully related to any of the cover crop biomass metrics we calculated (**Supplementary Material**).

Weed Seedbank Community

A total of 4,677 seedlings were counted, consisting of 16 identified species (**Table 3**). Seven seedlings were identified as belonging to the *Setaria* genus, but the species was unclear. The species common to the Midwest, *Setaria faberi* Herrm., *Setaria viridis* (L.) P. Beauv, and *Setaria pumila* (Poir.) Roem. & Schult., for this reason we combined the *Setaria* seedlings into one category for reporting and analysis. Unidentified dicotyledon seedlings were classified as “unknown dicotyledon (UD)” and accounted for 8 of the seedlings, respectively, representing $<0.2\%$ of the data (**Table 3**). Because they made up such a small contribution to the overall community we left them in the analysis, but labeled as unknown.

We note that using the germination-method is known to cause varying underestimation of species. The method can bias counts toward species responsive to the particular conditions used, for example by specifically underestimating species with



long seed dormancies or seeds that were not sufficiently stratified the previous winter (Gross, 1990).

The changes in seedbanks were driven by changes in the number of waterhemp (*Amaranthus tuberculatus*, AMATU) seeds (Figure 2).

Differences in seedbank composition were strongest at the trial level ($p < 0.01$) and were not statistically significant for cover crop treatment. In the trials with significant cover cropping effects, community changes were again driven mainly by a decrease in waterhemp (Figures 2, 3) which resulted in a slight shift toward a more grass-dominated community, but the effect was not strong. Community shifts were not consistently associated with an increase in the Shannon-Hill diversity index, species evenness, nor species richness (Table 4), but these results must be taken in context of the herbicide programs (Table 2).

Risk of Increasing Seedbanks

Results from the stochastic dominance analysis indicate that at low weed seedbank densities (<300 seeds m⁻²), cover cropping and control treatments did not differ, whereas at higher weed seedbank densities (>300 seeds m⁻²), cover cropping consistently exhibited lower densities than the no-cover treatments (Figure 4).

