

1 **Half-century history of crop nitrogen budget in the conterminous United States:**

2 **Variations over time, space and crop types**

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7 **Key Points:**

- 8 • We synthesized county-level crop nitrogen (N) budget data of eight major crops in the United  
9 States during 1970-2019
- 10 • The national crop NUE increased from 0.55 kg N kg<sup>-1</sup> N in the 1970s to 0.65 kg N kg<sup>-1</sup> N in  
11 the 2010s, with an average NUE of 0.59 kg N kg<sup>-1</sup> N
- 12 • N surplus accounts for 41% (27%-55%) of the national total N input, with a significant  
13 decline detected during the recent decade

## 14 **Abstract**

15 Spatiotemporal patterns of crop nitrogen (N) budget have important implications for agricultural  
16 N management and environmental policy. Previous studies examined crop N budget in different  
17 countries but often overlooked cross-crop differences at sub-national scales. In this study, we  
18 synthesize multiple databases to examine the N budget of eight major crops in the United States  
19 at the county scale during 1970-2019. Our analyses show that national crop NUE increased from  
20  $0.55 \text{ kg N kg}^{-1} \text{ N}$  in the 1970s to  $0.65 \text{ kg N kg}^{-1} \text{ N}$  in the 2010s. Four out of eight crops such as  
21 corn, rice, cotton, and sorghum demonstrated an increasing NUE trend during the study period,  
22 whereas the other crops overall presented a declining NUE trend. Nationwide, about 41% of the  
23 total N input was not used by these crops (N surplus) over the study period, of which temporal  
24 variation was mainly driven by corn due to its large planting area and high N input. The national  
25 N surplus first increased in the 1970s and remained relatively stable till the 2000s. Since the  
26 early 2010s, however, N surplus began to decline and approached the levels in the early 1970s—  
27 an encouraging development that may lead to decreased N pollution to the environment. The  
28 hotspots of national N surplus coincided with corn- and rice-producing counties. The sub-  
29 national variations and temporal dynamics in crop N budget revealed in this study highlight the  
30 urgent need to understand the farm-level crop N balance and the dominant factors controlling  
31 crop NUE for mitigating N pollution.

## 32 **1 Introduction**

33 The world agriculture system has been altered substantially to meet the rapid increase in global  
34 food demand (Lassaletta et al., 2016). Since the 1960s, global nitrogen (N) fertilizer use has  
35 increased nine-fold (Lu & Tian, 2017). During the period from 1984 to 2016, global inorganic N  
36 deposition increased by 8% from  $86.6$  to  $93.6 \text{ Tg yr}^{-1}$ , with a significant increase in East Asia and

37 South America and a large decline in Europe (Ackerman et al., 2019). An overall estimate of 50–  
38 70 Tg N yr<sup>-1</sup> is fixed by legume plants in global agricultural systems (Herridge et al., 2008). Over  
39 two-thirds of the total anthropogenic N inputs (e.g. N fertilizer, atmospheric deposition, and  
40 fixed N by legume crops) on the global land surface were reported to return to the atmosphere,  
41 percolate to the surface and subsurface water bodies, or transport to the ocean (Cao et al., 2020;  
42 Cassman et al., 2002; Schlesinger, 2009; Tilman et al., 2002). The proportion of N that is not  
43 utilized by crops (i.e., the difference between N inputs and N harvested in crops), usually defined  
44 as N surplus, was found to account for nearly 40% of global total N inputs (Liu et al., 2010).  
45 Tremendous N surplus induced by excessive N input can be potentially lost to the environment.  
46 It can be emitted into the atmosphere, causing green-house effect (i.e. N<sub>2</sub>O) and air pollution (i.e.  
47 NH<sub>3</sub> as precursors to PM<sub>2.5</sub>), re-deposited onto sensitive ecological communities, or delivered to  
48 groundwater and surface water, imposing detrimental impacts on the biodiversity and  
49 functioning of recipient ecosystems (Barak et al., 1997; Bouwman et al., 2002; Kumazawa, 2002;  
50 Van Meter et al., 2017).

51 Crop nitrogen use efficiency (NUE)—the harvested N in crop production per unit N  
52 input—has been proposed as an important indicator for agricultural sustainability in many  
53 countries (Norton et al., 2015). Improving crop NUE has been perceived as one of the most  
54 effective means for alleviating environmental degradation while increasing or maintaining crop  
55 productivity (Davidson & Kanter, 2014; Giller et al., 2004; Howarth et al., 2002). Therefore, it is  
56 of critical importance to understand crop NUE from both agronomic and socio-economic  
57 perspectives, addressing the challenges of food security, environmental degradation, and climate  
58 change (Zhang et al., 2015).

59 A wide variety of studies analyzed crop NUE trends at various spatial and temporal  
60 scales, revealing large disparities of crop NUE among countries (Lassaletta et al., 2014, 2016).  
61 Crop NUE was found to level off or even decline in developing countries as N input increases  
62 (Ciampitti & Vyn, 2014; Lassaletta et al., 2014; Zhang et al., 2015). This implies that increasing  
63 N input can hardly further stimulate crop yield gain, but instead leads to a greater N surplus. For  
64 example in China, the grain was only enhanced by 28% with an increase of N fertilizer  
65 consumption by 54% (Meng et al., 2016). However, the crop NUE in North America was  
66 reported to be higher than that in other regions around the globe and maintained an increasing  
67 trend as N input was raised (Lassaletta et al., 2016; Zhang et al., 2015). As of 2007, the United  
68 States national crop NUE was reported close to 55% with all intentional and unintentional N  
69 inputs entering food, livestock feed, biofuel, and industrial products counted (Houlton et al.,  
70 2013). Although previous studies have advanced our understanding of crop NUE, sub-national  
71 patterns and dynamics of crop-specific NUE are generally overlooked due to limited data  
72 availability. Besides, crop-specific tendencies in using N have not been explicitly examined  
73 because N input data for different crop types are either missing or discontinuous in the majority  
74 of countries. Using publicly available data sets, Sabo et al., (2019) compiled the inventories of N  
75 inputs and outputs as well as terrestrial N surpluses for all sub-basins of the contiguous U.S. and  
76 found that the national-scale N budget experienced little changes ( $\pm 6\%$ ) between 2002 and 2012.  
77 By using county-level data, a recent study developed an 88-year (1930–2017) data set of N mass  
78 balance across the contiguous U.S. (Byrnes et al., 2020), but cross-crop variations were not  
79 examined. As prior work to this study, Lu et al. (2019) investigated state-level crop-specific  
80 NUE, reporting that the NUE of corn and winter wheat had a large spatial heterogeneity across  
81 the U.S. and their responses to N enrichment slowed down in the recent two decades. Therefore,

82 it is important to understand the spatiotemporal patterns of N budget and NUE across various  
83 crop types for developing better crop nutrient management practices, creating a nutrient trading  
84 system, and reducing nutrient loss to the environment.

85 In this study, we compiled a multi-source database to characterize the crop-specific N  
86 budget for eight major crops in the U.S. from 1970 to 2019. Major crops of interest include  
87 barley, corn, cotton, durum wheat, rice, spring wheat, sorghum, and winter wheat. We extracted  
88 county-level data of major N input sources and crop yield to calculate crop-specific NUE and N  
89 surplus. We assessed how the crop NUE and N surplus changed along with crop yields and N  
90 input rates. We then examined the temporal changes of crop-specific NUE and N budget at the  
91 national scale. The spatial heterogeneity of crop-specific NUE and N surplus was also quantified  
92 at a county level over the past five decades. Specifically, we aimed at answering the following  
93 questions in this study: (i) How have crop yields and NUE responded to N input dynamics in the  
94 U.S. since 1970? (ii) How have the crop-specific NUE and N surplus changed over space and  
95 time? (iii) What are the relationships between the N surplus, crop yield, and N input across these  
96 eight crops?

## 97 **2 Data and Methods**

98 We obtained historical county-level crop yield data of barley (*Hordeum vulgare* L.), corn (*Zea*  
99 *mays* L.), cotton (*Gossypium hirsutum* L.), durum wheat (*Triticum durum* Desf.), rice (*Oryza*  
100 *sativa* L.), spring wheat (*Triticum aestivum* L.), sorghum (*Sorghum bicolor* L.), and winter wheat  
101 (*Triticum aestivum* L.) from the USDA National Agricultural Statistics Service (NASS) crop  
102 databases (<http://www.nass.usda.gov/index.asp>). Our published crop-specific fertilizer  
103 management database (Cao et al., 2018) has been updated and further refined from the state level  
104 to the county level. We also considered other anthropogenic N sources in this study, including

105 manure N application, atmospheric N deposition, and residual N fixation from previous-year  
106 soybean rotation for corn only. We analyzed the responses of yield and NUE to N input as well  
107 as the input-driven (i.e. N input-driven) and output-driven (i.e. crop yield-driven) N surplus  
108 trends from 1970 to 2019. Based on the county-level crop-specific N budget data sets, we further  
109 examined their spatial patterns and temporal dynamics.

## 110 **2.1 Crop yield and anthropogenic N input data sources**

111 ***Cropland planting area:*** We obtained the county-level planting areas of eight major crop types  
112 from 1970 to 2019 from the USDA-NASS crop database (<https://www.nass.usda.gov/>). Only the  
113 counties with tracked crop areas were included in this study. We gap-filled the planting area of  
114 the major crops by assuming the missing county-year samples follow the inter-annual variation  
115 in the state-level planting area of each crop (USDA-NASS, <https://www.nass.usda.gov/>). Except  
116 for the eight major crops plus soybean, the rest crop types (e.g., fruits, vegetables, and small  
117 grains) were grouped as “Others” and added up to keep the state-level cropland acreage  
118 consistent with the USDA-NASS survey.

119 ***Crop yield:*** The annual yield for each crop in each county from 1970 to 2019 was obtained from  
120 the USDA-NASS (<https://www.nass.usda.gov/>). Specifically, we used the year 1970 as the  
121 starting year to limit our analyses to a period when private sectors in the U.S. have heavily  
122 invested in advanced crop hybrids, agricultural management practices, and crop technologies to  
123 improve crop production (Evenson & Gollin, 2003; Tilman, 1998).

124 ***Chemical N fertilizer use:*** We developed annual county-level N fertilizer use data of these eight  
125 crops from 1970 to 2019 by combining the state-level crop-specific N fertilizer use rate (Cao et  
126 al., 2018; Lu et al., 2019) with the county-level N fertilizer consumption amount obtained from  
127 the Nutrient Use Geographic Information System (NuGIS, <https://nugis.tfi.org/>). Our newly

128 developed dataset differs from the NuGIS data by particularly characterizing the cross-crop  
 129 variations in N fertilizer use rates at the county level. Specifically, based on the most recent  
 130 surveys conducted by USDA-NASS, we updated our published state-level crop-specific N  
 131 fertilizer data to 2019 using the same approach adopted in Cao et al. (2018) and Lu et al. (2019).  
 132 Since the NuGIS county-level total N fertilizer consumption is only available from 1987 to 2014,  
 133 we used the inter-annual variations of state-level total N fertilizer consumption (Cao et al., 2018)  
 134 to gap-fill the periods of 1970-1986 and 2015-2019. The crop-specific N use rate generated  
 135 above was downscaled from the state- to county-level as Eq. 1:

$$136 \quad N rate_i^{ct} = \frac{N cons_{ct}}{\sum_{i=1}^{10} N rate_i^{st} \times Area_i^{ct}} \times N rate_i^{st} \quad (1)$$

137 where  $N rate_i^{ct}$  is N fertilizer use rate or manure N application rate of crop type  $i$  in county  $ct$ ,  
 138  $N cons_{ct}$  is annual county N fertilizer consumption or manure N consumption,  $N rate_i^{st}$  is N  
 139 fertilizer use rate or manure N rate of crop type  $i$  in state  $st$ , and  $Area_i^{ct}$  is county-level planting  
 140 area for crop type  $i$ . Crops include the abovementioned eight crops, soybean, and “Others”.

