

1 **Immunological evidence of variation in exposure and immune response**  
2 **to *Bacillus anthracis* in herbivores of Kruger and Etosha National**  
3 **Parks**

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23 **Keywords:** Anthrax, Adaptive immunity, Enzyme-linked immunosorbent assay (ELISA), *Equus*  
24 *quagga*, passive disease surveillance, Serology, Toxin neutralisation assay (TNA), *Tragelaphus*  
25 *strepsiceros*.

26 **Abstract**

27 Exposure and immunity to generalist pathogens differ among host species and vary across spatial  
28 scales. Anthrax, caused by a multi-host bacterial pathogen, *Bacillus anthracis*, is enzootic in Kruger  
29 National Park (KNP), South Africa and Etosha National Park (ENP), Namibia. These parks share many

30 of the same potential host species, yet the main anthrax host in one (greater kudu (*Tragelaphus*  
31 *strepsiceros*) in KNP and plains zebra (*Equus quagga*) in ENP) is only a minor host in the other. We  
32 investigated species and spatial patterns in anthrax mortalities, *B. anthracis* exposure, and the ability  
33 to neutralise the anthrax lethal toxin to determine if observed host mortality differences between  
34 locations could be attributed to population-level variation in pathogen exposure and/or immune  
35 response. Using serum collected from zebra and kudu in high and low incidence areas of each park  
36 (target of 20 samples/species/area), we estimated pathogen exposure from anti-protective antigen (PA)  
37 antibody response using enzyme-linked immunosorbent assay (ELISA) and lethal toxin neutralisation  
38 with a toxin neutralisation assay (TNA). Serological evidence of pathogen exposure followed mortality  
39 patterns within each system (kudus: 95% positive in KNP versus 40% in ENP; zebras: 83% positive in  
40 ENP versus 63% in KNP). Animals in the high-incidence area of KNP had higher anti-PA responses  
41 than those in the low-incidence area, but there were no significant differences in exposure by area  
42 within ENP. Toxin neutralizing ability was higher for host populations with lower exposure prevalence,  
43 i.e., higher in ENP kudus and KNP zebras than their conspecifics in the other park. These results  
44 indicate that host species differ in their exposure to and adaptive immunity against *B. anthracis* in the  
45 two parks. These patterns may be due to environmental differences such as vegetation, rainfall patterns,  
46 landscape or forage availability between these systems and their interplay with host behaviour  
47 (foraging or other risky behaviours), resulting in differences in exposure frequency and dose, and hence  
48 immune response.

## 49 **1 Introduction**

50 Disease dynamics may be shaped by the spatial structure of host-pathogen encounter rates, and by the  
51 frequency or dose of pathogen exposure affects host susceptibility and immunity to infection (1).  
52 Generalist pathogens can infect multiple host species and differ in their infection intensity or severity  
53 in those hosts, and many previous studies have strived to understand the risk of infection among  
54 different host species (2, 3). There is an abundance of knowledge on how multi-host pathogens evolve  
55 and how host species differ in their susceptibility and immune responses (2, 4-6), both spatially and  
56 within a particular environment, but there is little information on within-species variation in exposure  
57 and immune responses. It is therefore imperative to study within-species differences in exposure and  
58 immunity among populations for a better understanding of both disease progression as well as between  
59 transmission dynamics.

60 Anthrax, an archetypal multi-host disease, is a zoonosis that affects a wide range of species, although  
61 its most susceptible hosts are mammalian herbivores. Anthrax is caused by the gram-positive, capsule-  
62 and endospore-forming *Bacillus anthracis* bacterium. This pathogen must kill its animal host in a bid  
63 to further spread. Disease progression typically occurs either as acute or peracute septicaemia following  
64 incubation of 2-8 days (7). The variation in the incubation period could be due to the size of the  
65 infectious dose encountered and/or the exposure intervals (7-9). After the death of the host, blood oozes  
66 from the body orifices, exposing vegetative cells to oxygen, which triggers sporulation. The resulting  
67 endospores can survive in the soil for years until uptake (normally ingestion) by another susceptible  
68 host, within which the spore crosses the epithelium and can germinate forming vegetative cells. This  
69 germination followed by further propagation and an increase in cells producing toxins (10, 11) that  
70 ultimately leads to the death of the host (12). Due to the acute and peracute nature of anthrax, diagnosis  
71 is mainly based on detection of the pathogen post-mortem through molecular identification,  
72 microscopy and culture (19-21). The detection of specific antibodies in serum from live animals can,  
73 however, provide information on previous exposure to the pathogen.

74 For the development of immunity against anthrax, the host must be able to resist the establishment of  
75 disease or stall its progression (13). The virulence factors of *B. anthracis* are encoded on two plasmids  
76 namely pXO1, which is responsible for the production of the toxins, and pXO2, which codes for the  
77 poly-γ-D-glutamic acid capsule that helps the pathogen avoid detection by the host immune system  
78 (14, 15). The pXO1 plasmid encodes for the cell-binding protein protective antigen (PA), and two  
79 enzymes, the lethal factor (LF) and the oedema factor (EF) proteins. PA can combine with either LF  
80 or EF to form lethal toxin (LT) or oedema toxin (ET) respectively, which are responsible for the  
81 deleterious effects of *B. anthracis* (12, 16-18). These anthrax toxins can facilitate the establishment  
82 of infection and lead to host mortality (19). They also contribute to early and late-stage infection (19),  
83 thus, disrupting the effects of the toxin through toxin neutralisation which can both prevent the  
84 establishment or stall disease progression, promoting host survival.

85 Development of specific antibodies to PA, LF and EF proteins have been demonstrated using enzyme-  
86 linked immunosorbent assay (ELISA) following natural or experimental infection (20, 24-27). Toxin  
87 neutralizing antibodies also play an important role in conferring protection against anthrax in the host  
88 (20, 21). The toxin neutralization assay (TNA) is used to measure the capability of host serum to  
89 neutralise the cytotoxic effects of LT and ET on cells in vitro (20). The TNA quantifies only the  
90 functional subunit of the antibodies rather than the total anti-PA IgG antibodies detected by ELISA  
91 (20).

92 Antibody titres to *B. anthracis* diminish over time as reported in plains zebras (*Equus quagga*) that  
93 are naturally exposed, however it seems that frequent sublethal infections can boost antibody levels to  
94 maintain a detectable level of antibodies (22). The presence of neutralising antibodies against anthrax  
95 lethal toxin has been reported in vaccine studies, with neutralising antibodies positively correlated with  
96 anti-PA titres and increased survival rates (20, 23-25). Species differences in susceptibility to infection  
97 with anthrax have been reported (26). Some species like herbivores are highly susceptible, while  
98 carnivores and omnivores appear to be more resistant (27, 28). On the contrary, species that are resistant  
99 to spore challenge appear to be highly susceptible to intravenous toxin challenge and vice versa (29).  
100 However, no study has been conducted in free-living wild herbivores to see how toxin neutralisation  
101 ability varies across species or between areas of higher or lower risk of anthrax exposure.

102 The *Bacillus anthracis* lifecycle involves animal hosts, the external environment and potential  
103 mechanical vectors such as flies (30-33), vultures (e.g., *Gyps africanus*) (34-37), jackals (*Lupulella*  
104 *spp.*) and hyenas (*Crocuta crocuta*) (36, 38). Environmental factors influencing disease dynamics  
105 include soil properties such as calcium and pH, and weather factors such as rainfall, humidity, and  
106 temperature (39-43). Anthrax is endemic to Kruger National Park (KNP) in South Africa and Etosha  
107 National Park (ENP) in Namibia. Southern Africa, including KNP and ENP, is considered the origin  
108 of anthrax (44). These two parks vary in anthrax incidence, with high and low incidence areas  
109 documented. Anthrax primarily affects grazing herbivores in ENP with plains zebra contributing to  
110 most of the mortalities (45), while in KNP, the primary host species over time has been greater kudu  
111 (*Tragelaphus strepsiceros*), a browsing herbivore. In ENP, browsers such as kudu account for about  
112 1.7% of anthrax mortalities (46). In recent years in KNP, the seasonal timing and primary host species  
113 has shifted, to primarily wet season outbreaks affecting impala (*Aepyceros melampus*), a mixed  
114 grazing-browsing species (47).

115 The variation in anthrax ecology worldwide has served as an impediment for the blending of  
116 knowledge and outbreak forecasting (48) and therefore, identifying the variables that play a role in  
117 disease dynamics warrants substantial attention. Comparing two natural systems allows us to study the  
118 differences, patterns and pathways that may be unnoticed under the limited lens of a single system (49).  
119 In addition, comparing systems that differ in disease dynamics, but share the same potential host  
120 species, allows us to “control” for the large differences in ecology, behaviour and immunity between

121 different species, while exploring how exposure and immune response vary among populations of the  
122 same species. Before now, no research has been conducted to measure and compare the variability in  
123 *B. anthracis* exposure status or protection levels across different species and areas.

124 We investigated the variation in immune status among plains zebra and greater kudu in two different  
125 ecosystems (ENP, KNP) with different anthrax epidemiology. Specifically, we addressed the following  
126 questions: 1) Are serological patterns of host exposure to the anthrax bacterium concordant with spatial  
127 patterns of anthrax mortality from passive surveillance? 2) Does toxin neutralisation ability vary based  
128 on species and/or environmental factors, such as frequency or dose of pathogen exposure? If this toxin  
129 neutralisation is a species-level trait, then we would expect variation in the ability to tolerate or resist  
130 the effects of anthrax disease to be part of why species vary in their susceptibility to anthrax mortality,  
131 and that this ability would be consistent across study areas. However, if toxin neutralisation varies  
132 based on pathogen exposure, then we expect to observe differences in neutralisation ability for  
133 populations occurring in high or low anthrax incidence areas, where frequency of pathogen encounters  
134 by animals may vary, respectively. This study, therefore, investigated the immunological dynamics of  
135 anthrax infection in two national parks with a goal of understanding whether the rarity of disease  
136 mortality in an area is a function of low or no exposure or higher adaptive immune response. We  
137 examine the prevalence of exposure to the pathogen—as an index of exposure frequency—across host  
138 species and locations and evaluate how exposure relates to the ability of the host to mount an effective  
139 adaptive immune response, through the ability of hosts to neutralise the anthrax lethal toxin.

## 140 2 Materials and Methods

### 141 2.1 Study areas

142 This study compared serological evidence of *B. anthracis* exposure in host species in two large national  
143 parks. Etosha National Park (ENP; 22,915 km<sup>2</sup>), Namibia, and Kruger National Park (KNP; 19,485  
144 km<sup>2</sup>), South Africa, are located nearly 2,000 km apart in southern Africa (Figure 1), a region considered  
145 the origin of anthrax (50). The anthrax endemic regions of these ecosystems are classified as arid  
146 savannas, based on annual rainfall less than 650 mm (51). Central ENP has an average rainfall of  
147 358mm (Okaukuejo weather station 1954-2020; 19.1669° S, 15.9171° E), mostly an open shrubveld  
148 around a large salt pan. On the other hand, northern KNP is highly woody with grassland savannah  
149 (51), with an average rainfall of 430mm. ENP is largely flat with some mountains in the far western  
150 part of the park while KNP has varying elevations with Pafuri (found in the far northern part of KNP;  
151 22.4206° S, 31.2296° E) having lower elevation flood-plains surrounded by higher elevations. In both  
152 parks, there are areas of high and low anthrax incidence (defined here as regular or infrequent anthrax  
153 occurrence over time, respectively). In KNP the high incidence area extends from Pafuri to Shingwedzi  
154 (23.1167° S, 31.4333° E) in the north and the low incidence area extends from Skukuza (24.9948° S,  
155 31.5969° E) to Crocodile Bridge (25.3584° S, 31.8935° E) in the south. The high incidence area in  
156 ENP includes the central Okaukuejo management unit and the low incidence area include the western  
157 Otjovasandu (19.2300° S, 14.4800° E) management unit. These regions of low and high incidence  
158 were determined based on previous reports (36, 44) and the distribution of anthrax mortalities from  
159 historical data. Our study focused on plains zebra and greater kudu, sampled in high and low incidence  
160 areas of each park. For comparison, we included samples from a secondary anthrax host species in the  
161 high incidence area of each park: blue wildebeest (*Connochaetes taurinus*) in ENP and impala in KNP.