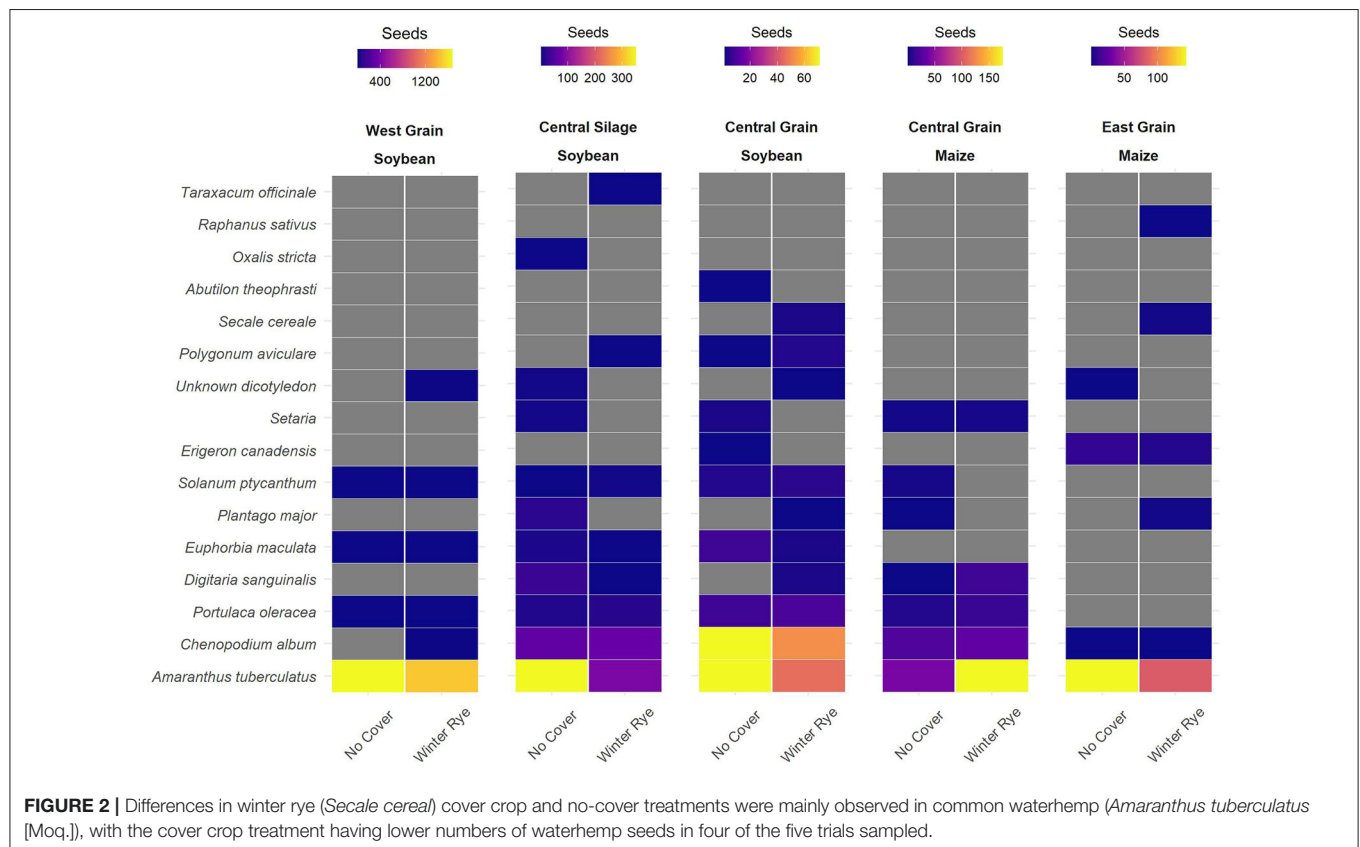
DISCUSSION

Cover Crop Interactions With Waterhemp

As is the case in many midwestern maize and soybean fields, waterhemp was the driver weed species in the locations sampled in this study (van Wyche, 2017, 2019). Due to the dominance of waterhemp in the weed communities of this study, the herbicide programs implemented at each trial may have provided contexts where cover crop effects on weeds would vary. Waterhemp populations with resistance to seven herbicide groups have been identified (Tranel, 2020), with populations resistant to one or more are prevalent in the midwest (Patzoldt et al., 2005; Chatham et al., 2015; Oliveira et al., 2017). While we did not measure the presence of resistance genes, resistances to herbicide groups 5, 9, 14, and 27 were likely present (Owen, 2017) with possible resistances to group 15 (Hager, 2019). The varying levels of waterhemp control via herbicides and the use of residuals in the different trials may have rendered cover cropping more or less effective (Table 2). The West site, where the largest absolute reduction in waterhemp with the use of cover cropping was observed (Figure 1), did not utilize chemistries that would reliably kill resistant waterhemp plants that had already emerged, nor a residual herbicide to suppress future waterhemp emergence. Of the sites included in this study, the West site was

TABLE 3 | Summary of weed species identified in this study in order of prevalence.

