

# Atmospheric degradation of cyclic volatile methyl siloxanes: Radical chemistry and oxidation products

3 *Mitchell W. Alton<sup>1,2</sup> and Eleanor C. Browne<sup>\*1,2</sup>*

4 <sup>1</sup>Department of Chemistry, University of Colorado, Boulder, Colorado, 80309, United States

7 *KEYWORDS:* *VMS, Siloxanes, Atmospheric Oxidation, Peroxy Radical, Oxidized VMS,*  
8 *Contaminants of Emerging Concern*

## 9 ABSTRACT

10 Cyclic volatile methyl siloxanes (cVMS) are anthropogenic chemicals that have come under  
11 scrutiny due to their widespread use and environmental persistence. Significant data on the  
12 environmental concentrations and persistence of these chemicals exists, but their oxidation  
13 mechanism is poorly understood, preventing a comprehensive understanding of the environmental  
14 fate and impact of cVMS. We performed experiments in an environmental chamber to characterize  
15 the first-generation oxidation products of hexamethylcyclotrisiloxane (D3),  
16 octamethylcyclotetrasiloxane (D4), and decamethylcyclopentasiloxane (D5) under different  
17 peroxy radical fates (unimolecular reaction or bimolecular reaction with either NO or HO<sub>2</sub>) that

18 approximate a range of atmospheric compositions. While the identity of the oxidation products  
19 from D3 changed as a function of the peroxy radical fate, the identity and yield of D4 and D5  
20 oxidation products remained largely constant. We compare our results against output from a  
21 kinetic model of cVMS oxidation chemistry. The reaction mechanism used in the model is  
22 developed using a combination of previously proposed cVMS oxidation reactions and standard  
23 atmospheric oxidation radical chemistry. We find that the model is unable to reproduce our  
24 measurements, particularly in the case of D4 and D5. The products that are poorly represented in  
25 the model help to identify possible branching points in the mechanism which require further  
26 investigation. Additionally, we estimated the physical properties of the cVMS oxidation products  
27 using structure activity relationships and found that they should not significantly partition to  
28 organic or aqueous aerosol. The results suggest that cVMS first-generation oxidation products are  
29 also long-lived in the atmosphere and that environmental monitoring of these compounds is  
30 necessary to understand the environmental chemistry and loading of cVMS.

## 31 1 INTRODUCTION

32 Cyclic volatile methyl siloxanes (cVMS) are high production volume chemicals<sup>1,2</sup> that are  
33 common components of consumer and industrial products such as deodorants, lotions, sealants,  
34 and lubricants. cVMS have high vapor pressures and low Henry's law constants, leading to  
35 preferential partitioning into the atmosphere<sup>3</sup> where they primarily degrade through reactions with  
36 OH.<sup>4</sup> cVMS lifetimes with respect to oxidation by OH are between 4 and 10 days (assuming a 24-  
37 hour average OH concentration of  $1.2 \times 10^6$  molecules  $\text{cm}^{-3}$ ),<sup>4-12</sup> and consequently, cVMS are  
38 globally distributed and have been identified in remote environments such as the arctic.<sup>13-15</sup> These  
39 abundant chemicals are also bioaccumulative and toxic.<sup>13,16-18</sup> Consequently, the European Union

40 placed restrictions on the use of cVMS in certain cosmetic products in 2016, with  
41 recommendations in 2021 to restrict the use in certain industrial processes.<sup>19–21</sup> Owing to their  
42 widespread use, toxicity, and environmental persistence, there has been significant interest in  
43 understanding the environmental fate of cVMS.<sup>5,12,13,16,17,22–38</sup>

44 Previous research has largely focused on quantifying ambient concentrations of parent  
45 cVMS<sup>13,35,38,39</sup> and understanding their oxidation kinetics;<sup>5,10,11</sup> significantly less is understood  
46 about the identity of cVMS oxidation products and their environmental chemistry.<sup>31,32</sup> In the  
47 atmosphere, cVMS oxidation is initiated by reactions with OH or Cl to produce an alkyl radical  
48 ( $\text{R}_3\text{SiCH}_2\cdot$ ) which quickly reacts with  $\text{O}_2$  to produce a peroxy radical ( $\text{R}_3\text{SiCH}_2\text{O}_2\cdot$ , or more  
49 generally,  $\text{RO}_2$ ). The subsequent reactions depend on atmospheric composition: reactions with NO  
50 will dominate under high NO mixing ratios, such as in urban areas; reactions with  $\text{HO}_2$  will  
51 dominate where high  $\text{HO}_2$  mixing ratios are found, such as in areas dominated by biogenic  
52 emissions; and unimolecular reactions can dominate in areas with low concentrations of NO or  
53  $\text{HO}_2$ , such as remote regions or even in urban areas for reactions with sufficiently fast unimolecular  
54 reactions.<sup>40</sup>

55 Laboratory investigations of cVMS oxidation have observed numerous oxidation products but  
56 typically identify the siloxanol (-CH<sub>3</sub> replaced with -OH) as the main product.<sup>4,7,8</sup> Two  
57 experimental studies have also reported the formate ester (-CH<sub>3</sub> replaced with -CH(O)H)  
58 product.<sup>4,5</sup> Additionally, oxidation products such as hydroperoxides, alcohols ( $\text{R}_3\text{SiCH}_2\text{OH}$ ), and  
59 products with an increased number of silicon atoms have been detected.<sup>8,41</sup> Siloxanol formation  
60 has often been ascribed to the hydrolysis of the formate ester,<sup>4,7</sup> a process that likely occurs on  
61 surfaces. Attempts to develop an oxidation mechanism based on these observations and known  
62 atmospheric peroxy radical chemistry have largely been unsuccessful, in part because the  $\text{RO}_2$  fate

63 has been unclear in many of the past experiments. It has generally been concluded that the Si in  
64 cVMS allows for unique chemistry to occur and several unusual reactions have been proposed  
65 based on the laboratory results.<sup>4,8,42</sup> Recent theoretical investigations of organosilicon chemistry  
66 provide evidence that organosilicon compounds undergo unique rearrangements inaccessible to  
67 carbon-based compounds.<sup>31,32</sup> The net effect of the mechanism proposed by the theoretical  
68 investigations oxidation under high NO<sub>x</sub> (= NO + NO<sub>2</sub>) conditions includes the production of HO<sub>2</sub>  
69 and the oxidation of two NO radicals to NO<sub>2</sub>.<sup>31,32</sup> This net effect is inconsistent with laboratory  
70 measurements showing that at most one NO is oxidized to NO<sub>2</sub> during cVMS oxidation and that  
71 cVMS reduces O<sub>3</sub> production and OH radical concentrations.<sup>18</sup> However, as cVMS oxidation  
72 products were not measured there is still opportunity for better online characterization of the  
73 chemistry.

74 Without an understanding of the cVMS oxidation mechanism and products, it is difficult to  
75 predict the fate of the oxidation products. For instance, one possibility is that the oxidation products  
76 contribute to aerosol mass. Based on observations of Si in urban nanoparticles it has been  
77 suggested that oxidation products of cVMS from personal care products are present in  
78 aerosol,<sup>24,39,43</sup> while other studies have suggested that the silicon in aerosol from organosilicon  
79 compounds is minimal.<sup>44,45</sup> There are substantial variations in the reported aerosol yields of cVMS,  
80 which range from approximately 0% to 50%.<sup>46</sup> The range in the reported yields is likely due to  
81 experimental conditions creating different oxidation products. Without knowledge of the oxidation  
82 mechanism and measurements of the oxidation products, however, this hypothesis is difficult to  
83 test.

84 In this study, we conducted a series of atmospheric chamber experiments designed to investigate  
85 the reaction mechanisms of hexamethylcyclotrisiloxane (D3), octamethylcyclotetrasiloxane (D4),

86 and decamethylcyclopentasiloxane (D5) under conditions of different RO<sub>2</sub> lifetimes and reaction  
87 partners. We also oxidized fully deuterated hexamethyldisiloxane (D<sub>18</sub>L2) with Cl atoms to  
88 provide some constraint on the identity of the oxidation products. Using a combination of our  
89 measurements, kinetic modeling, and constraints from the literature, we propose that a simplified  
90 oxidation mechanism for cVMS that can be used in modeling atmospheric cVMS chemistry. We  
91 identify points in the cVMS oxidation mechanism that require further investigation. The  
92 environmental fate of the observed products is investigated through estimations of vapor pressure  
93 and water solubility, with the results suggesting that the first-generation of cVMS oxidation does  
94 not create oxidation products that will participate in aerosol growth.

## 95 **2 MATERIALS AND METHODS**

96 Experiments were performed in a ~1 m<sup>3</sup> FEP Teflon® chamber at ambient lab temperature (295  
97 ± 3 K) and pressure (~860 mbar) at low relative humidity (<5% RH). A low RH was used to  
98 minimize the wall loss of the oxidation products. The chamber, previously described in Alton and  
99 Browne (2020),<sup>5</sup> was run in either batch mode, in which the chamber was slowly collapsed as air  
100 was sampled from it, or semi-batch mode, in which the sampled air was continuously replaced. To  
101 correct for dilution in semi-batch mode, the dilution rate was quantified by sampling for 30 minutes  
102 after completion of the experiment and determining the first order loss constant for each product.  
103 This rate constant represents both loss to the walls and dilution and was used to correct the  
104 measured values for individual species. Experiments for D3 were conducted in both batch and  
105 semi-batch modes and product distributions were determined to be consistent between the two  
106 methods of chamber operation on the timescale we are interested in (~3 hours). Experimental

conditions are summarized in Table 1.

**Table 1** Radical precursor concentrations, estimated oxidant concentrations, and estimated ratios of reaction between RO<sub>2</sub> and NO to HO<sub>2</sub> and HO<sub>2</sub> to other RO<sub>2</sub>.

cVMS <sup>a</sup>	Oxidant Precursor and Concentration (ppb <sub>v</sub> ) <sup>b</sup>	[OH] or [Cl] (molec cm <sup>-3</sup> )	[NO], [NO <sub>2</sub> ] (ppb)	NO:HO <sub>2</sub> <sup>c</sup>	τ <sub>RO2</sub> (s) <sup>d</sup>
D3* <sup>e</sup>	Cl <sub>2</sub> , 15	1.0×10 <sup>6</sup>	-	0.4:1	27
D3*	H <sub>2</sub> O <sub>2</sub> , 1000	8.0×10 <sup>7</sup>	-	0.04:1	4
D3*	HONO, 400	5.0×10 <sup>7</sup>	200,150	200:1	0.02
D3	Cl <sub>2</sub> , 15	1.0×10 <sup>6</sup>	-	0.4:1	27
D4	Cl <sub>2</sub> , 15	6.0×10 <sup>5</sup>	-	0.9:1	40
D4	H <sub>2</sub> O <sub>2</sub> , 1000	8.0×10 <sup>7</sup>	0.05, 8	0.04:1	4
D4	HONO, 400	5.0×10 <sup>7</sup>	230,190	100:1	0.02
D4	Cl <sub>2</sub> , 100	7×10 <sup>5</sup>	-	0.02:1	1.6
	CH <sub>2</sub> O, 1000				
D5	Cl <sub>2</sub> , 15	4.0×10 <sup>5</sup>	0.05, 0.05	1:1	46
D5	H <sub>2</sub> O <sub>2</sub> , 1000	8.0×10 <sup>7</sup>	0.05, 4	0.04:1	4
D5	HONO, 400	5.0×10 <sup>7</sup>	250,180	100:1	0.02

<sup>a</sup>Approximately 80 ppb<sub>v</sub> of cVMS was added to the chamber. <sup>b</sup>The initial mixing ratios of NO and NO<sub>2</sub> before the lights were turned on in the HONO experiments were ~50 ppb<sub>v</sub>, and ~100 ppb<sub>v</sub>, respectively, except for D3 which had ~100 ppb<sub>v</sub> of NO. Differences in mixing ratios are due to the inconsistency of HONO generation using HNO<sub>3</sub> and NaNO<sub>2</sub>. NO and NO<sub>2</sub> were only measured in experiments for which mixing ratios are reported. For other experiments, background mixing ratios of 50 ppt<sub>v</sub> were assumed. <sup>c</sup>Estimated radical concentrations were determined at the point that 10% of the cVMS has reacted, as calculated from the KinSim model.

