

Forum

Are we underestimating the ecological and evolutionary effects of warming? Interactions with other environmental drivers may increase species vulnerability to high temperatures



Elena Litchman¹✉ and Mridul K. Thomas³

¹Kellogg Biological Station, Michigan State Univ., Hickory Corners, MI, USA

²Dept of Global Ecology, Carnegie Inst. for Science, Stanford, CA, USA

³Dept F.-A. Forel for Environmental and Aquatic Sciences, Univ. of Geneva, Geneva, Switzerland

Correspondence: Elena Litchman (elitchman@carnegiescience.edu)

Oikos

2023: e09155

doi: [10.1111/oik.09155](https://doi.org/10.1111/oik.09155)

Subject Editor and

Editor-in-Chief: Dries Bonte

Accepted 16 August 2022



Warming, the most prominent aspect of global environmental change, already affects most ecosystems on Earth. In recent years, biologists have increasingly integrated the effects of warming into their models by capturing how temperature shapes their physiology, ecology, behavior, evolutionary adaptation and probability of extirpation/extinction. The more physiologically-grounded approaches to predicting ectotherms' responses use thermal performance curves (TPCs) obtained by measuring species performance (e.g. growth rate) under different temperatures. TPCs are typically measured while other factors are held constant at benign levels to 'isolate' the effects of temperature. Here we highlight that this practice paints a misleading picture because TPCs are functions of other factors, including global change stressors. We review evidence that resource limitation, pH, oxygen and CO₂ concentration, salinity, water availability, parasites and mutualists, all influence TPC shape and thermal traits such as optimum temperature for growth. Evidence from a wide variety of organisms – phytoplankton, protists, plants, insects and fish – points towards such interactions increasing organisms' susceptibility to high temperatures (reducing it in the case of mutualists). Failing to account for these interactions is likely to lead to erroneous predictions of performance in nature and an underestimation of the risks of warming. We discuss the general patterns and possible consequences of such interactions for ecological communities. But importantly, interactions with TPCs share common features that we can learn from. Incorporating these interactions into population and community models should lead to deeper insights and more accurate predictions of species' performance in nature – as well as strategies for managing natural and agricultural ecosystems in the face of warming.

Synthesis

Thermal performance curves (TPCs) are being increasingly used to predict species responses to rising temperatures. TPCs strongly depend on other environmental factors, but this dependence is rarely taken into account, which may make predictions inaccurate. Growth limitation by resources and other abiotic and biotic factors often lowers the optimum and maximum performance temperatures, thus making organisms more susceptible to negative effects of warming. We show that this dependence is widespread, if not universal, and was observed for microbes, plants, and animals. Therefore, predictions of ecological and evolutionary processes in a warming world should include temperature interactions with other drivers, otherwise the negative effects of rising temperatures may be underestimated.



www.oikosjournal.org

© 2023 The Authors. Oikos published by John Wiley & Sons Ltd on behalf of Nordic Society Oikos.

This is an open access article under the terms of the Creative Commons Attribution License, which permits use, distribution and reproduction in any medium, provided the original work is properly cited.

Keywords: global environmental change, optimum temperature, resource limitation, stressor interaction, thermal performance curve (TPC), warming

Introduction

The next 100 years are expected to see further increases in global mean temperature of 0.5–2°C. In some regions, mean and maximum temperatures are expected to increase by as much as 4–6°C (IPCC 2021). Understanding and predicting the consequences of this change has been a major, defining goal of biological research for nearly a generation – and will continue to be, for decades to come. A multitude of approaches has been used to understand what warming and associated environmental changes means for organisms, communities and ecosystems: from experiments in the lab, mesocosm and field, correlative analyses on expression patterns, genes, species, communities and ecosystems, from local to global scales, theoretical models ranging from the abstract and simple to the detailed and specific. To make this problem tractable, a substantial proportion of this work – especially the experimental and theoretical parts – has focused on the effects of temperature change alone on populations and communities and maintaining other environmental factors at benign (e.g. high nutrient/food concentration) levels. We argue that this approach must change. By ignoring or oversimplifying how temperature interacts with other factors to influence populations and communities, we draw conclusions and make projections that are likely to be heavily biased.