141 **Manure N application:** We reconstructed the county-level crop-specific manure N use rate in  
 142 the U.S. from 1970 to 2019. We obtained total manure consumption in each county during 1987-  
 143 2014 from the NuGIS (<https://nugis.tfi.org/>) and state-level crop-specific manure use rate from  
 144 the USDA-ERS (<https://data.ers.usda.gov/reports.aspx?ID=17883>). Assuming that county-level  
 145 missing data (1970-1986 and 2015-2019) follow the same inter-annual variations as the national  
 146 manure N use developed by the FAO (<http://www.fao.org/faostat/en/#data/EMN>), we gap-filled  
 147 the missing years in the county-level recovered manure data (i.e. the proportion that is applied in  
 148 fields).

149 We converted the state-level crop-specific manure use rate to manure N use rate from  
150 1996 to 2010 based on animal manure types and the manure moisture status from the USDA-  
151 ERS census data (<https://data.ers.usda.gov/reports.aspx?ID=17883>) (see manure N content  
152 conversion coefficients in Table S1). We downscaled the crop-specific manure N use rate  
153 generated above from state- to county-level by using Eq. 1. Unlike fertilizer data, we used the  
154 multi-year average of the crop-specific manure N use rate in each state instead of annual values  
155 owing to limited data.

156 ***Atmospheric N deposition:*** We obtained the total N deposition data at a 4-km resolution for the  
157 period 2000 to 2015 from the National Atmospheric Deposition Program (NADP) total  
158 deposition maps (TDEP, <http://nadp.slh.wisc.edu/committees/tdep/tdepmaps/>, (Schwede & Lear,  
159 2014)). To get the crop-specific N deposition rate for each county, we first resampled the TDEP  
160 N deposition into 5-arc-minute maps in ArcMap 10.7 (ESRI, 2018). We then extracted the N  
161 deposition data from 2000 to 2015 specifically for the eight crops using time-series crop  
162 distribution maps that have the same spatial resolution. The crop distribution maps were  
163 aggregated from 1 km land-use maps that were obtained from a recent study (Yu & Lu, 2018).  
164 The land-use maps were developed by keeping the county-level harvested area of each crop type  
165 in each year consistent with the county-level census records provided by NASS, USDA  
166 (<http://www.nass.usda.gov/index.asp>). Using the TDEP N deposition in 2000 as the endpoint, we  
167 back-calculated the gridded N deposition for the period 1970-1999 by taking the annual  
168 percentage change rate of the North American gridded N deposition data (Wei et al., 2014). The  
169 N deposition from 2016 to 2019 was assumed to remain at the same level as of 2015. Finally, we  
170 calculated the average N deposition rate of all the grids in which the same crop type was planted  
171 within each county in each year.

172 *N residues from previous-year soybean fixation*: Using the Crop Data Layer (CDL,  
173 [https://www.nass.usda.gov/Research\\_and\\_Science/Cropland/Release/index.php](https://www.nass.usda.gov/Research_and_Science/Cropland/Release/index.php)), we first  
174 identified the pixels in which soybean rotated with any of these eight studied crops in three  
175 consecutive years from 2008 to 2017. We included the rotation sequences such as soybean-other  
176 crop-soybean, soybean-other crop-other crop, or soybean-soybean-other crop (“other crop” here  
177 refers to one of the eight studied crops). We found that 74.4% of the national soybean area was  
178 used for rotation with the eight studied crops in 2017. Among them, 66.4% of soybean was  
179 rotated with corn, 4.3% with spring wheat, and 2% each with winter wheat and rice. The  
180 percentages of soybean rotating with the rest four crop types were all less than 1%, which  
181 together accounted for 1.7% of the national soybean used for rotation purposes. Therefore, in our  
182 study, we only considered the residual benefit of N fixation from the previous year's soybean  
183 cultivation for corn N budget. In addition, there are a few data points available to indicate the  
184 national acreage percentages of soybean rotated with corn, for example, 93% in 1993 (United  
185 States Department of Agriculture, 1994), 91% in 1997 (Padgitt et al., 2000). Soybean used in  
186 corn rotation became firmly established in the U.S. around 1940 (North Carolina Soybean  
187 Producers Association, 2019). Hence, we assumed that 100% soybean was used in corn rotation  
188 in 1940. Based on these available historical data points, we used linear interpolation to gap-fill  
189 the area percentages of soybean rotating with corn in the missing years from 1970 to 2008. Then  
190 the county-level soybean percentages and yields were used to calculate the residual N for corn in  
191 each county each year. The details of the residual N from previous-year soybean fixation can be  
192 found in Lu et al., (2019).

## 193 **2.2 Calculation of county-level crop-specific NUE and N surplus**

194 The NUE was calculated as a ratio of crop-recovered N to total N input in units of  $\text{kg N kg}^{-1} \text{N}$   
195 (also referring to a percentage (%) of total N input). The crop-recovered N, indicating how much  
196 N is retained per unit crop yield, was calculated as crop yield multiplied by a crop N recovery  
197 coefficient. We adopted the N recovery coefficients from the NuGIS, which is  $0.012 \text{ kg N kg}^{-1}$   
198 grain for corn and sorghum,  $0.0194 \text{ kg N kg}^{-1}$  grain for winter wheat, spring wheat, and durum  
199 wheat,  $0.011 \text{ kg N kg}^{-1}$  grain for rice,  $0.02 \text{ kg N kg}^{-1}$  grain for barley, and  $0.067 \text{ kg N kg}^{-1}$  harvest  
200 for cotton.

201 N surplus, defined as the portion of total N input that is not recovered by crops, was used  
202 to represent N that will be eventually lost to the air, soil and water systems. N surplus is widely  
203 used to assess the potential agricultural N loss (Sabo et al., 2019; Swaney et al., 2018;  
204 Wachendorf et al., 2006; P. Xu et al., 2006). We estimated N surplus (in units of  $\text{kg N ha}^{-1} \text{ yr}^{-1}$ )  
205 as total N input received by each crop minus the crop-recovered N.

## 206 **2.3 Spatial patterns of county-level NUE and N surplus**

207 We used metrics that summarize the decadal average crop NUE and N surplus to represent their  
208 spatial patterns. The unit of spatial NUE and N surplus is  $\text{kg N kg}^{-1} \text{N}$  and  $\text{Gg N per county yr}^{-1}$ ,  
209 respectively. Besides, we also applied Sen's slope to obtain the change magnitude of annual total  
210 N input and N surplus of the eight studied crops during two periods, 1970-1989 and 1990-2019,  
211 across space (Zhang et al., 2017). Sen's slope estimates monotonic trends and has low sensitivity  
212 to outliers (Sen, 1968). The statistical significance of the estimated Sen's slope was evaluated at  
213 a 5% level by using a non-parametric Mann-Kendall test (Kendall, 1948).

## 214 **3 Results**

### 215 **3.1 Crop-recovered N responses to N input of eight major crops**

216 Since we adopted constant crop-specific N recovery coefficients, we used crop-recovered N to  
217 represent crop yield when studying the trajectories of NUE and N surplus against crop yields.  
218 We calculated the planting area-weighted average crop-recovered N versus N input rates across  
219 the U.S. (Figure 1a). The national average crop-recovered N shows a distinct curvilinear  
220 relationship with the N input trajectory, indicating a strong N fertilization effect. However, it is  
221 noteworthy that the declines in crop-specific recovered N response are hidden by the national  
222 average. Specifically, the N inputs of corn and rice fall in the high end of the spectrum, while the  
223 rest of the crops cluster at the lower end. The recovered N of most crops, such as barley, corn,  
224 cotton, and winter wheat, first increases and then declines or levels off as N input rate increases.  
225 Spring wheat and durum wheat are consistent in showing an increasing response, followed by a  
226 slowed-down trend of recovered N after reaching the maximum. Rice recovered N near-linearly  
227 increases with N input rate. Among the eight crops, sorghum is the only crop showing a “V-  
228 shaped” response in recovered N with N input increases.

### 229 **3.2 Nitrogen use efficiency response to N input of eight crops**

230 The trajectories of crop NUE responses to N input (Figure 1b) are generally contrary to the  
231 above-described crop-recovered N trajectories. It is noteworthy that all crops except corn show a  
232 declining NUE trend as N input increases. Specifically, barley, cotton, spring wheat, winter  
233 wheat, and durum wheat show a similar decreasing trajectory. The decreases in sorghum NUE  
234 presents a smaller range, compared with other crops. Rice NUE shows a near-linear decreasing  
235 course with N input ranging from 100 to 250 kg N ha<sup>-1</sup>. Corn NUE is relatively stable along  
236 lower N input gradient (e.g., 100 to 150 kg N ha<sup>-1</sup>). When N input rises to a higher end (e.g., 175

237 to 225 kg N ha<sup>-1</sup>), corn NUE begins to slightly increase. The national crop NUE overall displays  
238 a slightly increasing trend (e.g., from 0.55 to 0.65 kg N kg<sup>-1</sup> N) as N input increases, which is  
239 dominated by corn NUE.

240 We also compared the temporal changes in NUE of the eight studied crops over the past  
241 50 years (Figure 2). Specifically, the NUE of barley and winter wheat experienced a large  
242 decline from 1970 to the mid-1990s and started increasing afterward. The NUE of spring and  
243 durum wheat declined from over 1.0 kg N kg<sup>-1</sup> N in the 1970s to 0.6 kg N kg<sup>-1</sup> N in the early  
244 2000s and leveled off thereafter. The NUE of cotton and sorghum was relatively stable (e.g., 0.6  
245 kg N kg<sup>-1</sup> N) from 1970 to the mid-1990s, followed by an increase thereafter. Corn NUE  
246 consistently increased from 0.45 kg N kg<sup>-1</sup> N in the 1970s to 0.6 kg N kg<sup>-1</sup> N in the 2010s. Rice  
247 had a lower NUE (0.4 kg N kg<sup>-1</sup> N) than other crop types and showed a marginal variation over  
248 the study period. During the first half of the study period (e.g., from 1970 to the mid-1990s),  
249 national crop NUE remained stable at ~0.5 kg N kg<sup>-1</sup> N. It slowly increased to ~0.65 kg N kg<sup>-1</sup> N  
250 by the end of the 2010s, with an average of 0.59 kg N kg<sup>-1</sup> N over the entire study period (Figure  
251 2).

### 252 **3.3 Relationship between crop-specific N Surplus, recovered N, and N input**

253 We examined the dynamics of annual N surplus along the gradient of crop-recovered N and N  
254 input for the eight crops (Figure 3). Generally, N surplus shows a curvilinear relationship with N  
255 input rates for all crops except that corn N surplus displays a “V-shape” when its N input ranges  
256 from 170 to 220 kg N ha<sup>-1</sup> (Figure 3a). Among the eight crops, corn and rice show the highest N  
257 surplus as well as the highest N input rates. Along with the N input gradient, the national average  
258 N surplus displays a three-stage trajectory: a first increasing stage (e.g., N surplus changes from  
259 25 to 50 kg N ha<sup>-1</sup>) and then a leveled-off stage (e.g., ~50 kg N ha<sup>-1</sup>), followed by a further rising

260 stage (50-62 kg N ha<sup>-1</sup>), the latter of which is likely driven by the trend of corn N surplus due to  
261 its large planting acreage.