162 Anthrax primarily affects grazing and mixed-feeding herbivores. In the high incidence region of ENP  
163 (Figure 1), deaths of plains zebra and other herbivores climax at the closing of the rainy season, while  
164 African elephant (*Loxodonta africana*) deaths climax during the late dry season, though cases in all

165 species can be observed sporadically throughout the year (36, 45, 52, 53). Seasonal outbreaks have  
166 been linked to differences in host foraging behaviour altering exposure rates (22, 45, 46) and seasonal  
167 immune trade-offs (54). Zebra and wildebeest are grazing herbivores, kudu are browsing herbivores,  
168 and impala and elephant are mixed-feeding herbivores, which graze or browse depending on  
169 conditions. Plains zebras contributed to most of the mortalities in ENP followed by blue wildebeest  
170 (45). browsers, which include kudu, contributed the least (46).

171 In KNP, the main host species over time has been greater kudu, a browser contributing up to 75% of  
172 recorded cases from 1960-1990s (55). Anthrax was historically associated with dry seasons or droughts  
173 in KNP, occurring in explosive outbreaks on a roughly decadal cycle (30, 37, 55-57). Since 2008,  
174 smaller outbreaks have occurred annually and mainly in the wet season, and primarily affecting impala,  
175 a mixed grazing-browsing species (47). Exposure of browsing species has been hypothesized to occur  
176 via blowflies (*Chrysomya* spp.) feeding on anthrax carcasses, and then depositing *B. anthracis* spores  
177 onto the leaves of trees/shrubs near the carcass (58, 59). Plains zebra have contributed only 4% (44/962)  
178 of cases in KNP outbreaks (anthrax mortality reports from 1988-2016 obtained from the Skukuza  
179 Veterinary Services).

## 180 2.2 Sample collection

181 Serum samples were obtained from live animal captures from the four study species. We sampled 20  
182 individuals per primary host species (zebra, kudu) per area, except for kudu in KNP (low incidence  
183 = 18, high incidence = 19). Twenty individuals per secondary host species were sampled only in high  
184 incidence areas of the parks where they occur (northern KNP: impala, n=20; central ENP: wildebeest,  
185 n=20). Negative and positive control serum samples were obtained by vaccinating two representative  
186 animals of each species (kudu, impala, zebra and wildebeest) in southern KNP. These animals were  
187 fitted with a satellite-GPS collar, sampled initially for the negative control, vaccinated with the Sterne  
188 Live Spore vaccine (Onderstepoort Biological Products, South Africa), and released. Each animal  
189 was vaccinated with 1 ml of Sterne spore vaccine intramuscularly as prescribed by the manufacturer.  
190 These animals were then recaptured after a month and serum samples were collected, which served  
191 as the positive controls. All ethical approvals were obtained from the University of Pretoria Research  
192 Ethics Committee, Animal Ethics Committee and the Department of Agriculture, Forestry and  
193 Fisheries (DAFF) in South Africa. Animals were immobilized following the “standard operating  
194 procedures (section 2.1.11) for the capture, transportation and maintenance in holding facilities of  
195 wildlife” by certified veterinarians and South African National Parks regulations. Also, approval was  
196 obtained from the University at Albany’s International Animal Care and Use Committee, approval  
197 numbers: 16-016, 18-013, 18-014, 18-015, 20-001. Permission to conduct research was obtained from  
198 the Namibian National Commission on Research, Science and Technology (authorization  
199 2017070704) and the Ministry of Environment, Forestry and Tourism, Namibia.

## 200 2.3 Mortality data

201 Mortality data were analysed to examine the distribution of *B. anthracis* positive cases and the  
202 distribution of mortality detection and reporting in each park (Table S1). These data were collected  
203 as part of the opportunistic passive mortality surveillance in these parks. The data for KNP ranged  
204 from 1990-2015 and for ENP from 1996-2015, after restricting the time series to cases with GPS  
205 coordinates. These data, however, excluded a substantial number of kudu anthrax mortalities from  
206 when kudu dominated the outbreak cases. Carcasses were identified as anthrax positive following a  
207 positive result from blood smear examination, bacterial isolation or molecular detection (11, 47).  
208 Other information obtained included the date, locality, species and sex. For analysis, each park was  
209 grouped into three regions: for KNP, these are the northern, central and southern regions while for

210 ENP, these are western, central and eastern regions as defined by the park management (Figure 1).  
211 The mortality data were grouped into two causes of death: anthrax or others (e.g., predation,  
212 unknown). Anthrax important species for this analysis include zebra, impala, kudu and wildebeest  
213 (wildebeest was excluded from KNP and impala from ENP as they did not contribute significantly),  
214 all other species both for anthrax mortality and other causes of death were categorized as “others”  
215 (for KNP other included mortalities from 57 different species, of which 21 species had anthrax  
216 mortalities and for ENP included mortalities from 27 species of which 6 species had anthrax  
217 mortalities). The mortality data were further used to confirm and distinguish between the high and  
218 low incidence areas of these parks.

#### 219 **2.4 Anti-protective antigen (PA) enzyme-linked immunosorbent assay**

220 In this study serum samples were assessed for the presence of specific antibodies against the anthrax  
221 PA as described by Yu, *et al.*, (60) and Ndumnego, *et al.*, (25). Briefly, microtiter plates (Thermo  
222 Scientific™ Pierce 96-well Plates-Corner, USA) were coated overnight with 0.5 µg/ml rPA (List  
223 Biological Laboratories Inc., USA) in bicarbonate buffer at 4°C. Plates were washed twice with  
224 Phosphate Buffered Saline (PBS) supplemented with 0.05% Tween-20 (Thermo Fisher Scientific,  
225 Waltham, MA USA) (PBST) using a Biorad PW40 washer (Mamesla-Coquette, France). Plates were  
226 blocked with PBST supplemented with 5% skimmed milk powder (PBSTM) and then incubated for  
227 1 h at room temperature. Plates were washed twice before the addition of duplicate test and control  
228 sera at a 1:40 dilution in PBSTM. This was followed by 30 min incubation on a rotatory incubator  
229 (Environmental Shaker-Incubator ES-20, Biosan Ltd, Germany). Afterwards, the plates were washed  
230 five times and recombinant protein A/G horseradish peroxidase (HRPO) conjugate (Pierce® Protein  
231 A/G, USA) for zebra and wildebeest (61) and protein G HRPO conjugate (Invitrogen Protein G, USA)  
232 for impala and kudu were added to respective wells and incubated for 30 min on the rotary incubator.  
233 The binding of protein G HRPO to impala and kudu was evaluated in Supplementary methodology  
234 Figure S1 and Table S2. The plates were washed five times, after which the substrate 2,2'-Azinobis[3-  
235 ethylbenzothiazoline-6-sulfonic acid]-diammonium salt (ABTS) (Thermo Scientific 1-Step ABTS,  
236 USA) was added and incubated in the dark for 45 min. The absorbance was read at 405 nm using the  
237 Biotek Powerwave XS2 reader (USA). The ELISA results were interpreted as binominal data  
238 (positive/negative) with the threshold set at the mean plus three standard deviations (SD) of the  
239 negative control for the respective species. The optical density (OD) values were normalised per  
240 species to reduce variations between plates. Normalisation between plates was done by calculating  
241 sample to positive (SP) ratios as the same positive control (for each species) was used on each of the  
242 plates. The binary outcome (positive/negative) was used to determine exposure while the SP ratios  
243 were used as a measure of the antibody response (25).

#### 244 **2.5 Toxin neutralisation assay (TNA)**

245 The TNA was used to estimate the variation of anthrax LT neutralising antibody amongst the different  
246 species in the two parks. The assay measures the ability of test sera to protect mouse macrophages  
247 from the cytotoxic effects of the toxin and is therefore not species-specific (20, 62).

248 The TNA was performed *in vitro* using J774A.1 mouse macrophage cell line (ECACC cat no  
249 91051511), with modifications as described by Hering, *et al.*, (23) and Ndumnego, *et al.*, (63). Flat-  
250 bottomed 96-well culture plates (Corning™, Corning incorporated, Germany) were seeded with 10<sup>5</sup>  
251 mouse macrophage cells in 200 µL Dulbecco's modified eagle media supplemented with 10% foetal  
252 bovine serum (TNA medium), and incubated at 37°C and 5% CO<sub>2</sub> for 24 h. Duplicate test sera were  
253 diluted two-fold (1:50 to 1:6400) in TNA medium containing 500 ng/mL PA and 400 ng/mL LF (List



254 Biological Laboratories Inc., USA). The sera and toxin were incubated for 1 h at 37°C and then  
 255 transferred to the previously seeded cells and incubated for 3 h. Each plate also included 3 wells without  
 256 cells as blanks, 3 wells for the toxin control and 2 wells for media control (used to calculate the  
 257 neutralisation titre). Each plate also contained a single dilution for the positive controls (to ensure  
 258 consistency and reproducibility of the assay) for each animal species. Twenty-five µL of 3-  
 259 (4,5dimethylthiazol-2-yl)-2,5-diphenyltetrazolium bromide (Invitrogen, USA) was added to every well  
 260 and incubated at 37°C and 5% CO<sub>2</sub> for 2 h. The cells were lysed using a 100 µL mixture of 90%  
 261 isopropyl alcohol, 0.5% sodium dodecyl sulphate (SDS), and 25 mM Hydrochloric acid (HCl) followed  
 262 by a 5 min incubation at room temperature.

263 The plates were read at an absorbance of 570 nm and the neutralisation titres (NT) were calculated as:

$$264 \quad NT = \frac{OD_{Sample} - OD_{Toxin\ control}}{OD_{Medium\ control} - OD_{Toxin\ control}} \times 100$$

265 The neutralisation titre 50 (NT<sub>50</sub>) was calculated, using the Gen5 analysis software (Biotek  
 266 Instruments, USA), as the highest titre that protected 50% of the macrophage cells. Samples that  
 267 could not protect 50% of the cells were assigned an arbitrary value of 0.1

## 268 2.6 Statistical analyses

269 Distributional patterns for total mortalities and anthrax mortalities were described for both parks to  
 270 evaluate how serological results match with what is known about anthrax mortalities, based on disease  
 271 surveillance. Anthrax mortalities for impala and wildebeest were only recorded in KNP and ENP,  
 272 respectively. Mortality data from each park were plotted in ArcGIS pro version 2.8 and summarized  
 273 as bar plots and maps.

274 We determined how host species differed in their immune responses (based on SP ratios) to *B.*  
 275 *anthracis* between the two parks and between high incidence and low incidence areas using  
 276 multivariable linear models coupled with the Tukey's Honestly Significance Difference (HSD) test for  
 277 multiple mean comparisons. Analyses were done separately for each species, and the SP ratios were  
 278 log-transformed to normalise the data. The predictor variables included national park (KNP or ENP),  
 279 area (high incidence or low incidence), LT neutralisation status (positive, negative) and the interaction  
 280 between national park and area. To compare exposure, we used logistic regression analysis with host  
 281 exposure (positive or negative for anti-PA antibodies) as the response variable and park, area, and  
 282 interaction between park and area as categorical independent variables.

283 To determine how the host species differed in their toxin neutralising ability, a multivariable linear  
 284 model with the Tukey's HSD test for multiple mean comparisons was performed to evaluate whether  
 285 national park (KNP, ENP), area (high or low incidence), host species (kudu, zebra), and level of anti-  
 286 PA immune response (ELISA ODs), significantly predicted LT neutralisation titres (NT<sub>50</sub>). Only TNA  
 287 positive animals were included in the analysis, and NT<sub>50</sub> and ELISA SP ratios were log-transformed  
 288 first to normalise the data. To determine the difference in proportions of animals that neutralised the  
 289 LT, logistic regression analysis was conducted to identify significant predictors for *B. anthracis* toxin  
 290 neutralisation ability (positive/negative status determined by TNA) in wild animal populations in ENP  
 291 and KNP. Wildebeest and impala were not included in the regression analyses because these were  
 292 sampled only from high incidence areas in ENP and KNP, respectively, but descriptive analyses for  
 293 these species were performed.

294 The extent of agreement between the binary outcomes of anti-PA ELISA and TNA results separately  
 295 for individual species (kudu=77, zebra=80, wildebeest=20 and impala=20) was determined using

296 Spearman's correlation and Cohen's kappa ( $k$ ) test (64). For this analysis,  $kappa \neq 0$ , means that the  
297 agreement between anti-PA ELISA and TNA is different from chance agreement. The strength of  
298 agreement was assessed based on the criterion by Landis, *et al.*, (64), where  $<0$  = poor; 0.01-0.20 =  
299 slight; 0.21-0.40 = fair; 0.41-0.60 = moderate; 0.61-0.80 = substantial; 0.81-1.00 = almost perfect.