<sup>d</sup>Rate constants used in calculating the RO<sub>2</sub> lifetimes were obtained from Ziemann and Atkinson (2012)<sup>52</sup>: k<sub>RO2+HO2</sub> = 1.5×10<sup>-11</sup> molec cm<sup>-3</sup> s<sup>-1</sup>, k<sub>RO2+NO</sub> = 9×10<sup>-12</sup> molec cm<sup>-3</sup> s<sup>-1</sup>, k<sub>RO2+RO2</sub> = 1×10<sup>-14</sup> molec cm<sup>-3</sup> s<sup>-1</sup>, and k<sub>RO2+OH/Cl</sub> = 2×10<sup>-10</sup> molec cm<sup>-3</sup> s<sup>-1</sup>. <sup>e</sup>Experiments marked with (\*) were performed in semi-batch mode, in which sampled air is continually replaced with clean air to maintain a constant chamber volume for the duration of the experiment.

109 D3 (98%, Acros Organics), D4 (98%, Acros Organics), D5 (97%, Sigma-Aldrich) were diluted  
110 in acetonitrile (99.8%, Sigma-Aldrich) to 5% (w/w) and injected into the chamber using a gently  
111 heated borosilicate glass tube with a stream of zero-air (AADCO Instruments, 737 series)  
112 transporting the evaporated cVMS to the chamber. We saw no evidence for thermal degradation  
113 of the cVMS from heating of the tube (maximum temperature estimated at less than 60°C). The  
114 oxidation chemistry of each precursor was investigated under three different RO<sub>2</sub> fates. Peroxy  
115 radical lifetimes were calculated using measured concentrations of NO or estimated concentrations  
116 of reactive species from a kinetic model, which will be discussed later in Section 3.3. Experiments  
117 using Cl<sub>2</sub> as an oxidant precursor were designed to favor unimolecular RO<sub>2</sub> reactions. For these  
118 experiments we calculate a RO<sub>2</sub> lifetime with respect to bimolecular reactions of about 0.4 minutes  
119 with ~55% of the RO<sub>2</sub> reacting with HO<sub>2</sub> and 20% undergoing unimolecular rearrangements  
120 (assuming a rate constant of  $8 \times 10^{-3}$  s<sup>-1</sup> for isomerization),<sup>31</sup> and the rest reacting with background  
121 NO. Experiments using H<sub>2</sub>O<sub>2</sub> as an oxidant precursor probed conditions where >90% of RO<sub>2</sub> react  
122 with HO<sub>2</sub> while experiments using HONO resulted in ~99% of RO<sub>2</sub> reacting with NO.

123 Cl<sub>2</sub> was added by a flow of N<sub>2</sub> over a gravimetrically calibrated permeation device (VICI  
124 Metronics). Hydrogen peroxide was added via a solution added into the heated borosilicate glass  
125 tube with a zero-air flow. HONO was generated by addition of 40 μL of a 1.5 M sodium nitrite  
126 solution (>99%, Sigma-Aldrich) to a 2 M nitric acid solution (70%, Fischer Scientific) while  
127 passing zero-air over the headspace of the solution, transporting any evolved gases into the  
128 chamber. Cl<sub>2</sub> and HONO were photolyzed into Cl atoms and OH + NO, respectively, with 370 nm  
129 fluorescent lights (General Electric, F40BL) positioned below the chamber. H<sub>2</sub>O<sub>2</sub> was photolyzed  
130 with 254 nm lights (General Electric, G36T8) to generate OH. Control experiments were  
131 performed to ensure the parent cVMS were stable in the presence of UV light. To test for

132 photostability of hydroperoxides generated from D4 oxidation, we performed a control experiment  
133 where we photolyzed Cl<sub>2</sub> with the 370 nm lights with ~1 ppm<sub>v</sub> of added formaldehyde to generate  
134 HO<sub>2</sub>. Partway through the experiment, the 254 nm lights were also turned on with the 370 nm  
135 lights. The production of the hydroperoxides increased at the same rate as the siloxanol, suggesting  
136 that more Cl<sub>2</sub> was photolyzed with the additional lights increasing total oxidation, but there was  
137 no additional loss of the hydroperoxide from photolysis.

138 As in our previous work,<sup>5</sup> a high-resolution long time-of-flight chemical ionization mass  
139 spectrometer (CIMS; Aerodyne Research, Inc. and Tofwerk AG; resolving power ~8000  
140  $m/z/\Delta m/z$ ) using protonated toluene as the reagent ion measured the cVMS parent compounds and  
141 oxidation products during the experiments. Compounds are detected as [M+H]<sup>+</sup> products. The  
142 ionization is relatively soft; we do not detect the methane fragment (R<sub>3</sub>Si<sup>++</sup>) from the parent cVMS  
143 that is generally seen in other proton transfer ionization schemes. We did observe R<sub>3</sub>Si<sup>+</sup> as a  
144 fragment of the siloxanol product, though less than 3% of the siloxanol signal was detected as this  
145 fragment. The CIMS sampled a total of 1.7 slpm, with 700 sccm of that flow consisting of  
146 humidified zero-air and the rest sampled from the chamber. The addition of water vapor enhances  
147 the instrument response (counts per second per ppb<sub>v</sub> normalized to the toluene reagent ion signal;  
148 ncps) for detection of cVMS and the oxidation products. Data was post-processed in Tofware  
149 v3.2.3 (Tofwerk AG) in the IGOR Pro environment (Wavemetrics, v8.0.4.2) using fully  
150 constrained peak fitting and allocation of isotope signals. An example of the peak fitting is shown  
151 in Section 1 of the Supporting Information (Figure S1). NO and NO<sub>2</sub> concentrations were  
152 measured using a chemiluminescence NO and NO<sub>2</sub> analyzer with a blue light converter for true  
153 NO<sub>2</sub> measurements (Teledyne, T200UP) with one minute resolution and 50 ppt<sub>v</sub> limit of detection.

154 We used KinSim v4.14<sup>47</sup> in Igor Pro to simulate gas-phase chemistry for each experiment.  
155 KinSim is open-source solver for kinetics modeling. Rate constants for non-cVMS species were  
156 obtained from Atkinson et al. (2004)<sup>48</sup> and Atkinson (2007).<sup>49</sup> The mechanism used in the KinSim  
157 simulations is presented in Section 2 of the Supporting Information. The lifetime of cVMS species  
158 with respect to oxidation by OH (~days) requires the use of elevated radical concentrations to  
159 ensure sufficient oxidation on the timescale of our experiment. However, the experiments were  
160 designed to avoid potentially unrepresentative reactions in the atmosphere (such as  $\text{RO}_2 + \text{RO}_2$  or  
161  $\text{RO}_2 + \text{OH}$ ). We assumed the chamber was well mixed. Wall partitioning calculations for all cVMS  
162 products were included using KinSim's built in wall-partitioning functions.<sup>50</sup> The reversible  
163 vapor–wall interactions were calculated upon mechanism compilation using the first-order vapor  
164 condensation rate coefficient ( $1 \times 10^{-3} \text{ s}^{-1}$ ),<sup>51</sup> the mass vapor saturation concentration of the  
165 partitioning molecule, and enthalpy of vaporization, with the latter two values estimated by the  
166 Melting Point, Boiling Point, and Vapor Pressure module (MPBPWIN v1.44) in the Environmental  
167 Protection Agency's (EPA) Estimations Programs Interface for Windows (EPIWIN v4.11). Rate  
168 constants for peroxy radical reactions with NO and  $\text{HO}_2$  were taken from Atkinson and Ziemann  
169 (2012)<sup>52</sup> and were kept constant. As the oxidation rate constants for the different parent cVMS are  
170 known,<sup>5</sup> the decay of the cVMS was used to constrain the concentration of oxidants, which was  
171 then used to determine the photolysis rates of the oxidant precursors. An average photolysis rate  
172 was used for all experiments using the same oxidant precursor ( $7 \times 10^{-5} \text{ s}^{-1}$ ,  $2 \times 10^{-4} \text{ s}^{-1}$ , and  $4 \times 10^{-4} \text{ s}^{-1}$   
173 for  $\text{H}_2\text{O}_2$ ,  $\text{Cl}_2$ , and  $\text{HONO}$ , respectively).

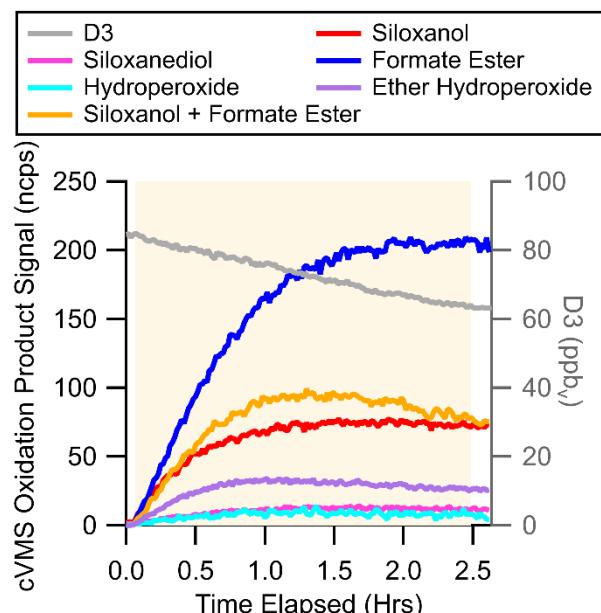
174 For estimating partitioning between condensed phases in the atmosphere we used EPIWIN to  
175 calculate the water solubility with the Water Solubility and Octanol-Water Partition Coefficient  
176 (WSKOW v1.43) and the Henry's Law constant with the HenryWin v3.21 module. In addition to

177 EPIWIN, the EPA's Toxicity Estimation Software Tool (TEST) was used to estimate the vapor  
178 pressures, boiling points, and water solubilities of the cVMS and their oxidation products. TEST  
179 does not estimate Henry's Law constants; however, a proxy of the ratio between the vapor pressure  
180 and the water solubility of the compound can be used to compare between the EPIWIN and  
181 TEST.<sup>53</sup> TEST reports a consensus value where it estimates the values using multiple methods  
182 then averages those results together. These estimations are further discussed in Section 3.4.

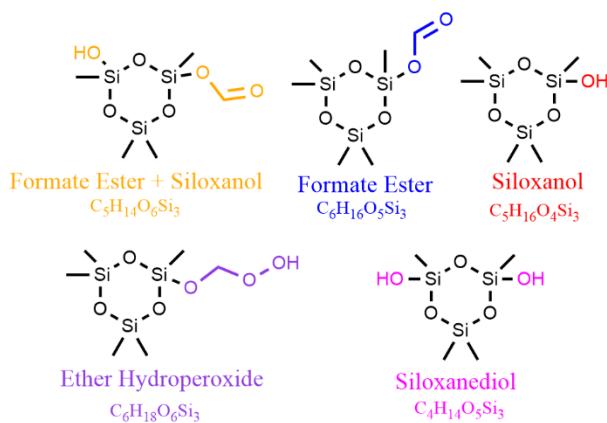
### 183 **3 RESULTS AND DISCUSSION**

#### 184 **3.1 cVMS Oxidation Products**

185 Figure 1 shows an example time series from a  
186 D3 oxidation experiment with H<sub>2</sub>O<sub>2</sub> as the oxidant  
187 precursor. The only product to exhibit significant  
188 loss after the lights are turned off is the  
189 difunctional siloxanol + formate ester product. As  
190 oxidants are continuously generated during these  
191 experiments, the first-generation oxidation  
192 products will undergo further oxidation. To  
193 minimize the impact of this oxidation, we analyze  
194 the product signals at the point where only 10% of  
195 the parent cVMS has been oxidized. This metric  
196 was chosen as only 4% of the oxidation products  
197 are expected to have reacted at this point due to the  
198 continued generation of oxidants. Additionally,  
199 although the oxidation products are expected to



**Figure 1** Example time series of D3 oxidation with H<sub>2</sub>O<sub>2</sub> as the oxidant precursors. The shading signifies when the lights are on, and the dashed vertical line signifies when 10% of D3 had been oxidized. Signals averaged to 1 min time resolution are shown.