262 By contrast, the crop-specific N surplus shows a non-linear relationship with crop-  
263 recovered N (Figure 3b). Unlike other crop types, rice demonstrates a consistently increasing N  
264 surplus trajectory as recovered N increases, indicating rice N input increase is more rapid than its  
265 recovered N increase. A few crops, such as corn and winter wheat, show a shift from a positive  
266 to a negative relationship between their recovered N and N surplus, indicating N surplus in these  
267 two crops shifts from an input-driven to an output-driven pattern. Durum wheat and spring wheat  
268 experience a first-declining-and-then-increasing trend, while barley, cotton, and sorghum show a  
269 consistent non-linear declining trend along the recovered N gradient. At the national scale, N  
270 surplus first increases from 30 to 55 kg N ha<sup>-1</sup> as crop-recovered N rises from 40 to 90 kg N ha<sup>-1</sup>,  
271 and it starts decreasing when crop-recovered N is above 90 kg N ha<sup>-1</sup>. Such results at the national  
272 scale suggest that increases in recovered N of output-driven crops likely helped reduce the  
273 national N surplus, among which corn was the dominant contributor due to its large planting area.

### 274 **3.4 Spatiotemporal patterns of NUE and N surplus of eight major crops across the U.S.**

#### 275 **3.4.1 Corn**

276 The cross-county corn NUE showed a distinct spatial pattern over decades (Figures 4a-e). During  
277 the 1970s and 1980s, corn NUE in most counties was less than 0.4 kg N kg<sup>-1</sup> N, except for the  
278 Great Plains and the Midwest regions (Figures 4a and 4b). Since the 1990s, the NUE in most  
279 counties started to increase, including major corn-producing states (e.g. Iowa, Nebraska,  
280 Wisconsin) (Figure 4c). During the 2000s, the area with corn NUE > 0.5 kg N kg<sup>-1</sup> N further  
281 extended to other Midwestern states, such as Illinois, Indiana, and Michigan, and the NUE in the  
282 southern and eastern U.S. increased from 0.2 to 0.5 kg N kg<sup>-1</sup> N (Figure 4d). During the 2010s,

283 corn NUE in most counties switched to a higher level ( $> 0.5 \text{ kg N kg}^{-1} \text{ N}$ ), with peaks at  $0.9 \text{ kg N}$   
284  $\text{kg}^{-1} \text{ N yr}^{-1}$  in some northeastern states like New York (Figure 4e). However, the NUE in North  
285 and South Dakota were as high as  $0.8 \text{ kg N kg}^{-1} \text{ N}$  in the 1970s and the 1980s, which possibly  
286 signals soil mining at low levels of N input, and NUE gradually declined to  $\sim 0.5 \text{ kg N kg}^{-1} \text{ N}$  in  
287 the most recent decades. In the most recent decades, our study shows many corn-producing  
288 counties in the western U.S. “disappeared”. It is likely related to the fact that USDA NASS  
289 decided to stop reporting crop yield in these areas if cropland areas declined to a lower-bound  
290 threshold. Some marginal corn-producing counties remained at a relatively high NUE level in the  
291 Central Valley in California, Washington, and Idaho (Figures 4a-e).

292 Corn N surplus experienced a consistent spatiotemporal pattern in the U.S. from the  
293 1970s to the 2010s (Figures 4f-j). The hot spots of the largest N surplus ( $> 5.0 \text{ Gg N yr}^{-1}$ ) were  
294 found in the Midwestern counties, with the peaks in both Illinois and Iowa (Figure 4b). Over the  
295 past five decades, high N surplus intruded westward to a larger spatial extent, including the  
296 Dakotas, Nebraska, and the southern parts of Minnesota. It showed a stable trend in the  
297 remaining regions over the study period, except that increasing corn N surplus was found in the  
298 Central Valley in California after the 2000s.

### 299 **3.4.2 Rice**

300 Compared with corn, rice NUE showed an overall lower value ( $< 0.5 \text{ kg N kg}^{-1} \text{ N}$ ), but a  
301 relatively stable trend over decades (Figures S1a-e). The number of rice-growing counties also  
302 decreased over time, concentrating on the northern and central counties within the Central Valley  
303 and the Lower Mississippi Alluvia Valley (LMAV). In the LMAV region, the NUE in the 1970s  
304 was the highest among the five decades, implying the possibility of soil mining when the  
305 anthropogenic N input rate was low. In the Central Valley, the lowest NUE occurred in the 1970s

306 and reached a peak in the 1990s, while it gradually decreased during the post-1990 period  
307 (Figures S1a-e).

### 308 **3.4.3 Wheat**

309 Winter wheat NUE overall showed a decreasing trend over the past five decades (Figure S2),  
310 reflecting the low levels of N input and possible soil mining at the beginning of the study period.  
311 Specifically, several states, including Arkansas, New Mexico, and Texas, were characterized by  
312 increasing NUE since the 2000s. The winter wheat NUE in a few states, including North Dakota,  
313 Oklahoma, and Missouri, decreased from  $\sim 1.0 \text{ kg N kg}^{-1} \text{ N}$  in the 1970s to  $< 0.5 \text{ kg N kg}^{-1} \text{ N}$  in  
314 the 1990s and then increased to  $\sim 0.6 \text{ kg N kg}^{-1} \text{ N}$  by the 2010s.

315 In contrast to winter wheat, the NUE of spring wheat and durum wheat showed a distinct  
316 spatial pattern over the five decades (Figures S3 and S4). For spring wheat, NUE reached  $1.0 \text{ kg}$   
317  $\text{N kg}^{-1} \text{ N}$  in most areas in the 1970s and largely decreased in Minnesota and the western area of  
318 North and South Dakota (Figure S3). The remaining states, including Idaho, Montana, Oregon,  
319 and Washington, demonstrated a slightly decreasing trend over the study period, although this  
320 decreasing trend became smaller or insignificant after the 2000s. The NUE dynamics of durum  
321 wheat over the study period was similar to that of spring wheat, decreasing from  $\sim 1 \text{ kg N kg}^{-1} \text{ N}$   
322 in the 1970s to  $0.4 \text{ kg N kg}^{-1} \text{ N}$  in the 2010s. However, durum NUE kept relatively stable in  
323 California and Arizona over decades (Figure S4).

324 The largest N surplus (above  $5.0 \text{ Gg N yr}^{-1}$ ) of winter wheat was found in the counties in  
325 Kansas, Oklahoma, and Northern Texas, and it kept relatively stable in the rest counties (Figures  
326 S2f-j). The N surplus of durum wheat and spring wheat also remained stable over decades, with  
327 the largest N surplus ( $6.5 \text{ Gg N yr}^{-1}$ ) found in the border between Montana and North Dakota  
328 (Figures S3 and S4).

#### 329 **3.4.4 Barley, Cotton, and Sorghum**

330 In terms of crop spatial distribution, not surprisingly, the number of counties where barley,  
331 cotton, and sorghum were planted was found to decrease over the past five decades (Figures S5-  
332 S7). The major producing areas of these three crops were found to respectively shrink into the  
333 northwest, the southern coastal regions, and the Great Plains. Barley NUE remained high ( $\sim 1$  kg  
334  $\text{N kg}^{-1}$  N) over the entire study period (Figure S5). For cotton, the highest increases in NUE were  
335 found in the southeastern coasts, and it remained stable ( $\sim 0.7$  kg  $\text{N kg}^{-1}$  N) in the southwestern  
336 coasts of the U.S. (Figure S6). Sorghum NUE was relatively low ( $\sim 0.4$  kg  $\text{N kg}^{-1}$  N) in most of  
337 its planting regions in the 1970s and the 1980s and peaked in the central Great Plains in the  
338 1990s ( $\sim 0.8$  kg  $\text{N kg}^{-1}$  N) (Figure S7). Since the 2000s and the 2010s, however, sorghum NUE  
339 has shown large variations across space.

340 The N surplus of these three crops was relatively homogeneous across the nation (Figures  
341 S5-S7), with less than 2.5 Gg  $\text{N yr}^{-1}$  N surplus found in most counties. The highest N surplus of  
342 them all occurred in the 1970s and the 1980s.

#### 343 **3.5 Crop-specific nitrogen budget across the U.S.**

344 We examined the national-level N budget of the eight crops from 1970 to 2019 (Figure 5). N  
345 fertilizer was the dominant N input for the eight crops and thus the dominant driver of N surplus  
346 dynamics. Among them, corn showed a consistently increasing trend in total N input amount and  
347 crop recovered N from 1970 to 2019 (Figure 5b). Corn N surplus first increased from 2.5 Tg  $\text{N}$   
348  $\text{yr}^{-1}$  in 1970 to 3.8 Tg  $\text{N yr}^{-1}$  in the late 1980s and then slightly decreased till the 2000s, and it  
349 overall experienced large variations in the 2010s and eventually decreased to a level observed in  
350 the 1970s (Figure 5b). The total N input and crop recovered N of barley and sorghum showed a  
351 similar decreasing trend from the early 1970s to the end of the 2010s (Figures 5a and 5g). Barley

352 N surplus was around zero in the 1970s and sharply increased to  $\sim 0.2 \text{ Tg N yr}^{-1}$  in the 1980s; it  
353 decreased to less than  $0.1 \text{ Tg N yr}^{-1}$  in the 2000s and to around zero in the 2010s (Figure 5a). The  
354 maximum N surplus of barley reached up to 62% of its total N input in 1985. Sorghum N surplus  
355 slowly decreased from  $0.3 \text{ Tg N yr}^{-1}$  in the 1970s to close to zero in the 2010s, accounting for 34%  
356 to 50% of its total N input over the study period (Figure 5h). Mainly driven by the increasing N  
357 fertilizer input, the N surplus of cotton, rice, durum wheat, spring wheat, and winter wheat all  
358 slowly increased from the 1970s to the 2000s and then gradually decreased till the end of the  
359 2010s (Figures 5c,d,e,f,h). Specifically, cotton N surplus ranged from  $0.1$  to  $0.4 \text{ Tg N yr}^{-1}$ ,  
360 accounting for 17% to 48% of the total N input (Figure 5c). Rice N surplus accounted for 53% of  
361 its total N input and remained relatively stable over time, with an average of  $0.1 \text{ Tg N yr}^{-1}$  (Figure  
362 5e). The maximum N surplus of durum wheat and spring wheat accounted for 60% ( $0.05 \text{ Tg N}$   
363  $\text{yr}^{-1}$ ) and 52% ( $0.26 \text{ Tg N yr}^{-1}$ ) of their total N inputs, respectively, while their negative N surplus  
364 in the 1970s was likely due to the small N fertilizer input and soil mining (Figures 5d and 5f).  
365 Winter wheat N surplus ranged from 8% to 26% of total N input and reached a maximum value  
366 of  $0.58 \text{ Tg N yr}^{-1}$  in 1989 (Figure 5h).

