

A framework for integrating microbial dispersal modes into soil ecosystem ecology

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Abstract

Dispersal is a fundamental community assembly process that maintains soil microbial biodiversity across spatial and temporal scales, yet the impact of dispersal on ecosystem function is largely unpredictable. Dispersal is unique in that it contributes to both ecological and evolutionary processes and is shaped by both deterministic and stochastic forces. The ecosystem-level ramifications of dispersal outcomes are further compounded by microbial dormancy dynamics and environmental selection. Here we review the knowledge gaps and challenges that remain in defining how dispersal, environmental filtering, and microbial dormancy interact to influence the relationship between microbial community structure and function in soils. We propose the classification of microbial dispersal into three categories, through vegetative or active cells, through dormant cells, and through acellular dispersal, each with unique spatiotemporal dynamics and microbial trait associations. This conceptual framework should improve the integration of dispersal in defining soil microbial community structure-function relationships.

Keywords: soil microbiology, dispersal, environmental filtering, dormancy, community assembly

¹ 1. Introduction

² The interplay between microbial dispersal, environmental filtering, and
³ microbial dormancy introduces ecoevolutionary dynamics to soil ecosystems

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4 that limit our ability to decipher, much less predict, community structure-
5 function relationships. Despite rapid development of high throughput molec-
6 ular methods, decreasing sequencing costs, and accelerating generation of
7 large ecological data, soils remain messy. This is because soils are massively
8 complex superorganisms with emergent functions that are not yet easily pre-
9 dicted. To inform our predictions of amassed scale effects that determine soil
10 behavior, we need a better understanding of how microbial processes, such
11 as dormancy and environmental filtering, compound dispersal outcomes that
12 transpire to whole ecosystems.

13 We suggest an adaptable framework for thinking about how microbial
14 dispersal across space and time influences soil biodiversity, and ultimately,
15 ecosystem function. Modes of microbial dispersal are categorized as cellular
16 (*i.e.* vegetative or dormant cells) and acellular (*i.e.* genetic material as-
17 sociated with viruses and/or gene flow independent of cellular life), where
18 dispersal outcomes for organisms and/or their genes happen across different
19 spatiotemporal scales (Figure 1). Here we discuss the mechanistic constraints
20 of microbial dispersal modes and the interplay between dispersal, environ-
21 mental filtering, and dormancy. Finally, we propose a traits-based approach
22 for quantifying dispersal outcomes, and suggest how this framework can be
23 used to evaluate soil microbial structure-function relationships.

24 *1.1. Microbial community assembly*

25 To predict ecosystem function from community composition, we first need
26 to understand the community assembly processes that create and maintain
27 patterns of microbial diversity. In both microbial ecology and the broader
28 field of ecology, niche theory and selection-based models have classically ex-
29 plained patterns of community assembly, looking to environmental selection
30 and biotic interactions to define niche space and determine what conditions
31 a species will persist (Chase and Leibold, 2003, Holt, 2009). Alternatively,
32 neutral theory relies on stochastic processes to explain community ecology
33 patterns (Hubbell, 2001, Chave, 2004). Few elements of ecology are an ab-
34 solute either-or, and mathematical frameworks unify both niche and neutral
35 theory (Harshey, 2003, Mutshinda and O'Hara, 2011). In reality, compre-
36 hensive theory explains that variations in community assembly arise through
37 both deterministic and stochastic processes, and that individual processes
38 exist somewhere along a continuum between selection and neutrality (Chase
39 and Myers, 2011).