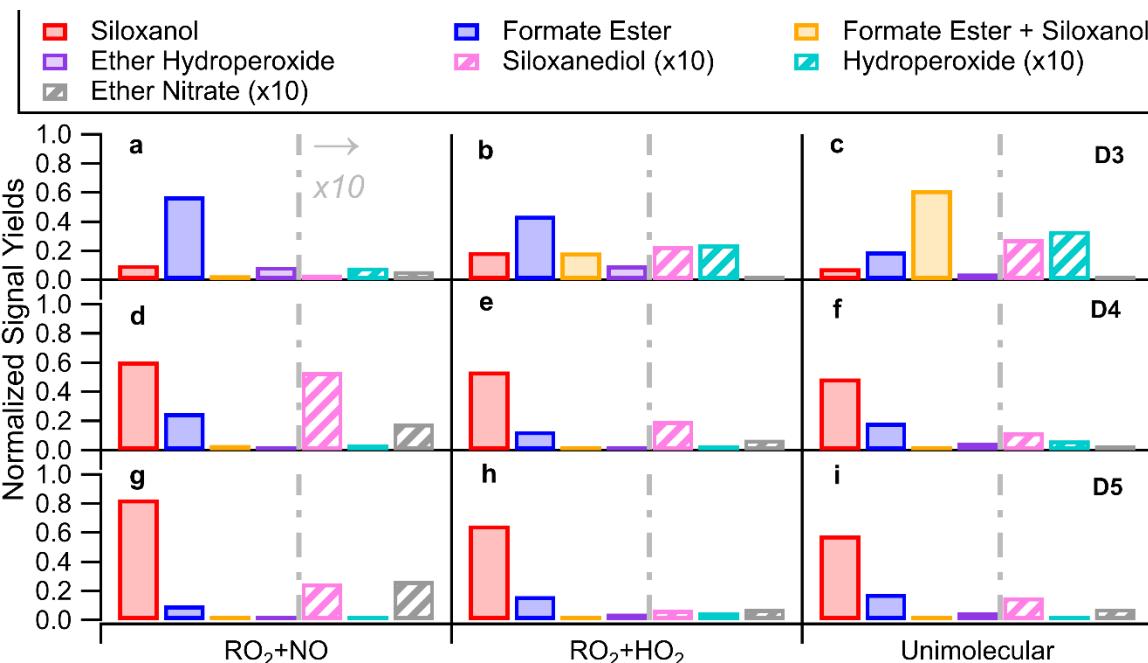
206  
**Figure 2** Example cVMS oxidation product  
 structures.

208

have lifetimes with respect to wall loss greater than one hour, measuring the products after only 10% of the cVMS reacted (~30 min) minimizes the impacts of wall loss on the measured concentrations. Comparing the results at the point where 10% of the cVMS has reacted allows for an easier intercomparison between the different experiments, as we are achieving the same oxidant exposures.

209 Possible isomers of select D3 oxidation products are shown in Figure 2. These assignments are  
 210 proposed based on previous works<sup>4,7,8</sup> and further informed through isotopically labeled linear  
 211 VMS experiments (described in Section 3 of the Supporting Information). We did not detect the  
 212 formation of any oxidation products that contained either fewer or greater silicon atoms than the  
 213 parent compound. For instance, we did not observe the formation of either D4 or a product with  
 214 10 silicon atoms during the oxidation of D5.

215 The CIMS was calibrated for the parent siloxanes. The sensitivity to the parent compounds in  
 216 these experiments was lower than reported in previous works<sup>5</sup> as the ion optics were tuned for  
 217 optimal resolving power, which sacrificed sensitivity. Authentic standards for cVMS oxidation  
 218 products are not commercially available and thus we are unable to quantify their concentrations.  
 219 However, after normalization to the reagent ion and when working in the linear response region,  
 220 the instrument response to a specific compound will be linear with respect to changes in



221

**Figure 3** Signal yield of the different cVMS oxidation products when the  $\text{RO}_2$  was most likely to react with (a,d,g) NO, (b,e,h)  $\text{HO}_2$ , or (c,f,i) when unimolecular reactions are favored, divided by the signal of the siloxane lost at that point. (a-c) are for D3, (d-f) for D4, and (g-i) for D5 oxidation. The signals for the siloxanediol, ether hydroperoxide, and ether nitrate are multiplied by 10 and striped for visual clarity.

222 concentration. As the reagent ion exhibited no significant depletion and the signals were  
 223 significantly above the limit of detection, these experiments are within the linear response region.  
 224 Thus, although the oxidation products cannot be quantified, differences across  $\text{RO}_2$  fate conditions  
 225 in the signal of a given product can be interpreted as changes in the relative amount of a product  
 226 being formed. Moreover, as shown in Section 4 of the Supporting Information, when 10% of the  
 227 cVMS had reacted, the total signal of the oxidation products detected normalized to the mixing  
 228 ratio of cVMS reacted, varied by <20% across the experimental conditions (Figures S3 and S4)  
 229 even when there are significant changes in the compositions of the ions formed (particularly for

230 D3). This finding suggests that variations in instrument response to different products were  
231 relatively small and that we were capturing most of the oxidation products.

232 Given the absence of authentic calibration standards, we elect to report the product abundance  
233 in terms of normalized signal yield. This quantity is calculated as the signal of an individual  
234 product divided by the sum of the total product signals and would be approximately equal to the  
235 molar yield if we had equal sensitivity to every oxidation product. Figure 3 shows the normalized  
236 signal yields of each of the products for the different cVMS studied.

237 The oxidation mechanism of D3 has been studied with theoretical calculations and will be  
238 discussed first. The most intense product signal formed in the D3 experiments under conditions  
239 that favored unimolecular reactions, as seen in Figure 3a, was  $(C_5H_{14}O_6Si_3)H^+$ . We assign this  
240 formula as a difunctional product, the siloxanol and formate ester. In the  $RO_2 + NO$  and  $RO_2 + HO_2$   
241 conditions, the product with the most intense signal was  $(C_6H_{16}O_5Si_3)H^+$ , which corresponds to  
242 the formate ester product. The next most intense signal in the experiments was  $(C_5H_{16}O_4Si_3)H^+$ ,  
243 which we identified as the siloxanol. We attribute  $(C_6H_{18}O_6Si_3)H^+$  to an ether hydroperoxide and  
244  $(C_6H_{18}O_5Si_3)H^+$  to the hydroperoxide. The hydroperoxide was identified by Sommerlade et al.  
245 (1993),<sup>8</sup> though the ether hydroperoxide was not. No nitrogen containing peaks were identified  
246 when D3 was oxidized in high  $NO_x$  ( $\sim 100$  ppb<sub>v</sub>  $NO_x$ ) conditions. This observation may suggest  
247 that the formation of organic nitrates was unfavorable. Alternatively, organic nitrates may be  
248 detected with low efficiency because of either fragmentation or low ionization efficiency. The  
249 presence of and instrument response to organic nitrates will be discussed further in the context of  
250 D4 and D5 oxidation.

251 The assignment of  $(C_6H_{16}O_5Si_3)H^+$  as a formate ester rather than a carboxylic acid was informed  
252 by previous experimental and theoretical work.<sup>4,32</sup> Additionally, we oxidized fully deuterated

253 hexamethyldisiloxane ( $D_{18}L2$ ) with Cl atoms in a similar experiment to Atkinson et al. (1995),<sup>4</sup>  
254 and observed  $(C_6D_{16}O_3Si_2)H^+$  as the major ion. No detectable signal existed for ions with  
255 exchangeable hydrogens. The  $^1H$  in the formula originated from the proton transfer reaction  
256 leading to ionization in the CIMS. If the product were instead a carboxylic acid, we would expect  
257 to observe  $(C_6D_{15}HO_3Si)H^+$ . Further details regarding this experiment are in Section 3 of the  
258 Supporting Information.

259 The product distribution in the D4 and D5 oxidation experiments differed substantially from the  
260 D3 experiments. Monofunctional products, the siloxanol and to a lesser extent the formate ester,  
261 dominated the product signals. Additionally, the signal yields of the siloxanol and formate ester  
262 showed little sensitivity to the  $RO_2$  lifetime and reaction partner. In the D4 and D5 experiments,  
263 signals for  $(C_8H_{23}O_7Si_4N)H^+$  and  $(C_{10}H_{29}O_8Si_5N)H^+$ , attributed to the organic nitrates, and  
264  $(C_8H_{23}O_8Si_4N)H^+$  and  $(C_{10}H_{29}O_9Si_5N)H^+$ , attributed to the ether organic nitrates, were detected.  
265 However, these compounds appeared to be formed with significantly smaller yields than the  
266 siloxanol and formate ester products. Only the ether nitrate yields are shown in Figure 3. Proton  
267 transfer ionization can lead to fragmentation of organic nitrates through nitric acid loss. As a result,  
268 instrument response for the protonated organic nitrate can be low with high limits of detection. We  
269 detected the nitric acid loss fragment,  $(C_{10}H_{28}O_6Si_5)H^+$ , for the ether nitrate product at ~25% of  
270 the ether nitrate signal for D5, which suggests that some fragmentation did occur. We did not  
271 observe ions consistent with either water- or  $NO_2$ -loss fragmentation pathways or charge transfer  
272 products. Although the reagent ion may be ineffective at detecting organic nitrates, we expect that  
273 the yield of organic nitrates in these experiments was small given that both the organic nitrate  
274 signal and the total product signals were similar under the different  $RO_2$  fates investigated which  
275 ranged from 99% of  $RO_2$  reacting with NO in the  $RO_2 + NO$  conditions to ~5% reacting with NO

276 in the RO<sub>2</sub> + HO<sub>2</sub> experiments. More discussion on the instrument response is in Section 4 of the  
277 Supporting Information. Additionally, previous work suggests that organic nitrate yields are low.  
278 Organic nitrates have not been detected before this work, even when cVMS was oxidized under  
279 high NO conditions.<sup>4,18</sup> Carter et al. (1992)<sup>54</sup> performed environmental chamber experiments  
280 investigating how siloxanes alter ozone formation in high NO<sub>x</sub> conditions. Through a model-  
281 measurement comparison, they determined that their results were inconsistent with organic nitrate  
282 formation, however, siloxane oxidation products were not measured.

