

# A framework for integrating microbial dispersal modes into soil ecosystem ecology

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## Abstract

Dispersal is a fundamental community assembly process that maintains soil microbial biodiversity across spatial and temporal scales, yet the impact of dispersal on ecosystem function is largely unpredictable. Dispersal is unique in that it contributes to both ecological and evolutionary processes and is shaped by both deterministic and stochastic forces. The ecosystem-level ramifications of dispersal outcomes are further compounded by microbial dormancy dynamics and environmental selection. Here we review the knowledge gaps and challenges that remain in defining how dispersal, environmental filtering, and microbial dormancy interact to influence the relationship between microbial community structure and function in soils. We propose the classification of microbial dispersal into three categories, through vegetative or active cells, through dormant cells, and through acellular dispersal, each with unique spatiotemporal dynamics and microbial trait associations. This conceptual framework should improve the integration of dispersal in defining soil microbial community structure-function relationships.

*Keywords:* soil microbiology, dispersal, environmental filtering, dormancy, community assembly

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## 1. Introduction

2 The interplay between microbial dispersal, environmental filtering, and  
3 microbial dormancy introduces ecoevolutionary dynamics to soil ecosystems

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4 that limit our ability to decipher, much less predict, community structure-  
5 function relationships. Despite rapid development of high throughput molec-  
6 ular methods, decreasing sequencing costs, and accelerating generation of  
7 large ecological data, soils remain messy. This is because soils are massively  
8 complex superorganisms with emergent functions that are not yet easily pre-  
9 dicted. To inform our predictions of amassed scale effects that determine soil  
10 behavior, we need a better understanding of how microbial processes, such  
11 as dormancy and environmental filtering, compound dispersal outcomes that  
12 transpire to whole ecosystems.

13 We suggest an adaptable framework for thinking about how microbial  
14 dispersal across space and time influences soil biodiversity, and ultimately,  
15 ecosystem function. Modes of microbial dispersal are categorized as cellular  
16 (*i.e.* vegetative or dormant cells) and acellular (*i.e.* genetic material as-  
17 sociated with viruses and/or gene flow independent of cellular life), where  
18 dispersal outcomes for organisms and/or their genes happen across different  
19 spatiotemporal scales (Figure 1). Here we discuss the mechanistic constraints  
20 of microbial dispersal modes and the interplay between dispersal, environ-  
21 mental filtering, and dormancy. Finally, we propose a traits-based approach  
22 for quantifying dispersal outcomes, and suggest how this framework can be  
23 used to evaluate soil microbial structure-function relationships.

### 24 1.1. *Microbial community assembly*

25 To predict ecosystem function from community composition, we first need  
26 to understand the community assembly processes that create and maintain  
27 patterns of microbial diversity. In both microbial ecology and the broader  
28 field of ecology, niche theory and selection-based models have classically ex-  
29 plained patterns of community assembly, looking to environmental selection  
30 and biotic interactions to define niche space and determine what conditions  
31 a species will persist (Chase and Leibold, 2003, Holt, 2009). Alternatively,  
32 neutral theory relies on stochastic processes to explain community ecology  
33 patterns (Hubbell, 2001, Chave, 2004). Few elements of ecology are an ab-  
34 solute either-or, and mathematical frameworks unify both niche and neutral  
35 theory (Harshey, 2003, Mutshinda and O’Hara, 2011). In reality, compre-  
36 hensive theory explains that variations in community assembly arise through  
37 both deterministic and stochastic processes, and that individual processes  
38 exist somewhere along a continuum between selection and neutrality (Chase  
39 and Myers, 2011).