Here we briefly describe how temperature shapes the growth of ectotherms, then discuss the available evidence on temperature interactions with other environmental factors, and the consequences of such interactions for predicting the effects of rising temperatures on species and communities. The strongest existing evidence is for the effects of resource limitation on the temperature response because this has received the most careful study, but we also provide examples of how other abiotic and biotic factors affect thermal performance curves.

Thermal performance curves (TPCs)

At a fundamental level, temperature affects organisms by changing chemical reaction rates. Accelerating reaction rates with increasing temperature from a low baseline tends to increase organismal performance and vital rates. As summarized in the Metabolic theory of ecology, increasing temperature drives exponential increases in rates of growth, death, movement, consumption, reproduction, mutation and more (Brown et al. 2004). This in turn causes global variation in a host of traits and life history strategies. For any particular biochemical reaction, however, the exponential increase in reaction rate with increasing temperature does not continue indefinitely: it slows, stops and reverses rapidly. At a high enough temperature, enzyme conformations begin to fail and they bind with unintended target molecules. In all

ectotherms, from bacteria to reptiles, this manifests at the organismal and population level as performance often being a left-skewed unimodal function of temperature (Fig. 1A). There are also additional thermodynamic, metabolic and physiological processes that control thermal limits (Tomanek 2008, Ritchie 2018).

This unimodal function describing the dependence of growth or other process on temperature is called the thermal performance curve (TPC) or thermal reaction norm. It has been at the core of attempts to mechanistically link physiology with species ranges, population dynamics and community composition. While the full TPCs can be incorporated into theoretical models, they can also conveniently be summarized using a few easily-understood parameters such as the optimum, maximum and minimum temperatures (T_{opt} , T_{max} and T_{min}). These parameters can be thought of as traits, and are often used to assess species' vulnerability to high or low temperatures and define their thermal niches (Fig. 1A, Addo-Beddiako et al. 2000, Deutsch et al. 2008, Sunday et al. 2011, Thomas et al. 2012). TPCs and these associated traits capture important patterns in – and constraints on – growth rates and geographic ranges (Sunday et al. 2012, Payne et al. 2016). Therefore, using these TPCs to project how warming will alter species performance and shift their ranges seems feasible with our present level of knowledge. At individual locations, temperature projections through time instead of space can be used to generate expectations of whether species would be able to persist (expected net population growth rate ≥ 0) and whether community composition would remain similar. There are complications that are difficult to address rigorously with this approach at present due to insufficient empirical data on TPC evolution and on how biotic interactions depend on species' TPCs (O'Donnell et al. 2018, Tüzün and Stoks 2018). We focus here on one complication that can and should be addressed: the dependence of TPCs and temperature traits on other environmental drivers such as nutrient/food availability.

The dependence of TPCs on environmental factors

The TPC is not a stable property of species, populations or even individuals

Temperature interacts with a number of other environmental factors to determine performance; or stated differently, the TPC is itself a function of other factors. Food/nutrient availability, pH, light (for photosynthetic organisms), salinity, water availability, oxygen concentration, as well as biotic interactions such as parasitism or mutualism, all can alter the shape of the TPC (Ern et al. 2016, Thomas et al. 2017, Aldea-Sánchez et al. 2021, Hector et al. 2021).