Code	Scientific name	Common name	Description	Percent of total found
AMATU	<i>Amaranthus tuberculatus</i> (Moq.) J. D. Sauer	Waterhemp	C ₄ forb	88.58%
CHEAL	<i>Chenopodium album</i> L.	Lamb's quarters	C ₃ forb	6.67%
POROL	<i>Portulaca oleracea</i> L.	Purslane	C ₄ grass	1.28%
DIGSA	<i>Digitaria sanguinalis</i> (L.) Scop.	Large crabgrass	C ₄ grass	1.00%
SETARIA ^a	<i>Setaria faberi</i> Herm., <i>Setaria viridis</i> (L.) P. Beauv., <i>Setaria pumila</i> (Poir.) Roem. & Schult., unknown species	Foxtail	C ₄ grass	0.56%
EPHMA	<i>Euphorbia maculata</i> L.	Spotted spurge	C ₄ forb	0.41%
PLAMA	<i>Plantago major</i> L.	Plantain	C ₃ forb	0.41%
SOPT7	<i>Solanum ptychanthum</i> Dunal	Eastern black nightshade	C ₃ forb	0.34%
ERICA	<i>Erigeron canadensis</i> L.	Horseweed	C ₃ forb	0.32%
UD ^b	-	-	-	0.17%
POLAV	<i>Polygonum aviculare</i> L.	Prostrate knotweed	C ₃ forb	0.11%
SECCE	<i>Secale cereale</i> L.	Cereal rye	C ₃ grass	<0.10%
ABUTH	<i>Abutilon theophrasti</i> Medik.	Velvet leaf	C ₃ forb	<0.10%
OXAST	<i>Oxalis stricta</i> L.	Yellow woodsorrel	C ₃ forb	<0.10%
RAPSR	<i>Raphanus sativus</i> L.	Radish	C ₃ forb	<0.10%
TAROF	<i>Taraxacum officinale</i> F. H. Wigg.	Dandelion	C ₃ forb	<0.10%

^aSeedlings identified as belonging to the *Setaria* genus were combined.^bUnknown dicot.**FIGURE 2** | Differences in winter rye (*Secale cereale*) cover crop and no-cover treatments were mainly observed in common waterhemp (*Amaranthus tuberculatus* [Moq.]), with the cover crop treatment having lower numbers of waterhemp seeds in four of the five trials sampled.