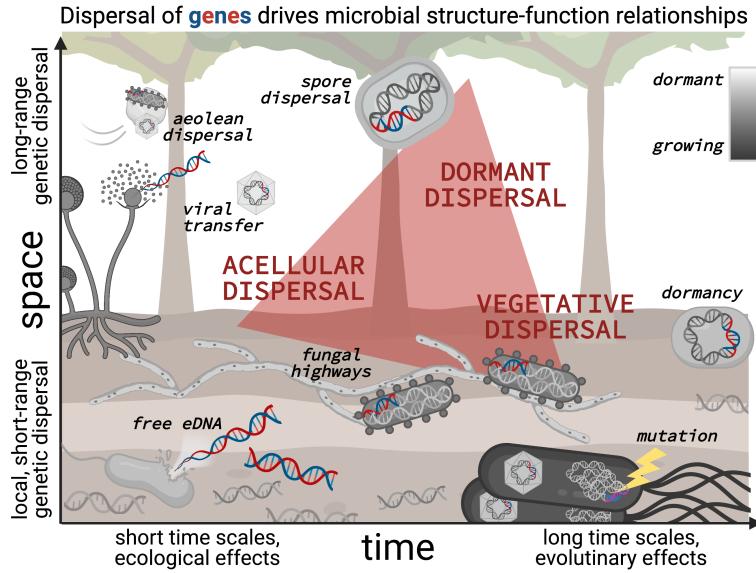


Figure 1: Microbial dispersal modes. Dispersal affects microbial structure-function relationships in soils by distributing genes (*i.e.* potential ecological functions, which are represented here by blue/red DNA fragments) in three primary ways: through active or vegetative cell dispersal, through dormant cell dispersal, and through acellular dispersal. Dispersal can occur at different scales over space and time, and can be independent of environmental filtering and ecological constraints that structure organism-level rules of microbial community assembly. Long-range spatial migration is likely dominated by dormant (*i.e.* spore) dispersal through aeolian deposition, though active cells and viruses also constitute the air microbiome. Local dispersal over short time scales includes viral-mediated genetic transfer as well as uptake of free environmental DNA (eDNA) from soil necromass pools. Cellular dispersal over intermediate spatial scales can occur via fungal highways or vectors including soil arthropods. Over longer time scales, dormancy shapes population genetics by effecting evolutionary diversification processes.

40 A useful synthesis describes community assembly as a function of the four
 41 fundamental ecoevolutionary processes of dispersal, selection or environmental
 42 filtering, ecological drift, and diversification (Vellend, 2010). How these
 43 same mechanisms extend to microbial biogeography has been eloquently sum-
 44 marized previously (Nemergut et al., 2013, Hanson et al., 2012, Martiny et al.,
 45 2006). Much research in the last few decades has quantified the relative con-
 46 tributions of these community assembly processes in microbial systems, see
 47 (Stegen et al., 2013, 2015, Caruso et al., 2011, Ofitseru et al., 2010, Liao et al.,
 48 2016). Of Vellend’s four fundamental processes, dispersal is the least under-

49 stood in terrestrial microbial systems and often assumed to be negligible.
50 Due to small cell size, large populations, high potential for dispersal, and
51 a bias for niche-based approaches, the influence of stochastic processes, in-
52 cluding dispersal, is historically under-explored in microbial ecology (Zhou
53 and Ning, 2017). We believe that a renewed focus of research efforts on un-
54 derstanding microbial dispersal will advance our understanding of microbial
55 structure-function relationships considerably.

56 **2. What is microbial dispersal?**

57 Dispersal is predominately defined as "the movement of individuals or
58 propagules with potential consequences for gene flow across space" (Ronce,
59 2007). Consequently, dispersal entails both dissemination and establishment
60 or colonization, each with unique constraints. But dispersal remains poorly
61 conceptualized for microbes. This is because theoretical frameworks in ecol-
62 ogy were historically built on observations of plants and animals, yet it is
63 crucial to apply, adapt, or develop theory that includes the microbial per-
64 spective (Prosser et al., 2007). While unification of micro- and macroecology
65 theory seems conceptually attainable, there remain unique aspects of mi-
66 crobial systems, including scaling, microbial species concepts, and gene flow
67 dynamics that continue to impose challenges to reconciliation (Shade et al.,
68 2018, Barberán, Casamayor and Fierer, 2014).

69 Current sampling methods limit our ability to accurately enumerate soil
70 microbes and their dispersal, though sampling challenges are certainly not
71 unique to microbes (Shade et al., 2018, Elphick, 2008). Perhaps the largest
72 hurdle to quantifying microbial dispersal is counting individuals and species
73 and identifying their presence-absence across sites. For census numbers, di-
74 rect counts using microscopy has both low feasibility and little resolutions
75 beyond basic cell morphology. Alternatively, quantification using molecu-
76 lar approaches like quantitative PCR of 16S ribosomal RNA (rRNA) genes
77 (or other marker genes) is preferred, although imperfect given biases in nu-
78 cleic acid extraction, amplification, and uneven distribution rRNA operon
79 copy numbers. More importantly, marker genes lack sufficient resolution to
80 address dispersal patterns of individual species (Choudoir et al., 2012).

81 Microbial species concepts are well supported with theoretical and empiri-
82 cal data (Achtman and Wagner, 2008, Rosselló-Móra and Amann, 2015, Ward
83 et al., 2008), but practical demarcations of microbial species remain challeng-
84 ing. Gene flow across space and time further obscures microbial population

boundaries. Gene exchange dynamics vary greatly between macro and microorganisms, making it difficult to apply macroecology dispersal theory to microbial systems. Furthermore, recombination patterns differ between microbial taxa, with microbes ranging from strictly clonal to wildly promiscuous (Gogarten et al., 2002, Didelot and Maiden, 2010, Jain et al., 2002). This is not to say that quantifying microbial dispersal is unattainable, but it does require careful experimental design and appropriate cultivation-based and/or molecular methods.