367 We also aggregated the N budget of the eight studied crops to represent the national crop  
368 N budget (Figure 6). Overall, the national total N input slowly increased from  $3.3 \text{ Tg N yr}^{-1}$  in  
369 1970 to  $8.2 \text{ Tg N yr}^{-1}$  in 2019. Over 90% of the national total N input was from fertilizer use,  
370 followed by the residual N of soybean rotation and atmospheric N deposition. Crop-harvested N  
371 increased by 120%, with a minimum value of  $2.6 \text{ Tg N yr}^{-1}$  found in 1970 and a maximum value  
372 of  $5.8 \text{ Tg N yr}^{-1}$  in 2016. Crop N surplus on average accounted for 41% (27% to 55%) of the  
373 total N input during the study period. A minimum crop N surplus of  $1.2 \text{ Tg N yr}^{-1}$  was found in  
374 1971, and a maximum of  $5.0 \text{ Tg N yr}^{-1}$  in 2012 (Figure 6). The 2012 peak of crop N surplus

375 might be related to the widespread drought in the U.S. this year. Additionally, the national N  
376 surplus showed a near-parabolic trend, with an annual increasing rate of  $0.09 \text{ Tg N yr}^{-1}$  ( $p < 0.05$ )  
377 over the period from 1970 to 1989 and an annual decreasing rate of  $0.02 \text{ Tg N yr}^{-1}$  ( $p = 0.1$ ) from  
378 1990 onward. The temporal dynamics of the national total N surplus were dominated by corn due  
379 to its large planting areas and high N input rates per unit area.

### 380 **3.6 Spatiotemporal trends of total N input and N surplus of the eight studied crops**

#### 381 **3.6.1 Corn**

382 During the period from 1970 to 1989, corn total N input in the Midwest U.S. showed a  
383 significantly increasing trend, with a maximum annual increasing rate  $> 0.2 \text{ Gg N yr}^{-2}$ , while it  
384 experienced a decreasing trend in the states in the eastern coast and the central Great Plains  
385 (Figure 7a). Since 1990, the significantly increasing trend in corn total N input has extended to a  
386 much larger area covering the entire Midwest U.S. and the northern Great Plains, with a peak  
387 increasing rate greater than  $0.25 \text{ Gg N yr}^{-2}$  found in Dakotas (Figure 7b).

388 The trend of corn N surplus from 1970 to 1989 demonstrated a similar spatial pattern but  
389 a smaller extent ( $p < 0.05$ ), compared with its total N input (Figure 7c). The maximum increasing  
390 rate ( $\sim 0.2 \text{ Gg N yr}^{-1}$ ) was found in the Corn-Belt region, including Illinois, Iowa, Indiana, Ohio,  
391 Michigan, and the eastern parts of Dakotas, while the N surplus in the remaining regions showed  
392 either a significantly decreasing or a flat trend during this period (Figure 7c). In the post-1990  
393 period, corn N surplus in most of the Corn-Belt region switched to a decreasing or flat trend;  
394 however, it significantly increased in the northern Great Plains (e.g., Dakotas and Kansas) due to  
395 the significant increases in its total N input (Figure 7d).

### 396 **3.6.2 Wheat**

397 The majority of the winter wheat growing areas experienced significant increases in total N input  
398 during the pre-1990 period (Figure S8a). From 1990 to 2019, however, it reversed to a  
399 significantly decreasing trend in most areas, except in Dakotas and Wyoming (Figure S8b).  
400 Winter wheat N surplus during these two periods followed a similar spatial pattern as its total N  
401 input (Figures S8c and S8d). Similar to winter wheat, spring wheat also experienced a positive  
402 increase in total N input over the past five decades, with a hotspot found in the northern Great  
403 Plains (Figure S9a). Spring wheat total N input in the eastern part of Dakotas had a significant  
404 decrease during the post-1990 period (Figure S9b). The spatial patterns of spring wheat N  
405 surplus trend during these two studied periods were largely driven by its total N input, with a  
406 peak ( $> 0.25 \text{ Gg N yr}^{-2}$ ) found in the northern Great Plains (Figures S9c and S9d). Durum wheat  
407 showed insignificant trends in total N input and N surplus during the two studied periods, and  
408 only some marginal counties in the northern states and Arizona demonstrated a significant trend  
409 over the past five decades (Figure S10).

### 410 **3.6.2 Barley, Cotton, Rice, and Sorghum**

411 The total N input of barley had an increasing trend in most counties from 1970 to 1989 (Figure  
412 S11). The largest total N input increase rate (e.g.,  $> 0.2 \text{ Gg N yr}^{-2}$ ) was found in North Dakota.  
413 However, most of the counties have been characterized by decreasing total N input since 1990  
414 (Figures 11a and 11b). The trends of barley N surplus mostly followed the spatial patterns of its  
415 total N input in the pre- and post-1990 periods (Figures S11c and S11d).

416 Cotton total N input showed a significantly decreasing trend in most counties (Figures  
417 S12a and S12b). Over the entire study period, cotton N surplus showed a decreasing trend, with

418 the peak annual decreasing rate (e.g.,  $-0.2 \text{ Gg N yr}^{-2}$ ) detected in central Texas since 1990  
419 (Figures S12c and S12d).

420 Rice total N input in the Rice-Belt demonstrated an overall increasing trend during the  
421 entire study period (Figures S13a and S13b). In other main rice-producing regions, such as  
422 California, the total N input experienced an insignificant decreasing trend from 1970 to 1989,  
423 and this decreasing trend became smaller or turned into an increasing trend after 1990. Not  
424 surprisingly, rice N surplus followed the spatial patterns of its total N input (Figures S13c and  
425 S13d), indicating an “input-driven” trajectory of N surplus.

426 Sorghum total N input displayed decreasing trends in most counties over the entire study  
427 period (Figures S14a and S14b). Counties in South Dakota and southeastern Missouri were  
428 identified as hotspots of increasing total N input from 1970 to 1989, whereas the centers of  
429 increasing N input moved to counties in eastern Kansas from 1990 to 2019. Similar to other  
430 above-mentioned crops, sorghum N surplus trend was also largely driven by its total N input  
431 over the entire study period (Figures S14c and S14d), demonstrating a smaller area with a  
432 significant changing trend after 1990 than N input ( $p < 0.05$ ).

#### 433 **4 Discussion**

434 Our analysis quantified yield responses, represented by crop-recovered N, to the N input of eight  
435 major crops in the U.S. using the long-term data. The largest yield response to N addition was  
436 found in corn nationally, while yield responses in the remaining seven crop types shifted to a  
437 lower rate as N input increases (Figure 1a). It likely indicates N saturation or limiting factor  
438 shifted from N to other resources (such as water, phosphorous, etc.) for these seven crops. We  
439 also examined the county-level crop-specific patterns of NUE and N surplus. A consistent

440 pattern that emerged for six out of the eight crops showed a decreasing NUE along the N input  
441 gradient, whereas the NUE of corn and rice barely changed (Figure 1b). Our analyses illustrated  
442 that the N surplus of corn and winter wheat was likely reduced by their yield increases, whereas  
443 the N surplus of the other studied crops was likely to be mitigated by a lower N input (Figure 3).

#### 444 **4.1 NUE responses to N inputs**

445 National crop NUE demonstrates a non-linear relationship with N input increases (Figure 1b).  
446 The non-linear NUE has been previously reported across the entire U.S., with a rapid yield  
447 response to medium-level N fertilizer use rates (125-175 kg N ha<sup>-1</sup> yr<sup>-1</sup>) (Lassaletta et al., 2014;  
448 Zhang et al., 2015). However, these previous studies did not provide information on crop-  
449 specific NUE patterns and yield responses to N input increase. Using state-level survey data, Lu  
450 et al., (2019) demonstrated that crop yield of corn and winter wheat in the U.S. plateaued with N  
451 fertilizer increases, and NUE declined when N input exceeded the threshold for these two crops.  
452 This finding generally agrees with our county-level analyses, suggesting that crop NUE declined  
453 along with N input increases (Figure 1b).

454 Within the U.S., increasing crop yields, resulting from advanced crop varieties, improved  
455 management practices and other technological improvements, largely increased crop NUE since  
456 the 1990s (van Grinsven et al., 2015; Xu et al., 2013). Additionally, N fertilizer input has  
457 remained steady in the U.S. since the 1990s (Cao et al., 2018). Among the eight studied crops,  
458 our analysis identified two crop groups showing different NUE dynamics along the N input  
459 gradients. The NUE of six crop types (e.g., barley, cotton, durum wheat, sorghum, spring wheat,  
460 and winter wheat) experienced an overall decreasing trajectory with a lower range of N input  
461 rates (e.g., 10-100 kg N ha<sup>-1</sup>, Figure 1b), while corn and rice NUE barely varied with N input  
462 rates that distributed in a higher range (e.g., 100-220 kg N ha<sup>-1</sup>). Hence, the flat national

463 relationship between crop NUE and N inputs should be used with caution for the nutrient  
464 management of specific crops at local scales because spatiotemporal variations in crop yield and  
465 NUE are likely to be smoothed when aggregated to the national scale. Our results also reflect  
466 that it is critical to developing sub-national analyses for understanding the impacts of N addition  
467 on yield gain of specific crops, especially in counties with a large amount of N input and large  
468 cropland areas like the U.S. Midwest.

#### 469 **4.2 Dynamics of U.S. crop NUE over space and time**

470 Our estimates showed that the national crop NUE ranged from 45% to 73% with an average of  
471 59% in the U.S over the past 50 years (Figure 2). A previous study reported the national crop  
472 NUE of 2007 was close to 55% in the U.S. (Houlton et al., 2013), which is slightly lower than  
473 our estimate. Using national-level data, Lassaletta et al., (2014) reported that U.S. crop NUE was  
474 in between 47% and 73% from 1961 to 2010, with a mean value of ~60% over those 50 years,  
475 agreeing well with our estimates. Nevertheless, we revealed more variations over space, time,  
476 and crops.

477       Among these crops, the spatiotemporal patterns of barley and three wheat types largely  
478 differ from other crops (Figures 2 and S2-S5). The initial higher NUE values of barley and three  
479 wheat types were likely due to the low levels of initial N input, and wheat might access extra N  
480 through soil mining at this stage, reducing soil fertility (De Klein et al., 2017). The NUE  
481 decreases in these crops were possibly caused by increases in N input that was used to alleviate  
482 soil degradation (De Klein et al., 2017). According to a recent study, N fertilizer input into  
483 winter wheat sharply increased from the 1970s to the 2000s (Cao et al., 2018). However, winter  
484 wheat yield increases were slower compared with its N input rates (Lu et al., 2019), leading to  
485 the decreasing NUE trend. The latter increasing trend of its NUE since the 2000s (Figure 2) was

486 possibly induced by the recent increases in yield and stagnated N fertilizer inputs. For example,  
487 winter wheat yields were reported to approach their maximal potential levels in the U.S. Great  
488 Plains, with less genetic advancement compared with corn (Ray et al., 2012). Cao et al. (2018)  
489 demonstrated that the N fertilizer input rates for winter wheat have leveled off since the 2000s.  
490 Another study shows that, for winter wheat from 1960 to 2008, a change point of its NUE  
491 (transitioning from an increase to a level-off) was observed in the western U.S. in 1993 and the  
492 central U.S. in 2003 (Lin & Huybers, 2012). Besides, the winter wheat-growing areas presented  
493 a broad range of agroclimatic zones, which might drive the spatial variability of its NUE.

494         Despite the increases in N fertilizer input rates for durum wheat and spring wheat in the  
495 U.S. from the 1960s to the most recent decade (Cao et al., 2018), the NUE of these two wheat  
496 types largely decreased from the 1970s to the 2000s and remained leveled-off since then. It is  
497 likely caused by the higher N input than the N demand for these two types of wheat. Besides,  
498 durum wheat and spring wheat were mainly distributed in the northwestern states, in which arid  
499 and semi-arid conditions might be one of the limiting factors for crop yield and NUE.

