

CREATING A CIRCULAR NITROGEN BIOECONOMY IN AGRICULTURAL SYSTEMS THROUGH NUTRIENT RECOVERY AND UPCYCLING BY MICROALGAE AND DUCKWEED: PAST EFFORTS AND FUTURE TRENDS

Highlights

- Aquatic vegetation-based nutrient recovery offers an alternate approach for treating agricultural wastewater
- Microalgae and duckweed can upcycle waste nutrients into valuable bio-based products
- Producing feed, fertilizer, and fuel from manure-grown aquatic vegetation promotes a circular N-bioeconomy

9 **Abstract.** The massive amounts of nutrients that are currently released into the environment as waste
10 have the potential to be recovered and transformed from a liability into an asset through photosynthesis,
11 industry insight, and ecologically-informed engineering design aimed at circularity. Fast growing
12 aquatic plant-like vegetation such as microalgae and duckweed have the capacity to enable local
13 communities to simultaneously treat their own polluted water and retain nutrients that underlie the
14 productivity of modern agriculture. Not only highly effective at upcycling waste nutrients into protein-
15 rich biomass, microalgae and duckweed also offer excellent opportunities to substitute or complement
16 conventional synthetic fertilizers, feedstocks in biorefineries, and livestock feed while simultaneously
17 reducing the energy consumption and greenhouse gas emissions that would otherwise be required for
18 their production and transportation to farms. Integrated systems growing microalgae or duckweed on
19 manure or agricultural runoff, and subsequent reuse of the harvested biomass to produce animal feed,
20 soil amendments, and biofuels presents a sustainable approach to advancing circularity in agricultural
21 systems. This article provides a review of past efforts made toward advancing the circular nitrogen
22 bioeconomy using microalgae- and duckweed-based technologies to treat, recover, and upcycle
23 nutrients from agricultural waste. The majority of the work with microalgae- and duckweed-based
24 wastewater treatment has been concentrated on municipal/industrial effluents with <50% of studies
25 focusing on agricultural wastewater. In terms of scale, more than 91% of the microalgae-based studies
26 and 58% of the duckweed-based studies were conducted at laboratory-scale. While the range of
27 nutrient removals achieved using these technologies depends on various factors such as species, light,
28 and media concentrations, 65-100% total N, 82-100% total P, 98-100% NO_3^- , and 96-100% NH_3/NH_4^+

29 can be removed by treating wastewater with microalgae. For duckweed, removals of 75-98% of total
30 N, 81-93% total P, 72-98% NH_3/NH_4^+ , and 57-92% NO_3^- have been reported. Operating conditions
31 such as hydraulic retention time, pH, temperature, and the presence of toxic nutrient levels and
32 competing species in the media should be given due consideration while designing these systems to
33 yield optimum benefits. In addition to in-depth studies and scientific advancements, policies
34 encouraging supply chain development, market penetration, and consumer acceptance of these
35 technologies are vitally needed to overcome challenges and to yield substantial socio-economic and
36 environmental benefits from microalgae- and duckweed-based agricultural wastewater treatment.

37 **Keywords.** Circular bioeconomy; Duckweed; Microalgae; Wastewater treatment; Nitrogen; Nutrient
38 recycling; Manure treatment

39 INTRODUCTION

40 Transitioning the current agricultural sector from a linear to a circular system is required to effectively
41 recycle valuable resources such as nitrogen (N). Considered one of the most important elements for
42 plant growth, N also forms a key component of amino acids that make up the proteins required by
43 humans and animals to meet their nutritional needs. Natural processes like atmospheric deposition, N-
44 fixation, plant and animal N uptake, nitrification, and denitrification, are all critical elements of the
45 complex N cycle that affects the availability of N in the environment (in the forms of organic N, nitrate
46 (NO_3^-), nitrite (NO_2^-), and ammonia (NH_3)) and its subsequent influence on air and water quality. In
47 agricultural systems, the relatively recent changes in agricultural practices, such as extensive soil tillage
48 and crop residue harvesting, and the increased use of chemical fertilizers have resulted in excessive N
49 applications and subsequent N leaching through groundwater infiltration and surface runoff
50 (Mazzoncini et al., 2011; Savci, 2012). Livestock farms that produce and release untreated manure are
51 another major source of N pollution to surface waters (Kleinman et al., 2018; Ribaldo, 2003). Excess
52 nutrients can be carried down gradient in streams and rivers, resulting in the growth of harmful algal
53 blooms that can cause eutrophication and hypoxia (oxygen depletion) in large water bodies such as the

54 Gulf of Mexico, Chesapeake Bay, Lake Erie, Lake Victoria, and other regions around the world
55 (Anderson et al., 2008; Kemp et al., 2005; Scavia et al., 2014). Agricultural wastewater thus often
56 necessitates treatment or nutrient recovery techniques before being released for reuse, or otherwise
57 long-lasting negative impacts on soil health, water quality, and biodiversity may result.

58 Although many N management strategies have been developed, full recovery of N from water
59 sources is typically challenging without significant energy and financial investment. For instance,
60 conventional N removal processes in wastewater treatment are known to cause serious environmental
61 impacts by contributing to the release of nitrous oxide (N_2O), a potent greenhouse gas (GHG)
62 (D'Odorico et al., 2018; Sutton et al., 2011); higher N removal from wastewater often requires higher
63 energy and chemical demands, and in turn leads to increased operational costs and more GHG emissions
64 (Hauck et al., 2016). Furthermore, most of the existing N removal technologies are focused on
65 municipal and industrial wastewater treatment with limited emphasis given to wastewater from
66 agricultural sources. Typically, agricultural wastewaters (especially those from livestock farms that
67 include manure, feedlot runoff, milking center wash water, etc.) are left untreated, spread on crop fields
68 to increase soil fertility, or occasionally treated using constructed wetlands (Dordio & Carvalho, 2013).
69 Untreated manure and agricultural soil mismanagement not only deteriorate stream water quality but
70 also increase N_2O emissions and overall N imbalances. Novel techniques and materials to remove and
71 recover N from agricultural wastewater without deleterious climate change effects are therefore
72 required to alleviate the environmental impacts from waste generation and improve soil, air, and water
73 quality. One promising set of options are photosynthesis-based technologies that incorporate the use of
74 aquatic vegetation to recover nutrients while simultaneously sequestering carbon dioxide (CO_2) from
75 the atmosphere and producing beneficial biomass. Evaluating the true impacts associated with these
76 techniques requires a cradle-to-grave analysis, or Life Cycle Assessment (LCA), of all processes and
77 products generated within the wastewater treatment system. Most of the LCA studies in this area have
78 focused on evaluating environmental impacts of microalgae-based municipal wastewater treatment with

79 concomitant biofuel production, with a few studies concentrating on the benefits of growing microalgae
80 on swine wastewater (Lopes et al., 2018; Maga, 2017; Wu et al., 2020). Although duckweed-based
81 municipal wastewater treatment is gaining popularity, and laboratory to full-scale experiments have
82 been conducted to demonstrate the plant's nutrient recovery efficiency (Cheng & Stomp, 2009;
83 Mohedano et al., 2012), LCA on this technique has only been done to a minimal extent (Roman &
84 Brennan, 2021a). Further, the concept of using microalgae and duckweed for treating agricultural runoff
85 and manure is still evolving and requires additional research to holistically evaluate potential
86 environmental impacts.

87 The transition to a resource recovery-focused approach for wastewater treatment over the past
88 decade parallels the global trend toward a circular bioeconomy which focuses on the conversion of
89 biomass and other bio-waste into useful products in an effort to transition away from the
90 overexploitation of fossil fuels (Ferreira et al., 2018; Nagarajan et al., 2020). A prime example is a
91 biorefinery that utilizes biomass to produce bioethanol as an alternative to conventional petroleum
92 refineries. Other examples include producing plant-based biodegradable plastics (Karan et al., 2019),
93 pharmaceuticals (Kesik-Brodacka, 2018), and construction materials (Shanmugam et al., 2021). A
94 circular N-bioeconomy specifically focuses on cycling N within the larger bioeconomy through
95 efficient N recovery techniques such as using biofertilizers and compost, making plant-based biofuels,
96 and producing animal feed from bio-waste. These techniques, when employed on a large-scale, are not
97 only environmentally sustainable, but also more economically viable than traditional fossil fuel-based
98 production processes (Awasthi et al., 2019; Nagarajan et al., 2020). Such a systems-level approach
99 further provides opportunities to conduct LCAs on several interconnected N-bioeconomy processes and
100 help address issues within the water-energy-food (WEF) nexus including, but not limited to: food
101 insecurity, GHG emissions, water pollution, and eutrophication (Del Borghi et al., 2020; Ubando et al.,
102 2020).