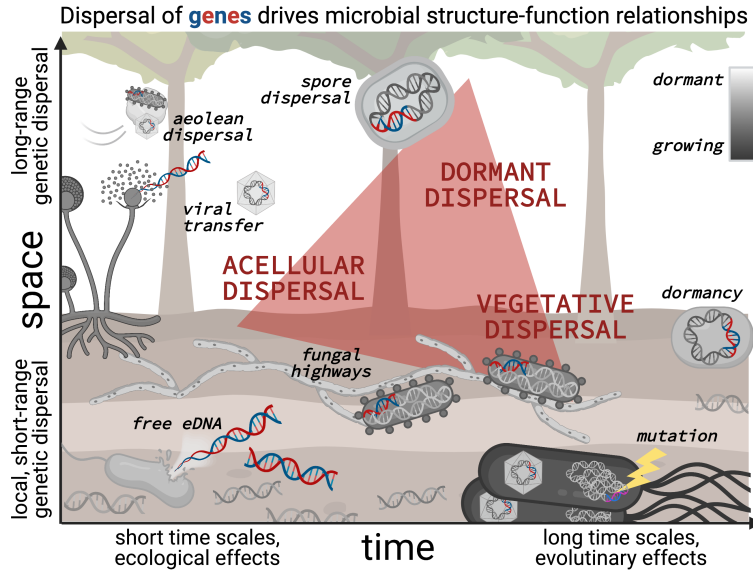


Figure 1: Microbial dispersal modes. Dispersal affects microbial structure-function relationships in soils by distributing genes (*i.e.* potential ecological functions, which are represented here by blue/red DNA fragments) in three primary ways: through active or vegetative cell dispersal, through dormant cell dispersal, and through acellular dispersal. Dispersal can occur at different scales over space and time, and can be independent of environmental filtering and ecological constraints that structure organism-level rules of microbial community assembly. Long-range spatial migration is likely dominated by dormant (*i.e.* spore) dispersal through aeolian deposition, though active cells and viruses also constitute the air microbiome. Local dispersal over short time scales includes viral-mediated genetic transfer as well as uptake of free environmental DNA (eDNA) from soil necromass pools. Cellular dispersal over intermediate spatial scales can occur via fungal highways or vectors including soil arthropods. Over longer time scales, dormancy shapes population genetics by effecting evolutionary diversification processes.

40 A useful synthesis describes community assembly as a function of the four  
 41 fundamental ecoevolutionary processes of dispersal, selection or environmen-  
 42 tal filtering, ecological drift, and diversification (Vellend, 2010). How these  
 43 same mechanisms extend to microbial biogeography has been eloquently sum-  
 44 marized previously (Nemergut et al., 2013, Hanson et al., 2012, Martiny et al.,  
 45 2006). Much research in the last few decades has quantified the relative con-  
 46 tributions of these community assembly processes in microbial systems, see  
 47 (Stegen et al., 2013, 2015, Caruso et al., 2011, Ofiteru et al., 2010, Liao et al.,  
 48 2016). Of Vellend’s four fundamental processes, dispersal is the least under-

49 stood in terrestrial microbial systems and often assumed to be negligible.  
50 Due to small cell size, large populations, high potential for dispersal, and  
51 a bias for niche-based approaches, the influence of stochastic processes, in-  
52 cluding dispersal, is historically under-explored in microbial ecology (Zhou  
53 and Ning, 2017). We believe that a renewed focus of research efforts on un-  
54 derstanding microbial dispersal will advance our understanding of microbial  
55 structure-function relationships considerably.

## 56 **2. What is microbial dispersal?**

57 Dispersal is predominately defined as "the movement of individuals or  
58 propagules with potential consequences for gene flow across space" (Ronce,  
59 2007). Consequently, dispersal entails both dissemination and establishment  
60 or colonization, each with unique constraints. But dispersal remains poorly  
61 conceptualized for microbes. This is because theoretical frameworks in ecol-  
62 ogy were historically built on observations of plants and animals, yet it is  
63 crucial to apply, adapt, or develop theory that includes the microbial per-  
64 spective (Prosser et al., 2007). While unification of micro- and macroecology  
65 theory seems conceptually attainable, there remain unique aspects of mi-  
66 crobial systems, including scaling, microbial species concepts, and gene flow  
67 dynamics that continue to impose challenges to reconciliation (Shade et al.,  
68 2018, Barberán, Casamayor and Fierer, 2014).

69 Current sampling methods limit our ability to accurately enumerate soil  
70 microbes and their dispersal, though sampling challenges are certainly not  
71 unique to microbes (Shade et al., 2018, Elphick, 2008). Perhaps the largest  
72 hurdle to quantifying microbial dispersal is counting individuals and species  
73 and identifying their presence-absence across sites. For census numbers, di-  
74 rect counts using microscopy has both low feasibility and little resolutions  
75 beyond basic cell morphology. Alternatively, quantification using molecu-  
76 lar approaches like quantitative PCR of 16S ribosomal RNA (rRNA) genes  
77 (or other marker genes) is preferred, although imperfect given biases in nu-  
78 cleic acid extraction, amplification, and uneven distribution rRNA operon  
79 copy numbers. More importantly, marker genes lack sufficient resolution to  
80 address dispersal patterns of individual species (Choudoir et al., 2012).

81 Microbial species concepts are well supported with theoretical and empiri-  
82 cal data (Achtman and Wagner, 2008, Rosselló-Móra and Amann, 2015, Ward  
83 et al., 2008), but practical demarcations of microbial species remain challeng-  
84 ing. Gene flow across space and time further obscures microbial population

boundaries. Gene exchange dynamics vary greatly between macro and microorganisms, making it difficult to apply macroecology dispersal theory to microbial systems. Furthermore, recombination patterns differ between microbial taxa, with microbes ranging from strictly clonal to wildly promiscuous (Gogarten et al., 2002, Didelot and Maiden, 2010, Jain et al., 2002). This is not to say that quantifying microbial dispersal is unattainable, but it does require careful experimental design and appropriate cultivation-based and/or molecular methods.