500         For corn, cotton, and sorghum, although NUE declines occurred in some counties (e.g.  
501 corn NUE declines in Dakotas and sorghum NUE declines in northern Texas), it overall  
502 increased over the past five decades (Figure 4, S1, S6, and S7). Even though an increasing trend  
503 was widely observed, corn in southern Great Plains and the eastern U.S. still showed a  
504 consistently low NUE (e.g., less than  $0.5 \text{ kg N kg}^{-1} \text{ N}$ ) over the past five decades (Figure 4). A  
505 previous study reported that the lower corn yield in Texas might be due to the common fall N  
506 application practice that likely increased soil N loss during the winter fallow period (Torbert et  
507 al., 2001). Here we used corn as an example to compare its annual N surplus against fall and  
508 spring N fertilizer application rates since these two application timings account for the majority

509 of annual total N input. For example, fall N fertilizer application accounted for 20% of annual  
510 total N fertilizer use in Texas and 31% in Iowa and Minnesota, and spring N application  
511 accounted for 34% in Texas, 43% in Iowa, and 59% in Minnesota in 2015. Across the U.S., corn  
512 fall N application from 2010 to 2019, ranging from 0-60 kg N ha<sup>-1</sup>, positively correlated with its  
513 N surplus (R=0.57, p<0.01, Figure 8a). On the other hand, a positive correlation was found  
514 between the spring N fertilizer application and N surplus with a higher statistical significance  
515 (Figure 8b). These findings suggested that fall N fertilizer application to corn could significantly  
516 contribute to its annual total N surplus, despite its lower input rates.

517         The increases in cotton NUE in the southeastern U.S. over the past five decades were  
518 more likely due to its yield increases because its total N input remained relatively stable (Figure  
519 4c). However, cotton NUE in the southern Great Plains (e.g., Texas) did not show significant  
520 changes although it remained relatively high (Figures S6 and S12). A recent study reported there  
521 was a difference in the adoption of conservation tillage between the southern Great Plains cotton  
522 systems (30%) and the southeastern U.S. (70%) (DeLaune et al., 2020). The residue retaining  
523 benefits of reduced till during the cotton rotation phase could significantly increase cotton lint  
524 yields in the southern Great Plains (Baumhardt et al., 2012), which might contribute to the  
525 different cotton yield responses to N input rates (Boquet et al., 2004). The 1990s-onward high  
526 sorghum NUE (> 0.6 kg N kg<sup>-1</sup> N) in the Great Plains was likely due to the reducing total N input  
527 (Figures 5g, S14a, and S14b). Rice NUE displayed a consistent spatiotemporal pattern over the  
528 past five decades (Figures 2 and S1), reflecting the linear relationship we found between its yield  
529 and N input over the entire study period (Figure 1a). More specific studies are needed to  
530 investigate whether other agricultural management practices can help reduce the N surplus of  
531 these three crops across regions.

### 532 **4.3 Regional hotspots of high NUE: a future increase in crop NUE?**

533 Regional variations in patterns of major crop NUE versus total N inputs from 1987 to 2012 were  
534 reported by a previous study that investigated ~17 crop species in the U.S. (Swaney et al., 2018).  
535 This study showed that crop NUE likely decreased with growing N fertilizer input in the  
536 majority of regions in the U.S. but did not report crop-specific crop NUE. Our results identified  
537 the critical areas with high crop NUE, well echoing the findings of a recent study (Sabo et al.,  
538 2021). High NUE in corn was found on the most fertile Midwest soils (e.g., Iowa, Illinois,  
539 Minnesota, and Nebraska), which received plentiful rainfall, and advanced management  
540 practices (e.g., rotation, tile drainage, and tillage). Furthermore, previous studies showed that  
541 earlier corn planting in some of these regions (e.g., Minnesota, Nebraska, and Iowa) contributed  
542 significantly to yield increases (Kucharik, 2006, 2008), which likely boosted the NUE increases.  
543 High crop NUE was also found in the Great Plains, where irrigation practice is widely applied to  
544 winter wheat and sorghum, as well as the south coasts, where advanced management practices  
545 (e.g. pest and disease control) are utilized for cotton production.

546 However, these high-NUE counties and regions were found to be the hotspots of N  
547 surplus for each crop type owing to the large planting area and high N input rates. For instance,  
548 the Midwest regions showed the largest corn N surplus (on average 5-8 Gg N yr<sup>-1</sup>). Nevertheless,  
549 we still found that the national total N surplus approximately declined at an annual rate of 0.8 Tg  
550 yr<sup>-1</sup> during the 2010s (Figure 6), which was significantly driven by corn N surplus decline  
551 (Figure 5b and S15b). Particularly during 2015-2019, we found that corn N surplus in the U.S.  
552 Midwest decreased to a level close to the 1970s (Figure 3f and Figure S15b). The national total  
553 N input for corn during 2015-2019 was averaged ~2 Tg N yr<sup>-1</sup> higher than that in the 1970s,  
554 while N surplus levels were close between these two periods (Figure 6). This indicates that crop

555 harvested N was much higher during the most recent years, likely driven by crop technology  
556 improvement and enhanced agricultural management practices. Corn yields were found to  
557 increase over the past decades (Kucharik et al., 2020), and better N management practices (e.g.,  
558 using cover crops and bioreactors) adopted by farmers might also help enhance crop yield and  
559 retain soil N (Christianson et al., 2018; Stuart et al., 2018). The most notable effort is the 4Rs of  
560 Nutrient Stewardship strategy (i.e., right N fertilizer application place, right application time,  
561 right N fertilizer use amount, and right N fertilizer use rate), which promotes proper nutrient  
562 management strategies to improve crop N use efficiency (IFA, 2009). In addition, policy-based  
563 efforts, such as the “avoid, control, and trap” strategy proposed by the USDA (Osmond et al.,  
564 2015), and state-based efforts, such as Iowa’s Nutrient Reduction Strategy, have set nutrient  
565 reduction goals and provided voluntary guidance to farmers to meet those goals (Lawrence,  
566 2013). On the other hand, we also realized that the 2012 drought in the U.S. was likely the main  
567 reason for the high average N surplus in the 2010s (Figure 5 and 6), causing a major increase in  
568 N surplus in Dakotas, Missouri, and Kansas (Figure 7). This indicates that extreme climatic  
569 events like drought could largely undermine the higher-input-efficiency strategy’s effectiveness  
570 in improving crop yield, which was proven in our follow-up analysis by using this database  
571 (Zhang, et al., 2021).

572 Our work also showed that among the eight crops, the counties with the lowest NUE  
573 generally corresponded to a much smaller total N surplus because the total N input amount in  
574 these counties was not as high as that in the hotspot regions. This finding indicates that future  
575 increases in N input in the low-NUE counties should be cautious because it requires much larger  
576 crop yield increases to take up excessive N in these areas, despite the low N surplus revealed.  
577 Boosting crop yield was shown to be difficult in these regions due to the low soil fertility or

578 climatic stresses (Hou et al., 2019). As climatic changes will continue, the low crop NUE regions  
579 may be at a higher risk than other regions. Compared with other key planting areas, the climate  
580 conditions in the U.S. Great Plains, southwestern U.S., and the western U.S. have a higher  
581 likelihood of extreme heat (i.e., days above 30 °C or extreme degree days) and drought combined,  
582 both of which have been documented as causes of crop yield loss (Basso & Ritchie, 2014; Lobell  
583 et al., 2013, 2014). The historical trends in climate and current low-NUE may be a precursor of  
584 continued N surplus increases in the next few decades in these counties.

585         Even in the counties with the highest crop NUE (e.g. the Midwest), extreme rainfall and  
586 flooding events likely drove a larger magnitude of N loss via leaching to surface water and  
587 groundwater bodies (Lu et al., 2020). Climatic warming and intensive land surface management  
588 (e.g. tillage) could largely increase N<sub>2</sub>O emission given the large input of N fertilizer (Gregorich  
589 et al., 2008). Additionally, N fertilizer shifting to the Urea-base fertilizer could enhance NH<sub>3</sub>  
590 emission, especially in areas with alkaline soils like in the northern Great Plains (Cao et al.,  
591 2020). Using county-level crop yield data and climate datasets, a recent study demonstrated that  
592 the magnitude of corn yield loss driven by excessive rainfall was comparable to that driven by  
593 extreme drought in the U.S. (Li et al., 2019). Therefore, farmers in these counties should be  
594 cautious when increasing N fertilizer use because the yield gap in these regions is likely the  
595 smallest nationwide. It is even more difficult to avoid potential N loss in these areas given the  
596 smaller yield gap and the higher likelihood of frequent extreme rainfall events and longer  
597 drought.

#### 598 **4.4 N loss risk and recommendations for future N management**

599 Achieving food security, sustainability, and ecosystem services at the regional and global scales  
600 is the ultimate goal of modern agriculture (Foley et al., 2011). Previous studies estimated that

601 global cereal grain demand will not be met without at least a 60% increase in global N use if  
602 crop NUE remains the same as the current level (Dobermann & Cassman, 2005; Ladha et al.,  
603 2016). A recent study developed spatially explicit typologies for components of crop N mass  
604 balance and indicated a large range of crop N trajectories across the U.S. (Byrnes et al., 2020),  
605 however, crop-specific N trajectories were not reported in this study. Using crop-specific N  
606 budget data, our analyses suggested that the N surplus of corn and winter wheat was likely  
607 output-driven and could be potentially reduced by enhancing the crop yield (Figure 3b). Besides,  
608 the national crop N surplus was largely driven by corn N surplus (Figures 5b and 6). Therefore, it  
609 will be more challenging to decrease the U.S. crop N surplus without further increasing corn  
610 yield that was detected linearly increasing over time (Kucharik et al., 2020). A recent study  
611 demonstrated that corn yield response likely leveled off when N fertilizer exceeded its threshold  
612 level at approximately 150 kg N ha<sup>-1</sup> (Lu et al., 2019). Furthermore, past trends may not  
613 necessarily indicate a continued future increasing trend of crop yield. Based on the above-  
614 mentioned evidence, it is still unclear whether further increases in N inputs can keep enhancing  
615 yields, and the yield response can be much more complex when the spatial heterogeneity of other  
616 environmental facts is considered. Hence, without a comprehensive understanding of crop yield  
617 performance under different management practices given various physical conditions, it is  
618 challenging to predict the future risk of N loss from cornfields in the U.S.

619         Among the remaining seven crop types, a few of them, such as barley and winter wheat,  
620 demonstrated a shift in N surplus from input-driven to output-driven as yield increases (Figure  
621 3). Interestingly, in the major winter wheat-producing states (e.g. Oklahoma and Kansas) and  
622 barley-producing states (e.g. North Dakota and California), we found that the N surplus showed  
623 an overall decreasing trend over decades, which might be due to the consistently increased NUE

624 observed in these states. Therefore, for the remaining counties where wheat and barley are  
625 widely planted, increasing crop yield could help to reduce N loss. Additionally, efficient N  
626 management of these specific crops is of particular importance for policy- and decision-making  
627 that aims at reducing N-related environmental pollution (Mitsch et al., 2001; Rabalais et al.,  
628 2002).