103 More than any other sector, agriculture has the largest impact on habitable land use (50%) and is the

104 second largest contributor to GHG emissions (24%) after energy production (IPCC, 2014; Ritchie,
105 2019). Additionally, the farming stage of the food supply chain accounts for 25% of global terrestrial
106 acidification and 74% of total freshwater and marine eutrophication (Poore & Nemecek, 2018). In
107 agricultural systems, one of the ways to promote a circular N-bioeconomy is by producing beneficial
108 byproducts from harvested or leftover biomass such as crop residues. For example, corn stover has been
109 widely recognized as a good candidate for lignocellulose-based biofuel production (Kim et al., 2019;
110 Qureshi et al., 2010), but corn stover-based biorefineries have not been yet been implemented on a
111 large-scale primarily due to the negative water quality impacts caused by the increased nutrient runoff
112 that occurs with the removal of crop residues from agricultural fields (Battaglia et al., 2021; Cibin et
113 al., 2012). Considering the tradeoffs between energy production and water quality deterioration, a
114 futuristic pathway to advance the circular N-bioeconomy in agriculture is to employ nutrient recovery
115 techniques which utilize fast-growing aquatic vegetation that naturally recover N from agricultural
116 runoff and enable the subsequent reuse the cultivated biomass for producing energy and other useful
117 products such as soil amendments and animal feed. With technological advancements and process
118 improvements, this practice could holistically tackle the issues within the larger WEF nexus, one such
119 example being the use of wastewater-grown aquatic vegetation to sustainably produce proteins for
120 animal consumption and to enhance food security.

121 The primary objective of this review is to identify past efforts made toward advancing the circular
122 N-bioeconomy in agricultural systems, with a specific focus on emerging sustainable methods of
123 treating and recovering nutrients from agricultural wastewater, and to understand the limitations and
124 future trends in this area. By reconciling the lessons learned from past studies, and through a
125 comprehensive analysis of improved N recovery techniques, the environmental and economic benefits
126 of adopting a circular N-bioeconomy approach in agricultural systems may be realized.

127 **PROMOTING A CIRCULAR N-BIOECONOMY IN AGRICULTURAL SYSTEMS**

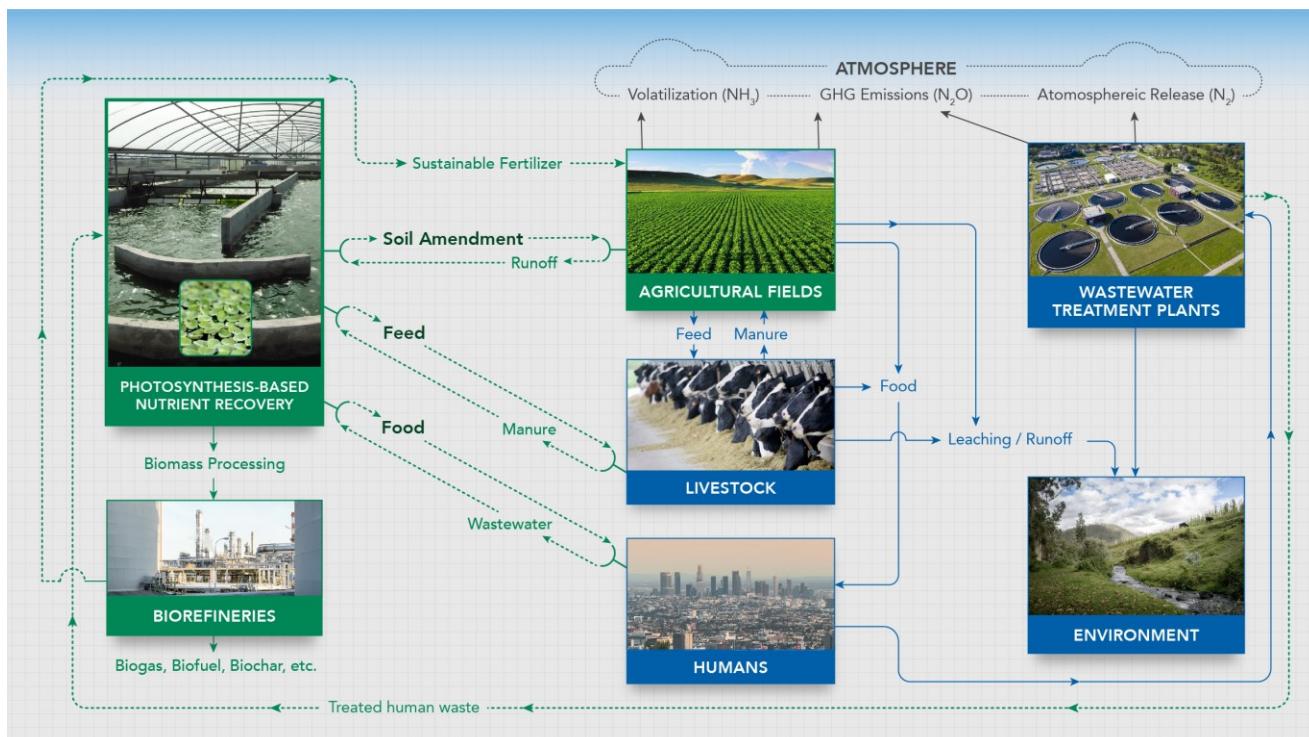
128 Traditionally, manure from livestock farms is stored in deep pits or on-site lagoons and subsequently

129 applied to crop fields which helps enrich the soil with nutrients, but can release NH₃ into the
130 atmosphere. Anaerobic digestion, a routine process used to treat manure prior to soil application, can
131 reduce CO₂ and methane (CH₄) emissions from manure through useful biogas production; however, the
132 remaining digestate, when applied on soil, still poses a risk of increased GHG emissions (Dietrich et
133 al., 2020). Livestock farms in general have been reported to be the major source of non-CO₂ GHG
134 emissions in the United States and China (Nagarajan et al., 2019). Although manure-fertilization of crop
135 fields has been recommended as a way to encourage circularity in agricultural systems, runoff from
136 these farms can cause pollution in adjacent water bodies if effective nutrient recovery techniques are
137 not implemented. Using manure as a biorefinery feedstock has been studied as another pathway to
138 promote the circular bioeconomy, but there are technical challenges associated with the conversion of
139 manure to biofuel and other useful byproducts owing to its heterogeneous composition (Chen et al.,
140 2005).

141 Cultivating protein-rich plant-like species including duckweed, azolla, seaweed, and microalgae on
142 wastewater has gained popularity in recent years as a novel method to recover nutrients before they are
143 released into the environment (Arumugam et al., 2018; Muradov et al., 2014; Nagarajan et al., 2020).
144 Duckweed (of family Lemnaceae), azolla (of family Salviniaceae), seaweed (a form of macroalgae),
145 and microalgae are all aquatic autotrophs with a wide-ranging diversity of species within each family.
146 These species require a smaller areal footprint to produce equivalent biomass when compared to
147 conventional land-grown crops, and are promising sources of biomass feedstock and animal feed
148 (Calicioglu et al., 2018; Hemalatha et al., 2019). In relation to the conventional lignocellulosic biomass,
149 both algae and duckweed have strong potential to be used in large-scale systems for upcycling N into
150 biomass due to their rapid growth rates. Their high protein content (of up to 50% by dry weight) and
151 the ability to be pumped for transportation are other benefits of using algae and duckweed for biomass,
152 feed, and food production. An LCA on a duckweed-based ecological wastewater treatment facility
153 indicated that without supplemental heating, such a facility can reduce energy consumption by a third

154 and GHG emissions by half when compared to a conventional wastewater treatment system (Roman &
155 Brennan, 2021b). A sustainable farming system promoting the circular N-bioeconomy concept could
156 involve growing these aquatic species on either diluted manure or bio-digestor effluents and harvesting
157 them to be used for: bioenergy production; as a fertilizer-substitute; or as a protein supplement in animal
158 feed. **Figure 1** illustrates the existing linear N economy in agricultural systems along with the
159 recommended pathways to transition towards a circular N-bioeconomy using aquatic vegetation for
160 nutrient recovery.