### 2.1. *Modes of microbial dispersal*

We propose classifying microbial dispersal into three categories, each with unique microbial trait associations and spatiotemporal dynamics: vegetative or active cells, dormant cells, and acellular or genetic dispersal (Figure 1). This conceptual framework is intended to better integrate microbial dispersal outcomes into community structure-function relationships. We note that molecular ecology methods (*e.g.* 16S rRNA gene amplicon surveys, shotgun metagenomics, whole genome sequencing of isolates) are often exclusively used to infer patterns of cellular dispersal. Thus, we encourage moving away from a strictly cellular framework and towards thinking about dispersal and its consequences for ecosystem function in terms of genetic dispersal, since ultimately, genes underlie functional potential. Finally, we acknowledge that these categories are not mutually exclusive, sometimes overlap, and exist on a multi-dimensional continuum.

#### 2.1.1. *Vegetative dispersal*

Vegetative dispersal is the movement of growing, physiologically-active microbial cells across space. Vegetative dispersal in soils can be passive or active and occurs at cellular, micro-habitat, and local spatial scales. Sporadic wetting events that saturate soils can induce passive cell dispersal via Brownian motion (Mitchell and Kogure, 2006), but most of the time the soil is an unsaturated and irregular matrix of solid particles and liquids connected by gaseous pores (Or et al., 2007). From the perspective of a single microbial cell the soil is cavernous, and movement across this habitat requires some evolutionary ingenuity. Bacteria and archaea have evolved diverse methods of motility and active dispersal across surfaces including flagellar and non-flagellar swimming, twitching, or gliding mechanisms (Jarrell and McBride, 2008). Social microbes have evolved multi-cellular modes of dispersal like biofilm or fruiting body formation (Harshey, 2003). Hyphal

121 growth in filamentous microorganisms, including some fungi and bacterial  
 122 actinomycetes, is another form of dispersal that creates mycelial networks  
 123 (Prosser and Tough, 1991). In addition to a filamentous developmen-  
 124 tal stage, some *Streptomyces* bacteria assume a newly discovered life stage  
 125 termed "exploratory growth" which allows cells to rapidly transverse surfaces  
 126 in response to environmental or biotic signals (Jones and Elliot, 2017, Jones  
 127 et al., 2017). Although similar in structure to filamentous bacterial hyphae,  
 128 fungal hyphae are much larger, and in fact bacteria can migrate along these  
 129 fungal highways (Kohlmeier et al., 2005, Warmink et al., 2011).

### 130 2.1.2. *Dormant dispersal*

131 Dormant dispersal is the movement of dormant microbial cells across  
 132 space. Dormancy is an organism's ability to reduce cell function to the min-  
 133 imum allowable energy expenditure, defined as maintenance energy (Pirt,  
 134 1987). Therefore, conduits of dormant dispersal are, by definition, passive.  
 135 Dormancy is reversible, which permits survival during periods of unfavorable  
 136 environmental conditions. In macroecology, the "temporal storage effect"  
 137 refers to a mechanism that contributes to species coexistence and depends  
 138 on varying environmental conditions, competition, and a persistent long-  
 139 lived state (Chesson and Warner, 1981, Warner and Chesson, 1985). In this  
 140 sense, we can also conceptualize microbial dormancy as the dispersal of cells  
 141 through time as well as space. Dormancy has recurrently evolved among  
 142 microorganisms, manifesting in diverse physiologies which may include mor-  
 143 phological differentiation and formation of spores, endospores, conidia, cysts,  
 144 or akinetes (Lennon and Jones, 2011). Dormancy is also surmised to include  
 145 "resting states" in which minimal energy is invested only in stopping cell  
 146 damage or decay, and is usually accompanied by a reduction in size, some-  
 147 times called viable-but-not-cultivable (Roszak and Cowell, 1987, Lennon and  
 148 Jones, 2011). For fungi engaging in sexual reproduction, dispersal of both  
 149 sexual and asexual spores may be crucial for successful establishment and  
 150 range expansion.