2.1. Modes of microbial dispersal

We propose classifying microbial dispersal into three categories, each with unique microbial trait associations and spatiotemporal dynamics: vegetative or active cells, dormant cells, and acellular or genetic dispersal (Figure 1). This conceptual framework is intended to better integrate microbial dispersal outcomes into community structure-function relationships. We note that molecular ecology methods (*e.g.* 16S rRNA gene amplicon surveys, shotgun metagenomics, whole genome sequencing of isolates) are often exclusively used to infer patterns of cellular dispersal. Thus, we encourage moving away from a strictly cellular framework and towards thinking about dispersal and its consequences for ecosystem function in terms of genetic dispersal, since ultimately, genes underlie functional potential. Finally, we acknowledge that these categories are not mutually exclusive, sometimes overlap, and exist on a multi-dimensional continuum.

2.1.1. Vegetative dispersal

Vegetative dispersal is the movement of growing, physiologically-active microbial cells across space. Vegetative dispersal in soils can be passive or active and occurs at cellular, micro-habitat, and local spatial scales. Sporadic wetting events that saturate soils can induce passive cell dispersal via Brownian motion (Mitchell and Kogure, 2006), but most of the time the soil is an unsaturated and irregular matrix of solid particles and liquids connected by gaseous pores (Or et al., 2007). From the perspective of a single microbial cell the soil is cavernous, and movement across this habitat requires some evolutionary ingenuity. Bacteria and archaea have evolved diverse methods of motility and active dispersal across surfaces including flagellar and non-flagellar swimming, twitching, or gliding mechanisms (Jarell and McBride, 2008). Social microbes have evolved multi-cellular modes of dispersal like biofilm or fruiting body formation (Harshey, 2003). Hyphal

121 growth in filamentous microorganisms, including some fungi and bacterial
122 actinomycetes, is another form of dispersal that creates mycelial networks
123 (Prosser and Tough, 1991). In addition to a filamentous developmen-
124 tal stage, some *Streptomyces* bacteria assume a newly discovered life stage
125 termed "exploratory growth" which allows cells to rapidly transverse surfaces
126 in response to environmental or biotic signals (Jones and Elliot, 2017, Jones
127 et al., 2017). Although similar in structure to filamentous bacterial hyphae,
128 fungal hyphae are much larger, and in fact bacteria can migrate along these
129 fungal highways (Kohlmeier et al., 2005, Warmink et al., 2011).

130 *2.1.2. Dormant dispersal*

131 Dormant dispersal is the movement of dormant microbial cells across
132 space. Dormancy is an organism's ability to reduce cell function to the min-
133 imum allowable energy expenditure, defined as maintenance energy (Pirt,
134 1987). Therefore, conduits of dormant dispersal are, by definition, passive.
135 Dormancy is reversible, which permits survival during periods of unfavorable
136 environmental conditions. In macroecology, the "temporal storage effect"
137 refers to a mechanism that contributes to species coexistence and depends
138 on varying environmental conditions, competition, and a persistent long-
139 lived state (Chesson and Warner, 1981, Warner and Chesson, 1985). In this
140 sense, we can also conceptualize microbial dormancy as the dispersal of cells
141 through time as well as space. Dormancy has recurrently evolved among
142 microorganisms, manifesting in diverse physiologies which may include mor-
143 phological differentiation and formation of spores, endospores, conidia, cysts,
144 or akinetes (Lennon and Jones, 2011). Dormancy is also surmised to include
145 "resting states" in which minimal energy is invested only in stopping cell
146 damage or decay, and is usually accompanied by a reduction in size, some-
147 times called viable-but-not-cultivable (Roszak and Cowell, 1987, Lennon and
148 Jones, 2011). For fungi engaging in sexual reproduction, dispersal of both
149 sexual and asexual spores may be crucial for successful establishment and
150 range expansion.

151 Soil microbes continuously fluctuate between active and dormant phys-
152 iological states (Stenström et al., 2001), and these varying stages of rest-
153 ing states is exemplified by the wide diversity of soil microbes that respond
154 within minutes to the first season's rain in a Mediterranean grassland (Pla-
155 cella et al., 2012). This is a demonstration of the taphonomic gradient (Lynch
156 and Neufeld, 2015), an idea which suggests that cellular metabolic state is
157 not a dichotomy of "active" or "dormant", but that microbial activity falls

158 along a gradient from active to dormant to fossilized. Measurements of soil
159 microbes being dormant at any one time range from most (Lennon and Jones,
160 2011) to almost none (Papp et al., 2018). Modern estimates of dormancy in
161 soils are largely based on the detection of rRNA, and in fact, many papers use
162 the absence of rRNA as an indication of dormancy (Aanderud et al., 2016,
163 Loeppmann et al., 2018, Kearns et al., 2016). The use of rRNA as a proxy
164 for active populations is problematic, as not all taxa degrade their rRNA as
165 they move into dormancy (Blazewicz et al., 2013). This means that certain
166 taxa will retain rRNA even when dormant, which can create a stochastic,
167 or worse, phylogenetically-conserved bias in discriminating between dormant
168 and active microbes. In other words, since dormant cells can include rRNA,
169 the use of rRNA as an indicator of an active state will under-estimate the
170 dormant population in natural systems.