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162

163 **Figure 1. Integrating wastewater-treatment and aquatic vegetation to promote a circular N-bioeconomy in**
164 **agricultural systems. Blue lines refer to the existing linear economy and green dashed lines show pathways to promote**
165 **a circular N-bioeconomy.**

166

167 The following section summarizes conventional farm nutrient management methods and reviews
168 emerging microalgae- and duckweed-based nutrient recovery technologies, highlighting the benefits
169 and challenges associated with each. Although a large share of published studies has been focused on
170 using microalgae and duckweed for treating municipal wastewater, there is growing trend toward

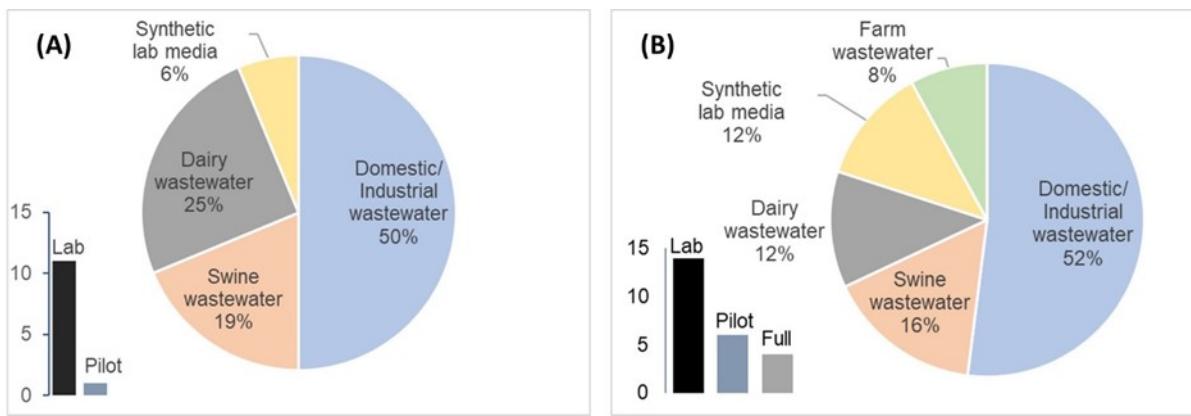
171 applying these technologies for treating agricultural runoff and manure. A circular N-bioeconomy can
172 be realized in agricultural systems by applying these practices to integrated farming systems to generate
173 value-added products.

174 **PAST EFFORTS IN MICROALGAE AND DUCKWEED-BASED NITROGEN RECOVERY METHODS**

175 Typically, wastewater treatment plants providing dedicated N removal processes are normally only
176 used to treat wastewater from domestic and industrial sources. Runoff from agricultural fields and
177 livestock farms are often left untreated, leading to surface and groundwater contamination. In certain
178 cases, manure and other organic waste from livestock farms are treated either using anaerobic digestors
179 or waste stabilization ponds that promote sedimentation of waste solids and anaerobic decomposition
180 to produce methane and other usable products such as biochar and compost. While anaerobic digestors
181 have better treatment efficiency than settling ponds due to added heating and mixing, they are a
182 comparatively expensive treatment option. Settling ponds on the other hand, while cost effective, can
183 contribute to high GHG and odor emissions (Craggs et al., 2014). Therefore, a cost-effective and
184 environmentally-friendly treatment method with high nutrient removal efficiency (e.g. using aquatic
185 vegetation such as microalgae or duckweed) would offer a sorely needed alternative for treating and
186 recovering N from farm wastewater. Existing practices to capture N from agricultural field runoff
187 involve the use of constructed wetlands, buffer strips, denitrification bioreactors, etc. (Husk et al., 2017;
188 Xia et al., 2020); there have been limited applications of using microalgae and duckweed-based N
189 recovery technologies to capture and treat runoff from crop fields due to the non-point source nature of
190 the runoff. Manure generated on livestock farms, however, is comparatively easier to collect and treat
191 than runoff; therefore, much of the work conducted in the past on microalgae and duckweed-based N
192 recovery from agricultural wastewater has been focused on manure from livestock farms. Theoretically,
193 these recovery methods could be adopted to treat cropland runoff if an on-farm treatment system (such
194 as a constructed wetland) is utilized to capture runoff from cropping areas.

195 The literature review for this study was carried out using Web of Science database

196 (https://www.webofknowledge.com) by finding articles with keywords ‘duckweed’, ‘microalgae’,
197 ‘bioeconomy’, ‘nutrient removal’, and ‘biomass production’. From the extensive list of papers found,
198 we shortlisted those in which microalgae and duckweed were used to treat wastewater. Studies
199 published between the years 1995-2020 are included in the review. **Table S-1** (Supplementary
200 Information) shows the complete list of selected papers. Of the reviewed studies that focused on
201 microalgae and duckweed-based wastewater treatment, more than half used wastewater from domestic
202 and industrial sources, and the majority of them were conducted at laboratory scale (**Figure 2**). For in-
203 depth review, only those papers focusing on agricultural wastewater treatment are summarized here
204 (**Table 1**).



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206 **Figure 2. Source of nutrients used in the reviewed articles that focused on (A) microalgae-based (n=12) and (B)**
207 **duckweed-based (n=24) wastewater treatment. The experimental scales used in the studies (lab, pilot, or full-scale)**
208 **are shown on the bottom left of each chart. Details of the studies reviewed are provided in the Supplementary**
209 **Information (Table S-1).**

210 **Microalgae-based Wastewater Treatment**

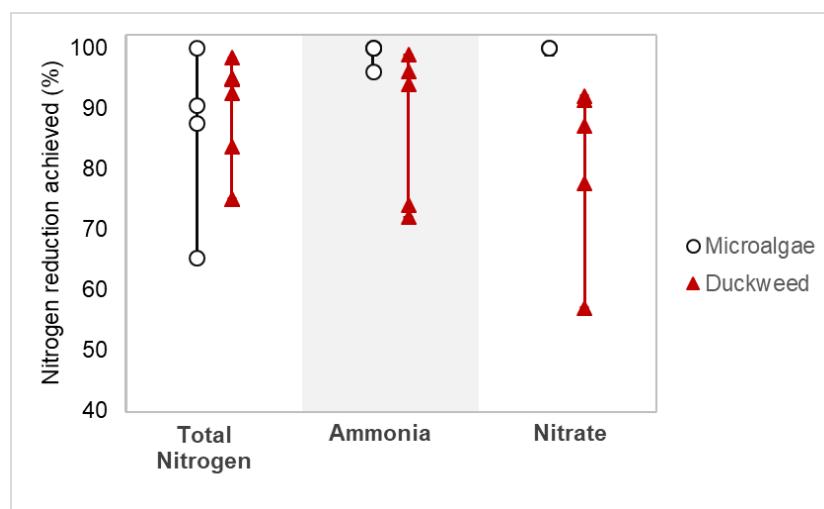
211 Microalgae are unicellular photosynthetic microorganisms that can grow in marine and freshwater
212 ecosystems, utilizing sunlight, CO₂ or organic carbon, water, and nutrients to build biomass with high
213 protein and lipid contents (40% and 30% by dry weight, respectively) (Acién Fernández et al., 2021;
214 Su, 2021). Microalgae can double in mass in less than a day and produce biomass yields as high as 100
215 ton dry mass/ha/yr (Acién Fernández et al., 2021). There are many strains of microalgae, with varying
216 effectiveness in removing nutrients and creating useful biomass; however, *Chlorella* and *Scenedesmus*
217 are the most commonly used genera for wastewater treatment applications (Su, 2021). Up to 1 kg of
218 microalgae can be produced per m³ of human sewage; however, with the elevated concentrations of

219 nutrients typically found in livestock manure, higher yields in the range of 10 to 100 kg/m³ of effluent
220 can be obtained (Acién Fernández et al., 2021), but this requires adequate dilution to avoid overloading
221 the treatment system.

222 Microalgae exhibit a higher removal rate of NH₄⁺ compared to NO₃⁻ and NO₂⁻ because the latter must
223 be reduced to NH₄⁺ (an energy intensive process) before being used for building amino acids and then
224 proteins in the cell (Cai et al., 2013; Maestrini et al., 1986). This is particularly important in the context
225 of treating livestock manure since it contains high levels of NH₄⁺. The uptake of NO₃⁻ by microalgae
226 can be partially reduced in the presence of ambient NH₄⁺, an inhibitory effect that is further enhanced
227 by factors such as limited light conditions and lower temperatures (Su, 2021). The phenomenon of NH₃
228 removal (but not recovery) is aided at elevated pH conditions since high pH causes NH₄⁺ to convert to
229 gaseous NH₃, which is then released into the air (Ferreira et al., 2018; Zimmo et al., 2003). Microalgae
230 can also remove N₂O from wastewater (Qie et al., 2019). Using microalgae, 65-100% total N, 82-100%
231 total P, 98-100% NO₃⁻, and 96-100% NH₃/NH₄⁺ removal has been achieved in treating farm, industrial,
232 and municipal wastewaters (**Figure 3, Table 1, Table S-1**). More studies concentrating on microalgal
233 treatment of agricultural wastewater are required to fully understand the range of nutrient reductions
234 that can potentially be achieved under different environmental conditions.