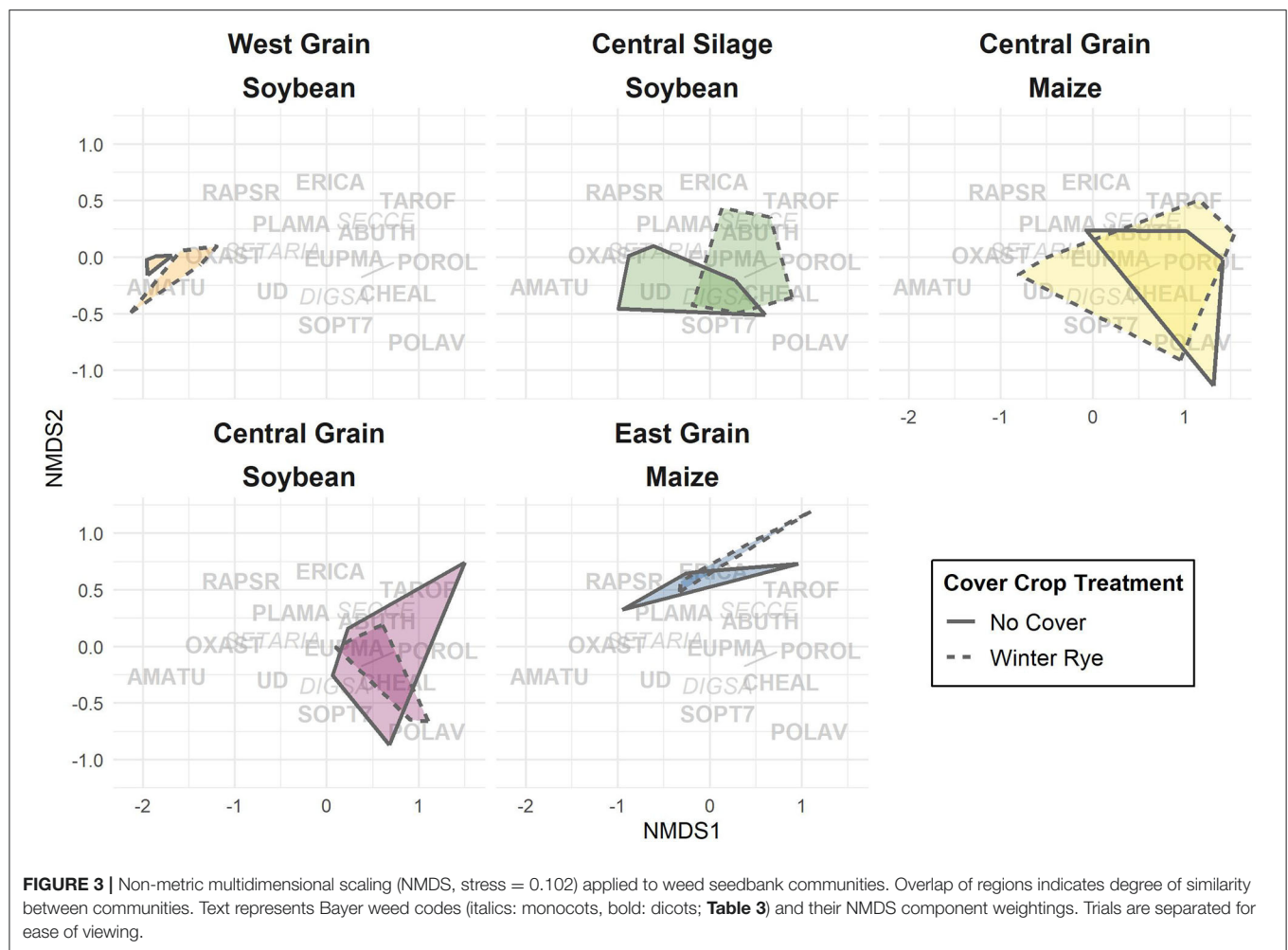


TABLE 4 | Estimated changes, standard errors (SE), and *p*-values for models of changes in diversity, richness, and evenness for each trial.

	Shannon Hill diversity		Richness		Evenness	
	Change (SE)	<i>P</i> -value	Change (SE)	<i>P</i> -value	Change (SE)	<i>P</i> -value
West-grain (Soybean)	0.05 (0.49)	0.92	1.5 (0.86)	0.09	0.03 (0.13)	0.81
Central-silage (Soybean)	−0.22 (0.43)	0.51	−2.2 (0.77)	0.01	0.18 (0.10)	0.10
Central-grain (Maize)	−0.33 (0.43)	0.45	0.6 (0.77)	0.44	−0.15 (0.10)	0.13
East-grain (Maize)	0.18 (0.48)	0.38	0.8 (0.87)	0.39	0.03 (0.11)	0.82
Central-grain (Soybean)	0.42 (0.43)	0.97	0.4 (0.77)	0.60	0.03 (0.10)	0.78

Estimate values show expected changes with the inclusion of a cover crop (ex. positive values indicate that metric increased with the inclusion of a cover crop). Trials are listed in descending order of absolute change in seedbank size with cover cropping. Significant differences at $p < 0.10$ are indicated with bold italics.

therefore most susceptible to waterhemp living and setting seed, and therefore may have provided the biggest opportunities for cover cropping effects to manifest. The Central trials utilized hand-weeding in soybean phases late in the season, which may have reduced the opportunities for cover crops effects to be expressed. The East site, which had the lowest average seedbank densities observed in this study, utilized a herbicide program that would control resistant waterhemp biotypes and included residual herbicide that would also reduce/delay waterhemp emergence, perhaps leaving little room for cover cropping effects.

While the previous crop of the individual trials may also play a role in dictating the weed responses to cover cropping, in the present study the previous crop is confounded with site effects, so it is difficult to draw conclusions from what may be spurious associations.