300 All statistical analyses were done in R Console version 3.2.1 (65) with significance assessed at a 5%  
301 level.

### 302 **3 Results**

#### 303 **3.1 Mortality Distributions**

304 In ENP, the highest number of all mortalities (76%) were recorded in the central region, followed by  
305 14.8% in the eastern region and 8.5% in the western region. In the central region, zebra contributed  
306 54.6% ( $N = 618$ ) of the total mortality, while wildebeest and kudu contributed 8.13% ( $N = 92$ ) and  
307 0.26% ( $N = 3$ ), respectively (Figure 2). Zebra had the highest total mortality in the western and eastern  
308 regions (11.1% and 37%, respectively), followed by kudu in the west (7.9%) and wildebeest in the east  
309 (11.4%).

310 Of the anthrax mortalities observed in ENP, the highest number was recorded in the central region  
311 (90.4%), followed by 7.4% in the eastern and 2.1% in the western regions. Considering anthrax  
312 mortalities by species in ENP, the contribution to mortality for zebra was 68.7%, which was higher  
313 than wildebeest (11.7%), and kudu (0.28%). In the central region, zebra similarly contributed 72%,  
314 followed by wildebeest (9%), with no anthrax mortality recorded for kudu. Of the anthrax mortality in  
315 the east and the west, zebra had 32.7% and 53.3%, respectively, kudu 1.9% and 6.7%, respectively,  
316 and wildebeest had 46.2% in the east (Figure 2).

317 In KNP, the highest number of mortalities (88.4%) were recorded in the north, with most mortalities  
318 clustered around the Pafuri region, followed by 8.5% in the central region and 3.1% in the southern  
319 region. Mortalities in the northern region among the species of interest were dominated by impala  
320 (27.3%), followed by kudu (22.6%) and zebra (5.6%). For the central and southern part of KNP, impala  
321 contributed 18.8% and 43.2%, kudu, 7.5% and 10.8%, and zebra 6.3% and 2.7% to the total mortality  
322 respectively (Figure 2). Of the total anthrax mortality in KNP, kudu contributed 35%, followed by  
323 impala (21.8%) and zebra only contributed 2.9% of cases. Kudu made up 35.6% of the anthrax-related  
324 mortalities in the northern region, followed by impala (22.3%) and zebra with 2.9%. In central KNP,  
325 impala contributed 1.9%, while kudu contributed 30.2% to the carcasses that were anthrax positive  
326 (Figure 2).

327 These patterns confirm our expectations that zebra in ENP and kudu in KNP are the primary host  
328 species these systems, and that they are minor hosts in the opposite park (i.e., zebra in KNP and kudu  
329 in ENP). The distribution of anthrax mortalities revealed that the central part of ENP (90.5%) and  
330 northern part of KNP (98.2%) are the most affected over the years, followed by the eastern part of ENP  
331 (7.4%) and the western part of ENP (2.1%) and central region (1.8%) in KNP, with no positive cases  
332 observed in the southern region of KNP (Figure 3).

333

#### 334 **3.2 Exposure to *B. anthracis***



335 Host species showed significant differences in exposure to *B. anthracis* between parks, based on anti-  
 336 PA antibody response. Kudus in KNP had significantly higher exposure to *B. anthracis* than kudus  
 337 in ENP ( $p = 0.005$ , Table 1 and Figure 4a). The kudus in KNP had significantly higher odds of  
 338 exposure to *B. anthracis* than those in ENP (odds ratios (OR) = 2.9,  $p = 0.005$ ). Zebra in ENP also  
 339 had higher odds of exposure to *B. anthracis*, with a higher proportion of anti-PA ELISA positives,  
 340 than those in KNP (OR = 3.1,  $p = 0.005$ , Table 1 and Figure 4c). Details for all four species are shown  
 341 in Table 1 and Figure 4a.

342 Host species had higher antibody response (based on SP values) in the park where they were  
 343 considered the primary anthrax host, but lower response where they were the secondary anthrax host.  
 344 Kudus in KNP had significantly higher (1.3 times,  $p = 0.047$ ) anti-PA ELISA response ( $1.24 \pm 0.74$ )  
 345 than those in ENP ( $0.65 \pm 0.46$ ), but exposure in the high versus low incidence areas (irrespective of  
 346 park) were not statistically different ( $p = 0.41$ , Table S3). Zebras in ENP had significantly higher (1.3  
 347 times,  $p = 0.034$ ) anti-PA ELISA response ( $0.69 \pm 0.53$ ) than those in KNP ( $0.53 \pm 0.40$ ), while  
 348 differences between the high versus low incidence areas (irrespective of park,  $p = 0.29$ ) were  
 349 statistically insignificant. The interaction between national park and area contributed significantly ( $p$   
 350 = 0.015) to the level of immune response in this study. When SPs were compared between incidence  
 351 areas within parks separately for each species, there was a significant difference for kudu (Table 1  
 352 and Figure 4b) and zebra (Table 1 and Figure 4c) in KNP, but not in ENP. In the high incidence areas,  
 353 the average anti-PA SPs for KNP impala and ENP wildebeest were  $0.52 \pm 0.23$  and  $0.48 \pm 0.19$ ,  
 354 respectively (Table 1).

### 355 3.3 Neutralisation of anthrax lethal toxin

356 The distribution of hosts (kudu and zebra) by park had a significant influence on the serum-LT  
 357 neutralisation titres (Table S4). Kudus in both parks and zebras in KNP all showed significantly  
 358 higher odds of *B. anthracis* toxin neutralisation (> 45%) than zebras in ENP (10%, 4/40;  $p = 0.001$ )  
 359 (Figure 5). The ability to neutralise the toxin pooled for all species across parks did not significantly  
 360 differ by area (high incidence area = 51.9% (41/79); low incidence = 46.2% (36/78),  $p = 0.47$ ). Further  
 361 analyses of the association of toxin neutralisation proportion and park (irrespective of incidence  
 362 status) showed a significantly higher proportion of zebra neutralizing the anthrax LT in KNP than in  
 363 ENP ( $p = 0.001$ ; Figure 5a). In contrast, a higher proportion of kudu in ENP was able to neutralise  
 364 the anthrax toxin than kudu in KNP, although the difference was not statistically significant ( $p = 0.15$ ;  
 365 Figure 5). Only 3/20 impala showed toxin neutralisation, while 9/20 wildebeest neutralised the toxin  
 366 (Figure 5a, Table 2).

367 When considering only TNA-positive animals (animals that showed neutralisation), kudus in ENP  
 368 also had significantly higher titres ( $110.0 \pm 84.6$ ) than those in KNP ( $64.2 \pm 56.3$ ;  $p = 0.03$ , Table 2  
 369 and Figure 5b). For zebra,  $NT_{50}$  were higher in KNP ( $75.3 \pm 45.9$ ) than ENP ( $38.3 \pm 10.947 \pm$ )  
 370 (regardless of area,  $p = 0.05$ ). The titres of the two host species were also compared within the national  
 371 parks, pooling across low and high incidence areas, which revealed that there was not a statistically  
 372 significant difference in neutralizing titres between kudu and zebra in KNP (64.2 vs. 75.3 ;  $p = 0.072$ );  
 373 in contrast, titres significantly differed between the two species in ENP (110.0 vs. 38.3, respectively;  
 374  $p = 0.03$ ) (Table 2 and Figure 5b). Details on  $NT_{50}$  levels for impala and wildebeest can be found in  
 375 Table 2 and were lower than zebra and kudu in all locations.

### 376 3.4 Relationship between pathogen exposure and toxin neutralisation

377 Kudu showed a statistically significant and moderate agreement between anti-PA and TNA (kappa =  
 378 0.47, 95% CI:0.28-0.66,  $p = 0.0001$ ). There was a slight agreement between these measures for zebra

379 (kappa = 0.096, 95% CI:0.089-0.19), but this was not significant ( $p = 0.213$ ). For wildebeest there  
380 was a fair agreement (kappa = 0.381, 95% CI: -0.02-0.78,  $p = 0.081$ ) and for impala no agreement  
381 (kappa = -0.195, 95% CI: -0.42-0.036,  $p = 0.253$ ), respectively, and neither species showed statistical  
382 significance (Table 3).

383 There was a medium and significant positive correlation between anti-PA titres and TNA values,  
384 using Spearman's correlation ( $\rho = 0.40$ ,  $p = 0.001$ ). A correlation in kudu in both parks and zebra  
385 in KNP (Figure 6) provided evidence for saturation in TNA values as SP values increased.

386 4 **Discussion**

387 In this study, we examined anthrax PA-specific and anthrax LT neutralising antibodies to compare  
388 immune exposure and response to *B. anthracis* in four wildlife species. This study reveals a wide  
389 presence of anti-PA antibodies in the various host species sampled. It was seen that the spatial patterns  
390 of anthrax mortality from passive surveillance from both parks reflect the serological patterns of  
391 exposure to *B. anthracis*. Interestingly, even though these parks share similar host species, there were  
392 significant differences in the proportions of animals that tested positive for anti-PA antibodies and  
393 level of antibody response (SP) between the two parks. Also, we noted that toxin neutralising ability  
394 is not necessarily a trait of species, but a product of environmental factors and exposure, which  
395 include access to the pathogen, frequency of exposure and/or the dose of exposure. This study also  
396 represents the first report of neutralising titres in wild herbivores.

397 **4.1 Spatial patterns in *B. anthracis* exposure and anthrax mortality**

398 Spatial patterns in *B. anthracis* exposure agreed with the anthrax mortality patterns in both parks. The  
399 mortality data from each park showed zebra and wildebeest in ENP and kudu and impala in KNP as  
400 the most affected species in their respective parks. This is in accordance with previous reports from  
401 both parks that show that these species have the highest anthrax mortality in these parks (45, 46, 55).  
402 Most of the anthrax mortalities in KNP were from the northern part of the park, agreeing with the  
403 high incidence status previously attributed to this region of the park (37). Fewer mortalities were  
404 found in the central region, but no positive anthrax cases were found in the southern part of KNP.  
405 This result strengthens the divide between the high incidence and low incidence areas of KNP.  
406 However, we found that >50% PA positive animals were reported in the low incidence area. The  
407 absence of anthrax mortality in the southern part of KNP could result from sampling bias, as a  
408 relatively low proportion of overall mortalities were from this part of the park. Also, it has been  
409 reported that relying on carcass discovery or passive surveillance might not give the true picture of  
410 exposure in a population (66). For ENP, most of the anthrax mortalities were found in the central part  
411 of the park, with very few cases found in the east and the west. Unlike in KNP, the western (low  
412 incidence) part of ENP had some anthrax cases, which suggests possible exposure in this part of the  
413 park as supported by the moderately high prevalence of anti-PA antibodies found in both zebra and  
414 kudu (50-60%) in this region.

415 The anti-PA antibodies reported in this study indicated that the animals in these parks are exposed to  
416 varying doses of *B. anthracis* spores and/or repeated exposures in the environment and can mount an  
417 effective adaptive immune response. These results build on existing evidence that herbivores exposed  
418 to sublethal doses of *B. anthracis* in the environment develop antibodies against the pathogen (22).  
419 Moreover, this claim contradicts previous studies suggesting herbivores in anthrax high incidence  
420 regions are susceptible and naïve to *B. anthracis* and die following severe and sudden exposure. These  
421 assumptions of previous studies were based on a lack of detectable anti-PA titres (27, 28). However,  
422 the current study and previous studies each used a different serological method, namely indirect anti-  
423 PA ELISA (this study), QuickELISA kit (Anthrax-PA kit, Immunetics, Incorporated, USA) (27) and  
424 competitive indirect anti-PA ELISA (28), which could account for the different results. The  
425 competitive indirect anti-PA ELISA, unlike the indirect anti-PA ELISA, requires a high quantity of  
426 antibodies for there to be a 0.2 OD difference between two consecutive dilutions due to the inhibited  
427 counterpart and are thus less sensitive than the latter (22). The Quick ELISA kit also lacks the  
428 sensitivity to detect animals with low antibody titres (22). The indirect ELISA used in our study is  
429 not without its limitations. The conjugate will only optimally bind for specific species for which they  
430 were developed and for closely related species (67, 68). In this study, protein A/G conjugate was used  
431 for zebra and wildebeest while protein G was used for kudu and impala, which were selected based

432 on a preliminary study (Figure S1). These differences in binding specificities make it unsuitable to  
433 compare antibody titres between species, but comparisons between locations within a species remain  
434 robust. There are varying reports of the binding ability of the commercially available conjugates in  
435 these wildlife species (67-69) and therefore species-specific conjugates to overcome this limitation  
436 are needed.