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238 **Figure 3. Ranges of nitrogen reductions achieved with microalgal- and duckweed-based wastewater treatment**
239 **(summarized from 22 papers). Each symbol represents the results reported by an individual study.**

Table 1. Summary of nitrogen removal and biomass production by microalgae and duckweed in selected agricultural wastewater treatment systems*.

Wastewater Type	Scale	Species used	Experimental Conditions/ Variables	Results	Citations
Microalgae-based					
Poultry, swine, brewery, cattle, dairy, and urban wastewater	Lab	<i>Scenedesmus obliquus</i>	Pre-treated cattle, dairy, and brewery wastewater	95-100% TN removal; 63-99% PO ₄ ³⁻ removal; Biomass produced with 31-53% protein content, 12-26% sugars, and 8-23% lipids	Ferreira et al. (2018)
Dairy wastewater	Lab	<i>Acutodesmus dimorphus</i>	Untreated dairy wastewater; Very low NO ₃ ⁻ concentration	100% NO ₃ ⁻ removal within 4 days; 100% NH ₃ removal within 6 days; 1 kg biomass is theoretically calculated to produce up to 273 g of biofuels	Chokshi et al. (2016)
Dairy wastewater	Lab	Algal consortium (<i>Chlorella saccharophila</i> UTEX 2911, <i>Chlamydomonas pseudococcum</i> UTEX 214, <i>Scenedesmus</i> sp UTEX1589 and <i>Neochloris oleoabundans</i> UTEX 1185)	Wastewater from collecting and holding tanks of dairy farm; Three different CO ₂ concentrations, irradiance of 80 mmol m ⁻² s ⁻¹ , 12 hr daylength, for 10 days	98% TKN removal; 99% NH ₃ removal; 86% NO ₃ ⁻ removal	Hena et al. (2015)
Swine wastewater	Lab	<i>Chlorella vulgaris</i>	12 days	90.51% TN removal and 91.54% TP removal	Wen et al. (2017)
Swine wastewater	Lab and computer model	<i>Chlorella</i> sp.	Optimizing dilution rate and HRT	Modeled optimal biomass yield and N removal at 2.26-day HRT and 8-fold dilution rate; experiment removal rates of 38.4 mg L ⁻¹ d ⁻¹ of TN and 60.4 mg L ⁻¹ d ⁻¹ of NH ₃	Hu et al. (2013)
Duckweed-based					
Swine wastewater	Lab	<i>Spirodelta oligorrhiza</i>	Two-week harvest and 6% wastewater to 94% tap water	83.7% TN removal and 89.4% TP removal	Xu & Shen (2011)
Swine wastewater	Lab	<i>Lemna minor</i>	12 hr light cycle, pre-treated swine wastewater at 4% dilution	74% NH ₃ removal; 0.14 g m ⁻² d ⁻¹ TN removal	Pena et al. (2017)
Diluted swine effluent	Lab	<i>Spirodelta</i> spp	Different N levels in growing media	Crude protein content increases from 15% at 1 to 4 mg N L ⁻¹ , to 37% at 10 to 15 mg N L ⁻¹ ; toxic effect above 60 mg N L ⁻¹	Leng et al. (1995)
Effluent and digested slurry of biorefinery processing cattle slurry	Lab	<i>Lemna minuta</i>	Various concentrations of effluent from biorefinery and digested slurry	75% TN removal; 81% TP removal; higher concentrations had toxic levels of sodium and potassium	Soíta et al. (2020)
Mixture of domestic and agricultural wastewater	Pilot	<i>Lemna japonica</i> 0234	Comparative study with water hyacinth (<i>Eichhornia crassipes</i>)	60% recovery of N over a year; 0.4 g m ⁻² d ⁻¹ TN removal	Zhao et al. (2014)
Mixture of domestic and agricultural wastewater	Pilot	<i>Lemna japonica</i> 0234	Combining duckweed and carrier biofilm	19.97% higher TN removal and 15.02% higher NH ₃ removal with duckweed	Zhao et al. (2015)
Swine wastewater	Full	<i>Landoltia punctata</i>	1 year duration at 30-day HRT	98.3% TN removal; 98.8% NH ₃ removal; 4.4 g m ⁻² d ⁻¹ TKN removal; 68 t ha ⁻¹ yr ⁻¹ biomass yield	Mohedano et al. (2012)

HRT: Hydraulic Retention Time; TN: Total Nitrogen; TKN: Total Kjeldahl Nitrogen; TP: Total Phosphorus.

*A complete list of studies that include municipal and industrial wastewater treatment with microalgae and duckweed are provided in Table S-1 (Supporting Information).

243 Some relatively new approaches, such as the addition of an organic carbon source to the growth
244 medium, have been proposed to increase the growth rates of microalgae (Ma et al., 2016). Generally,
245 higher growth rates are correlated with higher N removal efficiency (Ji et al., 2013). Due to its affinity
246 for NH₃ and the reduced metabolic cost to convert NH₄⁺ to organic matter compared to other nitrogen
247 forms, microalgae tend to grow faster in water with high NH₃ content. However, concentrations in
248 excess of 110 mg L⁻¹ NH₃ can be toxic and have detrimental effects on growth rate by disrupting the
249 thylakoid transmembrane proton gradient which is vital in supporting microalgal photosynthesis
250 (Salbitani & Carfagna, 2021; Zheng et al., 2019); however, some strains of microalgae, such as
251 *Chlorella vulgaris* and *Scenedesmus obliquus*, have been shown to grow in concentrations up to 360
252 mg L⁻¹ of NH₃ (Collos & Harrison, 2014; Morales-Amaral et al., 2015).

253 Although laboratory and pilot-scale studies have been conducted to explore how algal ponds
254 can be used to treat agricultural wastewater, limited studies have been conducted on its effectiveness
255 for treating and removing nutrients at the full scale. The varying concentrations of N and other elements
256 in wastewater can largely affect the performance of microalgae-based wastewater treatment and
257 therefore, supplementing with specific nutrients (C, N, P) may be required to achieve optimal C/N and
258 N/P ratios for enhanced N recovery (Su, 2021). For example, carbon in the form of CO₂ is supplied to
259 microalgal culture media to aid in the assimilation of inorganic N and P – an energy-intensive process
260 that drives high operating costs (Mohsenpour et al., 2021). Maintaining the C/N and N/P ratios of the
261 medium within an optimal range is specifically important to optimize biomass growth, which in turn
262 affects nutrient removal and treatment efficiency. For example, cattle-slaughterhouse wastewater with
263 a C/N ratio = 49.6 ± 9.4 and an N/P ratio = 6.7 ± 3.8 was found to be suitable for microalgal growth
264 (Maroneze et al., 2014).

265 To make microalgal wastewater treatment techniques sustainable, it is necessary to maximize
266 sunlight interception and to tailor the growing media specifically to the microalgae strain being
267 cultivated (Acién Fernández et al., 2021). When comparing microalgae-based treatment to conventional

wastewater treatment, there are lower energy demands, lower sludge production, reduced GHG emissions, and opportunities to convert biomass into useful products (**Table 2**). Primary drawbacks of algal-based wastewater treatment include: high retention time (7 – 10 days); increased land-use (10 m² of land per capita); the presence of competitive invasive species such as non-beneficial microalgae, parasites, and aquatic invertebrate predators; an inability to grow significant quantities of microalgae in highly turbid water due to low light penetration throughout the water column; and high harvesting costs (Nagarajan et al., 2020). Constructing algal ponds on marginal lands that are otherwise unsuitable for farming can mitigate the impacts arising from increased competition for arable land (Acién Fernández et al., 2021). Although synergistic relationships have been observed between microalgae and bacteria in wastewater, pilot studies have shown that the system can fail with excessive bacterial growth resulting in competition for nutrients and a subsequent reduction in algal growth (Su, 2021). Microalgal wastewater treatment is considered an expensive option owing mainly due to the harvesting operation that accounts for a large share of the total production cost. This cost can be reduced with coagulation-flocculation and sedimentation systems to help settle microalgae (Matamoros et al., 2015). The high capital cost associated with the installation of microalgal-wastewater treatment systems can be mitigated in part by enhancing profits through measures that increase algal biomass production such as the utilization of greenhouses and wavelength filters (Kang et al., 2015).