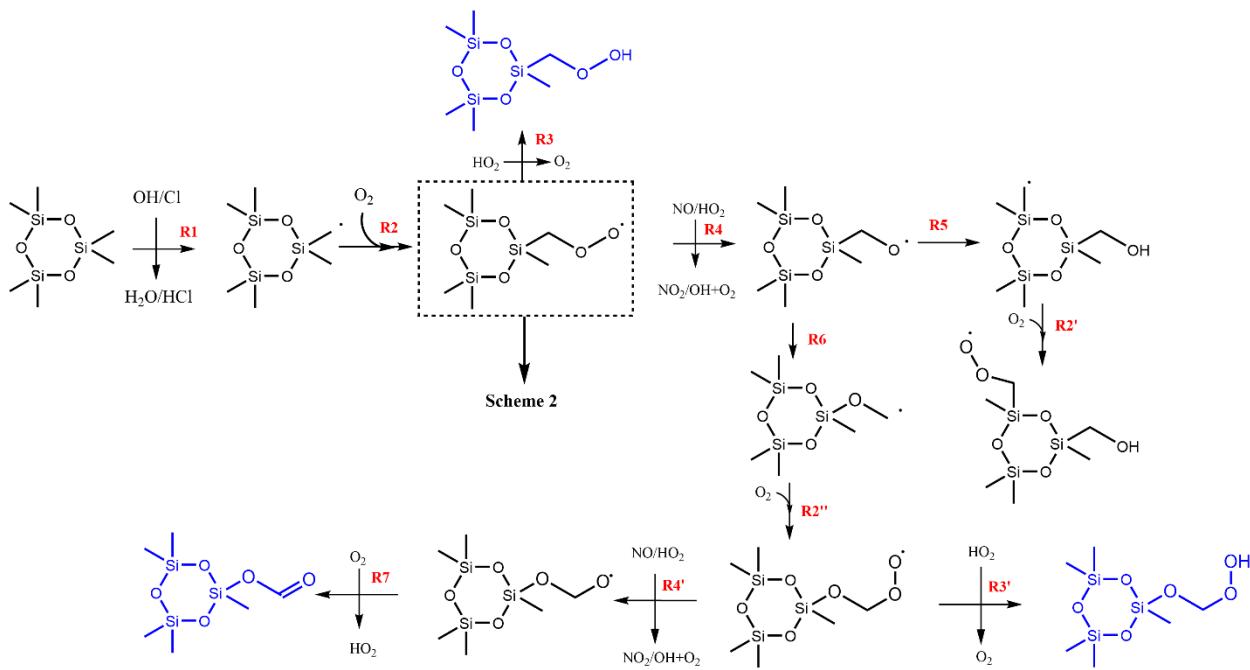
283 In works that identified products with an increased number of silicon atoms, RO<sub>2</sub> + RO<sub>2</sub>  
284 reactions or reactions within the condensed phase likely occurred. In our experiments, RO<sub>2</sub> + RO<sub>2</sub>  
285 reactions were minimal under all conditions (<1%) and no aerosol was formed, thus we did not  
286 expect to observe products with an increased number of silicon atoms.

### 287 **3.2 Discussion of Reaction Mechanism**

288 In most previous studies, the major oxidation products have typically been attributed to the  
289 siloxanol and the formate ester, though the formate ester has only been detected experimentally  
290 twice.<sup>4,5</sup> The siloxanol has been suggested to be a hydrolysis product of the formate ester.<sup>4</sup> Other  
291 works have suggested unusual reactions that could potentially explain the formation of the  
292 siloxanol and the formate ester products,<sup>8,31,32</sup> however, a lack of controlled and varied RO<sub>2</sub> fates  
293 in these experiments prevents a comprehensive assessment of the mechanism. In this work, we  
294 varied the lifetime and reaction partner of the RO<sub>2</sub> radical to gain insight into these reactions.  
295 Additionally, our real-time measurement technique has the advantage that conversion of the  
296 formate ester to the siloxanol on surfaces will be minimized. In this section we discuss the potential  
297 cVMS oxidation mechanism as compiled from previous experimental and theoretical  
298 studies.<sup>4,8,31,32,42</sup> The schemes presented in this section are based on these previously proposed

299 reactions as well as typical oxidation pathways in the atmosphere. In the following section, we  
 300 implement the reaction mechanism in a zero-dimensional kinetic box model and compare to our  
 301 measurements.

302 **Scheme 1** Potential reactions of cVMS in high NO<sub>x</sub>/HO<sub>2</sub> conditions



304 **3.2.1 RO<sub>2</sub> + NO**

305 As shown in Scheme 1, cVMS oxidation is initiated by OH or Cl abstracting a hydrogen from  
 306 one of the methyl groups (R1). The resulting alkyl radical quickly reacts with O<sub>2</sub> forming RO<sub>2</sub>  
 307 (R2). When the NO/HO<sub>2</sub> ratio is greater than 2, such as urban locations, the primary reaction of  
 308 RO<sub>2</sub> is with NO.<sup>52</sup> This reaction has two channels, one that forms an alkoxy radical (RO) and NO<sub>2</sub>  
 309 (R4), and one that forms an organic nitrate (-ONO<sub>2</sub>; not shown). Based on our measurements and  
 310 as discussed in Section 3.1, we suggest that the organic nitrate channel is minor and that reaction  
 311 with NO mainly proceeds through reaction R4. Possible fates for the R<sub>3</sub>SiCH<sub>2</sub>O<sup>·</sup> radical include  
 312 reaction with O<sub>2</sub>, decomposition, and isomerization. Isomerization reactions were expected to

313 dominate as reactions with  $O_2$  to form  $R_3SiCHO$  (not shown) and decomposition have been  
314 calculated to be at least 10 orders of magnitude slower than isomerization.<sup>31</sup> Possible isomerization  
315 pathways include a hydrogen shift from an adjacent methyl group or a unique rearrangement (R6)  
316 to form a carbon based radical. R6 is inaccessible to carbon-based volatile organic compounds  
317 (VOCs). The calculated lifetime (from theoretically determined rate constants) for  $R_3SiCH_2O^\bullet$  with  
318 respect to R6 was  $\sim 6 \times 10^{-12}$  seconds for D3, and this isomerization is expected to be the major RO  
319 fate.<sup>31</sup> The most intense hydroperoxide and nitrate signals detected during oxidation in this work  
320 are consistent with the formation of the ether hydroperoxide and ether nitrate, consistent with R6  
321 occurring rapidly. Note that the peroxy radicals formed from R5 followed by R2' and R6 followed  
322 by R2'' are isomers. However, based on theoretical calculations,<sup>31</sup> R5 is expected to be slow  
323 relative to R6 and thus further reactions and products from R2' were not considered. The carbon  
324 centered radical formed by R6,  $R_3SiOCH_2^\bullet$ , will subsequently add  $O_2$  (R2'') to form a new ether  
325 peroxy radical.<sup>31,32</sup> This radical can react with NO/ $HO_2$  again (R4') to form an ether alkoxy radical.  
326 The formate ester forms when the ether alkoxy radical reacts with  $O_2$  (R7). The ether peroxy  
327 radical can also react with  $HO_2$  to form the ether hydroperoxide (R3') or with NO to form the ether  
328 nitrate (not shown).

329     3.2.2  $RO_2 + HO_2$

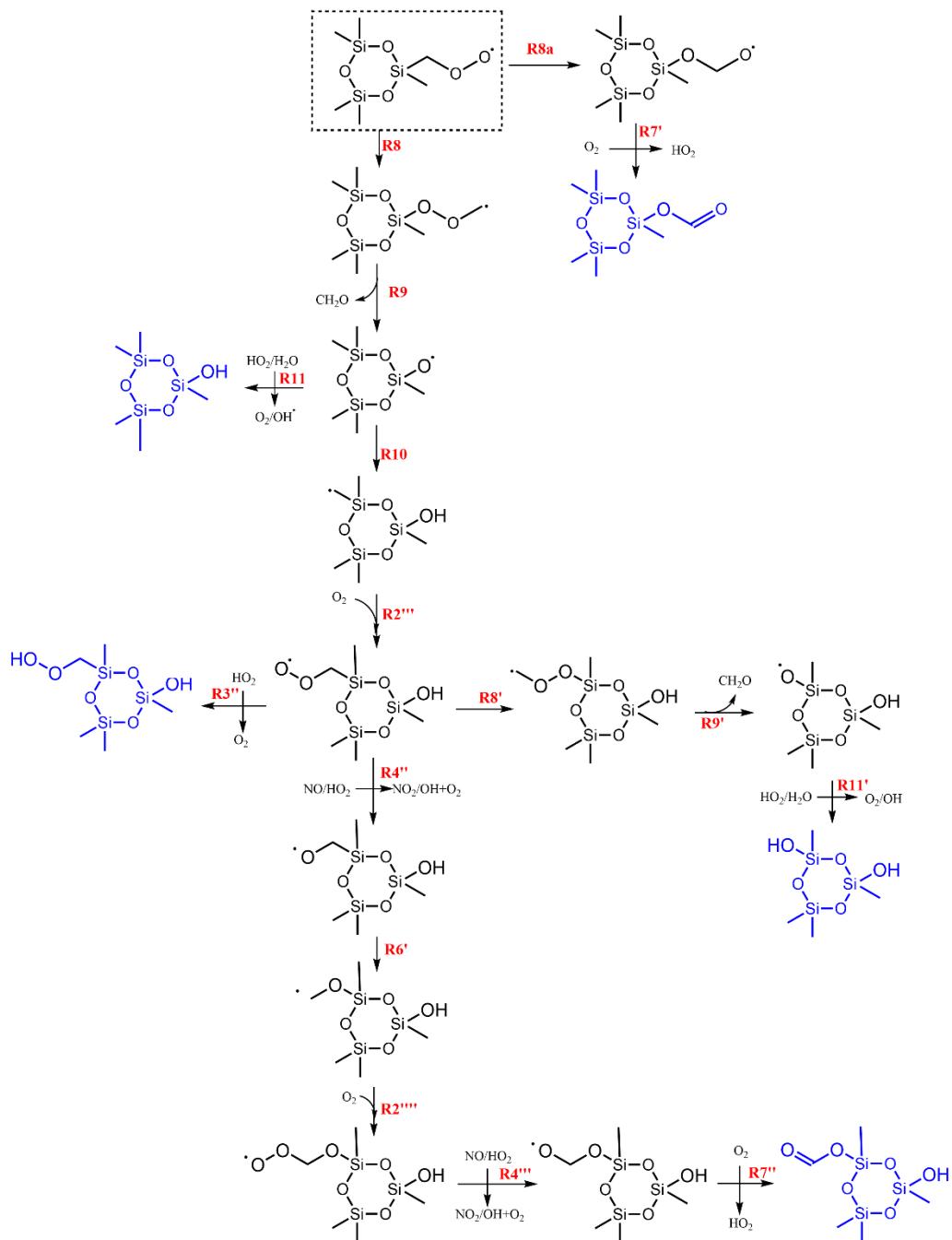
330     The main channel of the  $RO_2$  reaction with  $HO_2$  typically results in the formation of a  
331 hydroperoxide (R3).<sup>55,56</sup> In our experiments, the observed intensity of this expected product was  
332 low. While this may be due to a low instrument response to hydroperoxides when using proton  
333 transfer ionization, the fact that similar products in similar yields are formed under both  $RO_2 + NO$   
334 (<1% of  $RO_2$  reacting with  $HO_2$ ) and  $RO_2 + HO_2$  (>90% of  $RO_2$  reacting with  $HO_2$ ) conditions  
335 leads us to suggest that the  $RO_2$  radicals can react with  $HO_2$  to form the RO radical, OH, and  $O_2$

336 (R4). The formation of the RO radical from reaction of RO<sub>2</sub> with HO<sub>2</sub> has been observed  
337 previously in oxygenated organic molecules, leading to a radical recycling mechanism.<sup>56,57</sup> Once  
338 R6 occurs, the reactions are the same in the NO case with the addition of R3' to form the ether  
339 hydroperoxide.