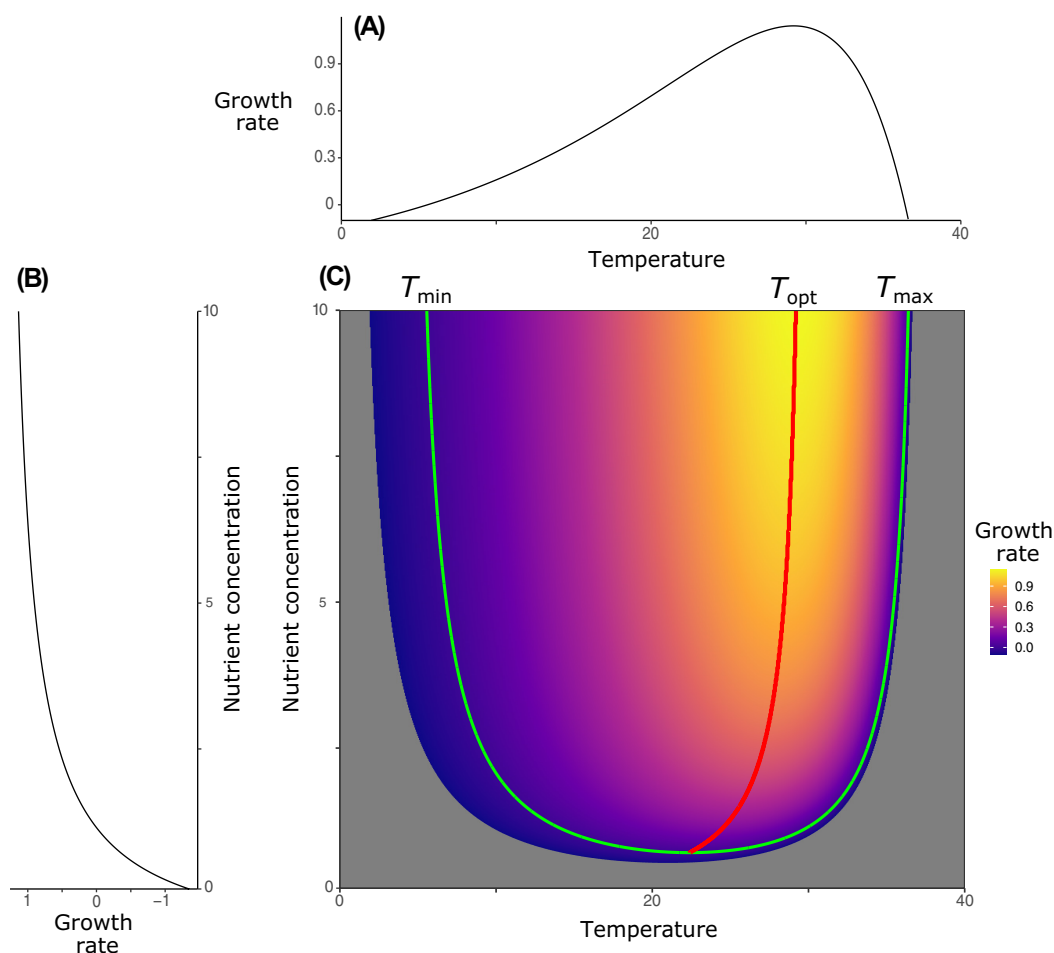


Figure 1. The dependence of population growth rate on temperature and nutrient concentration. (A) A typical thermal performance curve (TPC). (B) Growth dependence on nutrient concentration. (C) The growth rate surface as a function of temperature and nutrients, based on a model and data from [Thomas et al. \(2017\)](#). Growth rate is highest when the temperature is at T_{opt} and nutrient concentration is high. Growth rates below -0.1 are suppressed to highlight variation in positive values.