Table 2. Comparison of beneficial and detrimental impacts from conventional wastewater treatment and microalgae/duckweed-based wastewater treatment systems.

Criteria	Wastewater Treatment Impact		Sources
	Conventional	Microalgae/Duckweed-based	
N removal	Up to 99%	Up to 100% ^m Up to 93% ^d	Henze, 1991 ^c ; McCarty, 2018 ^c ; Samori et al., 2013 ^m ; Li et al., 2019 ^m ; Costa et al., 2016 ^d
GHG Emissions	High (0.005 to 0.8 kg CO ₂ eq./ m ³)	Low (1700 - 3300 mg CO ₂ m ⁻² day ⁻¹) ^d ; (8.3-14 g CO ₂ m ⁻³) ^m	Gupta & Singh, 2012 ^c ; Monteith et al., 2005 ^c ; Mohedano et al., 2019 ^d ; Sims et al., 2013 ^d ; Alcántara et al., 2015 ^m
Land Use	Low	High (5-6.5 m ² per capita) ^m	Acién Fernández et al., 2018 ^m ; Alcántara et al., 2015 ^m
Water Demand	Low	High	Soñta et al., 2020 ^d
Energy Demand	0.3-2.1 kWh m ⁻³	0.02- 1W m ⁻³ ^m	Capodaglio & Olsson, 2020 ^c ; Crawford & Sandino, 2010 ^c ; Pabi et al., 2013 ^c ; Alcántara et al., 2015 ^m ; Lopes et al., 2018 ^m
High value products	Fertilizers, bioenergy	Fertilizers, animal feed, human food, biogas, biofuel feedstock	Cassidy, 1998 ^c ; Spolaore et al., 2006 ^m ; Cheng et al., 2019 ^m ; Leng 1999 ^d ; Calicioglu et al., 2019 ^d

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289 **Duckweed-based Wastewater Treatment**

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Similar to microalgae, another sustainable technology to recycle N in the bioeconomy is to use duckweed to recover N from wastewater and subsequently use the harvested biomass to produce useful products. Duckweed is a free-floating aquatic plant in the Lemnaceae family with five genera and 36 known species (Bog et al., 2019). The macro-nutrient composition of different duckweed species are similar, though protein content can vary from 15-45% depending on the nutrient concentrations of the water in which they are grown (Chantiratikul et al., 2010). When compared to microalgae, duckweed's enhanced effectiveness to treat wastewater is mainly attributable to its easy harvesting (Culley & Epps, 1973) and ability to grow under a wide range of nutrient, temperature (5-33°C), and pH (5.5-8.5) conditions (Ceschin et al., 2019). With a doubling rate of every 1-2 days, an initial duckweed mat covering an area of 10 cm² has the potential to cover up to 1 hectare in under 50 days (Leng, 1999). However, the rate at which duckweed grows and accumulates biomass can depend heavily on the pH, temperature, and nutrient concentrations in the growth media, as well as on mat density, sunlight incidence, and day length.

Duckweed has been studied for removing N in swine, dairy, and municipal wastewaters, as well as dumpsite leachate and storm water, among others (**Table 1**). Like microalgae, which prefer NH₄⁺ uptake over NO₃⁻, duckweed has an affinity for NH₄⁺, which is typically seen in high concentrations in agricultural wastewaters such as those coming from livestock farms (Nagarajan et al., 2019). Factors such as the state of N, temperature, pH, salts, metal concentrations, bacterial presence, and mixing of growth media can affect duckweed's nutrient removal rates (**Table 1**, **Table S-1**). In treatment ponds, bacteria that become attached to duckweed fronds in the form of biofilm play a key role in increasing N removal through N fixation and aerobic degradation of complex compounds that make them easily available for plant uptake (Benjawan & Koottatep, 2007; Chen et al., 2019). Intermittent mixing of the growth media has been shown to promote nutrient removal but excess mixing can deteriorate duckweed growth and nutrient uptake (Chaiprapat et al., 2003). Past studies demonstrate that 75-98% of total N,

314 81-92% total P, 72-98% $\text{NH}_3/\text{NH}_4^+$, and 57-92% NO_3^- can be removed from wastewater treated with
315 duckweed (**Figure 3**). Although maximum nutrient reductions were similar for microalgae and
316 duckweed treatments, a wider range of removal rates were observed with duckweed, possibly due to
317 the higher number of duckweed studies reviewed here. The differences in removal rates are also
318 indicative of the wide range of growing conditions used in these studies which can have significant
319 impact on overall nutrient uptake.

320 Duckweed has previously been shown to have resistance to high levels of macro- and micronutrients
321 in its growth media; however, several studies have reported that high nutrient concentrations (in excess
322 of 60 mg N L^{-1}) can have negative impacts on duckweed growth (Iqbal & Baig, 2017; Sońta et al.,
323 2020). The optimum N concentration for supporting duckweed growth is around 60 mg L^{-1} , which is
324 within the concentration range of typical domestic wastewater sources, but far below many animal
325 wastewaters (Ferreira et al., 2018). Although duckweed has better resistance to high nutrient
326 concentrations compared to microalgae, both options require significant water demands for dilution,
327 increasing treatment costs (Sońta et al., 2020). For duckweed, N/P ratios of 4:1 to 5:1 have been found
328 to be suitable for growth, but little work has been done to optimize the C/N and N/P ratios for maximum
329 growth (Xu & Shen, 2011).

330 Through phytoremediation, duckweed can remove a wide range of contaminants, including
331 agricultural chemicals (such as ammonium, nitrate, phosphate, 2,4-dichlorophenol, dimethomorph, and
332 copper sulphate), nanomaterials (such as zinc oxide, alumina, and copper nanoparticles), and organic
333 pollutants (such as petroleum hydrocarbons) (Ekperusi et al., 2020). Duckweed was able to remove up
334 to 94% BOD and COD, 63-87% total suspended solids, 60-99% total P, 35-87% total dissolved solids,
335 and 40-100% heavy metals in studies across different scales (**Table 1**). Duckweed's ability to effectively
336 sequester up to three times more CO_2 than it emits (equaling 19,592 to 42,052 $\text{mg CO}_2 \text{ m}^{-2} \text{ d}^{-1}$ as
337 demonstrated in pilot-scale duckweed ponds) is particularly vital in addressing global warming
338 (Mohedano et al., 2019). Studies have contradicted on whether duckweed-based wastewater treatment

339 ponds are a source or a sink for CH₄ emissions due to complex reactions occurring at the soil-water
340 interface involving methane production by methanogens and oxidation by methanotrophs (Dai et al.,
341 2015).

342 Pilot and full-scale studies have been used to assess how duckweed can be used for sustainable
343 wastewater treatment while documenting the associated challenges. Like microalgae, the ideal HRT to
344 effectively treat wastewater using duckweed is too high (15-20 days) for making it profitable at the full-
345 scale and therefore technological advancements are needed to increase removal rates in these systems
346 (Acién Fernández et al., 2018; Shi et al., 2010). The toughest challenge in making duckweed an effective
347 treatment solution is its land use and dilution water requirement. With full-scale treatment ponds and
348 lagoons, there is an added challenge of adopting an appropriate harvesting regime for reliable biomass
349 recovery and ensuring that duckweed is the dominant organism in the water. **Table 3** lists the ideal
350 operating conditions for microalgae- and duckweed-based wastewater treatment systems, summarized
351 from the past studies reviewed in this section.