151 Soil microbes continuously fluctuate between active and dormant phys-  
 152 iological states (Stenström et al., 2001), and these varying stages of rest-  
 153 ing states is exemplified by the wide diversity of soil microbes that respond  
 154 within minutes to the first season's rain in a Mediterranean grassland (Pla-  
 155 cella et al., 2012). This is a demonstration of the taphonomic gradient (Lynch  
 156 and Neufeld, 2015), an idea which suggests that cellular metabolic state is  
 157 not a dichotomy of "active" or "dormant", but that microbial activity falls

158 along a gradient from active to dormant to fossilized. Measurements of soil  
159 microbes being dormant at any one time range from most (Lennon and Jones,  
160 2011) to almost none (Papp et al., 2018). Modern estimates of dormancy in  
161 soils are largely based on the detection of rRNA, and in fact, many papers use  
162 the absence of rRNA as an indication of dormancy (Aanderud et al., 2016,  
163 Loeppmann et al., 2018, Kearns et al., 2016). The use of rRNA as a proxy  
164 for active populations is problematic, as not all taxa degrade their rRNA as  
165 they move into dormancy (Blazewicz et al., 2013). This means that certain  
166 taxa will retain rRNA even when dormant, which can create a stochastic,  
167 or worse, phylogenetically-conserved bias in discriminating between dormant  
168 and active microbes. In other words, since dormant cells can include rRNA,  
169 the use of rRNA as an indicator of an active state will under-estimate the  
170 dormant population in natural systems.

### 171 *2.1.3. Acellular dispersal*

172 Genetic dispersal is the movement of genes across space that can be inde-  
173 pendent of cellular dispersal. Acellular dispersal can facilitate the expansion  
174 of functional capabilities with ecosystem-level ramifications. For example,  
175 genetic dispersal has long been observed for antibiotic resistance genes (Zhu  
176 et al., 2019) and microbial virulence factors (Wagner and Waldor, 2002).  
177 Viruses are ubiquitous with microbes and are a major source of genetic di-  
178 versity in natural systems (Correa et al., 2021). Viral-mediated horizontal  
179 gene exchange creates a model of dispersal that, while dependent on cellular  
180 machinery for replication and transmission, possesses unique spatial and tem-  
181 poral dynamics. New research has demonstrated that viruses are agents of  
182 genetic diversity that shape biogeochemical cycling (Starr et al., 2019, Trubl  
183 et al., 2018). Viruses direct carbon flows in ecosystems through a top-down  
184 manner, in which viral cell lysis increases organic matter concentrations.  
185 The 'viral shunt' as a source of fresh organic matter from viral predation has  
186 long been appreciated in marine systems, but is also important in terrestrial  
187 systems (Hungate et al., 2021). In a study of viral sequences from across  
188 a permafrost thaw gradient, authors found that many viruses encoded gly-  
189 coside hydrolases, some with confirmed activity, targeted at degradation of  
190 pectin, hemicellulose and starch. Further, modeling revealed that in almost  
191 every case viral abundance predicted pore water dissolved organic carbon,  
192 sometimes better than the host abundance (Emerson et al., 2018). Though  
193 evidence for viral-mediated genetic dispersal in soils remains somewhat lim-  
194 iting, there is clear precedent for viruses to act as agents of dispersal of genes

195 that can shape the functional capacity of soil microbial communities.

196 Extracellular relic DNA is abundant and stable in soils (Carini et al.,  
197 2016, Lennon et al., 2018), representing a large reservoir of genetic diversity  
198 uncoupled from cellular identity. Furthermore, transformation of free envi-  
199 ronmental DNA (eDNA) by naturally competent soil bacteria (Paget and  
200 Simonet, 1994) may represent an under appreciated mechanism of gene flow  
201 and introduction of new heritable traits in soil populations. While acellu-  
202 lar dispersal is not unique to microbes, it likely plays a much larger role in  
203 microbial ecology than it does in plant or animal ecology.

## 204 2.2. Long-distance dispersal

205 Aeolean deposition, or dispersal promoted by the action of wind, can  
206 cause dramatic changes in immigration rates of microbes in natural environ-  
207 ments. Microbes from terrestrial, marine, and glacial origins were found in  
208 the Arctic air microbiome (Šantl-Temkiv et al., 2018), indicating that the  
209 atmosphere represents a potentially important channel connecting Earth’s  
210 biospheres. Aerial dispersal shapes fungal community structure at local  
211 scales (500 m) with strong seasonal trends (Adams et al., 2013). At con-  
212 tinental scales, regional climactic and environmental variables shape the  
213 distribution of bacterial and fungal taxa associated with settled dust (Bar-  
214 berán et al., 2015). Wind and weather patterns have been connected to  
215 microbial migration at global scales (Kellogg and Griffin, 2006, Smith et al.,  
216 2013), and in particular, microbes on dust particles originating from seasonal  
217 desert storms are associated with transoceanic and intercontinental airborne  
218 dispersal routes (Kellogg and Griffin, 2006, Barberán, Henley, Fierer and  
219 Casamayor, 2014). Functional attributes related to dormancy are enriched  
220 in desert microbes (Fierer et al., 2012), supporting the hypothesis that air-  
221 borne dispersal is dominated by dormant cells.