171 *2.1.3. Acellular dispersal*

172 Genetic dispersal is the movement of genes across space that can be inde-
173 pendent of cellular dispersal. Acellular dispersal can facilitate the expansion
174 of functional capabilities with ecosystem-level ramifications. For example,
175 genetic dispersal has long been observed for antibiotic resistance genes (Zhu
176 et al., 2019) and microbial virulence factors (Wagner and Waldor, 2002).
177 Viruses are ubiquitous with microbes and are a major source of genetic di-
178 versity in natural systems (Correa et al., 2021). Viral-mediated horizontal
179 gene exchange creates a model of dispersal that, while dependent on cellular
180 machinery for replication and transmission, possesses unique spatial and tem-
181 poral dynamics. New research has demonstrated that viruses are agents of
182 genetic diversity that shape biogeochemical cycling (Starr et al., 2019, Trubl
183 et al., 2018). Viruses direct carbon flows in ecosystems through a top-down
184 manner, in which viral cell lysis increases organic matter concentrations.
185 The 'viral shunt' as a source of fresh organic matter from viral predation has
186 long been appreciated in marine systems, but is also important in terrestrial
187 systems (Hungate et al., 2021). In a study of viral sequences from across
188 a permafrost thaw gradient, authors found that many viruses encoded gly-
189 coside hydrolases, some with confirmed activity, targeted at degradation of
190 pectin, hemicellulose and starch. Further, modeling revealed that in almost
191 every case viral abundance predicted pore water dissolved organic carbon,
192 sometimes better than the host abundance (Emerson et al., 2018). Though
193 evidence for viral-mediated genetic dispersal in soils remains somewhat lim-
194 iting, there is clear precedent for viruses to act as agents of dispersal of genes

195 that can shape the functional capacity of soil microbial communities.

196 Extracellular relic DNA is abundant and stable in soils (Carini et al.,
197 2016, Lennon et al., 2018), representing a large reservoir of genetic diversity
198 uncoupled from cellular identity. Furthermore, transformation of free envi-
199 ronmental DNA (eDNA) by naturally competent soil bacteria (Paget and
200 Simonet, 1994) may represent an under appreciated mechanism of gene flow
201 and introduction of new heritable traits in soil populations. While acellu-
202 lar dispersal is not unique to microbes, it likely plays a much larger role in
203 microbial ecology than it does in plant or animal ecology.

204 *2.2. Long-distance dispersal*

205 Aeolian deposition, or dispersal promoted by the action of wind, can
206 cause dramatic changes in immigration rates of microbes in natural environ-
207 ments. Microbes from terrestrial, marine, and glacial origins were found in
208 the Arctic air microbiome (Šantl-Temkiv et al., 2018), indicating that the
209 atmosphere represents a potentially important channel connecting Earth’s
210 biospheres. Aerial dispersal shapes fungal community structure at local
211 scales (500 m) with strong seasonal trends (Adams et al., 2013). At con-
212 tinental scales, regional climactic and environmental variables shape the
213 distribution of bacterial and fungal taxa associated with settled dust (Bar-
214 berán et al., 2015). Wind and weather patterns have been connected to
215 microbial migration at global scales (Kellogg and Griffin, 2006, Smith et al.,
216 2013), and in particular, microbes on dust particles originating from seasonal
217 desert storms are associated with transoceanic and intercontinental airborne
218 dispersal routes (Kellogg and Griffin, 2006, Barberán, Henley, Fierer and
219 Casamayor, 2014). Functional attributes related to dormancy are enriched
220 in desert microbes (Fierer et al., 2012), supporting the hypothesis that air-
221 borne dispersal is dominated by dormant cells.