The available evidence suggests that TPC dependence on environmental factors is widespread. The pattern that emerges across taxa and environmental factors is that T_{opt} and T_{max} (as well as maximum growth rate) decline in stressful conditions such as resource limitation ([Fig. 1C](#)). In other words, organisms are more sensitive to high temperatures when deprived of resources or subjected to extremes in other environmental dimensions. In phytoplankton, major oceanic primary producers, nutrient limitation has been shown to not only decrease their maximum population growth rates but also lower their T_{opt} by 3–15°C ([Thomas et al. 2017](#), [Bestion et al. 2018](#), [Boyd 2019](#)). Light limitation also decreases T_{opt} in phytoplankton by about 4°C on average ([Edwards et al. 2016](#)) and as much as 18°C in well-resolved cases ([Kovács et al. 2016](#)); it also increases vulnerability to high temperature in seagrasses ([Kendrick et al. 2019](#)). In kelp, nitrogen limitation reduced high temperature tolerance ([Fernández et al. 2020](#)). In maize, one of the world's most important crops, high temperatures reduced yield three-fold more per °C with increasing water limitation (30% versus 10% at higher water availability) ([Anderson et al. 2015](#)). In another study,

irrigated maize was not negatively affected by high temperatures ([Carter et al. 2016](#)), though this is partly due to cooling by evaporation ([Siebert et al. 2017](#)).

This resource-dependence of temperature responses is not limited to photosynthetic organisms. Food reduction decreased T_{opt} and T_{max} by approximately 3–7°C in the freshwater ciliate *Urotricha farcta* ([Weisse et al. 2002](#)) and the marine flagellate *Oxyrrhis marina* ([Kimmance et al. 2006](#)). The decline in optimum temperature occurs in fish as well: at low food availability, T_{opt} for somatic growth declines by approximately 10°C in salmon ([Brett 1971](#)) and several degrees in coral reef damselfish larvae (precise values could not be quantified) ([McLeod et al. 2013](#)). The salmon study also showed a decrease in T_{max} of approximately 10°C. Food limitation also decreased T_{opt} and T_{max} for population growth rate by about 6°C in mosquitoes ([Huxley et al. 2021](#)). In some studies, however, a zooplankter *Daphnia*'s survival at high temperature was higher when fed low phosphorus algae or poor food quality cyanobacteria ([Starke et al. 2021](#), [Sarrazin and Sperfeld 2022](#)), suggesting that the effects of resource limitation may differ depending on the temporal scale of responses ([Sarrazin and Sperfeld 2022](#)).

Other environmental drivers and biotic interactions modify TPCs in a manner that is often similar to the effects of resource limitation. Salinity reduction lowered T_{opt} and T_{max} in phytoplankton from an estuary by 2–10°C (Bill et al. 2016). Frogs infected by chytrid pathogens had a reduced tolerance for high temperatures, experiencing spasms at temperatures ~4°C lower than uninfected frogs (Greenspan et al. 2017). Reductions in high-temperature tolerance were also found in *Daphnia* infected by bacterial pathogens, in both short- and long-term experiments (Vale et al. 2008, Hector et al. 2019, Laidlaw et al. 2020). We note that some of these biotic interaction experiments used short-term assays that are substantially different from those used when studying abiotic interactions, and so some caution in comparing measurements is warranted.

In contrast, mutualistic interactions increase heat tolerance in a wide range of taxa. In both wild and crop plants, fungal symbionts increase tolerance of high temperatures as well as other stresses such as drought (Redman et al. 2002, Rodriguez et al. 2008, Hubbard et al. 2014). In one extraordinary example, the host plant's heat tolerance is dependent on both a fungal endophyte and a virus that infects the fungus (Márquez et al. 2007). The gut microbiota increased heat tolerance of fruit flies (Jaramillo and Castañeda 2021) and tadpoles (Fontaine et al. 2022), and the symbionts of corals increased the thermal tolerance of holobionts (Berkelmans and van Oppen 2006, Pelosi et al. 2021). While the mechanisms of these interactions are not well understood, they alter high-temperature tolerance in a manner consistent with resource change: deprivation reducing T_{max} in the case of parasites/pathogens (due to reallocation towards combating infection) and supply increasing T_{max} in the case of mutualists.