352

353 **Table 3. Comparison of ideal operating conditions and variables affecting nutrient removal in microalgae- and**
354 **duckweed-based wastewater treatment systems.**

Variable	Units	Wastewater Treatment System		Citations
		Microalgae-based	Duckweed-based	
Typical Hydraulic Retention Time	days	7 to 10	15 to 20	Nagarajan et al., (2020); Shi et al. (2010)
Optimum Temperature	°C	15 to 30	5 to 33	Esbroeck, (2018)
Optimum pH	-	7 to 9	5.5 to 8.5	Esbroeck, (2018)
Biomass Doubling Rate	days	< 1	1 to 2	Acién Fernández et al. (2021); Leng (1999)
Biomass Yield	ton dry mass/ha/yr	100	73 to 180	Acién Fernández et al. (2021); Leng (1999)
Optimum C/N ratio	-	49.6 ± 9.4	-	Maroneze et al. (2014)
Optimum N/P ratio	-	6.7 ± 3.8	4 to 5	Maroneze et al. (2014); Xu & Shen (2011)
Ammonia Toxicity Level	mg NH ₄ ⁺ -N /L	> 110	> 60	Carfagna (2021); Soíta et al. (2020)

355

356 *Applications for Wastewater-grown Algae and Duckweed Biomass*

357 *Microalgae applications*

358 Biofuel has been effectively generated from microalgal biomass grown in swine and municipal

359 wastewater (Ma et al., 2014, 2016; Zhu et al., 2013). Besides producing biogas (CH₄ and CO₂) through
360 anaerobic digestion, digestate from microalgal biorefineries has the potential to be used as soil
361 amendment in place of synthetic fertilizers (Préat et al., 2020). When used as organic fertilizers,
362 microalgae can prevent nutrient leaching by slow release of N and P, and can even result in higher crop
363 yields (Coppens et al., 2016). Due to the high lipid content of some microalgae, it can be converted to
364 biodiesel (Samorì et al., 2013).

365 Microalgae have extensive applications in the food, feed, and health sectors. Within the last 50
366 years, the production of microalgae has increased due to its application in biochemicals, nutraceuticals,
367 human nutrition, aquafeed, and biofertilizers (Spolaore et al., 2006). Microalgae have a high protein
368 content and an essential amino acid composition similar to soybean and egg, making them suitable to
369 feed humans, livestock, and fish (Bleakley & Hayes, 2017). They can be substituted in 5-10% of poultry
370 feed and 33% of pig feed without causing any adverse health effects; replacing 1-5% of fish diet with
371 microalgae is even shown to promote health and aid in early growth (Acién Fernández et al., 2021;
372 Spolaore et al., 2006). Based on its high nutritional value and availability, microalgae can be used in
373 diets of malnourished people around the world (Christaki et al., 2011). Major limitations to future
374 research on microalgal applications include their high extraction cost and lack of widespread public
375 awareness on the health benefits of microalgae (Koyande et al., 2019).

376 Few studies have been completed on seaweed (a macroalgae) as an additional way to effectively
377 complete the circular N bioeconomy. Similar to microalgae and duckweed, seaweeds can remove N
378 from water and have a variety of applications in food, energy, and agricultural sectors. Bioethanol,
379 liquid fertilizers, and fish feed have been produced using seaweed biomass in pilot- and full-scale
380 studies (Seghetta et al., 2016). Seaweed also has nutraceutical, food, and neuroactive agent applications
381 (Barbosa et al., 2020). However, seaweed-based wastewater treatment projects are still in their
382 preliminary stages and need additional studies to measure feasibility and biomass availability for large-
383 scale use.

384 *Duckweed applications*

385 Several valuable uses of duckweed biomass grown on wastewater have been explored in the
386 past. The use of natural soil amendments that are produced by upcycling nutrient-rich duckweed
387 provides an economical and sustainable alternative to existing synthetic inorganic fertilizers which are
388 produced in a costly and energy-intensive processes using atmospheric N (e.g. producing ammonia
389 fertilizer using the Haber-Bosch process, Walsh et al., 2012). The potential effectiveness of duckweed
390 as a replacement for conventional fertilizers is primarily attributed to its high N content and increased
391 ability to retain that N in the soil (Kreider et al., 2019; Ma et al., 2015). Along with N-runoff into streams
392 during rain events, NH₃ volatilization typically accounts for a significant portion of N loss in agriculture
393 (Saggar et al., 2013); however, pairing duckweed with chemical fertilizer has been shown to
394 significantly reduce NH₃ volatilization by 36-52% and added 10-11% overall economic benefit in rice
395 fields compared to chemical fertilizer alone (Yao et al., 2017). The N and P bound within the duckweed
396 biomass makes it an ideal slow-release fertilizer and helps retain the nutrients in soil for longer,
397 effectively reducing nutrient runoff and pollution (Fernandez Pulido et al., 2021). In efforts to advance
398 the circular bioeconomy, the use of other aquatic plants, such as seaweed, have been explored as soil
399 amendments, especially for grain crops that have high N demands such as wheat, maize, and rice
400 (Sadeghi et al., 2018).

401 Using duckweed grown on agricultural wastewater for bioenergy production is another approach
402 to recycle otherwise untreated waste and close the N-bioeconomy cycle. Duckweed has potential for
403 ethanol production due to its high starch content when grown on low-nutrient waters (Calicioglu et al.,
404 2019; Cheng & Stomp, 2009). Using sequential fermentation and anaerobic digestion processes, an
405 ethanol yield of 0.07 to 0.15 g ethanol and 328 to 390 ml CH₄ per gram of total solids was achieved
406 with dried duckweed grown on treated wastewater, which was higher than lignocellulosic crops (such
407 as straw), and within the range reported for starch crops (such as corn and potatoes) (Calicioglu &
408 Brennan, 2018). After anaerobic digestion of duckweed to produce CH₄, the resulting digestate can be
409 utilized as an agricultural fertilizer (Calicioglu et al., 2019). A techno-economic analysis and LCA of a

410 hypothetical integrated wastewater-derived duckweed biorefinery indicated that duckweed pond
411 construction and operation account for the majority of capital and operating expenses, and that vertical
412 farming options should be investigated to reduce the detrimental impacts of land use (Calicioglu et al.,
413 2021).

414 One of the most important applications of duckweed in agriculture is its use as feed for livestock and
415 aquaculture. Besides being a key protein source, duckweed can also successfully accumulate micro-
416 minerals such as potassium, calcium, magnesium, sodium, and iron, which are typically not present in
417 adequate quantities in the livestock feed available to small farm owners (Leng et al., 1995). In Vietnam,
418 duckweed farming has been done for many years, where duckweed grown on ponds with diluted manure
419 and human waste is fed to ducks after mixing with cassava peelings (Leng, 1999). With overall protein
420 production rates at 10.1 tons ac⁻¹ yr⁻¹, duckweed can produce edible proteins 6-10 times faster than
421 soybeans per area (Landesman et al., 2005; Roman & Brennan, 2019). Under optimum growing
422 conditions, annual duckweed yield can range from 73 to 180 ton dry matter/ha/yr; however, even less
423 than optimal conditions can still provide an yield of 5 to 20 ton dry matter/ha/yr (Leng, 1999). This is
424 noticeably higher than the average yield for soybean (2.8 metric ton/ha) which is conventionally used
425 as a source of feed protein in livestock farms (Purdy & Langemeier, 2018), and on par or greater than
426 the 2 – 100 ton dry matter /ha/yr achievable with microalgae (Acién Fernández et al., 2021). In
427 aquaponics, both fresh and dried duckweed have been shown as effective feed in the production of
428 fishes such as carp and tilapia (Skillicorn et al., 1993).

429 Although the potential for the use of duckweed in animal feed is high, some researchers suggest
430 adding only a small fraction of duckweed to existing feeds until further research is conducted on
431 optimum inclusion rates so that any potential negative effects can be identified. A few feeding
432 experiments conducted with duckweed on pigs, poultry, ruminants, and fish indicate that duckweed can
433 be given as protein feed to these animals without any severe impact on health (Cheng & Stomp, 2009;
434 Hamid et al., 1993). However, other studies reported decreased weight gain and low intake of feed when

435 duckweed was added to animal diets (Sońta et al., 2019). This discrepancy in experimental outcomes
436 can most likely be attributed to the fact that duckweed species and growth medium composition can
437 highly influence the nutritional quality of the resulting duckweed biomass (Roman et al., 2021). A study
438 by Haustetn et al. (1990) on the potential of duckweed to replace soybean meal in poultry concluded
439 that *Lemna* and *Wolffia* species are as good as soybean as a source of essential amino acids and have no
440 effect on egg production. In ruminant animals, duckweed has a beneficial role in providing highly
441 soluble and readily fermentable protein, with 80-94% rumen degradation observed with proteins in
442 *Spirodela*, *Lemna*, and *Wolffia* species (Huque et al., 1996). A recent feeding trial conducted on mice
443 demonstrated that replacing up to 25% of dietary casein protein with duckweed protein had no adverse
444 effect on the growth and organ development (Roman et al., 2021). Additional research focusing on the
445 effect of a duckweed supplemented diet on animal health and organ development, giving due
446 importance to the type of duckweed used, is necessary to evaluate the feasibility of its large-scale
447 application and to increase farmer confidence in using duckweed as animal feed.