222 Atmospheric viral transmission of genetic material is possible considering  
223 estimates of viral particles in the air microbiome. By one account, viral-like  
224 particles and bacterial-like particles exist at concentrations of about  $10^5$   
225 per cubic meter of air, with similar concentrations inside and outside, and  
226 a viral to bacterial ratio of about 1.4–1 (Prussin et al., 2015). The enu-  
227 meration of bacterial and viral particles based on size may have resulted  
228 in over-estimating their abundances, but reliably quantifying airborne biotic  
229 particles is notoriously difficult (Judith et al., 2020). Further, it is unclear  
230 whether the viral constituents of the air microbiome are mostly human-  
231 derived, or whether the focus on human health has biased this estimation



232 (Prussin and Marr, 2015). Most of the work on the viral component of air  
233 microbiomes is focused on the built (*i.e.* indoor) environment with an effort  
234 to quantify pathogens, so the natural ecology of outdoor particles and their  
235 dispersal constraints remain under-explored.

### 236 2.3. Vector-mediated dispersal

237 Finally, microbes can disperse through animals vectors across varying  
238 spatial scales. Across intermediate to long-range distances, small mammals  
239 and birds are dispersers of arbuscular mycorrhizal (AM) spores (Correia  
240 et al., 2019, Mangan and Adler, 2000). At local scales, it's long been ap-  
241 preciated that soil arthropods assist fungal and bacterial dispersal (Ruddick  
242 and Williams, 1972, Lussenhop, 1992). A recent study demonstrates that  
243 geosmin, a volatile compound emitted by sporulating actinomycetes that  
244 smells like fresh soil after the rain, recruited arthropods and facilitated spore  
245 dispersal (Becher et al., 2020). Soil arthropods *Collembolans* accelerated the  
246 dispersal of antibiotic resistance genes in a controlled experiment, likely in-  
247 directly as a result of altered bacterial community structure in Collembolan-  
248 inhabited soils (Zhu et al., 2019).

### 249 2.4. Consequences of microbial dispersal

250 Dispersal is a key ingredient for spatial structuring of genetic diversity and  
251 population structure. Dispersal is also a unique mechanism as it impacts both  
252 ecological (Stegen et al., 2015) and evolutionary (Thompson and Fronhofer,  
253 2019) processes. Dispersal connects local populations with regional pools,  
254 and thus dispersal is the important glue connecting metacommunities and  
255 facilitating metacommunity dynamics (*e.g.* patch dynamics, species-sorting,  
256 and mass effects) (Leibold et al., 2004). For instance, cellular dispersal can  
257 influence community ecology by altering local abundance and distribution  
258 patterns of community members. As an evolutionary force related to gene  
259 flow, dispersal and can increase local diversity through the introduction of  
260 novel genetic material or can homogenize genetic diversity at high dispersal  
261 rates due to mass effects.

262 The prevalence of non-random distributions of bacterial species supports  
263 the idea that dispersal limitation is an important factor shaping community  
264 assembly (Martiny et al., 2006). Dispersal limitation refers to geographic  
265 or ecological constraints of dispersal, and in some cases can create distance-  
266 decay relationships. Distance-decay relationships are observed in patterns of

soil microbial community composition and structure across geographic distances ranging from micro to local to global scales (Albright and Martiny, 2018, Peay et al., 2007, Martiny et al., 2006). The taxa-area relationship is another illustration of dispersal limitation (Horner-Devine et al., 2004, Green and Bohannan, 2006). For instance, isolation by distance (IBD) describes a linear relationship between genetic variation and geographic distance (Wright, 1943), and this pattern is observed in the population structure of the soil microbe *Myxococcus xanthus* (Vos and Velicer, 2008). Biogeography studies in other microbial systems highlight the importance of dispersal limitation on spatial structuring of genetic and genomic diversity (Reno et al., 2009, Peay et al., 2010, Andam et al., 2016, Bottos et al., 2018).

### 3. Dispersal and its dependencies

The outcomes of dispersal on community function are interdependent on environmental filtering and dormancy dynamics acting at dispersal locations. Stronger environmental filtering reduces perceived rates of dispersal and shifts dispersal outcomes from more stochastic to more deterministic. Dormancy can mitigate environmental selection in heterogeneous or changing habitats, effectively increasing perceived rates of dispersal. In this way, environmental filtering and dormancy are opposing constraints related to dormancy in community assembly processes. However, the variables that dictate dispersal outcomes on soil community composition are still not mapped out to an extent that will facilitate prediction of structure-function relationships in soil.