Theoretical investigation of interactions between temperature and other drivers has been limited, but at least two recent models have examined temperature–resource (nutrient/food) interactions influence populations, or equivalently, how resource limitation alters TPCs (Thomas et al. 2017, Huey and Kingsolver 2019). Thomas et al. (2017) developed a simple model of temperature–resource interactions that separates the effects of the two factors on birth and death processes. Huey and Kingsolver (2019) formulated a bioenergetic model that focusses on the thermal sensitivities of energy gain and metabolism. Despite their structural differences, both models come to a similar conclusion: T_{opt} and T_{max} are saturating functions of resource concentration, consistent with the empirical findings described earlier. Both models also predict that T_{min} is altered as well, with low resources reducing cold tolerance in a similar manner. Although fewer studies have examined T_{min} , high N availability appears to increase cold tolerance in plants (Taulavuori et al. 2014, Toca et al. 2017).

Consequences of interactions of temperature with other environmental factors

Using TPCs obtained in otherwise benign conditions – with no resource limitation or other environmental stress – to

predict species survival and shifts in their geographic ranges is likely to underestimate the negative effects of warming. This is because in most habitats, environmental factors are at stressful levels at least part of the time. The observed dependence of thermal traits on other environmental factors has many consequences for organisms, populations and communities that need to be accounted for when predicting the effects of rising temperature and preparing for the future. Here we outline several such consequences that should be investigated.

- 1) Aquatic and terrestrial ecosystems with pronounced resource limitation may be more adversely affected by warming than ecosystems that are not resource-limited. Nutrient (nitrogen, phosphorus or iron) limitation is widespread in the oceans and is predicted to become even more prevalent in the future (Sarmiento et al. 2004, Hayashida et al. 2020). On land, vast regions are also limited by P, N or co-limited by more than one nutrient (Du et al. 2020, Hou et al. 2021). Aridification of the land surface is also increasing, especially in the subtropics, thus increasing areas with water limitation (Shi et al. 2021). Because resource limitation decreases T_{opt} and T_{max} (Thomas et al. 2017, Huey and Kingsolver 2019), a simultaneous reduction in resource availability alongside increasing temperatures is likely to be substantially worse than warming alone, affecting broad swathes of the globe (Busseni et al. 2020). Identifying areas that are undergoing changes in temperature as well as the type and degree of resource limitation (Hayashida et al. 2020) could help pinpoint communities that are especially vulnerable to climate change. Figure 2 shows global ocean nitrate concentration, temperature and the regions where the lowest nitrate concentration and highest temperatures overlap. Such areas appear predominantly in the tropics. T_{opt} values of tropical phytoplankton measured under replete nutrient conditions are very close to current ambient temperatures (Thomas et al. 2012), and so the declines of T_{opt} due to nutrient limitation are likely to be especially detrimental there, assuming temperatures rise or nutrients decline further in these regions. Tropical terrestrial organisms' T_{opt} values are also close to ambient temperatures (Deutsch et al. 2008), raising the possibility that food declines will reduce heat tolerance on land as well. In terrestrial plants, most studies focus either on the effects of single stressors, namely temperature and water limitation or on their interactions (Fahad et al. 2017, El Haddad et al. 2021). The next step should be investigating the effects of nutrient limitation on plant sensitivity to high temperatures, both at high and low water availability.
- 2) Heat stress may increase resource requirements while impairing the ability to acquire nutrients/food, such as by damaging nutrient transport mechanisms or reducing time available for foraging. This exacerbates both the harmful effects of resource limitation and high temperatures, causing a harmful positive feedback loop termed a 'metabolic meltdown' (Giri et al. 2017, Gerech et al. 2018, Huey and Kingsolver 2019). These feedback loops