222 Atmospheric viral transmission of genetic material is possible considering
223 estimates of viral particles in the air microbiome. By one account, viral-like
224 particles and bacterial-like particles exist at concentrations of about 10e5
225 per cubic meter of air, with similar concentrations inside and outside, and
226 a viral to bacterial ratio of about 1.4–1 (Prussin et al., 2015). The enu-
227 meration of bacterial and viral particles based on size may have resulted
228 in over-estimating their abundances, but reliably quantifying airborne biotic
229 particles is notoriously difficult (Judith et al., 2020). Further, it is unclear
230 whether the viral constituents of the air microbiome are mostly human-
231 derived, or whether the focus on human health has biased this estimation

232 (Prussin and Marr, 2015). Most of the work on the viral component of air
233 microbiomes is focused on the built (*i.e.* indoor) environment with an effort
234 to quantify pathogens, so the natural ecology of outdoor particles and their
235 dispersal constraints remain under-explored.

236 *2.3. Vector-mediated dispersal*

237 Finally, microbes can disperse through animals vectors across varying
238 spatial scales. Across intermediate to long-range distances, small mammals
239 and birds are dispersers of arbuscular mycorrhizal (AM) spores (Correia
240 et al., 2019, Mangan and Adler, 2000). At local scales, it's long been ap-
241 preciated that soil arthropods assist fungal and bacterial dispersal (Ruddick
242 and Williams, 1972, Lussenhop, 1992). A recent study demonstrates that
243 geosmin, a volatile compound emitted by sporulating actinomycetes that
244 smells like fresh soil after the rain, recruited arthropods and facilitated spore
245 dispersal (Becher et al., 2020). Soil arthropods *Collembolans* accelerated the
246 dispersal of antibiotic resistance genes in a controlled experiment, likely in-
247 directly as a result of altered bacterial community structure in Collembolan-
248 inhabited soils (Zhu et al., 2019).

249 *2.4. Consequences of microbial dispersal*

250 Dispersal is a key ingredient for spatial structuring of genetic diversity and
251 population structure. Dispersal is also a unique mechanism as it impacts both
252 ecological (Stegen et al., 2015) and evolutionary (Thompson and Fronhofer,
253 2019) processes. Dispersal connects local populations with regional pools,
254 and thus dispersal is the important glue connecting metacommunities and
255 facilitating metacommunity dynamics (*e.g.* patch dynamics, species-sorting,
256 and mass effects) (Leibold et al., 2004). For instance, cellular dispersal can
257 influence community ecology by altering local abundance and distribution
258 patterns of community members. As an evolutionary force related to gene
259 flow, dispersal and can increase local diversity through the introduction of
260 novel genetic material or can homogenize genetic diversity at high dispersal
261 rates due to mass effects.

262 The prevalence of non-random distributions of bacterial species supports
263 the idea that dispersal limitation is an important factor shaping community
264 assembly (Martiny et al., 2006). Dispersal limitation refers to geographic
265 or ecological constraints of dispersal, and in some cases can create distance-
266 decay relationships. Distance-decay relationships are observed in patterns of

267 soil microbial community composition and structure across geographic dis-
268 tances ranging from micro to local to global scales (Albright and Martiny,
269 2018, Peay et al., 2007, Martiny et al., 2006). The taxa-area relationship
270 is another illustration of dispersal limitation (Horner-Devine et al., 2004,
271 Green and Bohannan, 2006). For instance, isolation by distance (IBD) de-
272 scribes a linear relationship between genetic variation and geographic dis-
273 tance (Wright, 1943), and this pattern is observed in the population structure
274 of the soil microbe *Myxococcus xanthus* (Vos and Velicer, 2008). Biogeogra-
275 phy studies in other microbial systems highlight the importance of dispersal
276 limitation on spatial structuring of genetic and genomic diversity (Reno et al.,
277 2009, Peay et al., 2010, Andam et al., 2016, Bottos et al., 2018).

278 3. Dispersal and its dependencies

279 The outcomes of dispersal on community function are interdependent
280 on environmental filtering and dormancy dynamics acting at dispersal loca-
281 tions. Stronger environmental filtering reduces perceived rates of dispersal
282 and shifts dispersal outcomes from more stochastic to more deterministic.
283 Dormancy can mitigate environmental selection in heterogeneous or chang-
284 ing habitats, effectively increasing perceived rates of dispersal. In this way,
285 environmental filtering and dormancy are opposing constraints related to dor-
286 mancy in community assembly processes. However, the variables that dictate
287 dispersal outcomes on soil community composition are still not mapped out
288 to an extent that will facilitate prediction of structure-function relationships
289 in soil.