#### 3.1. Dispersal and environmental filtering

The Baas Becking hypothesis, "Everything is everywhere, but the environment selects" (Translated from the original Dutch: "Alles is overal: maar het milieu selecteert") (O'Malley, 2007) has persisted since its publication in the 1930s because of our continued and growing appreciation for microbial biodiversity and the rare biosphere, with modern high throughput methods still not plumbing the depths of the microbial species catalogue (Lynch and Neufeld, 2015). This hypothesis has been rejected (Papke et al., 2003, Telford et al., 2006) and accepted (Finlay, 2002, Finlay and Fenchel, 2004) for various ecosystems, scales, and populations. At its heart, the Baas Becking hypothesis is a direct test of the relative contributions of dispersal and environmental

301 selection in determining patterns of biogeography. Spatial scale plays an im-  
302 portant role, with niche selection functioning at smaller scales and dispersal  
303 at broader scales (Wisnoski et al., 2019). The hypothesis of cosmopolitan  
304 dispersal has been recently evaluated for genes, and authors found that gene  
305 pools show stronger evidence of environmental filtering and lower geographic  
306 constraints compared to whole organisms (Fodelianakis et al., 2019).

307 In a study modeling the interaction between dispersal rates and envi-  
308 ronmental filtering on microbial communities assembled on different litter  
309 qualities, dispersal limitation (defined as less than 25 percent turnover) re-  
310 sulted in high within-group and between-group distances, suggesting a preva-  
311 lence of stochastic processes (Evans et al., 2017). Community distance de-  
312 creased in simulations with higher dispersal rates, yet stochastic assembly  
313 was more prevalent under conditions of stronger selection, highlighting an  
314 important relationship between selection and dispersal. Conversely, drought  
315 stress shifted microbial community assembly to more deterministic processes  
316 (Chase, 2007). Under scenarios of environmental stress, we can imagine how  
317 the consequences of dispersal will also depend on what microbes and their  
318 associated traits are dispersing, their relative fitness, and their adaptive po-  
319 tentials.

### 320 3.2. *Dispersal and dormancy*

321 An accurate estimate of microbial dormancy in soils is critical to under-  
322 standing how community assembly processes shape soil biodiversity and to  
323 extrapolating the impact of dispersal on community function. Seed banks  
324 constructed of dormant microorganisms, many of which are members of  
325 the rare biosphere, are important contributors to generating and maintain-  
326 ing soil microbial diversity (Jones and Lennon, 2010, Lennon and Jones,  
327 2011, Aanderud et al., 2015). Furthermore, ecosystem models indicate dor-  
328 mancy dynamics are important for predicting biogeochemical nutrient cycling  
329 (Stolpovsky et al., 2011, Wang et al., 2015). Dormancy also has the poten-  
330 tial to shape population genetics and fundamental evolutionary processes  
331 (Shoemaker and Lennon, 2018).

332 There is a strong theoretical grounding for the hypothesis that dormancy  
333 shapes patterns of microbial biogeography by enhancing dispersal, but empir-  
334 ical evidence has been harder to come by (Epstein, 2009). Mestre and Höfer  
335 (Mestre and Höfer, 2020) outline a compelling conceptual framework, the  
336 Microbial Conveyor Belt, for surmising how dormancy, dispersal, and resus-  
337 citation interact to shape marine microbial community structure and function

at the global scale. The Theory of Island Biogeography likewise indirectly supports the link between dormancy and dispersal, where modeling exercises show that increasing the dormancy rate (expressed as a dampening of extinction rates over time) increases community richness (Lennon and Jones, 2011). Some of the first direct evidence linking microbial dormancy and dispersal limitation to microbial biogeography shows that dormancy dampens environmental and spatial distance-decay relationships for microbes in forested ponds (Locey et al., 2020). Another recent study found that both resuscitation of local dormant cells and regional dispersal of active cells contribute to soil community resilience following a period of thermal stress (Sorensen and Shade, 2020). This study poses the question, what are the long-term outcomes of dormant versus active cellular dispersal in natural systems?

### 3.3. *Dispersal outcomes on community function*

The outcomes of dispersal on community function depend on the interplay between microbial traits associated with dispersers and the strength of local environmental filtering. Dispersal-colonization tradeoffs may structure microbial trait distributions across the spatial and environmental landscape (Smith et al., 2018). In a recent wood decomposition study spanning sites along a forest/non-forest ecotone, dispersal limitation of traits associated with rapid wood-degradation shaped community composition and function such that fungal communities farther from forests decomposed wood blocks more slowly (Smith and Peay, 2021). Independent of forest proximity, there was also a significant negative relationship between alpha-diversity (shaped by stochastic dispersal) and decomposition due to interspecific competition, linking dispersal to independent drivers of community function in this system (Smith and Peay, 2021).