437 Sublethal exposure, and how frequently hosts encounter the pathogen, may have impacts on host  
438 immunity and disease dynamics (22, 26, 70). Kudu in the two parks showed a relatively high  
439 prevalence of pathogen exposure (65% in ENP and 84% in KNP), yet unlike KNP, kudu anthrax  
440 mortality in ENP is rarely observed. Thus, kudu in ENP may be commonly exposed to the pathogen,  
441 but in lower doses unlikely to cause mortality. Also, it has been shown in a previous study that an  
442 animal host may ingest a high number of spores that pass through the digestive tract without any  
443 invasion or that cause a sublethal infection (71). Our study reported kudu in KNP are significantly  
444 more likely to be exposed to the pathogen than their counterparts in ENP and make up about 75% of  
445 historical anthrax cases (55) and 35.6% of the recorded cases from 1990 in KNP. In contrast, kudu in  
446 ENP contribute only 0.28% of recorded cases (Figure 2) in this study and this was reflected in the  
447 anti-PA antibody prevalence.

448 Both kudu and zebra in the two parks had antibodies against *B. anthracis* PA, though differences in  
449 antibody prevalence corroborate species and regional differences in anthrax incidence. Anthrax  
450 outbreaks in kudu in KNP have been linked to dissemination by blowflies in the park (47, 72). Hugh-  
451 Jones, *et al.*, (37) indicated that *Chrysomya spp.* blowflies feeding on anthrax carcasses in KNP  
452 deposit *B. anthracis* spores onto the leaves of trees/shrubs near the carcass at the height that kudu  
453 feed, thereby creating a higher inoculum and exposure for the kudu in KNP. The increase in *B.*  
454 *anthracis* inoculum by the *Chrysomya* flies on shrubs eaten by browsers in KNP might cause the  
455 higher mortality rates reported for browsers in KNP compared to ENP. A blowfly transmission  
456 pathway has not been detected in ENP. While Nalisa (73) recorded the presence of *B. anthracis* in  
457 flies of the *Muscidae* and *Calliphoridae* families, these flies were observed in relative low density at  
458 carcass sites in ENP. However, because of high vertebrate scavenger activity in ENP, most carcasses  
459 are consumed before flies can reproduce (38). This suggests that kudus in ENP can be exposed to the  
460 pathogen, but possibly to a lesser extent due to a smaller amount of dissemination and bacterial  
461 inoculum, through other mechanical vectors depositing the spores onto the leaves of trees or shrubs  
462 (32, 59, 73).

463 In ENP, anthrax affects mainly grazers rather than browsers (45). Although this is supported by the  
464 low levels of anthrax mortality in ENP browsers, the anti-PA antibodies indicate that kudu in ENP  
465 are exposed to the *B. anthracis* spores in the environment and this may require further investigation.  
466 Furthermore, the low number of kudu cases reported in ENP over the years (46, 74) might be  
467 underreported as the species occurs primarily in inaccessible woodlands, which exist outside of the  
468 central open plains region (75, 76), leading to reduced mortality surveillance in these habitats. Despite  
469 these differences in surveillance effort, Huang, *et al.*, (77) reported that open habitats in ENP have  
470 higher anthrax risk than the woodland habitats preferred by kudu.

471 Zebra in ENP had significantly higher antibody responses, as indicated by the anti-PA ELISA than  
472 zebra in KNP (Figure 5d, Table 1). The high proportion of zebra (82.5%) in ENP testing positive for  
473 anti-PA antibodies in this study was similar to Cizauskas, *et al.*, (22) who demonstrated a 52-87%  
474 prevalence of anti-PA antibodies in ENP zebra. This prevalence is reflected by zebra making up  
475 68.7% of the anthrax mortalities in ENP compared to only 2.9% in KNP (Figure 2 and 3). In previous  
476 studies conducted in ENP and Serengeti National Park, Tanzania, none of the zebras tested positive

477 (27, 28). The difference between the exposures and antibody levels in the two populations of zebras  
478 could be associated with the spore concentration in the soil ingested during grazing (45, 78), or  
479 interactions between zebra diet and foraging behaviour altering exposure risk over time (46).

480 Based on our results, kudu in KNP and zebra in ENP encounter lethal doses of the pathogen in the  
481 environment more often than other species in these parks resulting in the higher mortality rates as  
482 seen in the mortality reports. These exposure differences may arise from behavioural and ecological  
483 factors as well as climate extremes such as droughts and flooding (79). Furthermore, the season of  
484 anthrax outbreaks between the two parks (45, 55) may contribute towards the difference observed  
485 between animal species in the two parks. The mortality and exposure results confirm that kudu in  
486 KNP and zebra in ENP are the most affected species in each park, followed by impala for KNP and  
487 wildebeest in ENP (45, 47, 55) (Figure 4a and Table 1).

488 The animals in the high incidence region of KNP had higher antibodies titres as reflected by their  
489 anti-PA antibody response than animals in the low incidence region of the park. These animals are  
490 2.8 times more likely to be seropositive for *B. anthracis* anti-PA antibodies than animals in the low  
491 incidence region of the park. The presence of physical barriers such as rivers restrict the long-range  
492 movement of animals (personal communication, Skukuza State Veterinary Services, O. Louis van  
493 Schalkwyk) and may explain the difference in exposure. Also, home range sizes may be much smaller  
494 in KNP (Huang, unpublished data). We speculate that animal movement may restrict spore  
495 distribution and therefore may be responsible for the difference noted. Also, differences in animal  
496 densities and wild ungulate community composition could influence the variation seen in this study,  
497 and this requires further study. The finding of seropositive zebras and kudu in southern KNP indicates  
498 that animals are also exposed in the “low incidence” area. Steenkamp, *et al.*, (80) identified the ‘low  
499 incidence’ area in KNP as a region of high *B. anthracis* spore suitability. Also, previous anthrax  
500 reports from KNP show that large anthrax outbreaks in the 1960s spread from the north south to the  
501 central part of KNP (57). There was an obvious bias in the passive surveillance of KNP as seen in the  
502 disparity between samples submitted from the north and south (Figure 2). Also, a similar bias was  
503 noticed in ENP where mortalities, in general, were underreported in both the western and eastern  
504 regions of the park (Figure 2).

505 In ENP there was no significant difference in anti-PA antibodies in animals in the high and low  
506 incidence regions. The absence of spatial patterns in exposure could be because ENP does not have  
507 physical barriers (such as rivers) that would prevent or slow movement between the west and central  
508 regions of the park, and thus animals can move across regions (81, 82). Secondly, animals in ENP  
509 have relatively large home ranges, and animals sometimes move between the western and central  
510 parts of the park (Huang, unpublished data). A study suggested that spores could concentrate more in  
511 the waterholes dispersed in the western part of ENP, as 26% of waterholes in the western part tested  
512 positive for anthrax spores (36), although Turner, *et al.*, (83) found that spore concentrations in  
513 waterhole sediments are too low for lethal exposures. Cloete (84) reported that there was no  
514 significant difference in spore survival by soil types sampled from different regions of the park.  
515 Together, these results suggest that the whole park may be a suitable habitat for *B. anthracis*  
516 especially when there are no physical barriers (beyond the salt pan) to restrict herbivore movement  
517 or spore distribution. Thus, most of ENP could potentially be high incidence for anthrax, but cases in  
518 the west may be underreported due to lower surveillance effort over time. Surveillance could be more  
519 evenly applied in both parks, to examine whether the serological patterns observed here are evidence  
520 of unreported anthrax cases/outbreaks or sublethal exposures to spores that do not lead to mortalities.

521 Based on results of this study, different herbivore species in the same ecosystem could be affected at  
522 different times and different rates, based on differences in their ecology or behaviour. Outbreaks in

523 zebra populations have been shown to occur mostly during the wet season or towards the end of the  
524 rainy season, with some cases occurring during droughts or extended dry periods (85, 86). In contrast,  
525 outbreaks in kudu occur largely during the dry season as seen in KNP and other parks (30, 55-57,  
526 87). The grazing versus browsing transmission pathways occur at different timescales, which may  
527 have important effects on disease dynamics, pathogen diversity, and host resistance. Browsing-based  
528 transmission should occur shortly after host death before rainfall or leaf loss by deciduous  
529 trees/shrubs reduces exposure(74). Grazing-based transmission occurs only upon the regeneration of  
530 vegetation at a carcass site, and continues for years, with exposure dose decaying over time (78, 83).

#### 531 4.2 Species and spatial patterns in toxin neutralisation ability

532 Spatial patterns in toxin neutralisation suggest that environment (affecting exposure frequency or  
533 dose) and the presence of neutralizing antibodies are the major determinants of the animal's tolerance  
534 to the LT. Kudu and zebra demonstrated interesting variation in levels of neutralisation. Kudu in ENP  
535 had a higher TNA response than kudu in KNP. Similarly, zebra in KNP had a higher TNA response  
536 than zebras in ENP. These results agree with the mortality records of these species in the two parks  
537 (45, 46, 55). Based on mortality patterns and exposure prevalence, we can assume that zebra in ENP  
538 and kudu in KNP are exposed more often, and to larger doses, than in the other park. Thus, those host  
539 populations with lower mortality (kudu in ENP and zebra in KNP) are more likely to be exposed to  
540 sublethal amounts of the pathogen based on their foraging behaviour and the relative risk of that  
541 behaviour in the two landscapes (45, 46, 55), yet show greater toxin neutralisation than their  
542 counterparts in the other park. A previous study showed that animals that were immunized with  
543 antigens of spore origin conferred protection against *B. anthracis* through the production of antibodies  
544 that reduced spore germination (88). This type of sublethal passive natural "vaccination" may have  
545 induced anti-spore antibodies and reduced germination in zebra in KNP and kudu in ENP, but this  
546 would need further investigation (89).

547 The production of high-affinity memory B-cells during affinity maturation in the germinal centres is  
548 very important in the stimulation of an effective immune response (90, 91). When the concentration  
549 of the antigen is high or encountered more frequently, this leads to low competition among B-cells  
550 and the germinal centres become occupied with producing antibodies that have a very low affinity  
551 (90-92). Dumas, *et al.*, (93) also suggested that a higher immune response is derived from severe  
552 disease caused by exposure to a high amount of antigen over longer periods. Zebra in KNP and kudu  
553 in ENP could be better protected from the effect of the LT (94, 95), which may be due to their ability  
554 to develop antibodies of high affinity (20). As discussed earlier, a relationship has been established  
555 between antigen dose, "immunization" (exposure) interval and development of antibodies with high  
556 affinity (90, 92). This relationship may play a role in animals with higher neutralisation that may have  
557 moderate doses and at longer intervals. It is important to note that no study has been conducted on  
558 affinity maturation with relation to dose in natural systems. Verma, *et al.*, (96) suggested that  
559 characteristics of the antibodies (factors such as the species of origin, subclasses and isotype) being  
560 examined in the test could largely affect the measure to which neutralisation can be influenced. As  
561 such, we suspect that species idiosyncrasies could have also played a role in the differences observed.  
562 For the above-mentioned reasons, variability in the kinetics of the antibody affinity maturation  
563 process, anti-spore activities and species idiosyncrasies in the animals sampled may add to the  
564 diversity of the neutralising ability observed (Ngundi *et al.*, 2010).

565 Another hypothesis for why species have anti-PA antibodies without toxin neutralizing titres (e.g.  
566 ENP zebra) or in areas with few anthrax mortalities recorded (e.g. southern KNP) might be due to



567 cross-reaction with closely related antigens to *B. anthracis* PA (93), which needs further  
568 investigation. Cross-reactivity will affect the specificity of the technique (PA-ELISA). *Bacillus*  
569 *cereus* biovar *anthracis* and atypical *B. cereus* have been reported to cause anthrax-like infections in  
570 humans and animals (97-99). Furthermore, members of *B. cereus sensu stricto* have been reported to  
571 be closely related to *B. anthracis* (100). Since TNA quantifies only the neutralizing antibodies in  
572 serum, the *B. cereus* isolates with similar pag genes may account for the anti-PA positive samples  
573 that were negative for TNA (20). Kudu in ENP (46) and zebra in KNP (Figure 2) are considered less  
574 susceptible (not major hosts) species in these parks. We suggest that their ability to mount neutralizing  
575 immune responses against the toxin could be, to an extent, responsible for their protection (94, 95,  
576 101). This hypothesis is based on laboratory studies that reported LT neutralizing antibodies post-  
577 vaccination correlated with survival rates in rabbits (*Oryctolagus cuniculus*) (95, 102), guinea pigs  
578 (*Cavia porcellus*) (101) and mice (*Mus musculus*) (94).