340 3.2.3 Conditions Favoring Unimolecular Reactions

341 Under conditions of low HO<sub>2</sub> or NO concentrations, the lifetime of the RO<sub>2</sub> radical with respect  
342 to bimolecular reactions is long and unimolecular reactions may become important. Possible  
343 unimolecular reactions of cVMS derived RO<sub>2</sub> identified through quantum chemical calculations  
344 include 1,3-, 1,5-, and 1,7-hydrogen shifts to form QOOH with a propagated radical. However, the  
345 more favorable (lower energy barrier) isomerization was predicted to be the unusual pathway  
346 shown in R8 (Scheme 2).<sup>31</sup>

347 **Scheme 2** Potential reactions of cVMS in low NO<sub>x</sub>/HO<sub>2</sub> condition.



348

349 After undergoing R8, the resulting radical is expected to quickly proceed through R9, generating  
 350 a siloxy radical and formaldehyde. We suggest that the siloxy radical can undergo reaction R11 to  
 351 make the siloxanol product after reaction with H<sub>2</sub>O or HO<sub>2</sub>. The reaction of RO with H<sub>2</sub>O (R11)

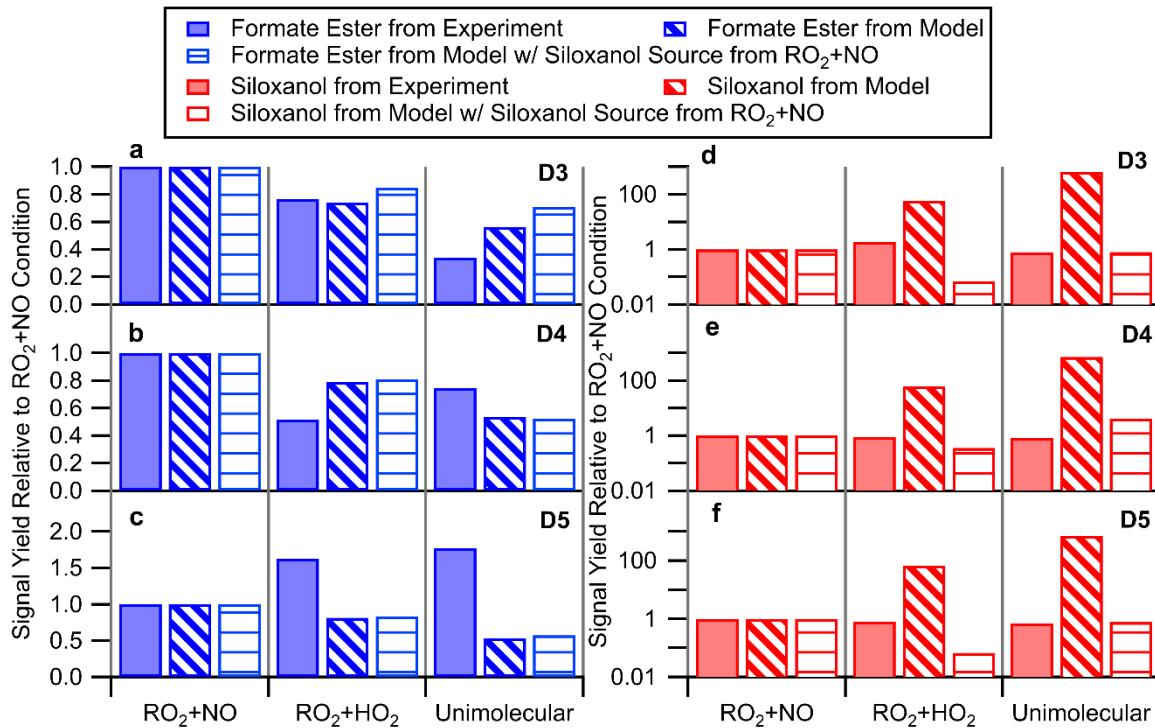
352 has been proposed to occur on silica surfaces,<sup>58</sup> while Carter et al. (1992)<sup>54</sup> previously proposed  
353 the reaction of RO with HO<sub>2</sub> to make the siloxanol. Fu et al. (2020)<sup>31</sup> suggested RO could undergo  
354 a reaction and abstract a hydrogen from a methyl group on a nearby silicon atom (R10, 1,5 H-  
355 shift), or on the same silicon atom (1,3 H-shift), which is less favorable than the 1,5 H-shift. Both  
356 reactions propagate the radical to create another functional group on the cVMS molecule, and our  
357 measurement technique cannot distinguish which methyl groups have been modified, therefore we  
358 did not consider this difference. Another unusual isomerization of the RO<sub>2</sub> radical (R8a) that was  
359 proposed initially by Atkinson et al. (1995)<sup>4</sup> and used to explain observation of the formate ester  
360 product was previously investigated with quantum chemical calculations and determined to be  
361 energetically unfavorable.<sup>31,32</sup>

### 362 **3.3 Kinetic Model Using the Proposed Mechanism**

363 We implement the reactions in Schemes 1 and 2, with rate constants informed from previously  
364 published quantum chemical calculations,<sup>31</sup> in a kinetic model (KinSim in Igor Pro) to investigate  
365 if the proposed mechanism can reproduce the changing intensity of different products under  
366 different RO<sub>2</sub> fates. The full mechanisms with rate constants for the three VMS studies are  
367 presented in Section 2 of the Supporting Information. Reactions to form the organic nitrate with  
368 NO were excluded from the model. RO<sub>2</sub> + RO<sub>2</sub> reactions were minimal in all experiments (<0.5%  
369 of the RO<sub>2</sub> reactions were with another RO<sub>2</sub>), and therefore were not considered in the mechanism.  
370 In the model, we held bimolecular reaction rates (RO<sub>2</sub> + HO<sub>2</sub> and RO<sub>2</sub> + NO) constant. As reaction  
371 R8a was previously determined to not be energetically favorable,<sup>31</sup> this reaction was not used. We  
372 included the oxidation of the first-generation products with rate constants equal to the parent  
373 cVMS. As there is uncertainty in the molar yields due to a lack of calibration standards, we focus  
374 on comparing the relative changes in signal yields as a function of RO<sub>2</sub> fate. Figure 4 shows the

375 signal yields of the siloxanol and formate ester products from our experiments and the kinetic  
376 model at the point 10% of the cVMS was oxidized, normalized to the signal yield in the RO<sub>2</sub> +  
377 NO conditions. By observing the relative changes between the conditions, the absolute calibration  
378 of the oxidation products should not affect the interpretation of the data.

379 In our experiments with D3, we detect that the signal corresponding to the formate ester  
380 decreases with an increasing RO<sub>2</sub> lifetime, which is consistent with the formate ester being formed  
381 after RO<sub>2</sub> reacts with NO/HO<sub>2</sub>. Consequently, the model captures this decrease albeit the measured  
382 decrease in formate ester yield with increasing RO<sub>2</sub> lifetime is greater than the decrease predicted  
383 by the model (Figure 4). For D5, we observed an increase in the formate ester with increasing RO<sub>2</sub>  
384 lifetime, which contrasts with the model prediction of a decreasing trend. The model also fails to  
385 capture the trends in the relative siloxanol yield for all three cVMS species; the model predicts an  
386 increase in siloxanol yield with increasing RO<sub>2</sub> lifetime while the measurements suggest that the  
387 siloxanol yield is essentially independent of RO<sub>2</sub> lifetime and fate. In fact, the model predicts  
388 negligible siloxanol yield under RO<sub>2</sub> + NO conditions (Figure S5), which causes the modeled  
389 siloxanol yield normalized to the RO<sub>2</sub> + NO conditions to change by orders of magnitude as the  
390 RO<sub>2</sub> lifetime increases. The model also predicts significantly less siloxanol (~100× less) formation  
391 than in our experiments, if we assume that we detect all the products with equal sensitivity.  
392 Siloxanol formation has previously been attributed to hydrolysis of the formate ester prior  
393 investigations of siloxane chemistry.<sup>4,7,18</sup> Our high time-resolution measurements of the formate  
394 ester and siloxanol are inconsistent with this hypothesis since we observed simultaneous formation  
395 of both the formate ester and the siloxanol.



396

**Figure 4** The signal yields of the formate ester (a-c) and siloxanol (d-f) for all three cVMS from the experimental results and kinetic modeling using the mechanism presented, normalized to the results from the RO<sub>2</sub> + NO conditions. (a) and (d) are for D3 oxidation, (b) and (e) are for D4 oxidation, and (c) and (f) for D5. Note the log axis on the siloxanol results. Only the rate constants for reactions with OH/Cl and wall loss constants were changed between cVMS. In the model results that had a source of siloxanol from the RO<sub>2</sub> + NO reaction, 5% of the reaction product made the siloxanol directly, and the other 95% made RO.

397 The model-measurement gap could potentially be explained by an unknown formation mechanism  
 398 of the siloxanol in high NO/HO<sub>2</sub> conditions. As a thought experiment, if a branching ratio is added  
 399 to produce 5% of the siloxanol and 95% of the RO after the reaction of RO<sub>2</sub> with NO, then the  
 400 amount of siloxanol formed between experiments is slightly closer to the experimental results,  
 401 though it underestimates the siloxanol formation in the RO<sub>2</sub> + HO<sub>2</sub> conditions (Figure 4). This

402 formation of the siloxanol could be achieved through a process such as decomposition of a  
403 chemically activated alkoxy radical. Another possibility would be to have a faster rate constant for  
404 the isomerization in R8, as the siloxy radical formed after that isomerization could be a source of  
405 siloxanol. However, for this reaction to be important in all conditions (as the siloxanol is the largest  
406 product in the high NO<sub>x</sub> experiments with D4 and D5), the rate constant would need to be increased  
407 by multiple orders of magnitude to compete with the RO<sub>2</sub> + NO reaction. Overall, the siloxanol  
408 formation mechanism remains unclear and more investigation to better understand why RO<sub>2</sub> fates  
409 do not significantly affect the products formed from D4 and D5 oxidation is required.

410 The model also fails to accurately capture the evolution of the largest product in the conditions  
411 that favor unimolecular reactions from D3 oxidation, the formate ester and siloxanol difunctional  
412 product (not shown in Figure 4). Because this difunctional product was the largest signal during  
413 the conditions that favor unimolecular reactions, it was anticipated that D3 R<sub>3</sub>SiO<sup>•</sup> radicals can  
414 readily undergo the auto-oxidation reaction shown in R10 after being formed from R8 and R9.  
415 Since the siloxanol was formed in R10, the next steps need to preferentially make the formate ester  
416 over another siloxanol. Therefore, R4'' needs to dominate over R8', though to form the initial  
417 siloxanol, R8 needs to dominate over R4. This pathway requires that the ratio of the R4:R8 rate  
418 constants is less than R4'':R8', which stands in contrast to theoretical calculations suggesting that  
419 the isomerization rate increases as the molecules become more functionalized.<sup>31</sup> However, we did  
420 not detect any evidence in these experiments of the siloxanetriol product, which would likely be  
421 formed if the isomerization reactions R8 and R8' both dominated over bimolecular reactions, as  
422 the next RO<sub>2</sub> isomerization branching would likely not be different. Another possible explanation  
423 is that concentrations of NO or HO<sub>2</sub> were higher later in the experiment, pushing the mechanism

424 down R4''. The model, however, indicates that NO and HO<sub>2</sub> concentrations do not change enough  
425 to alter the product distribution.

426 Overall, it is evident that the oxidation mechanism is unable to replicate our experimental results,  
427 particularly for D4 and D5. The finding that D3 oxidation produces different products than D4 and  
428 D5 suggests that results, either from laboratory-based experiments or theoretical calculations, from  
429 smaller VMS and Si-containing molecules (i.e., D3, hexamethyldisiloxane [L2], and  
430 tetramethylsilane) may not hold for larger Si compounds. In particular, the lack of sensitivity on  
431 the oxidation product yields on RO<sub>2</sub> fate for D4 and D5 is unusual. As the siloxanol was the most  
432 abundant product for D4 and D5 under all conditions, our results suggests that there is more unique  
433 chemistry occurring that requires further investigation.