An intuitive hypothesis is that dispersal can mitigate microbial responses to environmental stress by introducing stress-tolerant microbes, but this prediction depends on the regional pool of microbial traits, their adaptive potential, and the extent of functional redundancy. In an experimental evolution experiment, dispersal elevated community growth under ambient conditions but hindered growth in a warming treatment (Lawrence et al., 2016), suggesting that dispersal may dampen the ability of microbial communities to adapt to environmental change by introducing maladapted individuals. In another study looking at the interaction between dispersal and drought, dispersal altered the community composition to a greater extent under drought conditions but also resulted in loss of community function, which was contrary

375 to the hypothesis that dispersal could mitigate drought stress by introduc-  
376 ing tolerant microbes (Evans et al., 2020). In a common garden experiment  
377 across a natural precipitation gradient, enhanced dispersal had no effect on  
378 community composition, which was the strongest predictor of functional re-  
379 sponses to changes in moisture (Waring and Hawkes, 2018). However, under  
380 certain conditions, dispersal can enhance community stability in the face of  
381 environmental change for both acute and more gradual disturbances (Evans  
382 et al., 2019, Sorensen and Shade, 2020).

383 The order and timing of dispersal events can also influence dispersal out-  
384 comes. For instance, the release of fungal spores during day versus night  
385 influences dispersal longevity and survival (Oneto et al., 2020). Historical  
386 contingencies are past biological interactions or environmental conditions,  
387 whose order and timing impact the trajectory of a community response. Pri-  
388 ority effects are a specific example of a biotic historical contingency where  
389 the early or late arrival of a species determines community assembly out-  
390 comes (Fukami, 2015). The important of historical contingencies (Hawkes  
391 and Keitt, 2015) and priority effects (Sprockett et al., 2018, Hiscox et al.,  
392 2015, Svoboda et al., 2018) on microbial community assembly has been  
393 demonstrated across a range of habitats.

#### 394 **4. Traits-based approach for predicting dispersal outcomes**

395 Traits are increasingly invoked as the key parameters to understand ecosys-  
396 tem function. Traits include the physiological, life history, and behavioral  
397 characteristics of organisms that underlie ecosystem function (Martiny et al.,  
398 2015). Because traits more directly relate to ecosystem function, and most  
399 traits are phylogenetically conserved to some degree, traits are a valuable  
400 tool in linking microbial biogeography to ecosystem function (Green et al.,  
401 2008, Nelson et al., 2016, Fierer et al., 2012). Quantification of traits related  
402 to dormancy and dispersal should also be valuable to understanding their  
403 interaction, but current attempts are hampered by the breadth of traits that  
404 contribute to these processes.

405 For example, range size correlates to genomic and phenotypic attributes  
406 of dust-associated microbes, suggesting that these traits may be related to  
407 dispersal capabilities (Choudoir et al., 2018). For AM fungi, a recent study  
408 showed that small spore size was positively associated with aerial dispersal  
409 (Chaudhary et al., 2020), while another study found spore size to be a poor  
410 predictor of AM fungal range size (Kivlin, 2020). It’s clear we are far from

understanding physiological traits that determine dispersal outcomes. Since atmospheric dispersal is important to both plants and microbes, looking to decades of studies in plant ecology for inspiration about traits related to dispersal will likely yield fruitful insights (Thomson et al., 2010, Tamme et al., 2014). Once traits are identified, analyses developed for genome-wide association studies (GWAS) (Eriksson et al., 2010, San et al., 2020) may offer useful insights for identifying genetic variation related to common traits associated with dormancy and/or dispersal.

Ultimately, we need to develop a predictive framework for implementing dispersal traits into changes in ecosystem function. One suggestion is implementation of the response-effect framework (Lavorel and Garnier, 2002), where response traits determine community structure (indirect drivers) and effect traits influence ecosystem function (direct drivers). This framework has been previously applied to fungal systems (Crowther et al., 2014, Koide et al., 2014), and while it can be challenging to parse indirect versus direct effects on ecosystem function, there is predictive power when response and effect traits are correlated. Using this framework, microbial dormancy and dispersal traits are response traits which control microbial community structure directly (and ecosystem function indirectly), as separate from effect traits that govern ecosystem function directly. Another approach could implement a tradeoff framework, such as the yield-resource acquisition-stress (Y-A-S) traits framework developed as a microbial analog to Grime’s competitor-stress tolerator-ruderal (C-S-R) framework (Grime, 1977, Malik et al., 2020). For example, dormancy could be invoked as a measure of community stress response. A third approach could implement dormancy or dispersal as a performance filter along an ecological gradient overlaying other system traits (Webb et al., 2010). These frameworks could be high-level conceptualizations to be combined with statistical modeling. For example, the relative contribution of dispersal to ecosystem function can be estimated using generalized nonlinear models, with microbial traits as potential fixed effects screened in model selection. Structural equation modeling (SEM) can be employed downstream to define direct and indirect drivers of ecosystem function.