### 579 **4.3 Association between anti-PA and TN antibodies**

580 Some studies have demonstrated a correlation between anti-PA antibody titres and toxin neutralizing  
581 titres (25, 103). Ndumnego, *et al.*, (25) quantified the anti-PA IgG titres and reported a high  
582 correlation with neutralizing antibodies in vaccinated goats (*Capra aegagrus hircus*). Parreiras, *et al.*,  
583 (103) compared anti-PA ELISA and TNA in mice vaccinated with PA. In our study, a significant  
584 positive correlation was found between the anti-PA ELISA antibody response (SP) and the NT<sub>50</sub> in  
585 animals that naturally acquired the antigen, despite differences between zebra in ENP and kudu in  
586 KNP. Although it was seen that anti-PA immune response had an effect on toxin neutralisation status  
587 in kudu, this was not true in zebra. This result was largely influenced by the zebra in ENP as only a  
588 few showed neutralisation. However, the correlation noticed was expected as neutralising antibodies  
589 are subsets (functional) of the total anti-PA IgG antibodies (20). Not all seropositive animals based  
590 on anti-PA ELISA showed neutralizing activities, but most animals with neutralizing activities had a  
591 high anti-PA titre. Most studies previously conducted were controlled laboratory studies, with  
592 animals vaccinated with a predetermined dose and at planned frequencies, which allow for the  
593 production of antibodies with high affinity. This is in contrast to this study, where animals were free-  
594 roaming and as such, they encounter pathogen at varying doses and frequencies. This study further  
595 confirms the presence of the *B. anthracis* LT antibodies in animal sera.

## 596 **5 Conclusions**

597 Results of this study suggest that immune responses against multi-host pathogens are influenced by  
598 several factors (environment, species idiosyncrasies, frequency of exposure, exposure dose), which can  
599 be missed from a narrow focus of a single system or species. In this study, the host species from both  
600 parks varied in their exposure to *B. anthracis* and immune response to its LT. These patterns may be  
601 due to environmental differences between these systems and how they relate to host behaviour, which  
602 may lead to variation in the frequency of exposure and dose and, in turn, a corresponding  
603 immunological trade-off between exposure and tolerance (or resistance) to the anthrax LT.  
604 Furthermore, this study revealed that animals in both regions of the parks are exposed to anthrax spores  
605 in the environment, which in some cases (e.g., KNP zebra) was inconsistent with anthrax mortality  
606 data. As such, our study provides valuable insight into the mechanisms driving variation in anthrax  
607 dynamics observed in these parks, with implications for anthrax variation globally.

608

## 609 **6 Recommendations for future research**

610 Future studies examining the role of environmental conditions such as landscape, rainfall, and forage  
611 availability on host behaviour are needed to establish mechanisms behind the variation in the exposure  
612 status of a given host species across locations. Secondly, because of the varying reports in the binding  
613 ability of commercially available conjugates, we recommend the development of species-specific  
614 conjugates to overcome this limitation. Thirdly, we recommend increased surveillance effort,  
615 especially in the “low incidence areas,” to improve the quality of data currently available. We also  
616 recommend that investigation into the role of anthrax risky behaviours or other mechanical vectors in  
617 the transmission of *B. anthracis* is needed in ENP to allow comparison to KNP. Further work could  
618 investigate the detection of *B. anthracis* in the high versus low incidence regions of these parks as well  
619 as the detection of closely related *B. cereus* species in the parks. Future studies could also investigate  
620 how exposure frequency and dose affect the correlation between anti-PA antibodies and NT<sub>50</sub>.

## 621 **7 Author Contributions**

622 SOO, HvH and WCT conceived the ideas of the study. SOO, HvH, WCT and PLK designed the study.  
623 SOO, AH, OLVS, EHD, Y-HH and AH collected the data; SOO and HvH designed the methodology;  
624 SOO, CB and Y-HH analyzed the data; SOO and HvH wrote the first draft of the manuscript. All  
625 authors contributed significantly to manuscript revision, read, and gave approval for publication.

## 626 **8 Acknowledgements**

627 We wish to express our appreciation to SANParks Veterinary Wildlife Services and the Skukuza State  
628 Veterinary office, South Africa, as well as the Etosha Ecological Institute under the Namibian Ministry  
629 of Environment, Forestry and Tourism for their immense support throughout the period of this study.  
630 In South Africa we thank Lin-Mari de Klerk-Lorist, Rudi Lorist and Schalk van Dyk for assistance  
631 with animal captures & sampling. In Namibia we thank Werner Kilian, Claudine Cloete and John  
632 Mfuno for logistical support. Hendrina Joel, Gabriel Shatumbu, Carly Dickson, Amélie Dolfi, Seth  
633 Guim, Jason Iiyambo, Naftali Iiyambo, Mark Jago, Johannes Kapner, Kantana Mathews, Carl-Heinz  
634 Moeller, Paulus Namholo and Janine Sharpe provided assistance with animal captures and sample  
635 collection.

## 636 **9 Funding**

637 This work was supported by NSF Division of Environmental Biology (DEB-1816161/DEB-2106221)  
638 to W.C.T., P.L.K. and H.v.H. Any use of trade, product, or firm names is for descriptive purposes only  
639 and does not imply endorsement by the U.S. Government.

640 -

641 Table 1. Differences in host exposure to *Bacillus anthracis* by species and location, assessed through anti-protective antigen (PA) antibodies.  
 642 Optical density (OD) values were measured using an anti-PA ELISA, and mean sample to positive (SP) ratios were estimated for all sampled  
 643 animals in a given location. SD is the standard deviation. Areas of high and low incidence in each park (ENP = Etosha National Park and  
 644 KNP = Kruger National Park) are shown in Figure 1. The species of study included greater kudu (*Tragelaphus strepsiceros*), plains zebra  
 645 (*Equus quagga*), impala (*Aepyceros melampus*), and blue wildebeest (*Connochaetes taurinus*).

646  
647

Animal species	National park	Location	No. of animals	% of positive animals (N)	Mean SP $\pm$ SD for positive animals	Odds ratio of exposure	p-value
<b>Kudu</b>	ENP	High incidence	20	70 (14)	0.59 $\pm$ 0.29		0.94 <sup>a</sup>
		Low incidence	20	60 (12)	0.71 $\pm$ 0.58		0.51 <sup>b</sup>
		Whole park	40	65 (26)	0.65 $\pm$ 0.46		
	KNP	High incidence	19	94.7 (18)	1.54 $\pm$ 0.29		0.04 <sup>a</sup>
		Low incidence	18	72.2 (13)	1.02 $\pm$ 0.39		0.09 <sup>b</sup>
		Whole park	37	83.8 (31)	1.24 $\pm$ 0.74	2.9 <sup>c</sup>	0.005 <sup>a</sup>
<b>Zebra</b>	ENP	High incidence	20	95 (19)	0.73 $\pm$ 0.48		0.97 <sup>a</sup>
		Low incidence	20	70 (14)	0.66 $\pm$ 0.58		0.10 <sup>b</sup>
		Whole park	40	82.5 (33)	0.69 $\pm$ 0.53	3.1 <sup>c</sup>	0.07 <sup>s</sup>
	KNP	High incidence	20	75(15)	0.66 $\pm$ 0.33		0.04 <sup>b</sup>
		Low incidence	20	50 (10)	0.41 $\pm$ 0.46		0.03 <sup>a</sup>
		Whole park	40	62.5 (25)	0.53 $\pm$ 0.40		0.09 <sup>b</sup>
<b>Wildebeest</b>	ENP	High incidence	20	35 (7)	0.52 $\pm$ 0.23		
<b>Impala</b>	KNP	High incidence	20		0.48 $\pm$ 0.19		

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<sup>a</sup> p-value for comparison of mean anti-PA OD.

<sup>b</sup> p-value for comparison of proportion of positive animals

<sup>c</sup> Odds ratio comparing national parks for each species

652 Table 2. Lethal toxin (LT) neutralisation titres and proportion of herbivores that neutralised anthrax LT in Kruger National Park (KNP),  
 653 South Africa, and Etosha National Park (ENP) in Namibia. The location of high and low incidence areas in each park are shown in Figure 1.  
 654 TNA is the toxin neutralisation assay; SD is the standard deviation, and the species of study included greater kudu (*Tragelaphus*  
 655 *strepsiceros*), plains zebra (*Equus quagga*), impala (*Aepyceros melampus*), and blue wildebeest (*Connochaetes taurinus*). The neutralisation  
 656 titre 50 (NT<sub>50</sub>) was the highest titre that protected 50% of mouse macrophage cells.

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Animal species	National park	Location	No. of animals sampled	% of positive animals (N)	Mean NT <sub>50</sub> ± SD of TNA positive animals	Odds ratio	p-value
<b>Kudu</b>	ENP	High incidence	20	70 (14)	92.0 ± 66.7	1.96 <sup>c</sup>	0.56 <sup>a</sup>
		Low incidence	20	60(12)	130.9 ± 100.5		0.95 <sup>b</sup>
		Whole park	40	65(26)	110.0 ± 84.6		0.03 <sup>a</sup>
	KNP	High incidence	19	52.6(10)	100.9 ± 73.5		0.61 <sup>a</sup>
		Low incidence	18	44.4(48)	59.6 ± 16.1		0.06 <sup>b</sup>
		Whole park	37	48.6(18)	64.2 ± 56.3		
<b>Zebra</b>	ENP	High incidence	20	15(3)	35.2 ± 10.9	23.7 <sup>c</sup>	0.29 <sup>a</sup>
		Low incidence	20	5(15)	47.8 ± NA		NA <sup>b</sup>
		Whole park	40	10(4)	38.3 ± 10.9		
	KNP	High incidence	20	70(14)	85.0 ± 47.6		0.45 <sup>b</sup>
		Low incidence	20	75(15)	66.3 ± 43.9		0.72 <sup>a</sup>
		Whole park	40	72.9(29)	75.3 ± 45.9		0.05 <sup>a</sup>
<b>Wildebeest</b>	ENP	High incidence	20	40(8)	38.5 ± 40.9	NA	NA
<b>Impala</b>	KNP	High incidence	20	15(3)	22.3 ± 8.3	NA	NA

658 <sup>a</sup> p-value for comparison of mean of neutralisation titre 50 (NT<sub>50</sub>).

659 <sup>b</sup> p-value for comparison of the proportion of animals that showed neutralisation

660 <sup>c</sup> Odds ratio comparing national parks for each species

661 Table 3. Comparison of anti-protective antigen (PA) enzyme-linked immunosorbent assay (ELISA)  
 662 and toxin neutralisation assay (TNA) for the detection of immune exposure to *B. anthracis* in  
 663 kudu, zebra, wildebeest and impala from Kruger (KNP) and Etosha (ENP) National Parks in  
 664 South Africa and Namibia, respectively. The species of study included greater kudu  
 665 (*Tragelaphus strepsiceros*), plains zebra (*Equus quagga*), impala (*Aepyceros melampus*), and  
 666 blue wildebeest (*Connochaetes taurinus*).