448 In addition to protein, duckweed has high amounts of antioxidants that can be especially useful when
449 incorporated into human diets (Sońta et al., 2020). Due to its ability to accrue micronutrients such as
450 iodine, duckweed can be used in human diets to alleviate the problem of malnutrition in countries
451 around the world (Vladimirova & Georgiyants, 2014). Duckweed's role in controlling mosquito
452 populations has also been studied to some extent, with certain species such as *Lemna minor* being
453 reported to release compounds that repelled female mosquito's oviposition and affected larval
454 development in mosquitos (Eid et al., 1992; Marten et al., 1996). This can have a widespread impact on
455 public health in many regions around the world that are especially vulnerable to mosquito-borne
456 diseases. Advances in duckweed genomics have resulted in three different genomes sequenced to date
457 (*S. polystachya* 9509, *L. minor* 5500, and *W. australiana* 8730) (Acosta et al., 2021). Genomic studies
458 open up a wide range of opportunities within the plant microbiology community by providing valuable
459 information on species identification and traits present in these species. Moving forward, techniques

460 like gene editing and genetic transformations can be used to identify duckweed lines with superior traits
461 that are most effective in nutrient recovery and use in beneficial downstream applications such as
462 combating malnutrition, controlling mosquito populations, and serving as a sustainable alternative to
463 conventional fertilizers, feeds, and fuels.

464 **FUTURE TRENDS AND CHALLENGES**

465 ***Localized sustainable feed and fertilizer production***

466 The growing demand for animal-derived food products and the extensive use of conventional animal
467 feed such as corn and soybean has caused the current livestock production system to become
468 unsustainable. Alternate feed materials are therefore required to overcome this challenge and to
469 transition from a linear to a circular system in the livestock industry. The potential of using algae and
470 duckweed as animal feed has already been studied to some extent as discussed in the sections above.
471 Compared to fish and soy sectors that produce 7000 kt/year of fish-based feed and 200,000 kt/year of
472 soy-based feed (costing \$1.8 and \$0.6 per kg, respectively), microalgal production is still a small-scale
473 industry producing 100 kt/year biomass and offers an expensive feed option costing \$17-\$30 per kg
474 (Acién Fernández et al., 2021).

475 Importing feed products from off-site leads to increased expenses for farmers and greater GHG
476 emissions compared to localized feed production on farms (Sasu-Boakye et al., 2014). In agriculture,
477 especially in dairy farms, developing an integrated on-farm wastewater treatment-N recovery practice
478 by growing protein-rich aquatic vegetation on diluted manure could result in sustainable localized feed
479 production for the livestock. Another pathway to recycle the N contained in the manure-grown algae or
480 duckweed is to use them as fertilizer alternatives for crops or as a useful soil amendment material to
481 improve soil fertility. This approach would be especially useful on large-scale farms consisting of mixed
482 livestock and cropping systems if the algae- or duckweed-based fertilizers are processed on-site and
483 applied to the crop fields within the same farm. Such an on-site system would not only increase farmer
484 profit by decreasing feed and fertilizer imports and transportation requirements, but also be a more
485 environmentally-friendly option by reducing emissions and overall carbon and water footprints (Sasu-

486 Boakye et al., 2014). Considering the low nutrient content of microalgal biofertilizers (<5% N and <1%
487 P), a better way to utilize microalgae may be as a fertilizer additive or biostimulant, which at very low
488 dosages of 2 L/ha has been shown to reduce chemical fertilizer use by over 10% (Acién Fernández et
489 al., 2021).

490 ***Integrated farm wastewater treatment-biorefinery systems***

491 Research focusing on integrated models that combine microalgae- or duckweed-based domestic
492 wastewater treatment and biorefinery systems has gained major attention in recent years, especially
493 with the growing trend to transition from fossil fuels to renewable energy sources (Calicioglu et al.,
494 2019; Nagarajan et al., 2019). However, this approach still needs to be studied in detail for biomass
495 grown on agricultural wastes. Besides offering a promising sustainable solution to upcycle farm
496 wastewater nutrients into biomass, these approaches can help curb the long-term issue of food/feed
497 versus fuel competition arising from the conventional use of corn grains for producing biofuel.
498 Biorefineries based on wastewater-grown microalgae and duckweed are largely in their initial stages,
499 with several processes and technologies still being developed.

500 ***Sustainable protein-sources for humans***

501 Animal-derived protein currently accounts for approximately 45% of total human protein
502 consumption and this share is expected to increase significantly by 2050 (Boland et al., 2013). Human
503 consumption of animal-based proteins is increasing at a high rate, aggravating global warming and
504 calling for the need for alternative plant-based protein substitutes. A report by United Nations Food and
505 Agricultural Organization (FAO) estimated that global livestock production releases 7.1 gigatonnes of
506 CO₂ eq. per annum, accounting for 14.5% of anthropogenic GHG emissions in the form of CO₂, CH₄,
507 and N₂O, and these emissions are expected to increase substantially in the coming years (Gerber et al.,
508 2013). Animal-derived metabolic waste further contributes to other environmental impacts such as
509 eutrophication, acidification, and GHG emissions (Wu et al., 2014). Livestock production also causes
510 land use change impacts and subsequent soil erosion, with deforestation typically accounting for 85%
511 of livestock-related GHG emissions (FAO, 2006). According to the FAO, 26% of the world's ice-free

512 land is utilized for livestock grazing and one-third of the arable land is used for cultivating livestock
513 feed. A shift to a low-meat diet and plant-based proteins is not only recommended to alleviate the
514 environmental impacts discussed above, but also to benefit human health (Appenroth et al., 2018;
515 Koyande et al., 2019). Edible versions of seaweed have long been consumed by people in the Asia-
516 Pacific region but have recently gained popularity in other parts of the world such as Europe. The global
517 seaweed cultivation market is projected to be worth USD 30.2 billion by 2025 (MarketsandMarkets,
518 2021). Duckweed and microalgae have been consumed in the past predominantly by people in some
519 developing regions, but are now also rising in popularity as sustainable food sources in developed
520 countries (Appenroth et al., 2018; Kusmayadi et al., 2021). Duckweed's ability to accumulate toxic
521 heavy metals (such as cadmium, nickel, and lead) and carcinogens (such as arsenic) warrant careful
522 monitoring and treatment technologies to curb excessive accumulation of these chemicals into the food
523 chain (Khan et al., 2020). Similar to other vascular plants, duckweed has the potential to adsorb
524 microplastics to their fronds and roots, which when consumed by humans can cause long-term harmful
525 health effects. Pre-treatment methods such as density-driven separation, flocculation, and sedimentation
526 that can remove up to 88% microplastics from wastewaters may be used in conjunction with duckweed-
527 wastewater treatment if high levels of microplastics are identified in the growing media (Vivekanand et
528 al., 2021). Given that the nutritional composition and protein accumulation of algae and duckweed
529 depends heavily on the type of growth media, the concept of growing them on wastewater merits further
530 research to evaluate their nutritional value and safety for human consumption.

531 ***Challenges within the circular N-bioeconomy***

532 The two biggest challenges behind using wastewater-grown algae or duckweed to advance the
533 circular N-bioeconomy concept are: (1) the production costs associated with cultivation and frequent
534 harvesting; and (2) the sociological resistance to consuming vegetation grown on wastewater. The issue
535 concerning high production cost can be addressed to a great extent by implementing this approach on a
536 large scale and producing a combination of valuable products, for instance, animal feed, protein
537 supplements, and crop fertilizer. Pond construction accounts for a major share of the production cost

538 associated with duckweed-based biorefinery models (Calicioglu, 2019). Constructing the ponds on land
539 inappropriate for agricultural purposes will avoid major competition for arable land (Kreider, 2015).
540 Further, the emerging trend of vertical farming (using stacked trays of plants in growth media
541 illuminated by LED lights) can be utilized to significantly reduce the land requirements for duckweed
542 cultivation, which is anticipated to make the system more economical and sustainable compared to
543 employing the typical pond-grown approach (Roman & Brennan, 2021a).