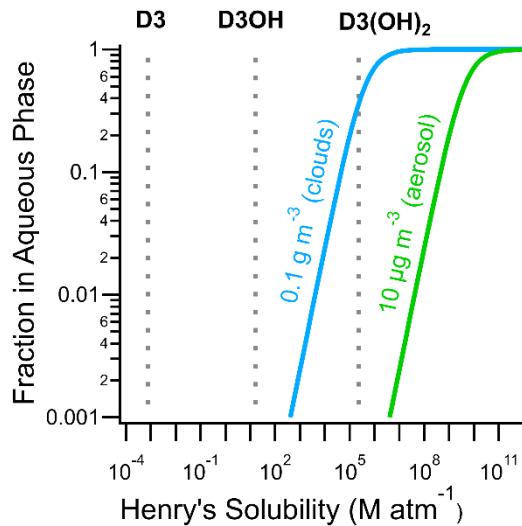
434 **3.4 Possible Fates of the First-Generation Products**

435 As cVMS are globally distributed due to their long atmospheric lifetimes, their oxidation  
436 products will also be globally distributed. Previous research has focused on the parent compounds,  
437 though the oxidation products should not be discounted. Here, we use various structure activity  
438 relationships (SAR) to investigate possible loss pathways for the main cVMS oxidation products  
439 we measured and discussed (siloxanol, formate ester, multi-substituted siloxanols/formate esters,  
440 hydroperoxides, and organic nitrates). The estimated values of vapor pressure, water solubility,  
441 and Henry's Solubility (M atm<sup>-1</sup>) constants of all the products are presented in Table S6 of the  
442 Supporting Information.<sup>59</sup> Note that the Henry's Law Constant (atm M<sup>-1</sup>) is the inverse of the  
443 Henry Solubility.

444 Using the estimated vapor pressures, and assuming that absorption into organic aerosol is the  
445 main gas-particle partitioning method for these compounds, less than 1% of the least volatile  
446 oxidation product (the D5 siloxanediol) mass would partition to aerosol at a moderate organic

447 aerosol loading of  $10 \mu\text{g m}^{-3}$ .<sup>60</sup> This value was  
448 determined using the vapor pressure estimated by  
449 MPBPWIN, which predicts vapor pressures about  
450 one order of magnitude lower than TEST, thus  
451 giving an upper limit to aerosol partitioning. It has  
452 been determined that cVMS heterogenous  
453 reactions with components of mineral dust aerosol  
454 can lead to significant removal of the cVMS and  
455 its oxidation products, but due to the low typical  
456 dust loadings, the total loss to dust is expected  
457 to be minimal.<sup>27</sup>

458 The oxidation products have estimated  
459 Henry's Solubility constants that can be up to  $\sim 8$   
460 orders of magnitude higher (higher partitioning  
461 into the aqueous phase) than the parent cVMS.  
462 Out of the various methods compared,  
463 HenryWIN gave the largest estimated Henry's  
464 Solubility coefficient of  $\sim 10^5 \text{ M atm}^{-1}$  for the D3  
465 siloxanediol. Even using this Henry's Solubility  
466 constant, at most 30% of the siloxanediol will  
467 partition to cloud droplets, while there is no  
468 significant partitioning into aqueous aerosol, as  
469 shown in Figure 5, adapted from Daumit et al.<sup>61</sup> We note that this estimate is uncertain as Henry's



**Figure 5** Estimated Henry's Solubility, estimated with HenryWIN v3.21, for D3 and the siloxanol oxidation products (dashed lines). The mass concentration labels on the solid lines correspond to the amount of liquid water in clouds or ambient aerosol. Only siloxanol products are shown as the hydroxyl groups change the Henry's Solubility most significantly. Larger cVMS oxidation products are estimated to have lower Henry's solubility than D3. All values are listed in Table S6 of the Supporting Information. Adapted with permission from Daumit, K. E.; Carrasquillo, A. J.; Hunter, J. F.; Kroll, J. H. Laboratory Studies of the Aqueous-Phase Oxidation of Polyols: Submicron Particles vs. Bulk Aqueous Solution. *Atmos. Chem. Phys.* 2014, 14 (19), 10773–10784. <https://doi.org/10.5194/acp-14-10773-2014>.

470 Solubility and Law constants measured in the laboratory vary by orders of magnitude between  
471 studies for the parent cVMS.<sup>62</sup> and the Henry's Law constants of the different oxidation products  
472 have not been measured.<sup>59,63,64</sup> However, as we use the most extreme values from the structure  
473 activity relationships, we assume these to be the upper limits for partitioning and thus conclude  
474 that absorptive and aqueous partitioning will have minor impacts on the mixing ratios of the  
475 oxidation products.

476 The oxidized products may also be transformed in the atmosphere via oxidation. Atkinson et al.  
477 (1995)<sup>4</sup> measured the rate constants of the tetramethylsilane and trimethylsiloxanol reactions with  
478 Cl atoms and OH radicals which showed the rate constant for the siloxanol oxidation reactions  
479 were 2 times faster than the parent compounds for Cl and 10 times faster for OH. However, it is  
480 uncertain how the rate constants will change with larger cVMS. Using the Atmospheric Oxidation  
481 Program for Microsoft Windows (AOPWIN) in EPIWIN with the adjustments to the Si containing  
482 group contributions suggested by Alton and Browne (2020),<sup>5</sup> the rate constant for reaction with  
483 OH is estimated to be approximately a factor of 5 times faster for D3OH compared to D3, and 3  
484 times faster for D5OH compared to D5. This lowers the lifetime of D3OH and D5OH to  
485 approximately 2 days for both of the products, compared to 11 and 4.4 days.<sup>5</sup> Based on this work,  
486 it is probable that, like the parent VMS compounds, these first-generation oxidation products will  
487 have atmospheric lifetimes of days and thus it is necessary to better understand multiple-generation  
488 oxidation products and their potential chemistry and deposition to completely understand cVMS  
489 environmental fates.

490 **4 CONCLUSIONS**

491 As ~90% of cVMS emitted into the environment partitions into the atmosphere,<sup>3</sup> understanding  
492 the atmospheric degradation of these compounds is critical for understanding their environmental

493 impacts. In this work, we oxidized three cVMS under conditions of different RO<sub>2</sub> fates and  
494 measured the oxidation products to gain insight into the cVMS oxidation mechanism. We observed  
495 that the main oxidation product for D4 and D5 is the siloxanol, regardless of the fate of the peroxy  
496 radical. As D4 and D5 are the most abundant cVMS in the atmosphere, we suggest that in chemical  
497 transport models cVMS oxidation products can be adequately represented as the siloxanol, similar  
498 to previous representations of this chemistry.<sup>25</sup> Due to the high vapor pressure and low water  
499 solubility of the cVMS and oxidation products, it is predicted that not only is the parent cVMS  
500 globally distributed,<sup>13,65</sup> the oxidation products are likely also globally present, requiring multiple  
501 generations of oxidation before significant removal will occur. Because the oxidation products are  
502 also likely to be long lived in the atmosphere, more measurements of the oxidation products in the  
503 atmosphere and environmental matrices are necessary to better understand the environmental  
504 processing of these anthropogenic chemicals.

## 505 ASSOCIATED CONTENT

### 506 **Supporting Information.**

507 Detailed information of the peak fitting in TOFWARE, KinSim mechanism and inputs,  
508 isotopically labeled D<sub>18</sub>L2 oxidation experiment, instrument response in different conditions, and  
509 the estimated physical parameters of cVMS and the oxidation products are presented in the  
510 Supporting Information (PDF) available free of charge.

### 511 **Data Availability**

512 Upon acceptance, the experimental data will be made publicly available through the Index of  
513 Chamber Atmospheric Research in the United States website (<https://icarus.ucdavis.edu/>).

514 **Safety**

515 **Caution!** Ultraviolet light is damaging to biological tissues. Caution is required when working  
516 with the any lights that emit ultraviolet wavelengths and protective eyewear must be used at all  
517 times.

518 **AUTHOR INFORMATION**

519 **Corresponding Author**

520 \*Eleanor C. Browne, Department of Chemistry, University of Colorado, Boulder, Colorado 80309,  
521 United States and Cooperative Institute for Research in Environmental Sciences, University of  
522 Colorado, Boulder, Colorado 80309, United States; orcid.org/0000-0002-8076-9455; Phone: 303-  
523 735- 7685; Email: Eleanor.Browne@Colorado.edu

524 **Author Contributions**

525 Mitchell W. Alton – Department of Chemistry, University of Colorado, Boulder, Colorado  
526 80309, United States and Cooperative Institute for Research in Environmental Sciences,  
527 University of Colorado, Boulder, Colorado 80309, United States; orcid.org/0000-0002-7119-3706

528 MWA and ECB designed the experiments. MWA performed all the experiments that contributed  
529 to this work. The analysis of the data was performed mainly by MWA with support and guidance  
530 from ECB. The manuscript was written through contributions of both authors. Both authors have  
531 given approval to the final version of the manuscript.

532 **ACKNOWLEDGMENTS**

533 This research was supported by the National Science Foundation under Grant CHE-1808606.  
534 Additional funding to support MWA was provided by the Cooperative Institute for Research in  
535 Environmental Sciences Graduate Student Research Fellowship Grant.

536 REFERENCES

537 (1) Organisation for Economic Co-Operation and Development. The 2004 OECD List of High  
538 Production Volume Chemicals. **2004**.

539 (2) U.S. Environmental Protection Agency. CompTox Chemicals Dashboard  
540 <https://comptox.epa.gov/dashboard/DTXSID102718> (accessed Oct 3, 2021).

541 (3) Mackay, D.; Cowan-Ellsberry, C. E.; Powell, D. E.; Woodburn, K. B.; Xu, S.; Kozerski, G.  
542 E.; Kim, J. Decamethylcyclopentasiloxane (D5) Environmental Sources, Fate, Transport,  
543 and Routes of Exposure. *Environ. Toxicol. Chem.* **2015**, *34* (12), 2689–2702.  
544 <https://doi.org/10.1002/etc.2941>.

545 (4) Atkinson, R.; Tuazon, E. C.; Kwok, E. S. C.; Arey, J.; Aschmann, S. M.; Bridier, I. Kinetics  
546 and Products of the Gas-Phase Reactions of (CH<sub>3</sub>)<sub>4</sub>Si, (CH<sub>3</sub>)<sub>3</sub>SiCH<sub>2</sub>OH,  
547 (CH<sub>3</sub>)<sub>3</sub>SiOSi(CH<sub>3</sub>)<sub>3</sub> and (CD<sub>3</sub>)<sub>3</sub>SiOSi(CD<sub>3</sub>)<sub>3</sub> with Cl Atoms and OH Radicals. *J. Chem.*  
548 *Soc. Faraday Trans.* **1995**, *91* (18), 3033. <https://doi.org/10.1039/ft9959103033>.

549 (5) Alton, M. W.; Browne, E. C. Atmospheric Chemistry of Volatile Methyl Siloxanes:  
550 Kinetics and Products of Oxidation by OH Radicals and Cl Atoms. *Environ. Sci. Technol.*  
551 **2020**, *54* (10), 5992–5999. <https://doi.org/10.1021/acs.est.0c01368>.

552 (6) Atkinson, R. Kinetics of the Gas-Phase Reactions of a Series of Organosilicon Compounds  
553 with Hydroxyl and Nitrate (NO<sub>3</sub>) Radicals and Ozone at 297 +/- 2 K. *Environ. Sci. Technol.*

554 1991, 25 (5), 863–866. <https://doi.org/10.1021/es00017a005>.