Species	National Park	Sample number	ELISA Status	TNA		<i>p</i> -value
				No. negative (%)	No. positive (%)	
<b>Kudu</b>	ENP	40	Negative	12 (30.0)	9(22.5)	0.002 669
			Positive	2(5.0)	17(42.5)	
	KNP	37	Negative	8(21.6)	1(2.7)	0.012 670
			Positive	11(29.7)	17(45.9)	
Total	77	Negative	20(26.0)	10(13.0)	0.001 671	
		Positive	13(16.9)	34(44.1)		
<b>Zebra</b>	ENP	40	Negative	7(17.7)	0	0.437 672
			Positive	29(72.5)	4(10.0)	
	KNP	37	Negative	9(22.5)	5(12.5)	<0.001 673
			Positive	2(5.0)	24(60.0)	
Total	80	Negative	16(20.0)	5(6.3)	0.049 674	
		Positive	31(38.7)	28(35.0)		
<b>Wildebeest</b>	ENP	20	Negative	9(45.5)	2(10.0)	0.022 675
			Positive	2(10.0)	7(35.0)	
<b>Impala</b>	KNP	20	Negative	10(50.0)	2(10.0)	0.656 676
			Positive	7(35.0)	1(5.0)	

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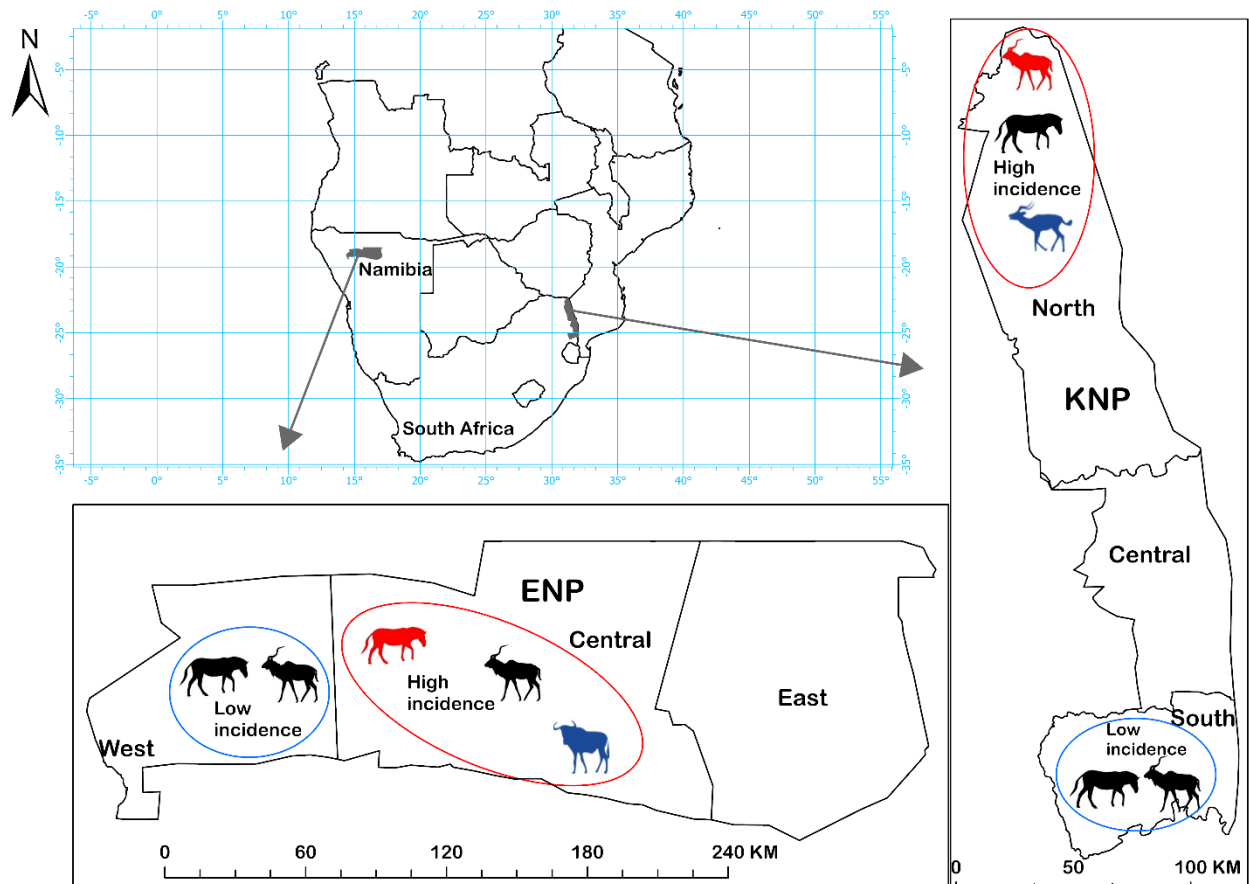
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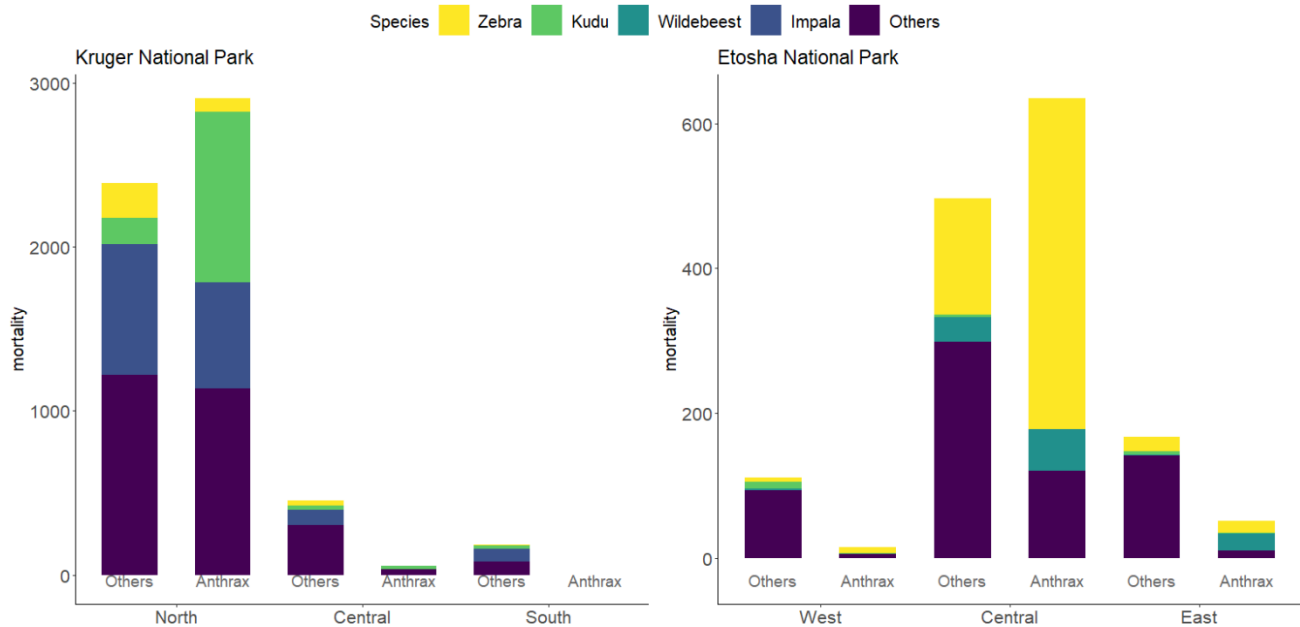
941 Figure 1 Etosha National Park (ENP) and Kruger National Park (KNP) in southern Africa, showing  
 942 the study areas where anthrax outbreaks occur with high (red circles) or low (blue circles)  
 943 incidence. Host species sampled for this study in different areas are shown with animal  
 944 silhouettes. Kudu (*Tragelaphus strepsiceros*) and zebra (*Equus quagga*) were sampled in all  
 945 four areas. Secondary host species were sampled in high incidence areas of each park; impala  
 946 (*Aepyceros melampus*) in KNP, and wildebeest (*Connochaetes taurinus*) in ENP. The primary  
 947 anthrax host species in a high incidence area is shown in red and the secondary host species in  
 948 blue, otherwise, silhouettes are black. Assignment of areas as high or low incidence was  
 949 based on anthrax mortality patterns from each park, and where anthrax occurs most  
 950 commonly or least commonly, respectively.



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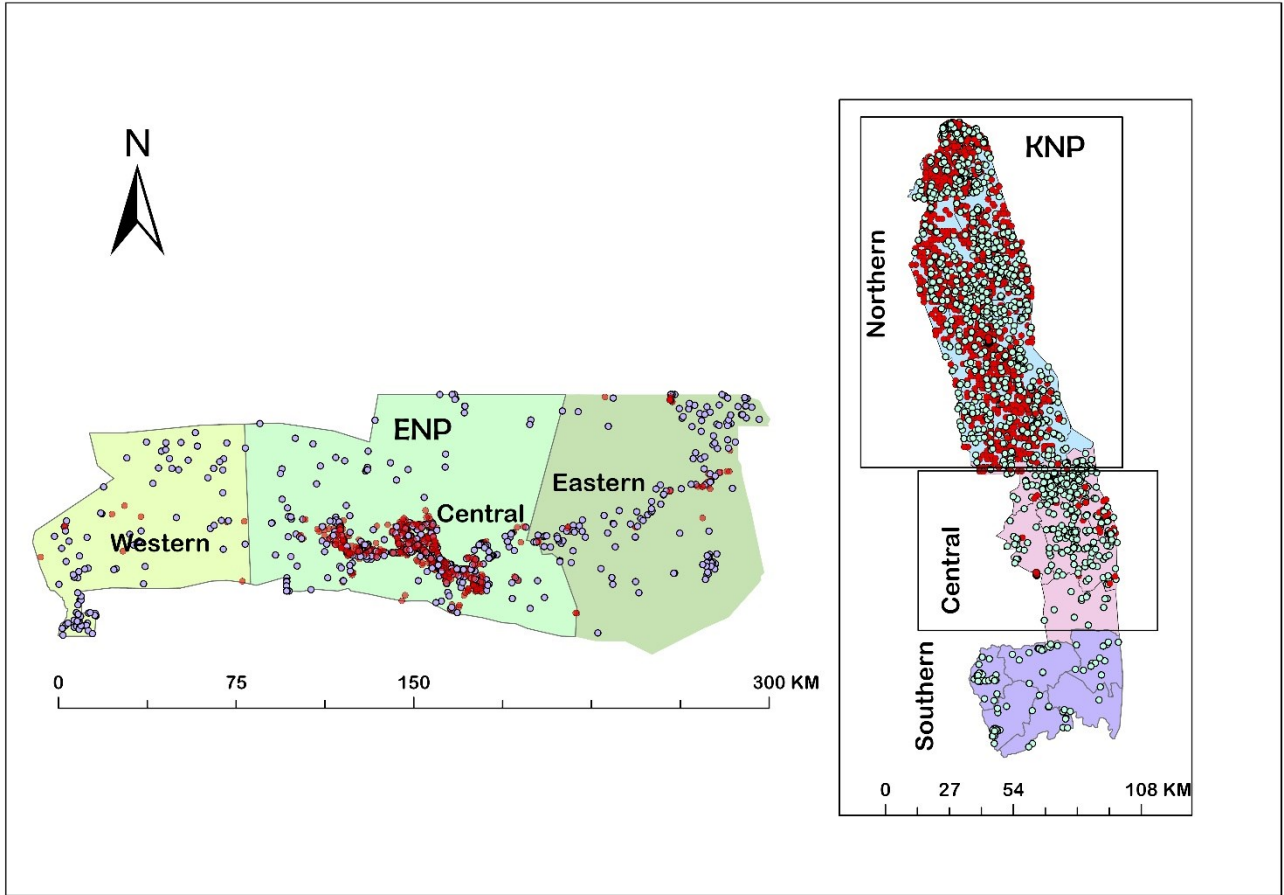
953 Figure 2 Bar charts of the distribution of mortalities by region and species from 1990-2016 in  
 954 Kruger National Park (KNP) and from 1996-2016 in Etosha National Park (ENP). Mortalities  
 955 are group into anthrax or other causes of death. Species of study included greater kudu  
 956 (*Tragelaphus strepsiceros*), plains zebra (*Equus quagga*), impala (*Aepyceros melampus*), and  
 957 blue wildebeest (*Connochaetes taurinus*). Species that fell into the “other” category included  
 958 21 species for anthrax mortalities and 57 species for other mortalities in KNP and 6 species for  
 959 anthrax mortalities and 27 species for other mortalities in ENP. Data for KNP were provided  
 960 by Skukuza Veterinary Services and for ENP from the Etosha Ecological Institute.