629         Compared with the crops discussed above, sorghum and durum wheat showed distinct  
630 relationships among N surplus, N input, and crop yield. Recent studies pointed out that the  
631 optimum N application rates might be much less if sorghum was grown under limited irrigation  
632 and dryland areas in the Great Plains (Hao et al., 2014; Tamang et al., 2011). Our analyses  
633 showed that N surplus decreased along with the sorghum yield gradient (Figure 3). Therefore, to  
634 maintain a lower N surplus, N input to sorghum should be cautiously estimated with the  
635 considerations of local environmental conditions (e.g., climate and soil wetness) and  
636 management practices (e.g., irrigation availability) that jointly affect its yield. For durum wheat,  
637 our analyses presented an overall positive correlation between its N surplus and N input rate as  
638 well as its yield. This indicates that N input stimulated durum wheat yield, but its yield increase  
639 was slower than the N input growth. An experimental study showed that, as the N fertilizer rate  
640 increased from 73 to 403 kg N ha<sup>-1</sup>, although grain yield and total N uptake of durum wheat  
641 increased, its NUE linearly decreased (Liang et al., 2014). Besides, N fertilizer effects were  
642 much more significant than cultivar effects on its NUE, and so growers' decisions on N input  
643 rate would be more efficient in enhancing the NUE than their cultivar selection (Liang et al.,  
644 2014). Liang et al., (2014) also pointed out that grain yield and total N uptake would plateau  
645 when its N fertilizer input reached a threshold. Therefore, the loose coupling between elevated N

646 fertilizer input and yield gain implies that durum wheat N surplus is more regulated by its N  
647 fertilizer input.

#### 648 **4.5 Uncertainty**

649 The data and methods adopted in our study have several limitations. First, we used the NuGIS N  
650 fertilizer consumption ratio of the county to state in 1987 and 2014 to gap-fill the county total N  
651 consumption for the periods of 1970-1986 and 2015-2019, respectively. The gap-filled county-  
652 specific N fertilizer consumption had little impact on the state- and national-level analyses.  
653 However, it might lead to over- or under-estimation of N fertilizer use rates and N surplus  
654 amount in some counties for these periods, if significant changes occurred in county-level crop  
655 acreage or crop-specific fertilizer use rate. This uncertainty can be reduced if the county-level  
656 total N fertilizer consumption data become available in the entire study period. Second, due to  
657 the lack of time-series data of county-level crop-specific manure application rate, we assumed  
658 that manure N applied per acre of each crop type was consistent with the multi-year crop-specific  
659 average values reported in each state. Such an assumption likely led to a less variable manure  
660 application rate for a given crop over time and space. Third, we used harmonized cropland  
661 distribution database to extract the N deposition inputs for each crop each year. However, before  
662 2008 (the earliest year when CDL-derived crop type maps are available for the entire nation), the  
663 crop-specific N deposition rates we extracted in this study were likely biased by limited  
664 knowledge of crop type and distribution patterns within a county. Finally, the crop N recovery  
665 coefficient and the coefficient determining the soil residual N from soybean N fixation were  
666 assumed to not vary with time and space because there are no systematic methods that account  
667 for these variations. Future research may provide more insights and efforts in gap-filling,

668 reducing uncertainties in data and methods, and assess the roles of these limitations in  
669 characterizing spatiotemporal patterns of crop N budget in the U.S.

## 670 **5 Conclusion**

671 This study investigated the N budget of eight major crops in the U.S., including barley, corn,  
672 cotton, durum wheat, spring wheat, winter wheat, sorghum, and rice. This study differed from  
673 previous analyses by using county-scale long-term data from 1970 to 2019 for the eight major  
674 crop types and comparing the responses of crop recovered N, NUE, and N surplus to N input  
675 over space and time. Our results showed that national crop NUE estimated from the eight major  
676 crops slowly increased from 0.5 kg N kg<sup>-1</sup> N in the 1970s to 0.65 kg N kg<sup>-1</sup> N in the 2010s, with  
677 an average of 0.59 kg N kg<sup>-1</sup> N. The N surplus for these eight major crops on average remained  
678 at around 41% (27% to 55%) of the total N input during the study period, which is consistent  
679 with other estimates (Houlton et al., 2013; Lassaletta et al., 2014). The corn N surplus in the  
680 2010s decreased to a level close to the 1970s (Figure 5b and S15b), which drove the large  
681 decline of the national total N surplus (Figure 6). This was likely caused by policy-driven and  
682 state-oriented efforts on reducing agricultural N loss. Spatial patterns in NUE and N surplus over  
683 the past half-century suggest some regions had historically improved crop N utilization but were  
684 still facing a high risk of losing N when N input increases. Increases in N surplus in some  
685 regions, including portions of the Great Plains (e.g., corn N surplus) and Rice-Belt (e.g., rice N  
686 surplus) over the most recent decades, may be a cause of concern. Furthermore, N input has been  
687 a major factor in driving yield increase. Yet, for some crops such as wheat and barley, the crop  
688 yield has likely plateaued despite increasing N inputs. With more extreme weather events and  
689 changing climate, future crop NUE is less certain, and some regions like the Midwest and the  
690 Great Plains appear to be at higher risk than other regions for future N-related pollution if crop

691 yield does not increase or extra N is applied. Extensive farm-level data collections across diverse  
692 conditions of climate, soil, and management regimes are expected to reveal the dominant factors  
693 controlling NUE variations in different crop types across the nation.

#### 694 **Data availability**

695 The annual crop-specific NUE and N surplus data can be found via  
696 <https://doi.org/10.6084/m9.figshare.13030436> (Zhang et al., 2021).

#### 697 **Competing interests**

698 The authors declare no competing interest.

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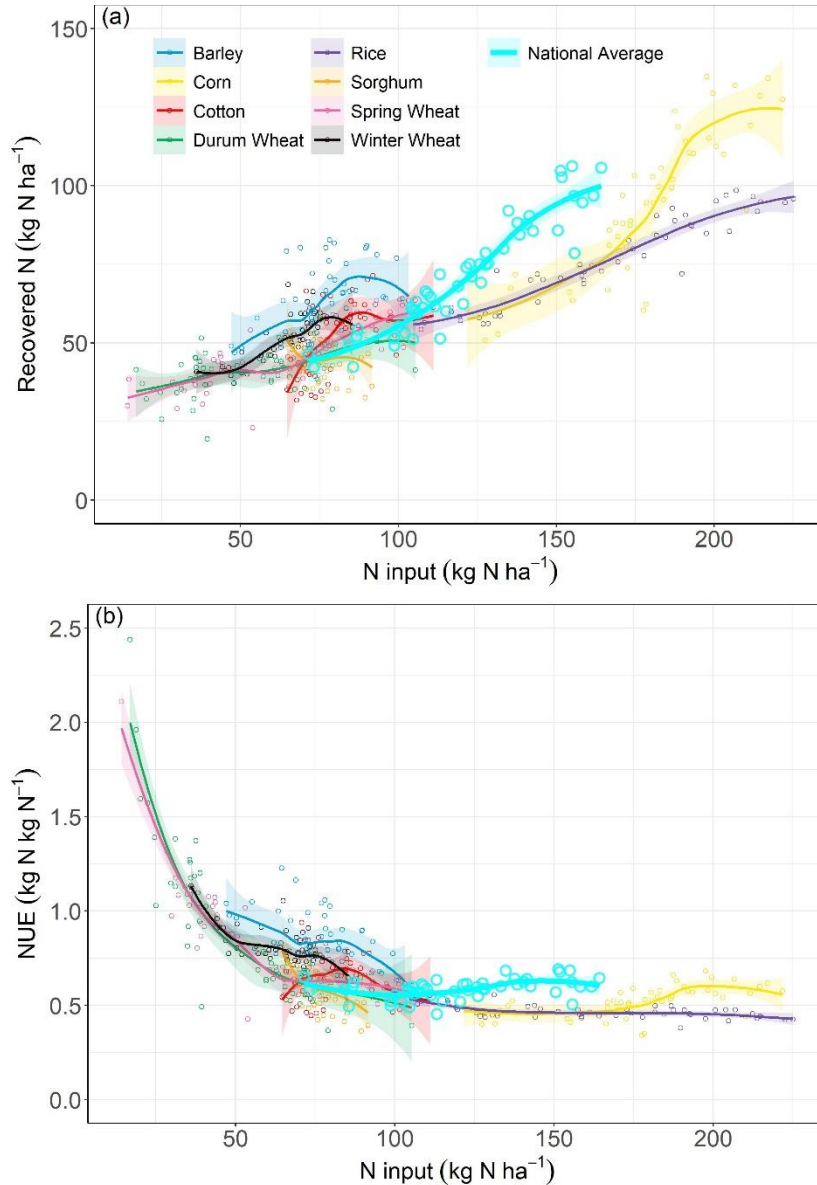
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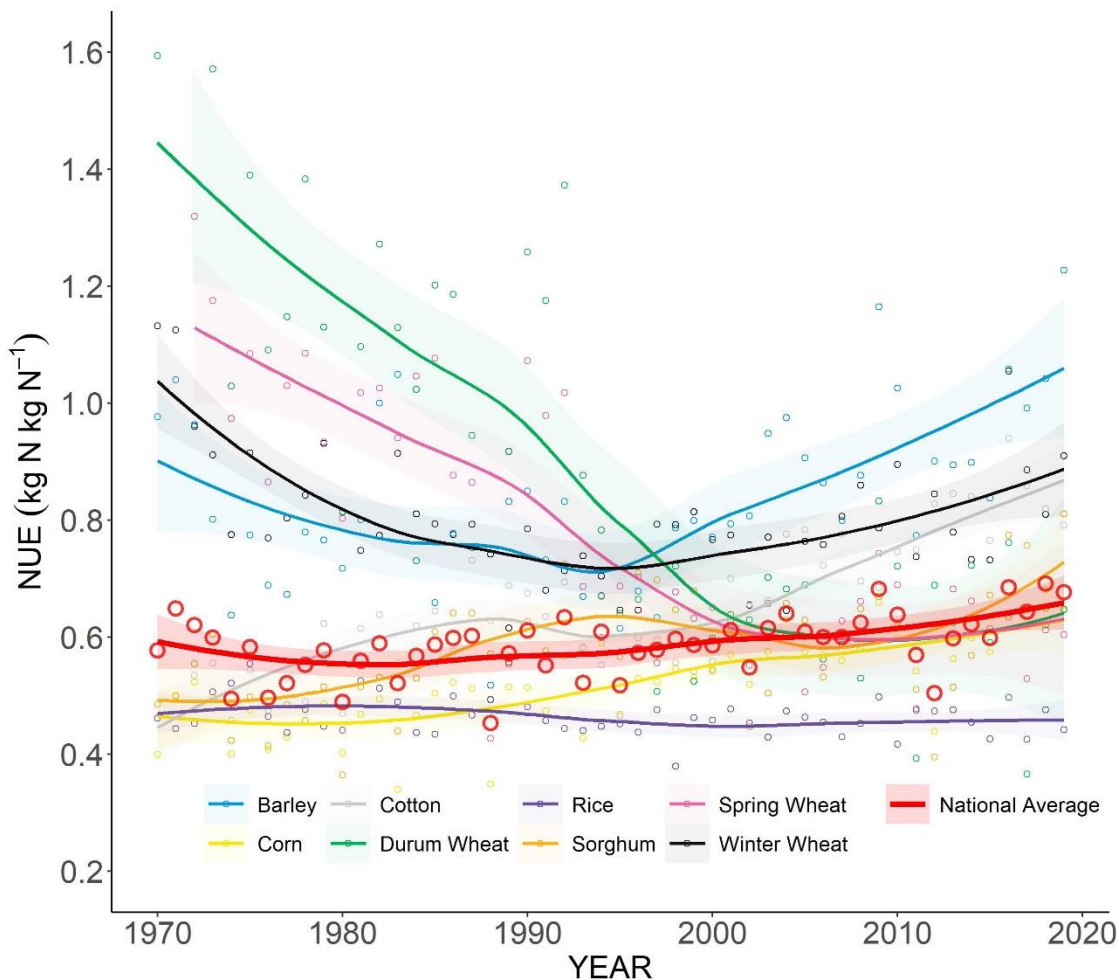
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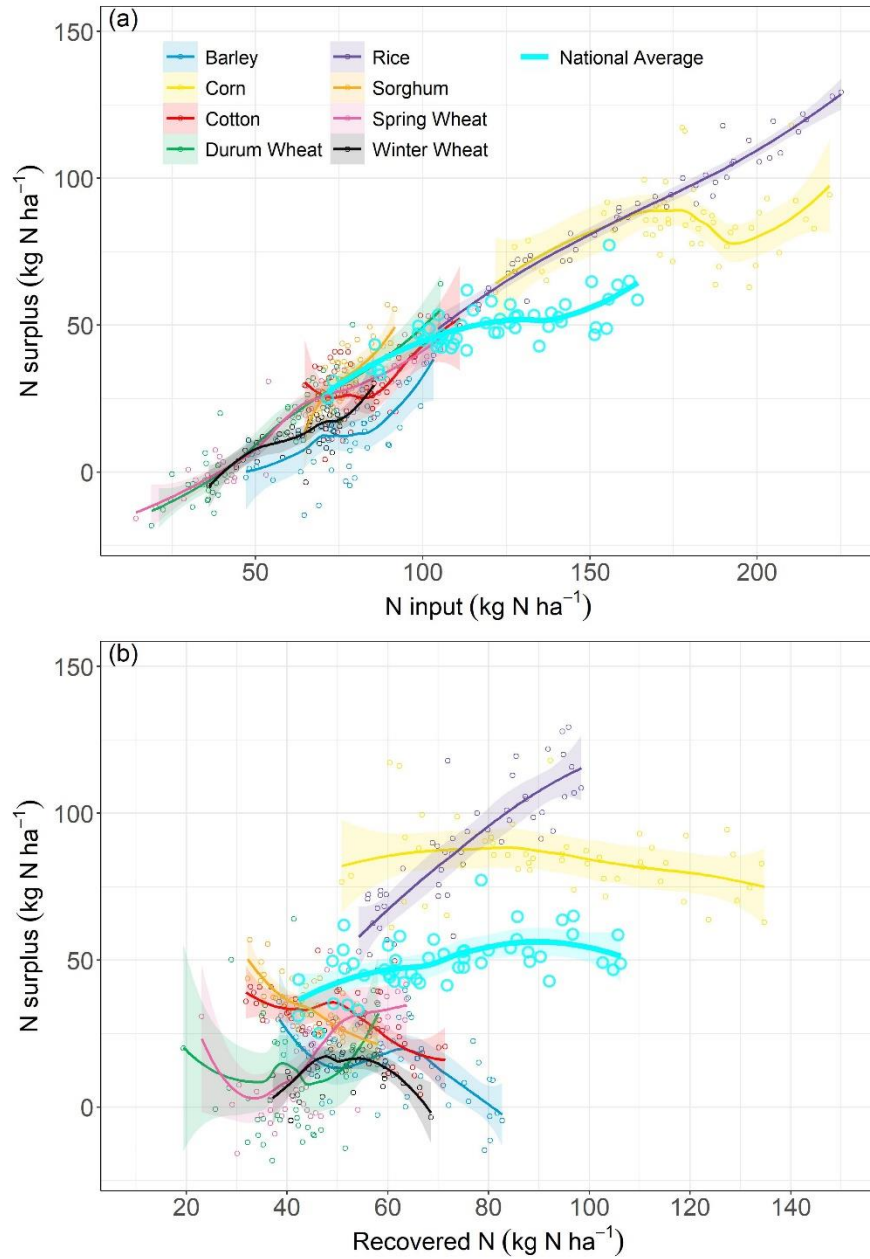
913 Figure 1. The relationship between crop yield and total nitrogen (N) input rate (a) and the  
 914 relationship between crop nitrogen use efficiency (NUE) and total N input rate (b) of eight major  
 915 crops from 1970 to 2019 in the U.S. The solid curves are the “loess” fitted correlations for each  
 916 crop type over the analyzed years (1970-2019). The shaded areas around the fitted curves  
 917 indicate the 95% confidence interval. The national average yield and N input data from 1970 to  
 918 2019 are aggregated from the eight crops based on the planting area. The national crop NUE