Regardless of the mechanisms involved, we believe our results regarding reductions in waterhemp seed densities are robust. Under no-till management and in the absence of new inputs to the seedbank, waterhemp seed densities can decline >99% after 5 years (Steckel et al., 2007). Our plots

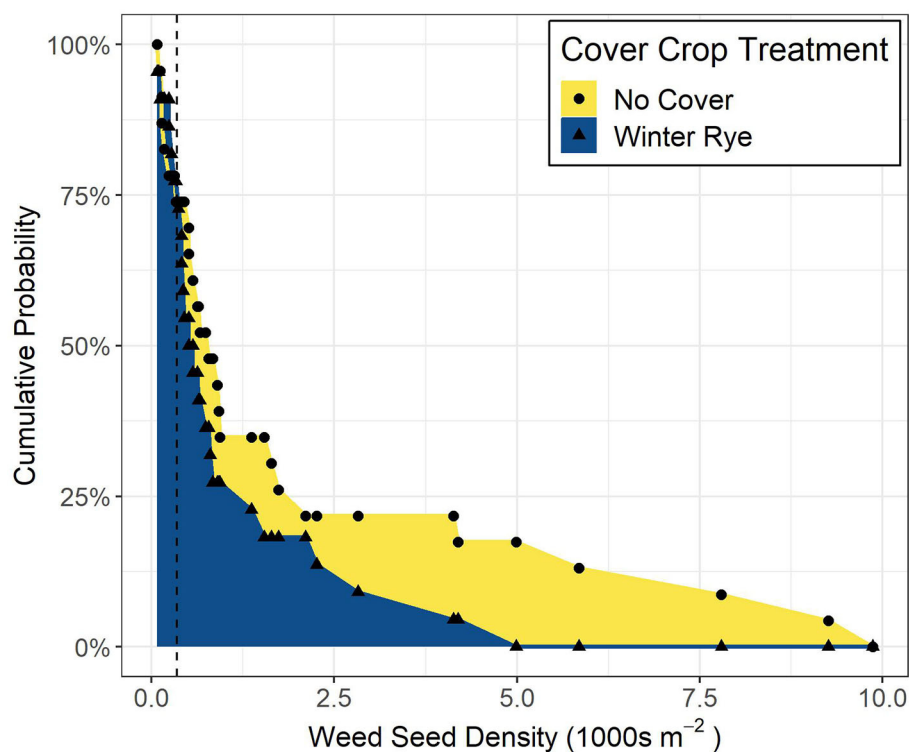


FIGURE 4 | Each point represents the cumulative probability of having a weed seedbank of that density or higher; as seedbank densities increase above 300 seeds m^{-2} (dashed line), the probability of having a larger seedbank is higher for no-cover systems compared to cover-cropped systems; the area under the curve is proportional to the risk of increasing weed seedbank densities.

have been in place long enough (>10 years) for treatment effects to be detected despite possible legacy effects of the original seedbanks.

Cover Crop Mechanisms of Weed Suppression

Previous research indicates a cover crop's potential for in-season weed suppression is strongly related to the cover crop's biomass production (Baraibar et al., 2018; MacLaren et al., 2019; Nichols et al., 2020a; Smith et al., 2020). Cover crops might reduce weed seedbank densities via several mechanisms, all of which could conceivably be intensified with increases in the quantity of cover crop biomass produced. It is thus surprising that in the present study neither the absolute nor relative effects of cover cropping on the weed seedbank was related to any of the cover crop biomass production metrics we evaluated. For example, while the West trial consistently produced <1 $Mg\ ha^{-1}$ of cover crop biomass, it exhibited the largest absolute decreases in weed seedbank size from cover cropping (Figure 1). Even considering the herbicide program, it is surprising such small amounts of cover crop biomass could have meaningful effects on the weed seedbank.

The pattern in the present study might also be related to the emergence timing of waterhemp, which can extend well-beyond the time the cover crop is killed while maintaining high reproductive success (Wu and Owen, 2014). In other

studies, weed communities may not have been dominated by late germinating weed species such as waterhemp, and the communities might have therefore been more directly responsive to cover crop biomass.

It is also possible cover crops affected weeds in ways less directly dependent upon the amount of cover crop biomass produced. The act of planting the cover crop itself may have provided some weed control. Additionally, even modest amounts of biomass present over the winter may provide enough ground cover to promote seed mortality through granivore activity (Carmona and Landis, 1999; Heggenstaller et al., 2006), and in the spring cover crop mulch can provide habitat for seed-eating invertebrates (Pullaro et al., 2006). Allelopathic compounds from rye residue may catalyze pathogen attack on seeds and reduce the vigor of germinated seeds (Barnes and Putnam, 1983; Mohler et al., 2012), and production of these compounds may be more dependent upon growing conditions compared to rye biomass production *per se* (Mwaja et al., 1995). While our study did not test these effects directly, our data suggest these mechanisms should be considered when assessing the effects of cover cropping on weed communities.