544 The circular bioeconomy is heavily dependent on the availability of ample biomass to produce
545 bio-based energy and products, especially for large-scale systems. For instance in Belgium, the
546 implementation of innovative conversion technologies to produce fertilizers and other valuable products
547 from bio-based products was constrained due to the lack of sufficient biomass (Maes & Van Passel,
548 2019). Logistical aspects related to the collection and transport of biomass products should be given
549 high importance in a biorefinery system since it is a direct measure of operational costs as well as
550 environmental impact in terms of carbon emissions (Ubando et al., 2020). Utilizing pipes for pumping
551 instead of ground transport for conveying biomass material (such as duckweed), and using natural sun-
552 drying techniques, are ways to encourage sustainability in this context. Studies in Vietnam have
553 demonstrated that duckweed can successfully be used in small-scale farms, and that a major share of
554 the costs derived from drying and transporting the duckweed can be mitigated by using inexpensive
555 sun-drying methods (Leng, 1999).

556 ***Supply chain model***

557 A robust supply chain must be designed for microalgae- or duckweed-based wastewater treatment
558 systems to be simultaneously economical and sustainable (Mohseni & Pishvaee, 2016). Considering
559 that these systems have the potential to influence multiple sectors such as energy, agriculture, and food
560 processing, an efficient supply chain model is essential to upscale the locally developed practices to the
561 national level, and to eventually enter global markets. Inevitably, the end-use products of these systems
562 should substitute the existing products (e.g. generating duckweed-based biofuel instead of petroleum-
563 based fuel, substituting existing chemical fertilizers with duckweed-based soil amendments,

564 supplementing livestock diet with duckweed-based proteins instead of soybeans, etc.). Systematically
565 designing the supply chain to make the byproducts and end-use products available to consumers is also
566 equally important to make the system resilient. In addition, optimizing the processes and products
567 within the entire value chain is also required to develop a system that is cost-effective, beneficial to
568 society, and has minimal environmental impacts. This provides growing opportunities to use multi-scale
569 modeling tools, optimization methods, and LCA to help policy-makers and other stakeholders quantify
570 benefits and risks, and make decisions regarding the emerging practices within the circular N-
571 bioeconomy.

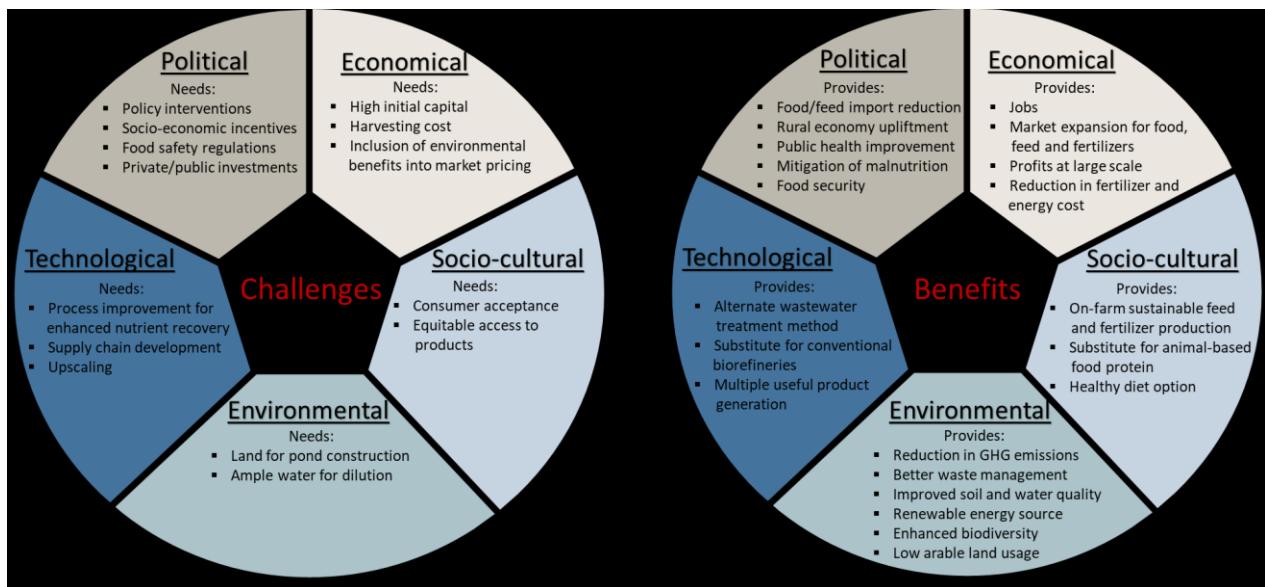
572 ***Policy interventions and socio-economic development***

573 Effective policies have to be designed to encourage investments in technologies and products
574 that advance the circular N-bioeconomy (Maes & Van Passel, 2019). Additionally, subsidizing these
575 products and offering economic and social incentives for processing and/or using these products will
576 make the production system more profitable. For instance, providing economic incentives for growing
577 duckweed on manure waste and re-using it as feed or fertilizer would encourage more farmers to
578 implement this technique, which would have a critical influence on the entire duckweed market. These
579 incentives will not only help overcome the cultural resistance of farmers in cultivating duckweed in
580 place of traditional crops, but also encourage them to develop the skills required to implement such
581 integrated farming systems, which is usually a major constraint in establishing these practices.
582 Additionally, supporting the development of a local duckweed market, as seen in the case of Vietnam,
583 will be useful in promoting duckweed as a cash crop and encouraging farmers in rural communities to
584 engage in duckweed farming (Leng, 1999). Creating more revenue streams through successful policy
585 implementation would attract more private and public investments in the near- and long-term.
586 Environmental externalities (i.e. uncompensated environmental effects of production and consumption
587 of a particular product) have to be incorporated into the true market pricing of the emerging alternative
588 products to achieve reasonable profits and to run the system sustainably.

589 Designing new methods to reuse microalgae or duckweed grown on agricultural wastewater would

590 not only influence the current livestock and fertilizer market, but also expand the sustainable food, feed,
591 and energy markets. The algae products market is projected to grow by 5.2% from 2016 to 2023; and
592 with more use in cosmetics and natural colorants, the compounded annual growth rate of a single algal
593 species (*Spirulina spp.*) is expected to be 10% by 2026 with a market value of US \$2000 million
594 (Credence Research, 2017). The microalgae market in particular is currently valued at 50 million euros
595 and predicted to be worth 70 million euros by 2025 (Acién Fernández et al., 2021). Emerging
596 applications of microalgae, besides using them for biofuel, include the production of biomaterials,
597 biofertilizers, biostimulants, and biopesticides (Acién Fernández et al., 2021). Market expansion of
598 wastewater treatment and biomass production technologies using aquatic vegetation would create more
599 job opportunities and improve the rural economy, allowing further research into developing sustainable
600 products and methods in agricultural systems. Socio-economic and techno-economic analyses would
601 provide further insight into the long-term social and economic impacts triggered by these systems.

602 **Figure 4** summarizes the economical, socio-cultural, political, environmental, and technological
603 challenges and benefits linked to using aquatic vegetation for fostering a circular N-bioeconomy in
604 agricultural systems.



605
606 **Figure 4. Challenges and benefits associated with using aquatic vegetation for wastewater treatment in the circular**
607 **N-bioeconomy.**

608 **CONCLUSIONS**

609 Growing either microalgae or duckweed on manure and agricultural runoff and subsequently using
610 the harvested plant biomass for the production of biofuels, animal feed, or soil amendments provides a
611 promising opportunity to recycle N and promote a circular N-bioeconomy in agricultural systems.
612 However, its ease of harvesting and tested ability to grow under a wider range of environmental
613 conditions gives duckweed some advantages over microalgae. Although over half of the reviewed
614 studies utilized microalgae and duckweed for municipal/industrial wastewater treatment, there is a
615 growing trend toward using this approach for capturing nutrients in livestock manure, which has
616 promising potential. With a capacity of over 90% nitrate and ammonia removal, various applications of
617 these aquatic organisms are being explored in the form of biofeedstock, fertilizers, animal feed, and
618 human food as a way to transition from a linear to circular bioeconomy. Additional in-depth
619 experimental trials are required to fully understand the nutrient interactions, uptake dynamics, and
620 toxicity risks in microalgae and duckweed-based wastewater treatment systems. LCA studies and
621 techno-economic analyses specifically focusing on agricultural wastewater treatment are necessary to
622 evaluate the environmental impacts and economic feasibility of using these technologies in the
623 agricultural sector. With the help of effective policies and technological advancements, several of the
624 political, socio-cultural, and infrastructural challenges that hinder large-scale implementation of these
625 sustainable practices can be overcome.

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