919 from 1970 to 2019 is calculated as the ratio of total crop harvested N to the total N input of the  
920 eight crops. The national-level data are shown by the cyan dots and cyan solid line for each pair  
921 of the variables at a national level.



922  
923 Figure 2. National-scale nitrogen use efficiency (NUE) time series of eight studied crops and the  
924 national average crop NUE from 1970 to 2019 in the United States. The solid curves are the  
925 “loess” fitted curves for each crop over the analyzed years (1970-2019). The shaded areas around  
926 the fitted curves indicate the 95% confidence interval. The red solid curve indicates the non-  
927 linear NUE trend at a national level. The crop NUE, including the crop-specific and national

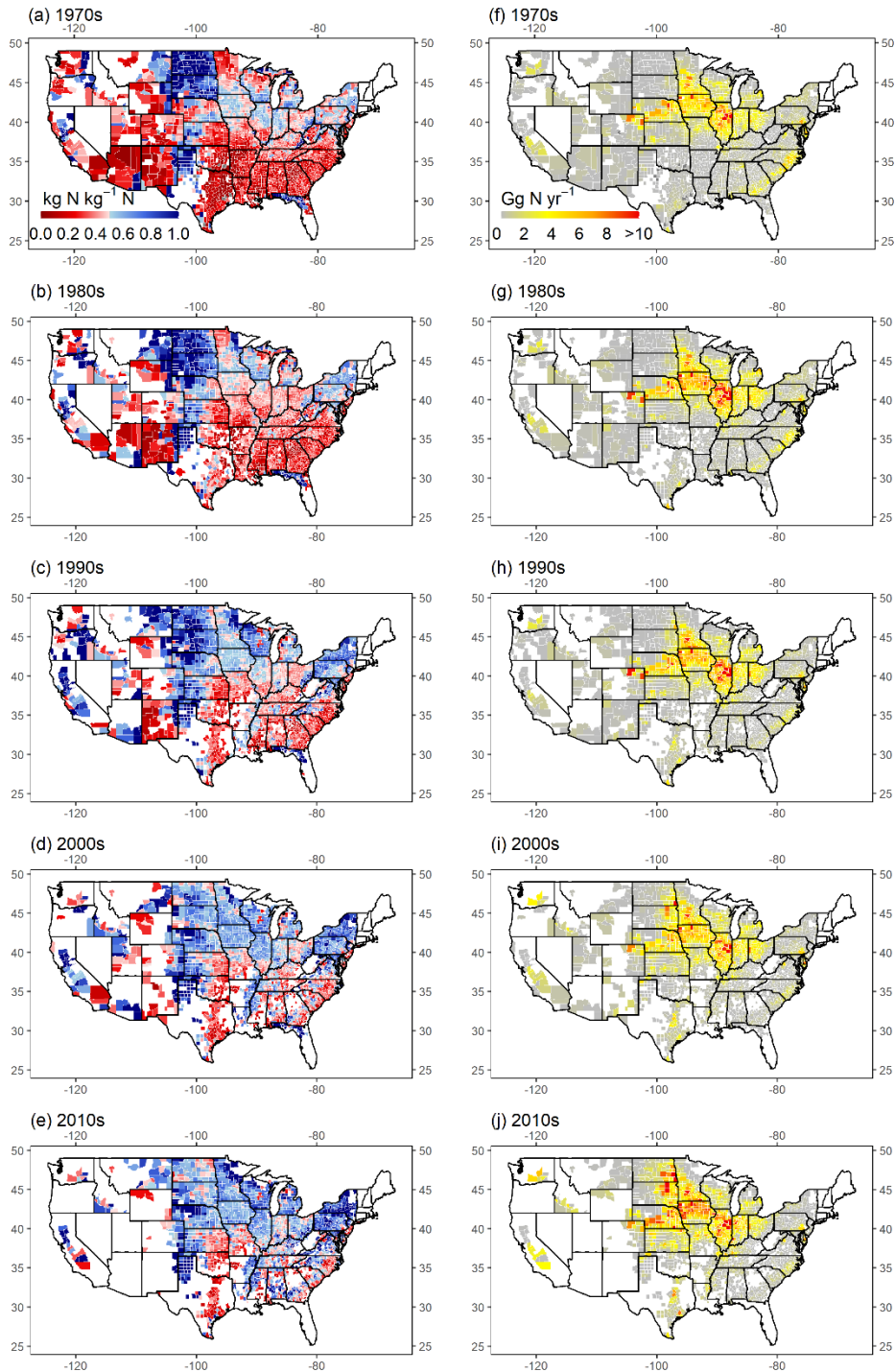
928 average, is calculated as the ratio of total crop harvested N to the total N input for each crop type  
929 or the eight crops counted together.



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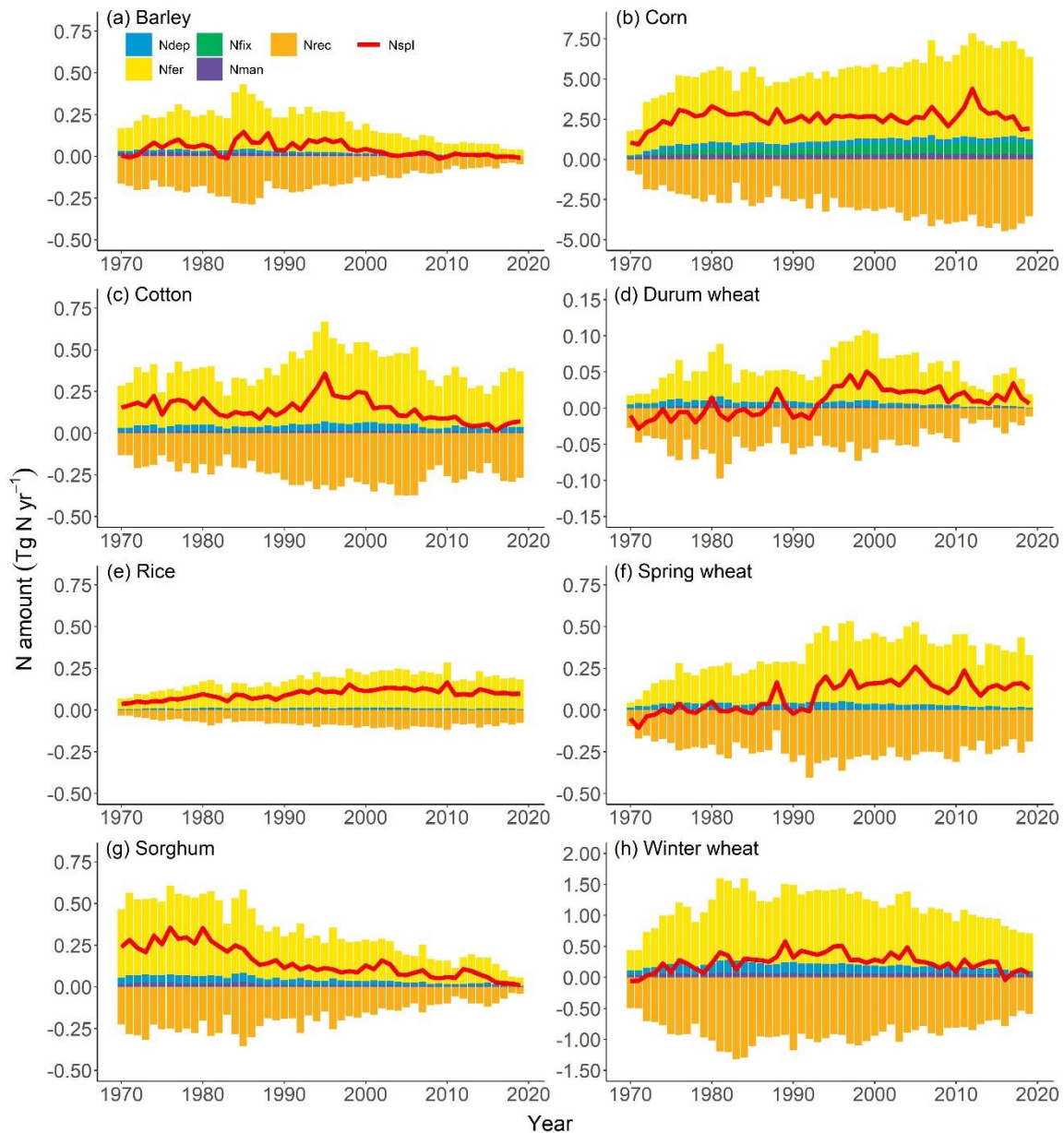
931 Figure 3. Dynamics of N surplus over the gradient of total N input rate (a), and crop yield (b) for  
932 eight major crops in the U.S. from 1970 to 2019. The solid curves are the “loess” fitted curves  
933 for each crop type over the analyzed years (1970-2019). The shaded areas around the fitted

934 curves indicate the 95% confidence interval. The cyan solid line indicates the non-linear trend  
935 for each pair of variables at the national level. The national average N surplus, N input, and crop  
936 yield are aggregated from the eight crops based on the planting area, shown by the cyan dot.