290 3.1. Dispersal and environmental filtering

291 The Baas Becking hypothesis, "Everything is everywhere, but the envi-
292 ronment selects" (Translated from the original Dutch: "Alles is overal: maar
293 het milieu selecteert") (O'Malley, 2007) has persisted since its publication in
294 the 1930s because of our continued and growing appreciation for microbial
295 biodiversity and the rare biosphere, with modern high throughput methods
296 still not plumbing the depths of the microbial species catalogue (Lynch and
297 Neufeld, 2015). This hypothesis has been rejected (Papke et al., 2003, Telford
298 et al., 2006) and accepted (Finlay, 2002, Finlay and Fenchel, 2004) for various
299 ecosystems, scales, and populations. At its heart, the Baas Becking hypothe-
300 sis is a direct test of the relative contributions of dispersal and environmental

301 selection in determining patterns of biogeography. Spatial scale plays an im-
302 portant role, with niche selection functioning at smaller scales and dispersal
303 at broader scales (Wisnioski et al., 2019). The hypothesis of cosmopolitan
304 dispersal has been recently evaluated for genes, and authors found that gene
305 pools show stronger evidence of environmental filtering and lower geographic
306 constraints compared to whole organisms (Fodelianakis et al., 2019).

307 In a study modeling the interaction between dispersal rates and envi-
308 ronmental filtering on microbial communities assembled on different litter
309 qualities, dispersal limitation (defined as less than 25 percent turnover) re-
310 sulted in high within-group and between-group distances, suggesting a preva-
311 lence of stochastic processes (Evans et al., 2017). Community distance de-
312 creased in simulations with higher dispersal rates, yet stochastic assembly
313 was more prevalent under conditions of stronger selection, highlighting an
314 important relationship between selection and dispersal. Conversely, drought
315 stress shifted microbial community assembly to more deterministic processes
316 (Chase, 2007). Under scenarios of environmental stress, we can imagine how
317 the consequences of dispersal will also depend on what microbes and their
318 associated traits are dispersing, their relative fitness, and their adaptive po-
319 tentials.

320 *3.2. Dispersal and dormancy*

321 An accurate estimate of microbial dormancy in soils is critical to under-
322 standing how community assembly processes shape soil biodiversity and to
323 extrapolating the impact of dispersal on community function. Seed banks
324 constructed of dormant microorganisms, many of which are members of
325 the rare biosphere, are important contributors to generating and maintain-
326 ing soil microbial diversity (Jones and Lennon, 2010, Lennon and Jones,
327 2011, Aanderud et al., 2015). Furthermore, ecosystem models indicate dor-
328 mancy dynamics are important for predicting biogeochemical nutrient cycling
329 (Stolpovsky et al., 2011, Wang et al., 2015). Dormancy also has the poten-
330 tial to shape population genetics and fundamental evolutionary processes
331 (Shoemaker and Lennon, 2018).

332 There is a strong theoretical grounding for the hypothesis that dormancy
333 shapes patterns of microbial biogeography by enhancing dispersal, but empir-
334 ical evidence has been harder to come by (Epstein, 2009). Mestre and Höfer
335 (Mestre and Höfer, 2020) outline a compelling conceptual framework, the
336 Microbial Conveyor Belt, for surmising how dormancy, dispersal, and resus-
337 citation interact to shape marine microbial community structure and function

338 at the global scale. The Theory of Island Biogeography likewise indirectly
339 supports the link between dormancy and dispersal, where modeling exercises
340 show that increasing the dormancy rate (expressed as a dampening of extinc-
341 tion rates over time) increases community richness (Lennon and Jones, 2011).
342 Some of the first direct evidence linking microbial dormancy and dispersal
343 limitation to microbial biogeography shows that dormancy dampens envi-
344 ronmental and spatial distance-decay relationships for microbes in forested
345 ponds (Locey et al., 2020). Another recent study found that both resusci-
346 tation of local dormant cells and regional dispersal of active cells contribute
347 to soil community resilience following a period of thermal stress (Sorensen
348 and Shade, 2020). This study poses the question, what are the long-term
349 outcomes of dormant versus active cellular dispersal in natural systems?

350 *3.3. Dispersal outcomes on community function*

351 The outcomes of dispersal on community function depend on the inter-
352 play between microbial traits associated with dispersers and the strength of
353 local environmental filtering. Dispersal-colonization tradeoffs may structure
354 microbial trait distributions across the spatial and environmental landscape
355 (Smith et al., 2018). In a recent wood decomposition study spanning sites
356 along a forest/non-forest ecotone, dispersal limitation of traits associated
357 with rapid wood-degradation shaped community composition and function
358 such that fungal communities farther from forests decomposed wood blocks
359 more slowly (Smith and Peay, 2021). Independent of forest proximity, there
360 was also a significant negative relationship between alpha-diversity (shaped
361 by stochastic dispersal) and decomposition due to interspecific competition,
362 linking dispersal to independent drivers of community function in this system
363 (Smith and Peay, 2021).