555 (7) Markgraf, S. J.; Wells, J. R. The Hydroxyl Radical Reaction Rate Constants and  
556 Atmospheric Reaction Products of Three Siloxanes. *Int. J. Chem. Kinet.* **1997**, 29 (6), 445–  
557 451. [https://doi.org/10.1002/\(SICI\)1097-4601\(1997\)29:6<445::AID-KIN6>3.0.CO;2-U](https://doi.org/10.1002/(SICI)1097-4601(1997)29:6<445::AID-KIN6>3.0.CO;2-U).

558 (8) Sommerlade, R.; Parlar, H.; Wrobel, D.; Kochs, P. Product Analysis and Kinetics of the  
559 Gas-Phase Reactions of Selected Organosilicon Compounds with OH Radicals Using a  
560 Smog Chamber-Mass Spectrometer System. *Environ. Sci. Technol.* **1993**, 27 (12), 2435–  
561 2440. <https://doi.org/10.1021/es00048a019>.

562 (9) Safron, A.; Strandell, M.; Kierkegaard, A.; Macleod, M. Rate Constants and Activation  
563 Energies for Gas-Phase Reactions of Three Cyclic Volatile Methyl Siloxanes with the  
564 Hydroxyl Radical. *Int. J. Chem. Kinet.* **2015**, 47 (7), 420–428.  
565 <https://doi.org/10.1002/kin.20919>.

566 (10) Bernard, F.; Papanastasiou, D. K.; Papadimitriou, V. C.; Burkholder, J. B. Temperature  
567 Dependent Rate Coefficients for the Gas-Phase Reaction of the OH Radical with Linear  
568 (L<sub>2</sub>, L<sub>3</sub>) and Cyclic (D<sub>3</sub>, D<sub>4</sub>) Permethylsiloxanes. *J. Phys. Chem. A* **2018**, 122 (17), 4252–  
569 4264. <https://doi.org/10.1021/acs.jpca.8b01908>.

570 (11) Kim, J.; Xu, S. Quantitative Structure-Reactivity Relationships of Hydroxyl Radical Rate  
571 Constants for Linear and Cyclic Volatile Methylsiloxanes. *Environ. Toxicol. Chem.* **2017**,  
572 36 (12), 3240–3245. <https://doi.org/10.1002/etc.3914>.

573 (12) Tuazon, E. C.; Aschmann, S. M.; Atkinson, R. Atmospheric Degradation of Volatile  
574 Methyl-Silicon Compounds. *Environ. Sci. Technol.* **2000**, 34 (10), 1970–1976.

575 https://doi.org/10.1021/es9910053.

576 (13) Genualdi, S.; Harner, T.; Cheng, Y.; MacLeod, M.; Hansen, K. M.; Van Egmond, R.;  
577 Shoeib, M.; Lee, S. C. Global Distribution of Linear and Cyclic Volatile Methyl Siloxanes  
578 in Air. *Environ. Sci. Technol.* **2011**, *45* (8), 3349–3354. <https://doi.org/10.1021/es200301j>.

579 (14) Xu, S.; Warner, N.; Bohlin-Nizzetto, P.; Durham, J.; McNett, D. Long-Range Transport  
580 Potential and Atmospheric Persistence of Cyclic Volatile Methylsiloxanes Based on Global  
581 Measurements. *Chemosphere* **2019**, *228*, 460–468.  
582 <https://doi.org/10.1016/j.chemosphere.2019.04.130>.

583 (15) Warner, N. A.; Evenset, A.; Christensen, G.; Gabrielsen, G. W.; Borgä, K.; Leknes, H.  
584 Volatile Siloxanes in the European Arctic: Assessment of Sources and Spatial Distribution.  
585 *Environ. Sci. Technol.* **2010**, *44* (19), 7705–7710. <https://doi.org/10.1021/es101617k>.

586 (16) Brooke, D.; Crookes, M.; Gray, D.; Robertson, S. *Environmental Risk Assessment Report:*  
587 *Decamethylcyclopentasiloxane*; Environment Agency, 2009.

588 (17) Brooke, D. N.; Brooke, M. J.; Gray, D.; Robertson, S.; Crookes, M.; Gray, D.; Robertson,  
589 S. *Environmental Risk Assessment Report: Octamethylcyclotetrasiloxane*; Environment  
590 Agency, 2009.

591 (18) Allen, R. B.; Annelin, R. B.; Atkinson, R.; Carpenter, J. C.; Carter, W. L. P.; Chandra, G.;  
592 Fendinger, N. J.; Gerhards, R.; Grigoras, S.; Hatcher, J. A.; Hobson, J. F.; Kochs, P.;  
593 Lehmann, R. G.; Maxim, L. D.; Mazzoni, S. M.; Mihaich, E. M.; Miyakawa, Y.; Pohl, E.  
594 R.; Powell, D. E.; Roy, S.; Sawano, T.; Slater, G. S.; Spivack, J. L.; Stevens, C.; Wischer,  
595 D. *Organosilicon Materials*; Chandra, G., Ed.; The Handbook of Environmental Chemistry;

596 Springer Berlin Heidelberg: Berlin, Heidelberg, 1997; Vol. 3. <https://doi.org/10.1007/978-3-540-68331-5>.  
597

598 (19) European Chemicals Agency. Recommendation of the European Chemicals Agency of 20  
599 December 2011 for the Inclusion of Substances in Annex XIV to REACH (List of  
600 Substances Subject to Authorisation) of Regulation (EC) No 1907/2006. **2011**, *1* (April), 1–  
601 7.

602 (20) United Kingdom Health & Safety Executive. Annex XV Restriction Report Proposal for a  
603 Restriction. **2015**, No. June 2015, 1–89.

604 (21) European Chemicals Agency. Recommendation of the European Chemicals Agency of 14  
605 April 2021 for the Inclusion of Substances in Annex XIV to REACH (List of Substances  
606 Subject to Authorisation). **2021**, *1* (April), 1–7.

607 (22) US EPA. Enforceable Consent Agreement for Environmental Testing for  
608 OCTAMETHYLCYCLOTETRASILOXANE (D4) (CASRN 556-67-2) Docket No . EPA-  
609 HQ-OPPT-2012-0209. **2014**.

610 (23) Dudzina, T.; Von Goetz, N.; Bogdal, C.; Biesterbos, J. W. H.; Hungerbühler, K.  
611 Concentrations of Cyclic Volatile Methylsiloxanes in European Cosmetics and Personal  
612 Care Products: Prerequisite for Human and Environmental Exposure Assessment. *Environ.*  
613 *Int.* **2014**, *62*, 86–94. <https://doi.org/10.1016/j.envint.2013.10.002>.

614 (24) Bzdek, B. R.; Horan, A. J.; Pennington, M. R.; Janecek, N. J.; Baek, J.; Stanier, C. O.;  
615 Johnston, M. V. Silicon Is a Frequent Component of Atmospheric Nanoparticles. *Environ.*  
616 *Sci. Technol.* **2014**, *48* (19), 11137–11145. <https://doi.org/10.1021/es5026933>.

617 (25) Janechek, N. J.; Hansen, K. M.; Stanier, C. O. Comprehensive Atmospheric Modeling of  
618 Reactive Cyclic Siloxanes and Their Oxidation Products. *Atmos. Chem. Phys.* **2017**, *17*(13),  
619 8357–8370. <https://doi.org/10.5194/acp-17-8357-2017>.

620 (26) Chandramouli, B.; Kamens, R. M. The Photochemical Formation and Gas-Particle  
621 Partitioning of Oxidation Products of Decamethyl Cyclopentasiloxane and Decamethyl  
622 Tetrasiloxane in the Atmosphere. *Atmos. Environ.* **2001**, *35* (1), 87–95.  
623 [https://doi.org/10.1016/S1352-2310\(00\)00289-2](https://doi.org/10.1016/S1352-2310(00)00289-2).

624 (27) Navea, J. G.; Xu, S.; Stanier, C. O.; Young, M. A.; Grassian, V. H. Heterogeneous Uptake  
625 of Octamethylcyclotetrasiloxane (D4) and Decamethylcyclopentasiloxane (D5) onto  
626 Mineral Dust Aerosol under Variable RH Conditions. *Atmos. Environ.* **2009**, *43* (26), 4060–  
627 4069. <https://doi.org/10.1016/j.atmosenv.2009.05.012>.

628 (28) Janechek, N. J.; Marek, R. F.; Bryngelson, N.; Singh, A.; Bullard, R. L.; Brune, W. H.;  
629 Stanier, C. O. Physical Properties of Secondary Photochemical Aerosol from OH Oxidation  
630 of a Cyclic Siloxane. *Atmos. Chem. Phys.* **2019**, *19* (3), 1649–1664.  
631 <https://doi.org/10.5194/acp-19-1649-2019>.

632 (29) Fairbrother, A.; Woodburn, K. B. Assessing the Aquatic Risks of the Cyclic Volatile Methyl  
633 Siloxane D4. *Environ. Sci. Technol. Lett.* **2016**, *3* (10), 359–363.  
634 <https://doi.org/10.1021/acs.estlett.6b00341>.

635 (30) Tang, X.; Misztal, P. K.; Nazaroff, W. W.; Goldstein, A. H. Siloxanes Are the Most  
636 Abundant Volatile Organic Compound Emitted from Engineering Students in a Classroom.  
637 *Environ. Sci. Technol. Lett.* **2015**, *2* (11), 303–307.  
638 <https://doi.org/10.1021/acs.estlett.5b00256>.

639 (31) Fu, Z.; Xie, H.-B.; Elm, J.; Guo, X.; Fu, Z.; Chen, J. Formation of Low-Volatile Products  
640 and Unexpected High Formaldehyde Yield from the Atmospheric Oxidation of  
641 Methylsiloxanes. *Environ. Sci. Technol.* **2020**, *54* (12), 7136–7145.  
642 <https://doi.org/10.1021/acs.est.0c01090>.

643 (32) Ren, Z.; da Silva, G. Auto-Oxidation of a Volatile Silicon Compound: A Theoretical Study  
644 of the Atmospheric Chemistry of Tetramethylsilane. *J. Phys. Chem. A* **2020**, *124* (32),  
645 6544–6551. <https://doi.org/10.1021/acs.jpca.0c02922>.

646 (33) Coggon, M. M.; McDonald, B. C.; Vlasenko, A.; Veres, P. R.; Bernard, F.; Koss, A. R.;  
647 Yuan, B.; Gilman, J. B.; Peischl, J.; Aikin, K. C.; DuRant, J.; Warneke, C.; Li, S.; de Gouw,  
648 J. A. Diurnal Variability and Emission Pattern of Decamethylcyclopentasiloxane (D5) from  
649 the Application of Personal Care Products in Two North American Cities. *Environ. Sci.  
650 Technol.* **2018**, *52* (10), 5610–5618. <https://doi.org/10.1021/acs.est.8b00506>.

651 (34) Wu, Y.; Johnston, M. V. Molecular Characterization of Secondary Aerosol from Oxidation  
652 of Cyclic Methylsiloxanes. *J. Am. Soc. Mass Spectrom.* **2016**, *27* (3), 402–409.  
653 <https://doi.org/10.1007/s13361-015-1300-1>.

654 (35) Yucuis, R. A.; Stanier, C. O.; Hornbuckle, K. C. Cyclic Siloxanes in Air, Including  
655 Identification of High Levels in Chicago and Distinct Diurnal Variation. *Chemosphere*  
656 **2013**, *92* (8), 905–910. <https://doi.org/10.1016/j.chemosphere.2013.02.051>.

657 (36) Milani, A.; Al-Naiema, I. M.; Stone, E. A. Detection of a Secondary Organic Aerosol Tracer  
658 Derived from Personal Care Products. *Atmos. Environ.* **2021**, *246* (July), 118078.  
659 <https://doi.org/10.1016/j.atmosenv.2020.118078>.