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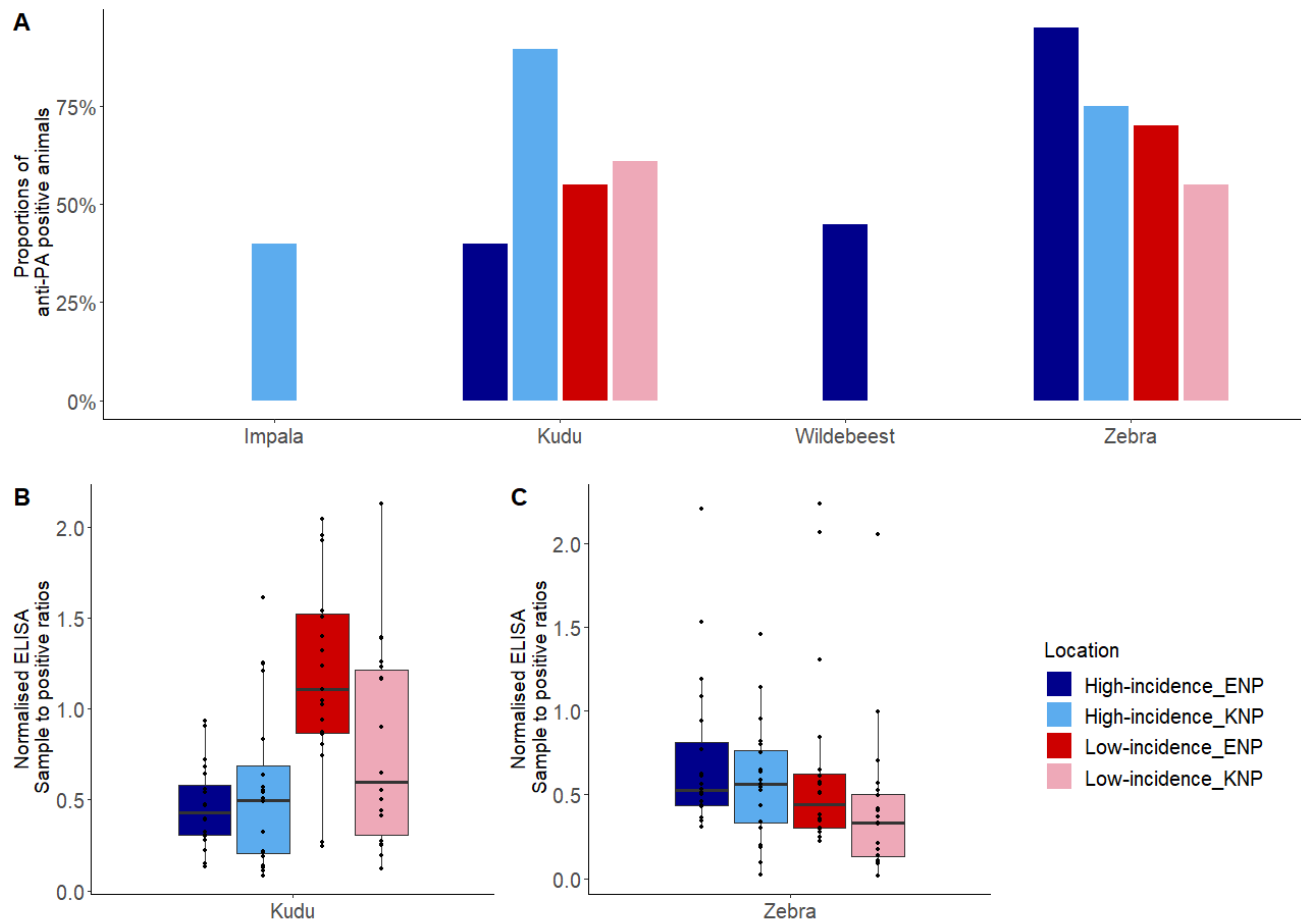
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963 Figure 3 Map showing distributions of mortalities from 1990-2015 in the three regions of Kruger  
964 National Park (KNP), South Africa and from 1996-2016 in the three regions of Etosha  
965 National Park (ENP), Namibia. Red dots indicate anthrax positive mortalities and the white  
966 dots indicate non-anthrax mortalities.



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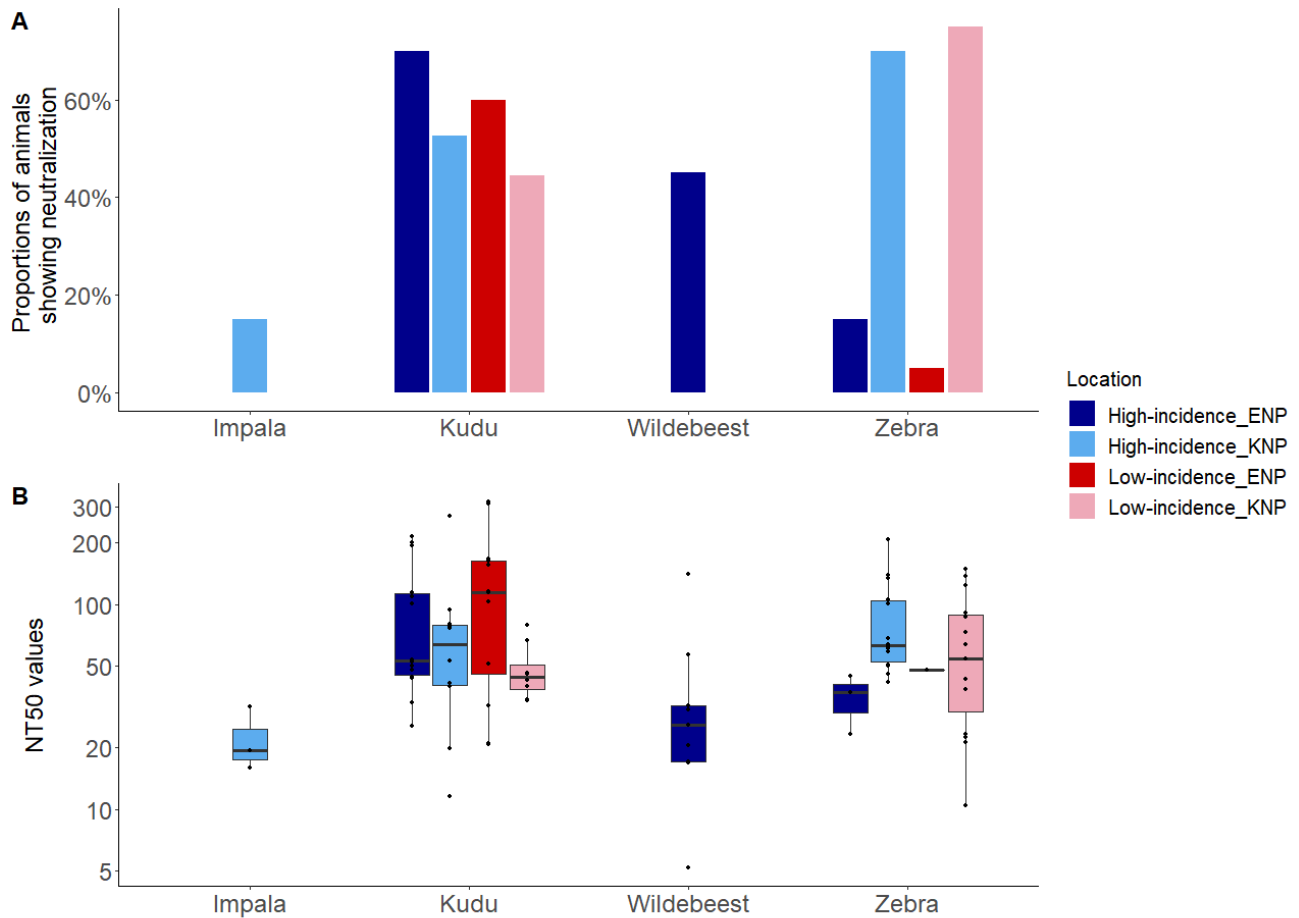
969 Figure 4 Host exposures to *Bacillus anthracis* assessed based on anti-Protective Antigen (PA)  
 970 antibody titres. (a) The proportion of each host species that was seropositive for anti-PA  
 971 antibodies, as determined using enzyme-linked immunosorbent assay (ELISA), by area.  
 972 (b,c) Box plots showing sample to positive (SP) ratios for antibodies against PA for (b)  
 973 kudu (*Tragelaphus strepsiceros*) in each park and area, and (c) zebra (*Equus quagga*) in  
 974 each park and area. Kudu and zebra were sampled from high incidence and low incidence  
 975 areas of Kruger National Park (KNP) in South Africa and Etosha National Park (ENP) in  
 976 Namibia, while impala (*Aepyceros melampus*) in KNP, and wildebeest (*Connochaetes*  
 977 *taurinus*) were sampled from only the high incidence area of KNP and ENP, respectively.  
 978 Box plots b and c were separated to avoid comparison between species as the technique  
 979 utilised is species-specific. The locations of high and low incidence areas in each park are  
 980 shown in Figure 1.



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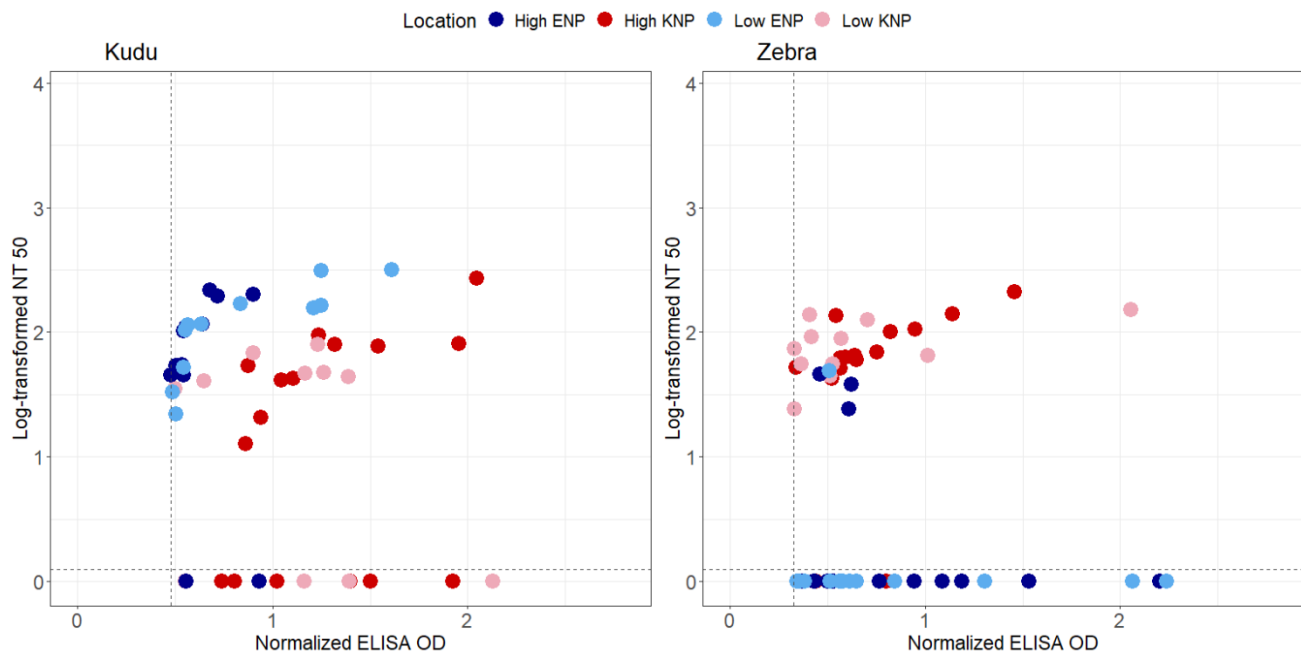
982

983 Figure 5 Host toxin neutralisation against the *Bacillus anthracis* lethal toxin for four wild herbivore  
 984 species sampled in Kruger National Park (KNP), South Africa, or Etosha National Park (ENP), Namibia,  
 985 showing (a) the proportion of animals showing neutralization, and (b) the neutralisation titre 50 (NT<sub>50</sub>).  
 986 The NT<sub>50</sub> was the highest titre that protected 50% of mouse macrophage cells. The y-axis of plot b  
 987 represents log<sub>10</sub> transformed NT<sub>50</sub>. Species of study included greater kudu (*Tragelaphus strepsiceros*),  
 988 plains zebra (*Equus quagga*), impala (*Aepyceros melampus*), and blue wildebeest (*Connochaetes taurinus*).  
 989 The locations of high and low incidence areas in each park are shown in Figure 1.  
 990



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994 Figure 6: Scatter plots representing log-transformed neutralisation titre 50 (NT<sub>50</sub>) and normalised anti-  
 995 Protective Antigen (PA) enzyme-linked immunosorbent assay (ELISA) optical  
 996 densities (ODs) for greater kudu (*Tragelaphus strepsiceros*) and plains zebra (*Equus  
 997 quagga*) from Kruger National Park (KNP) in South Africa and Etosha National Park  
 998 (ENP) in Namibia. The NT<sub>50</sub> was calculated as the highest titre that protected 50% of  
 999 the mouse macrophage cells. Shapes and colours of marker points represent different  
 1000 parks (blue circle: ENP, red triangle: KNP), and variation in colour shades indicate  
 1001 study area that differs by anthrax incidence (dark: high incidence, light: low incidence)  
 1002 in each park. TNA negative sera (samples that could not protect 50% of the  
 1003 macrophages) are seen below the dotted horizontal lines in each plot. Only samples that  
 1004 were anti-PA positive are shown as others were assumed to not have been exposed to  
 1005 BA (the threshold for anti-PA positive is shown with the dotted vertical line). The  
 1006 locations of high and low incidence areas in each park are shown in Figure 1.