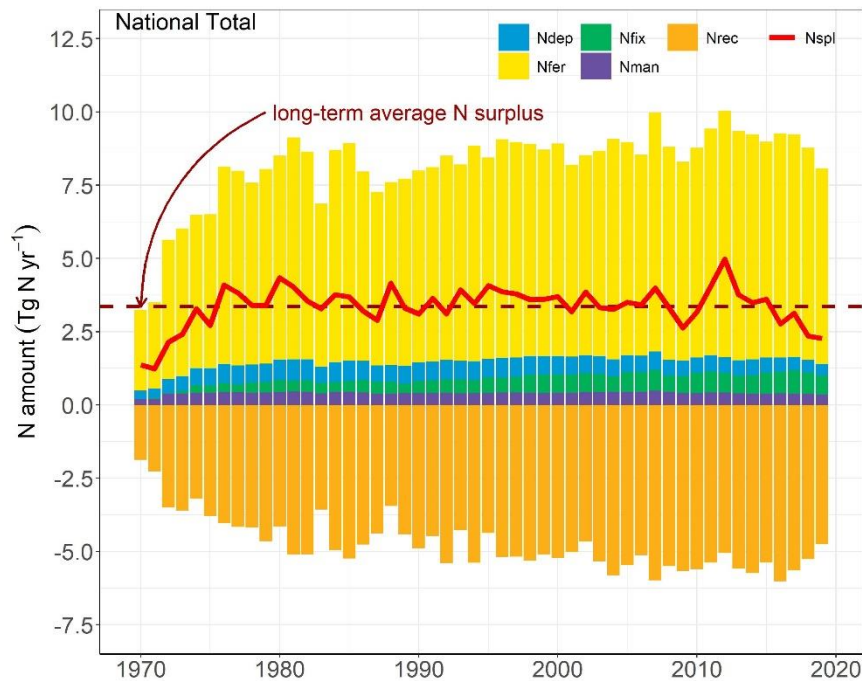
937

938 Figure 4. Decadal average NUE (a-e, unit in  $\text{kg N kg}^{-1} \text{N yr}^{-1}$ ) and N surplus (f-j, unit in  $\text{Gg N yr}^{-1}$ )  
 939 <sup>1</sup>) of corn in each county of the contiguous U.S. Annual planting area in each county has been  
 940 used to calculate the county-level N surplus.

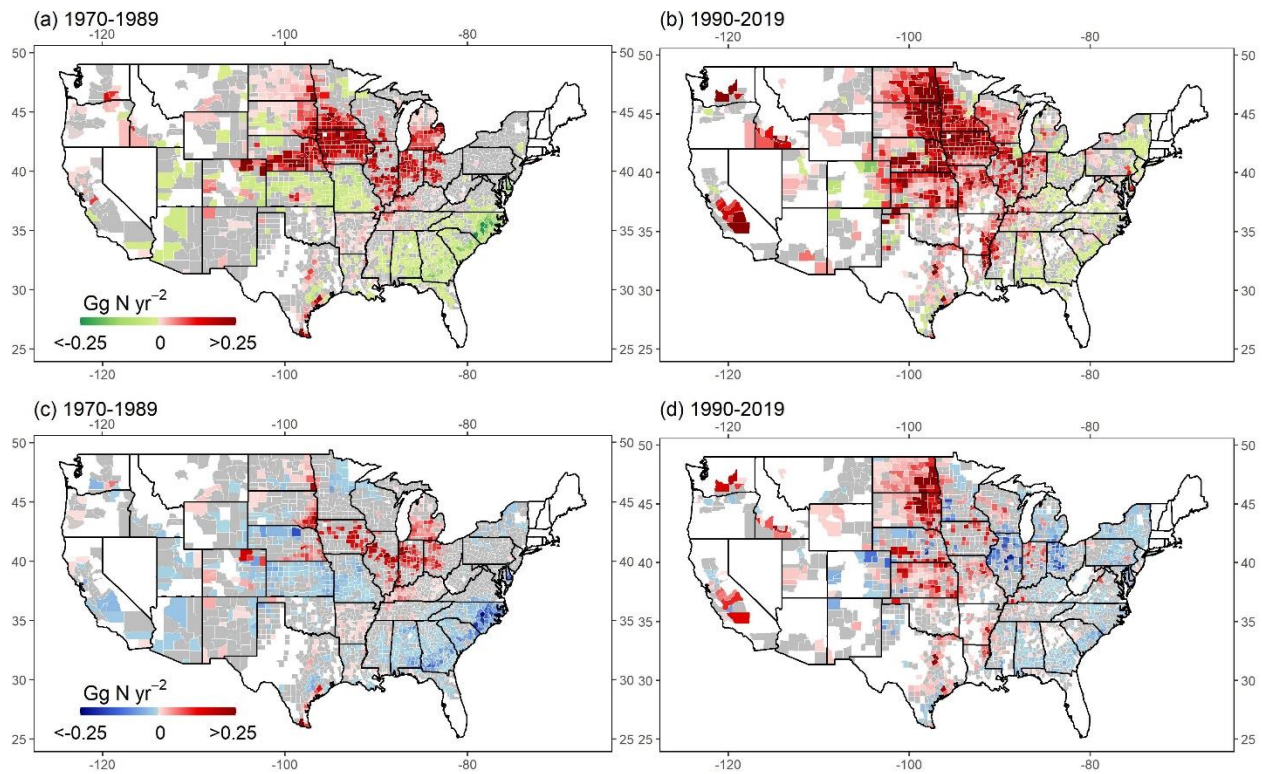


941  
 942 Figure 5. Crop-specific nitrogen budget in the U.S. from 1970 to 2019. N inputs include manure  
 943 (Nman), atmospheric deposition (Ndep), fertilizer (Nfer), and biological N fixation (Nfix), and N  
 944 output includes crop recovered N (Nrec). The time series of nitrogen surplus (Nspl) is shown by

945 the red curve. The data are aggregated from county-level crop-specific N budget. The y axis  
946 scales in corn, durum, and winter wheat differ from other crops.

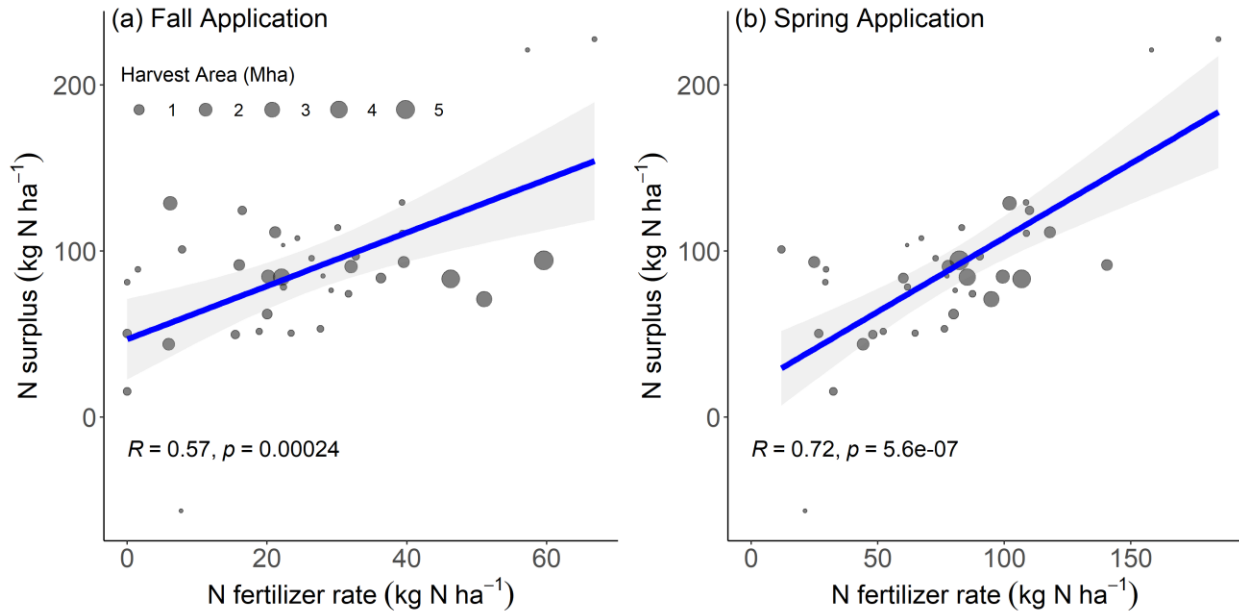


947  
948 Figure 6. Nitrogen budget aggregated from the eight crops in the U.S. from 1970 to 2019. N  
949 inputs include manure (Nman), atmospheric deposition (Ndep), fertilizer (Nfer), and biological N  
950 fixation (Nfix), and N output includes crop recovered N (Nrec). The time series of nitrogen  
951 surplus (Nspl) is shown by the red curve. The dashed line is the long-term average (1970-2019)  
952 N surplus.



953

954 Figure 7. The total nitrogen (N) input trend (a and b, unit in  $\text{Gg N yr}^{-2}$ ) and N surplus trend (c  
 955 and d, unit in  $\text{Gg N yr}^{-2}$ ) of corn in each county of the contiguous U.S. during the periods 1970-  
 956 1989 and 1990-2019. The green colors and the red colors in (a) and (b) represent the  
 957 significantly decreasing and increasing trends ( $p < 0.05$ ) in total N input, respectively. The blue  
 958 colors and the red colors in (c) and (d) represent the significantly decreasing and increasing  
 959 trends ( $p < 0.05$ ) in N surplus, respectively. Grey area indicates no significant trend at 0.05 level  
 960 according to the Mann-Kendall test.



961

962 Figure 8. Relationship between state-level corn annual total nitrogen surplus and N fertilizer  
 963 rates at four application timings from 2010 to 2019. Each dot represents an individual state. The  
 964 letter “*R*” is the Pearson correlation coefficient. Based on the state-level corn-specific N fertilizer  
 965 application timing survey published by USDA-ERS, we extract the N fertilizer use during fall  
 966 (previous year) and spring (before planting). Due to limited data, we adopt the timing  
 967 information from the latest survey for each state. More details can be found in Cao et al. (2018).