660 (37) Ahrens, L.; Harner, T.; Shoeib, M. Temporal Variations of Cyclic and Linear Volatile  
661 Methylsiloxanes in the Atmosphere Using Passive Samplers and High-Volume Air  
662 Samplers. *Environ. Sci. Technol.* **2014**, *48* (16), 9374–9381.  
663 <https://doi.org/10.1021/es502081j>.

664 (38) Gallego, E.; Perales, J. F.; Roca, F. J.; Guardino, X.; Gadea, E. Volatile Methyl Siloxanes  
665 (VMS) Concentrations in Outdoor Air of Several Catalan Urban Areas. *Atmos. Environ.*  
666 **2017**, *155*, 108–118. <https://doi.org/10.1016/j.atmosenv.2017.02.013>.

667 (39) Tran, T. M.; Abualnaja, K. O.; Asimakopoulos, A. G.; Covaci, A.; Gevao, B.; Johnson-  
668 Restrepo, B.; Kumosani, T. A.; Malarvannan, G.; Minh, T. B.; Moon, H. B.; Nakata, H.;  
669 Sinha, R. K.; Kannan, K. A Survey of Cyclic and Linear Siloxanes in Indoor Dust and Their  
670 Implications for Human Exposures in Twelve Countries. *Environ. Int.* **2015**, *78*, 39–44.  
671 <https://doi.org/10.1016/j.envint.2015.02.011>.

672 (40) Praske, E.; Otkjær, R. V.; Crounse, J. D.; Hethcox, J. C.; Stoltz, B. M.; Kjaergaard, H. G.;  
673 Wennberg, P. O. Atmospheric Autoxidation Is Increasingly Important in Urban and  
674 Suburban North America. *Proc. Natl. Acad. Sci. U. S. A.* **2018**, *115* (1), 64–69.  
675 <https://doi.org/10.1073/pnas.1715540115>.

676 (41) Wu, Y.; Johnston, M. V. Aerosol Formation from OH Oxidation of the Volatile Cyclic  
677 Methyl Siloxane (CVMS) Decamethylcyclopentasiloxane. *Environ. Sci. Technol.* **2017**, *51*  
678 (8), 4445–4451. <https://doi.org/10.1021/acs.est.7b00655>.

679 (42) Allen, R. B.; Annelin, R. B.; Atkinson, R.; Carpenter, J. C.; Carter, W. L. P.; Chandra, G.;  
680 Fendinger, N. J.; Gerhards, R.; Grigoras, S.; Hatcher, J. A.; Hobson, J. F.; Kochs, P.;  
681 Lehmann, R. G.; Maxim, L. D.; Mazzoni, S. M.; Mihaich, E. M.; Miyakawa, Y.; Pohl, E.

682 R.; Powell, D. E.; Roy, S.; Sawano, T.; Slater, G. S.; Spivack, J. L.; C. Stevens, D. W.  
683 *Organosilicon Materials*; Chandra, G., Ed.; The Handbook of Environmental Chemistry;  
684 Springer Berlin Heidelberg: Berlin, Heidelberg, 1997; Vol. 3. <https://doi.org/10.1007/978-3-540-68331-5>.

686 (43) Cheng, Z.; Qiu, X.; Shi, X.; Zhu, T. Identification of Organosiloxanes in Ambient Fine  
687 Particulate Matters Using an Untargeted Strategy via Gas Chromatography and Time-of-  
688 Flight Mass Spectrometry. *Environ. Pollut.* **2021**, *271*, 116128.  
689 <https://doi.org/10.1016/j.envpol.2020.116128>.

690 (44) Lu, D.; Tan, J.; Yang, X.; Sun, X.; Liu, Q.; Jiang, G.; Tan, J.; Sun, X.; Lu, D.; Jiang, G.;  
691 Yang, X. Unraveling the Role of Silicon in Atmospheric Aerosol Secondary Formation: A  
692 New Conservative Tracer for Aerosol Chemistry. *Atmos. Chem. Phys. Discuss.* **2018**, *19*  
693 (3), 1–19. <https://doi.org/10.5194/acp-2018-914>.

694 (45) Lu, D.; Liu, Q.; Yu, M.; Yang, X.; Fu, Q.; Zhang, X.; Mu, Y.; Jiang, G. Natural Silicon  
695 Isotopic Signatures Reveal the Sources of Airborne Fine Particulate Matter. *Environ. Sci.*  
696 *Technol.* **2018**, *52* (3), 1088–1095. <https://doi.org/10.1021/acs.est.7b06317>.

697 (46) Charan, S.; Huang, Y.; Buenconsejo, R.; Li, Q.; Cocker III, D.; Seinfeld, J. Secondary  
698 Organic Aerosol Formation from the Oxidation of Decamethylcyclopentasiloxane at  
699 Atmospherically Relevant OH Concentrations. *Atmos. Chem. Phys. Discuss.* **2021**, No.  
700 May, 1–17. <https://doi.org/10.5194/acp-2021-353>.

701 (47) Peng, Z.; Jimenez, J. L. KinSim: A Research-Grade, User-Friendly, Visual Kinetics  
702 Simulator for Chemical-Kinetics and Environmental-Chemistry Teaching. *J. Chem. Educ.*  
703 **2019**, *96* (4), 806–811. <https://doi.org/10.1021/acs.jchemed.9b00033>.

704 (48) Atkinson, R.; Baulch, D. L.; Cox, R. A.; Crowley, J. N.; Hampson, R. F.; Hynes, R. G.;  
705 Jenkin, M. E.; Rossi, M. J.; Troe, J. IUPAC Task Group on Atmospheric Chemical Kinetic  
706 Data Evaluation, *Http://Iupac.Pole-Ether.Fr. Atmos. Chem. Phys.* **2004**, *1* (4), 1461–1738.

707 (49) Atkinson, R.; Baulch, D. L.; Cox, R. A.; Crowley, J. N.; Hampson, R. F.; Hynes, R. G.;  
708 Jenkin, M. E.; Rossi, M. J.; Troe, J. Evaluated Kinetic and Photochemical Data for  
709 Atmospheric Chemistry: Volume III - Gas Phase Reactions of Inorganic Halogens. *Atmos.*  
710 *Chem. Phys.* **2007**, *7* (4), 981–1191. <https://doi.org/10.5194/acp-7-981-2007>.

711 (50) Liu, X.; Day, D. A.; Krechmer, J. E.; Brown, W.; Peng, Z.; Ziemann, P. J.; Jimenez, J. L.  
712 Direct Measurements of Semi-Volatile Organic Compound Dynamics Show near-Unity  
713 Mass Accommodation Coefficients for Diverse Aerosols. *Commun. Chem.* **2019**, *2* (1), 1–  
714 9. <https://doi.org/10.1038/s42004-019-0200-x>.

715 (51) Krechmer, J. E.; Pagonis, D.; Ziemann, P. J.; Jimenez, J. L. Quantification of Gas-Wall  
716 Partitioning in Teflon Environmental Chambers Using Rapid Bursts of Low-Volatility  
717 Oxidized Species Generated in Situ. *Environ. Sci. Technol.* **2016**, *50* (11), 5757–5765.  
718 <https://doi.org/10.1021/acs.est.6b00606>.

719 (52) Ziemann, P. J.; Atkinson, R. Kinetics, Products, and Mechanisms of Secondary Organic  
720 Aerosol Formation. *Chem. Soc. Rev.* **2012**, *41* (19), 6582–6605.  
721 <https://doi.org/10.1039/c2cs35122f>.

722 (53) Boethling, R.; Meylan, W. How Accurate Are Physical Property Estimation Programs for  
723 Organosilicon Compounds? *Environ. Toxicol. Chem.* **2013**, *32* (11), 2433–2440.  
724 <https://doi.org/10.1002/etc.2326>.

725 (54) Carter, W. L. P.; J, P.; Malkina, I. L.; D, L. *Investigation of the Ozone Formation Potential*  
726 *of Selected Volatile Silicone Compounds; Final Report to Dow Corning Corporation;*  
727 Midland, MI, 1992.

728 (55) Seinfeld, J. H.; Pandis, S. N. *Atmospheric Chemistry and Physics: From Air Pollution to*  
729 *Climate Change*, 2nd ed.; John Wiley & Sons, Inc: Hoboken, New Jersey, 2006.

730 (56) Orlando, J. J.; Tyndall, G. S. Laboratory Studies of Organic Peroxy Radical Chemistry: An  
731 Overview with Emphasis on Recent Issues of Atmospheric Significance. *Chem. Soc. Rev.*  
732 **2012**, *41* (19), 6294–6317. <https://doi.org/10.1039/c2cs35166h>.

733 (57) Jenkin, M. E.; Valorso, R.; Aumont, B.; Rickard, A. R. Estimation of Rate Coefficients and  
734 Branching Ratios for Reactions of Organic Peroxy Radicals for Use in Automated  
735 Mechanism Construction. *Atmos. Chem. Phys. Discuss.* **2019**, No. 2, 1–46.  
736 <https://doi.org/10.5194/acp-2019-44>.

737 (58) Narayanasamy, J.; Kubicki, J. D. Mechanism of Hydroxyl Radical Generation from a Silica  
738 Surface: Molecular Orbital Calculations. *J. Phys. Chem. B* **2005**, *109* (46), 21796–21807.  
739 <https://doi.org/10.1021/jp0543025>.

740 (59) Xu, S.; Kropscott, B. Evaluation of the Three-Phase Equilibrium Method for Measuring  
741 Temperature Dependence of Internally Consistent Partition Coefficients (KOW, KOA, and  
742 KAW) for Volatile Methylsiloxanes and Trimethylsilanol. *Environ. Toxicol. Chem.* **2014**,  
743 *33* (12), 2702–2710. <https://doi.org/10.1002/etc.2754>.

744 (60) Donahue, N. M.; Robinson, A. L.; Pandis, S. N. Atmospheric Organic Particulate Matter:  
745 From Smoke to Secondary Organic Aerosol. *Atmos. Environ.* **2009**, *43* (1), 94–106.

746 https://doi.org/10.1016/j.atmosenv.2008.09.055.

747 (61) Daumit, K. E.; Carrasquillo, A. J.; Hunter, J. F.; Kroll, J. H. Laboratory Studies of the  
748 Aqueous-Phase Oxidation of Polyols: Submicron Particles vs. Bulk Aqueous Solution.  
749 *Atmos. Chem. Phys.* **2014**, *14* (19), 10773–10784. https://doi.org/10.5194/acp-14-10773-  
750 2014.

751 (62) Sander, R. Compilation of Henry's Law Constants (Version 4.0) for Water as Solvent.  
752 *Atmos. Chem. Phys.* **2015**, *15* (8), 4399–4981. https://doi.org/10.5194/acp-15-4399-2015.

753 (63) Xu, S.; Kozerski, G.; Mackay, D. Critical Review and Interpretation of Environmental Data  
754 for Volatile Methylsiloxanes: Partition Properties. *Environ. Sci. Technol.* **2014**, *48* (20),  
755 11748–11759. https://doi.org/10.1021/es503465b.

756 (64) Xu, S.; Kropscott, B. Method for Simultaneous Determination of Partition Coefficients for  
757 Cyclic Volatile Methylsiloxanes and Dimethylsilanediol. *Anal. Chem.* **2012**, *84* (4), 1948–  
758 1955. https://doi.org/10.1021/ac202953t.

759 (65) Krogseth, I. S.; Kierkegaard, A.; McLachlan, M. S.; Breivik, K.; Hansen, K. M.; Schlabach,  
760 M. Occurrence and Seasonality of Cyclic Volatile Methyl Siloxanes in Arctic Air. *Environ.*  
761 *Sci. Technol.* **2012**, *47* (1), 502–509. https://doi.org/10.1021/es3040208.

762

763

764

765

TOC/Abstract Art