1007

1008

## *Supplementary Material*

### 1009 **10 Supplementary Data**

1010 Supplementary Methodology: Briefly, a pooled sera for each species (impala and kudu) was coated at  
1011 a dilution of 1:2000 in coating buffer (bicarbonate buffer) per well and incubate overnight at 4°C. This  
1012 was followed by a blocking step where coated plates were blocked with the blocking buffer (200 µL)  
1013 containing PBST and 5% skimmed milk powder (PBSTM) and then incubated at room temperature for  
1014 1 hour. The three commercially available conjugates were tested against kudu (*Tragelaphus*  
1015 *strepsiceros*) and impala (*Aepyceros melampus*) each species. Each conjugate was added to 12 wells  
1016 of each plate for each species at a dilution of 1:10000. The plates were incubated at room temperature  
1017 for 30 minutes. Subsequently, the plates were washed after which the ABTS substrate (2,2'-Azinobis  
1018 [3-ethylbenzothiazoline-6-sulfonic acid]-diammonium salt; Thermo Scientific 1-Step ABTS, USA)  
1019 was added and allowed in the dark for the colour to develop for 45 minutes. The absorbance was read  
1020 at 405 nm using a Biotek Powerwave XS2 reader (USA). A one-way ANOVA with a Tukey's test was  
1021 performed to determine if there was a statistically significant difference between conjugates. Results  
1022 are shown in Supplementary Table S2 and Supplementary Figure 1.

1023



1024 **11 Supplementary Figures and Tables**

1025 **Supplementary Table S1:** Mortality data from Kruger National Park (KNP), South Africa, and  
 1026 Etosha National Park (ENP) in Namibia showing the distribution of carcass detection and anthrax  
 1027 mortality. Data for KNP ranged from 1990-2015 and for ENP, from 1996-2015 and the species of  
 1028 study included greater kudu (*Tragelaphus strepsiceros*), plains zebra (*Equus quagga*), impala  
 1029 (*Aepyceros melampus*), and blue wildebeest (*Connochaetes taurinus*) and the count (n) for the  
 1030 mortality.

<b>Park</b>	<b>Part</b>	<b>Species</b>	<b>Cause</b>	<b>n</b>
ENP	Central	Kudu	Others	3
ENP	Central	Others	Others	299
ENP	Central	Wildebeest	Others	34
ENP	Central	Zebra	Others	161
ENP	Central	Others	Anthrax	120
ENP	Central	Wildebeest	Anthrax	58
ENP	Central	Zebra	Anthrax	457
ENP	West	Kudu	Others	9
ENP	West	Others	Others	94
ENP	West	Wildebeest	Others	2
ENP	West	Zebra	Others	6
ENP	West	Kudu	Anthrax	1
ENP	West	Others	Anthrax	6
ENP	West	Zebra	Anthrax	8
ENP	East	Kudu	Others	4
ENP	East	Others	Others	142
ENP	East	Wildebeest	Others	1
ENP	East	Zebra	Others	20
ENP	East	Kudu	Anthrax	1
ENP	East	Others	Anthrax	10
ENP	East	Wildebeest	Anthrax	24
ENP	East	Zebra	Anthrax	17
KNP	North	Impala	Others	796
KNP	North	Kudu	Others	160
KNP	North	Others	Others	1221
KNP	North	Zebra	Others	211
KNP	North	Impala	Anthrax	647
KNP	North	Kudu	Anthrax	1037
KNP	North	Others	Anthrax	1136
KNP	North	Zebra	Anthrax	85
KNP	Central	Impala	Others	94
KNP	Central	Kudu	Others	22
KNP	Central	Others	Others	305
KNP	Central	Zebra	Others	32
KNP	Central	Impala	Anthrax	1
KNP	Central	Kudu	Anthrax	16
KNP	Central	Others	Anthrax	36

KNP	Central	Zebra	Anthrax	0
KNP	South	Impala	Others	80
KNP	South	Kudu	Others	20
KNP	South	Others	Others	80
KNP	South	Zebra	Others	5

1031 All other species both for anthrax mortality and other causes of death were categorized as “others” (21 different species  
 1032 for anthrax mortality and 57 for other mortality for KNP and 6 different species for anthrax mortality and 27 species for  
 1033 other mortalities in ENP). Mortality data was acquired from Skukuza Veterinary Services and Etosha Ecological Institute.  
 1034

1035 **Supplementary Table S2:** Optical Density (OD) values of each conjugate against each species at a  
 1036 dilution of 1:20000 for kudu (*Tragelaphus strepsiceros*) and impala (*Aepyceros melampus*)

Species	Protein A	Protein G	Protein AG
Impala	0.25	3.982	2.912
Impala	0.589	3.982	2.494
Impala	0.611	3.971	2.348
Impala	0.594	3.977	2.42
Impala	0.226	3.791	2.744
Impala	0.602	3.789	2.455
Impala	0.622	3.799	2.393
Impala	0.608	3.858	2.367
Impala	0.27	3.887	2.568
Impala	0.585	3.896	2.162
Impala	0.6	3.885	2.134
Impala	0.628	3.796	2.211
Kudu	1.392	3.872	2.657
Kudu	1.737	3.683	2.821
Kudu	1.775	3.911	2.795
Kudu	1.769	3.872	2.783
Kudu	1.257	3.728	2.509
Kudu	1.764	3.837	2.845
Kudu	1.837	3.82	2.861
Kudu	1.795	3.863	2.86
Kudu	1.233	3.892	2.197
Kudu	1.755	3.677	2.721
Kudu	1.833	3.793	2.53
Kudu	1.843	3.781	2.414

1037

1038 **Supplementary Table S3.** A generalised linear model (Gaussian distribution) for the significance of  
 1039 anti-protective antigen (PA) antibodies. Optical density (OD) values were measured using an anti-PA  
 1040 ELISA, and mean sample to positive (SP) ratios were estimated for all sampled animals in a given  
 1041 location. SD is the standard deviation. Areas of high and low incidence in each park (ENP = Etosha  
 1042 National Park and KNP = Kruger National Park) are shown in Figure 1. The species of study  
 1043 included greater kudu (*Tragelaphus strepsiceros*), plains zebra (*Equus quagga*), impala (*Aepyceros*  
 1044 *melampus*), and blue wildebeest (*Connochaetes taurinus*). Separate multivariable models were  
 1045 performed for kudu and zebra

1046

Variable	Category	No. of animals sampled	Mean SP $\pm$ SD	Coefficient	<i>p</i> -value
<b>Kudu</b>					
<b>National park</b>	ENP	40	0.65 $\pm$ 0.07		
	KNP	37	1.28 $\pm$ 0.12	1.33	0.047
<b>Location</b>	Low incidence	38	0.85 $\pm$ 0.11		
	High incidence	39	1.06 $\pm$ 0.11	0.89	0.41
<b>TNA status</b>	Negative	33	0.79 $\pm$ 0.13		
	Positive	44	1.08 $\pm$ 0.09	1.35	0.004
<b>Interaction: Park*location status (KNP*incidence area)</b>				1.64	0.015
<b>Zebra</b>					
<b>National park</b>	KNP	40	0.53 $\pm$ 0.06		
	ENP	40	0.69 $\pm$ 0.08	1.33	0.034
<b>Incidence status</b>	Low incidence	40	0.55 $\pm$ 0.08		
	High Incidence	40	0.67 $\pm$ 0.07	1.12	0.29
<b>TNA status</b>	Negative	47	0.60 $\pm$ 0.08		
	Positive	33	0.62 $\pm$ 0.07	1.22	0.15
<b>Wildebeest</b>	ENP (High incidence)	20	0.32 $\pm$ 0.05	NA	NA
<b>Impala</b>	KNP (Low incidence)	20	0.31 $\pm$ 0.04	NA	NA

1047

1048 **Supplementary Table S4.** A generalised linear model (Gaussian distribution) for the significance of  
 1049 *Bacillus anthracis* lethal toxin (LT) neutralisation scores from wildlife species sampled in two national  
 1050 parks in southern Africa. The neutralisation titre 50 (NT<sub>50</sub>) is the highest titre that protected 50% of  
 1051 mouse macrophage cells. ELISA is the enzyme linked immunosorbent assay; SD is the standard  
 1052 deviation, and mean sample to positive (SP) ratios were estimated for all sampled animals in a given  
 1053 location. Areas of high and low incidence in each park (ENP = Etosha National Park and KNP = Kruger  
 1054 National Park) are shown in Figure 1. Serum samples were collected from kudu (*Tragelaphus*  
 1055 *strepsiceros*), and zebras (*Equus quagga*) in Kruger National Park (KNP) in South Africa and Etosha  
 1056 National Park (ENP) in Namibia (see Figure 1 for parks and sub-locations).

1057

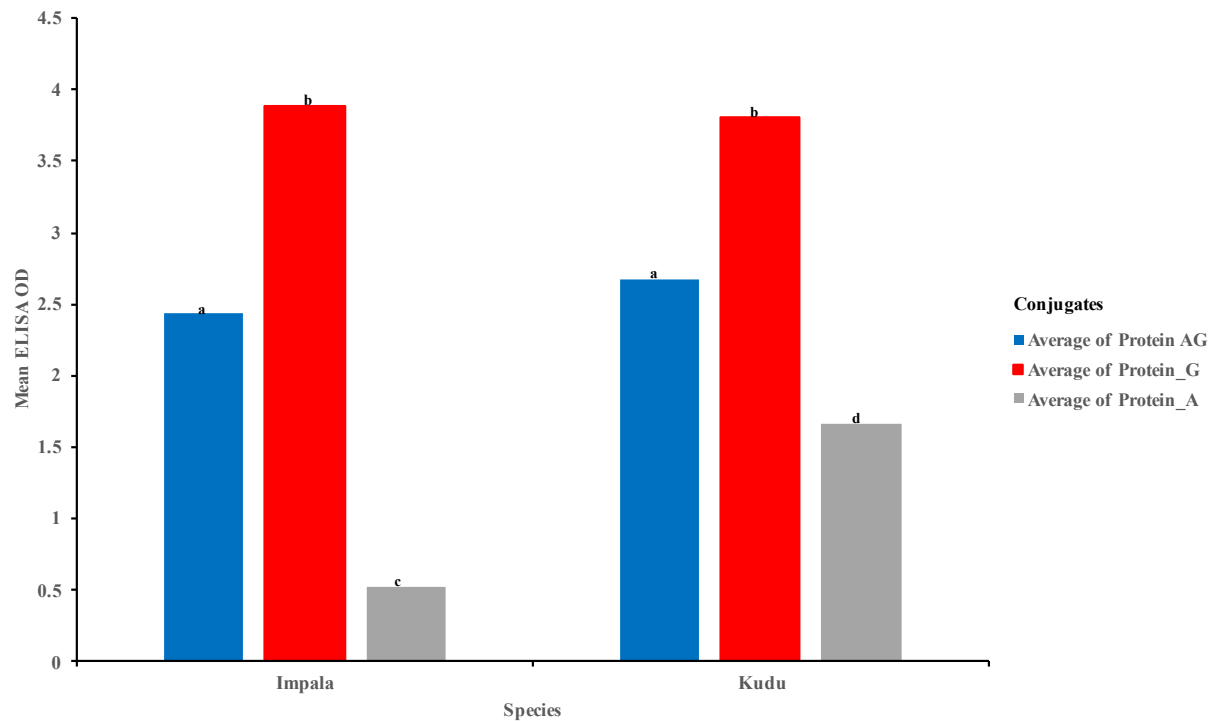
Variable	Category	No. of animals sampled	Mean NT <sub>50</sub> ± SD	Exp (coefficient)	p-value
<b>Host species</b>	Zebra	33	70.85 ± 7.8		
	Kudu	44	91.25 ± 11.6	0.68	<0.0001
<b>National park</b>	KNP	47	71.07 ± 7.3		
	ENP	30	100.43 ± 15.0	0.83	0.11
<b>Sub-location<sup>a</sup></b>	High incidence	41	81.72 ± 9.4		
	Low incidence	36	83.41 ± 12.0	NA	NA
<b>Host species x National park</b>				1.96	<0.0001
<b>ELISA SPs (Log)</b>				1.76	<0.0001

1058 <sup>a</sup> Sub-location was not included in the final Gaussian model because the variable was not significant

1059

1060

## 1061 11.1 Supplementary Figures



1062

1063 **Supplementary Figure 1.** Bar graph showing the mean optical densities (OD) of three different  
1064 protein conjugates for kudu (*Tragelaphus strepsiceros*) and impala (*Aepyceros melampus*). Different  
1065 lower-case letters above each bar indicate statistically significant differences ( $p < 0.05$ ) between the  
1066 different conjugates across the two species.

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