364 An intuitive hypothesis is that dispersal can mitigate microbial responses
365 to environmental stress by introducing stress-tolerant microbes, but this pre-
366 diction depends on the regional pool of microbial traits, their adaptive poten-
367 tial, and the extent of functional redundancy. In an experimental evolution
368 experiment, dispersal elevated community growth under ambient conditions
369 but hindered growth in a warming treatment (Lawrence et al., 2016), suggest-
370 ing that dispersal may dampen the ability of microbial communities to adapt
371 to environmental change by introducing maladapted individuals. In another
372 study looking at the interaction between dispersal and drought, dispersal
373 altered the community composition to a greater extent under drought con-
374 ditions but also resulted in loss of community function, which was contrary

375 to the hypothesis that dispersal could mitigate drought stress by introducing
376 tolerant microbes (Evans et al., 2020). In a common garden experiment
377 across a natural precipitation gradient, enhanced dispersal had no effect on
378 community composition, which was the strongest predictor of functional re-
379 sponds to changes in moisture (Waring and Hawkes, 2018). However, under
380 certain conditions, dispersal can enhance community stability in the face of
381 environmental change for both acute and more gradual disturbances (Evans
382 et al., 2019, Sorensen and Shade, 2020).

383 The order and timing of dispersal events can also influence dispersal out-
384 comes. For instance, the release of fungal spores during day versus night
385 influences dispersal longevity and survival (Oneto et al., 2020). Historical
386 contingencies are past biological interactions or environmental conditions,
387 whose order and timing impact the trajectory of a community response. Pri-
388 ority effects are a specific example of a biotic historical contingency where
389 the early or late arrival of a species determines community assembly out-
390 comes (Fukami, 2015). The important of historical contingencies (Hawkes
391 and Keitt, 2015) and priority effects (Sprockett et al., 2018, Hiscox et al.,
392 2015, Svoboda et al., 2018) on microbial community assembly has been
393 demonstrated across a range of habitats.

394 **4. Traits-based approach for predicting dispersal outcomes**

395 Traits are increasingly invoked as the key parameters to understand ecosys-
396 tem function. Traits include the physiological, life history, and behavioral
397 characteristics of organisms that underlie ecosystem function (Martiny et al.,
398 2015). Because traits more directly relate to ecosystem function, and most
399 traits are phylogenetically conserved to some degree, traits are a valuable
400 tool in linking microbial biogeography to ecosystem function (Green et al.,
401 2008, Nelson et al., 2016, Fierer et al., 2012). Quantification of traits related
402 to dormancy and dispersal should also be valuable to understanding their
403 interaction, but current attempts are hampered by the breadth of traits that
404 contribute to these processes.

405 For example, range size correlates to genomic and phenotypic attributes
406 of dust-associated microbes, suggesting that these traits may be related to
407 dispersal capabilities (Choudoir et al., 2018). For AM fungi, a recent study
408 showed that small spore size was positively associated with aerial dispersal
409 (Chaudhary et al., 2020), while another study found spore size to be a poor
410 predictor of AM fungal range size (Kivlin, 2020). It's clear we are far from

411 understanding physiological traits that determine dispersal outcomes. Since
412 atmospheric dispersal is important to both plants and microbes, looking to
413 decades of studies in plant ecology for inspiration about traits related to
414 dispersal will likely yield fruitful insights (Thomson et al., 2010, Tamme
415 et al., 2014). Once traits are identified, analyses developed for genome-wide
416 association studies (GWAS) (Eriksson et al., 2010, San et al., 2020) may
417 offer useful insights for identifying genetic variation related to common traits
418 associated with dormancy and/or dispersal.

419 Ultimately, we need to develop a predictive framework for implementing
420 dispersal traits into changes in ecosystem function. One suggestion is im-
421 plementation of the response-effect framework (Lavorel and Garnier, 2002),
422 where response traits determine community structure (indirect drivers) and
423 effect traits influence ecosystem function (direct drivers). This framework has
424 been previously applied to fungal systems (Crowther et al., 2014, Koide et al.,
425 2014), and while it can be challenging to parse indirect versus direct effects
426 on ecosystem function, there is predictive power when response and effect
427 traits are correlated. Using this framework, microbial dormancy and disper-
428 sal traits are response traits which control microbial community structure
429 directly (and ecosystem function indirectly), as separate from effect traits
430 that govern ecosystem function directly. Another approach could implement
431 a tradeoff framework, such as the yield-resource acquisition-stress (Y-A-S)
432 traits framework developed as a microbial analog to Grime's competitor-
433 stress tolerator-ruderal (C-S-R) framework (Grime, 1977, Malik et al., 2020).
434 For example, dormancy could be invoked as a measure of community stress
435 response. A third approach could implement dormancy or dispersal as a
436 performance filter along an ecological gradient overlaying other system traits
437 (Webb et al., 2010). These frameworks could be high-level conceptualizations
438 to be combined with statistical modeling. For example, the relative contri-
439 bution of dispersal to ecosystem function can be estimated using generalized
440 nonlinear models, with microbial traits as potential fixed effects screened
441 in model selection. Structural equation modeling (SEM) can be employed
442 downstream to define direct and indirect drivers of ecosystem function.