We note that with endpoint sampling, as we did in the present study, it is difficult to link the cumulative effect of 10 years of biomass production with one season's weed seedbank. Sampling weed seedbanks yearly would enable a more direct connection to be drawn between cover crop biomass production

and seedbank densities. However, our data show that in the contexts we sampled, the weed suppressive potential of cover crops in the long-term was not directly related to cover crop biomass production.

Cover Crop Effects on Weed Seed Communities

The lack of a consistent and significant effect of cover cropping on the structure of the weed seed communities in the present study is consistent with the findings of other studies (Moonen and Bàrberi, 2004; Smith et al., 2015; Alonso-Ayuso et al., 2018). It is unsurprising that an over-wintering cover crop would be a weak filter in systems dominated by summer annuals that are well-adapted for regeneration in maize-soybean rotations (Tranel, 2020). In systems with more diverse cropping systems or seedbanks, cover crops might create more marked shifts in weed communities.

While the germination method may have failed to identify or underestimated weed species resulting in an underestimation of weed species richness, the number of weed species found in each plot (ranging from 1 to 8) matches field-based observations of maize-soybean rotations (Hirsh et al., 2013). Additionally, the dominance of common waterhemp rendered the Shannon Hill diversity and evenness metrics insensitive to small contributions by other species. Accordingly, our results may be due to the already-simplified nature of the communities, where random variation easily obscures subtle signals in the less prevalent species.

CONCLUSIONS

Our study is the first we are aware of that quantifies the long-term impacts of cover cropping on weeds in the midwestern United States. We found evidence that cover cropping can meaningfully reduce the size of the weed seedbank compared to a no-cover control in certain contexts. More research in long-term plots comparing cover crop effects in various cropping systems and management regimes is needed to identify conditions in which cover crops are most effective at reducing and/or preventing weed seed deposits. Endpoint sampling, used in our study, is useful in assessing whether systems merit more attention, but longitudinal samplings of weed seedbanks in long-term studies are needed to better assess the seedbank trajectories of these systems.

Changes in seedbanks were driven by change in densities of common waterhemp, a weed resistant to multiple herbicide modes of actions. We found that when weed seed densities are above 300 seeds m^{-2} , cover cropping exhibits no risk of enlarging weed seedbanks compared to no-cover systems.

In the production contexts examined, the amount of cover crop biomass produced was not associated with the magnitude of cover crop effects on weed seedbanks. The lack of relationship

suggests cover crop biomass may not be the best metric for predicting long-term impacts of cover-cropping on weeds in all systems, particularly those dominated by late-germinating species such as waterhemp. Cover crops may suppress weeds through a combination of mechanisms, and the relative contribution likely varies by site and/or year. Parsing out these effects could aid in the design of systems better able to take advantage of cover crop weed suppression.

DATA AVAILABILITY STATEMENT

The datasets presented in this study can be found in csv format in an online repository accessed via <https://doi.org/10.25380/iastate.12762011.v1> and as an R package at <https://github.com/vanichols/PFIweeds2020>.

AUTHOR CONTRIBUTIONS

VN, ML, and SC designed the study. SG and SC contributed data and project management. VN facilitated data collection and wrote the original draft of the manuscript. VN, ML, and LE analyzed the data. All authors contributed to editing of the final manuscript. All authors contributed to the article and approved the submitted version.

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SUPPLEMENTARY MATERIAL

The Supplementary Material for this article can be found online at: <https://www.frontiersin.org/articles/10.3389/fagro.2020.591091/full#supplementary-material>

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Conflict of Interest: The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

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