## 5. Road map and research recommendations

To improve our predictions of structure-function relationships in soils, we need to apply and evaluate a more precise, yet adaptable conceptualization of microbial dispersal. We propose a reframing of microbial dispersal into

447 active, dormant, and acellular modes. The ecological, spatial, and temporal  
448 restraints vary between cellular and acellular dispersal (Figure 1), with genetic  
449 dispersal potentially decoupled from environmental filtering and organismal  
450 identity. We are not the first to frame patterns of biogeography through the  
451 lens of genetic dispersal. For example, Baltrus (Baltrus, 2020) discusses the  
452 dynamics of genetic element dispersal and the contributions to the biogeog-  
453 raphy microbial pathways. We also encourage researchers to contemplate the  
454 limits of this thinking. Can microbial dispersal always be sufficiently repre-  
455 sented as genetic dispersal, or is it sometimes more important to consider  
456 the individual organisms harboring these genes?

457 To close current knowledge gaps, we recommend starting with these re-  
458 search directions. First, we need to develop a quantitative theoretical frame-  
459 work that integrates microbial dispersal, dormancy, and environmental fil-  
460 tering. A new model describes the interactions between dispersal and dor-  
461 mancy and outcomes on community diversity across scales (Wisnoski and  
462 Shoemaker, 2021). This model considers multiple dormancy traits (*i.e.* sur-  
463 vival and germination rates) and how these processes interact with dispersal  
464 to create nonlinear effects on metacommunity diversity across local and re-  
465 gional spatial scales. Models that capture dispersal and dormancy dynamics  
466 will continue to improve as we better measure these phenomena, both as  
467 ecosystem processes and as microbial traits.

468 Second, we need to develop more accurate methods for quantifying, and  
469 accounting for, microbial dispersal and dormancy. Golan and Pringle (Golan  
470 and Pringle, 2017) provide a comprehensive framework for considering fungal  
471 long distance dispersal that entails mathematical models, genetic inference,  
472 and direct quantification based on spore capture. We also need improved  
473 tools to quantify genetic dispersal (Brito, 2021). We recommend incorpo-  
474 rating dispersal and dormancy explicitly into soil structure-function studies.  
475 A typical structure-function analysis neglects the influence of dispersal from  
476 a regional pool, and also assumes that all recovered DNA sequences are  
477 representative of active (or potentially active) organisms. Incorporating dis-  
478 persal might mean a no-dispersal or enhanced-dispersal treatment as part of  
479 the experimental design, or accounting for new taxa from the atmosphere  
480 at regional scales or along fungal highways at local scales. Incorporating  
481 dormancy might mean including a resuscitation treatment (*e.g.* bacterial  
482 resuscitation factor Rpf, see (Kuo et al., 2021)), or filtering taxa based on  
483 microbial activity. Although at present our methods for quantifying active  
484 versus non-active fractions are imperfect. If we knew how to differentiate ac-

485 tive versus inactive rRNA, we'd be a lot closer to estimating true dormancy  
486 rates. Cell-resolved metabolomics might offer a promising solution (Walsh  
487 et al., 2018).

488 Finally, we need to better qualify and quantify traits related to dormancy  
489 and dispersal. This may require different strategies for different microbial  
490 lineages. For example, identifying a taxon-specific sporulation gene, or an  
491 environmental signal that is associated with entering dormancy or the re-  
492 suscitation of a particular microbe. New research on the homeostasis of  
493 ribosomes in *Methanococcus* during energy limitation underscores the need  
494 for alternative traits that accompany activity and dormancy (Müller et al.,  
495 2021). We as microbial ecologists need to support research that focuses on  
496 specific members of a microbial community (*e.g.* revitalization of microbial  
497 cultivation efforts (Carini, 2019)). In understanding the trees, we may finally  
498 be able to see the forest.

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## 503 7. Author Contributions

504 M.J.C and K.M.D conceptualized and investigated the perspective. M.J.C  
505 and K.M.D wrote and revised the manuscript.

## 506 8. Declaration of Interests

507 The authors declare no competing interests.

## 508 9. Inclusion and Diversity

509 One or more of the authors of this paper self-identifies as a member of the  
510 LGBTQ+ community. While citing references scientifically relevant for this  
511 work, we also actively worked to promote gender balance in our reference  
512 list.



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