443 **5. Road map and research recommendations**

444 To improve our predictions of structure-function relationships in soils, we
445 need to apply and evaluate a more precise, yet adaptable conceptualization
446 of microbial dispersal. We propose a reframing of microbial dispersal into

447 active, dormant, and acellular modes. The ecological, spatial, and temporal
448 restraints vary between cellular and acellular dispersal (Figure 1), with genetic
449 dispersal potentially decoupled from environmental filtering and organismal
450 identity. We are not the first to frame patterns of biogeography through the
451 lens of genetic dispersal. For example, Baltrus (Baltrus, 2020) discusses the
452 dynamics of genetic element dispersal and the contributions to the biogeog-
453 raphy microbial pathways. We also encourage researchers to contemplate the
454 limits of this thinking. Can microbial dispersal always be sufficiently repre-
455 sented as genetic dispersal, or is it sometimes more important to consider
456 the individual organisms harboring these genes?

457 To close current knowledge gaps, we recommend starting with these re-
458 search directions. First, we need to develop a quantitative theoretical frame-
459 work that integrates microbial dispersal, dormancy, and environmental fil-
460 tering. A new model describes the interactions between dispersal and dor-
461 mancy and outcomes on community diversity across scales (Wisniski and
462 Shoemaker, 2021). This model considers multiple dormancy traits (*i.e.* sur-
463 vival and germination rates) and how these processes interact with dispersal
464 to create nonlinear effects on metacommunity diversity across local and re-
465 gional spatial scales. Models that capture dispersal and dormancy dynamics
466 will continue to improve as we better measure these phenomena, both as
467 ecosystem processes and as microbial traits.

468 Second, we need to develop more accurate methods for quantifying, and
469 accounting for, microbial dispersal and dormancy. Golan and Pringle (Golan
470 and Pringle, 2017) provide a comprehensive framework for considering fungal
471 long distance dispersal that entails mathematical models, genetic inference,
472 and direct quantification based on spore capture. We also need improved
473 tools to quantify genetic dispersal (Brito, 2021). We recommend incorpo-
474 rating dispersal and dormancy explicitly into soil structure-function studies.
475 A typical structure-function analysis neglects the influence of dispersal from
476 a regional pool, and also assumes that all recovered DNA sequences are
477 representative of active (or potentially active) organisms. Incorporating dis-
478 persal might mean a no-dispersal or enhanced-dispersal treatment as part of
479 the experimental design, or accounting for new taxa from the atmosphere
480 at regional scales or along fungal highways at local scales. Incorporating
481 dormancy might mean including a resuscitation treatment (*e.g.* bacterial
482 resuscitation factor Rpf, see (Kuo et al., 2021)), or filtering taxa based on
483 microbial activity. Although at present our methods for quantifying active
484 versus non-active fractions are imperfect. If we knew how to differentiate ac-

485 tive versus inactive rRNA, we'd be a lot closer to estimating true dormancy
486 rates. Cell-resolved metabolomics might offer a promising solution (Walsh
487 et al., 2018).

488 Finally, we need to better qualify and quantify traits related to dormancy
489 and dispersal. This may require different strategies for different microbial
490 lineages. For example, identifying a taxon-specific sporulation gene, or an
491 environmental signal that is associated with entering dormancy or the re-
492 suscitation of a particular microbe. New research on the homeostasis of
493 ribosomes in *Methanococcus* during energy limitation underscores the need
494 for alternative traits that accompany activity and dormancy (Müller et al.,
495 2021). We as microbial ecologists need to support research that focuses on
496 specific members of a microbial community (e.g. revitalization of microbial
497 cultivation efforts (Carini, 2019)). In understanding the trees, we may finally
498 be able to see the forest.

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503 **7. Author Contributions**

504 M.J.C and K.M.D conceptualized and investigated the perspective. M.J.C
505 and K.M.D wrote and revised the manuscript.

506 **8. Declaration of Interests**

507 The authors declare no competing interests.

508 **9. Inclusion and Diversity**

509 One or more of the authors of this paper self-identifies as a member of the
510 LGBTQ+ community. While citing references scientifically relevant for this
511 work, we also actively worked to promote gender balance in our reference
512 list.

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