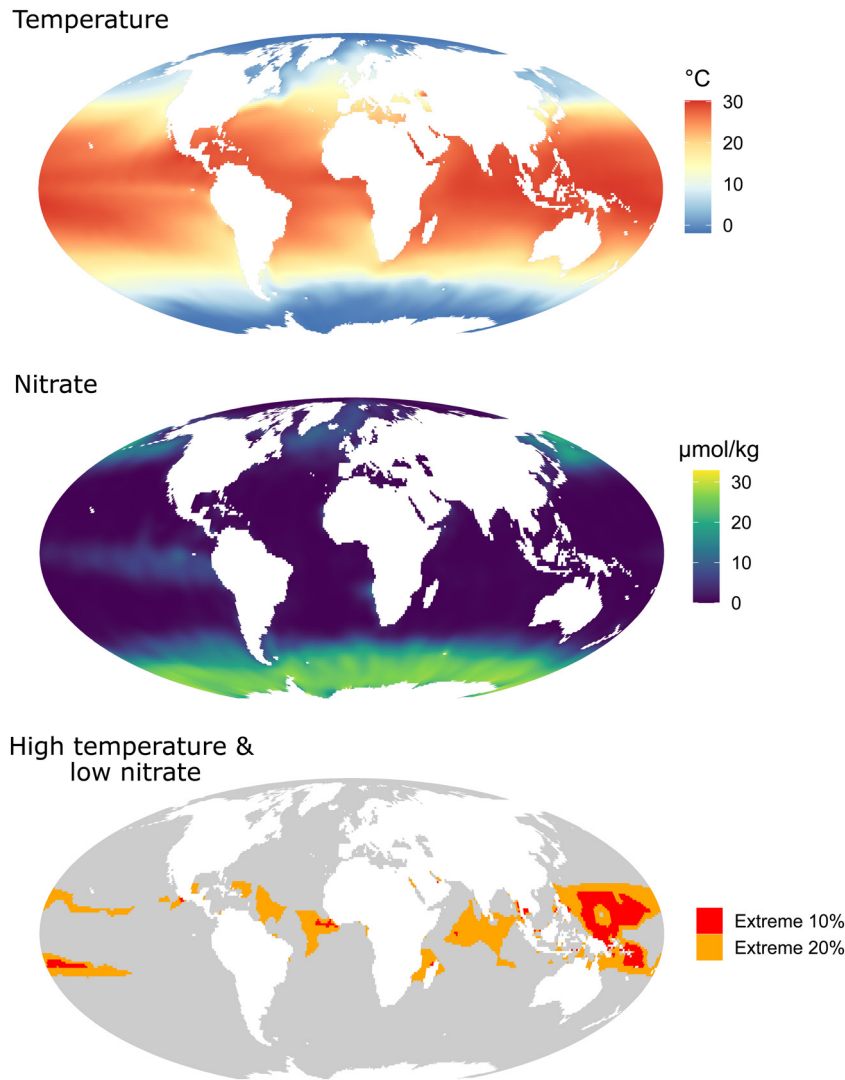


Figure 2. Oceanic regions where temperature–nutrient interactions are most likely to be limiting phytoplankton growth and shaping ecosystem dynamics. The bottom map highlights oceanic locations where temperatures are near their maximum and nitrate concentrations near their minimum. Red indicates regions where temperature is in the top 10% and nitrate in the bottom 10%. Orange uses a 20% threshold for both instead. For both variables, we use annual mean values and ignore other factors that also shape growth. Data source: World Ocean Atlas 2018 (Garcia et al. 2018, Locarnini et al. 2019).

are underexplored, and may already play an important role in organismal performance in the tropics and during heat waves.

- 3) The available evidence is consistent with a simple prediction: environmental conditions (both abiotic and biotic) that reduce growth rate also reduce T_{opt} and T_{max} relative to benign conditions. This remains to be tested rigorously across a wide range of conditions. But if true, it would offer us a simple and powerful tool because quantifying changes in growth rate is often cheaper, faster and easier than quantifying T_{opt} and T_{max} directly. Predicting how other environmental changes will shape tolerance of high temperatures would become easier if general eco-physiological patterns such as this exist.
- 4) Because species differ in resource requirements, the same resource levels are limiting to some species and

not others (Grover 1997, Edwards et al. 2012). These differences may increase the differences in vulnerability to high temperatures and therefore change community composition. Good nutrient competitors may have their TPCs relatively unchanged by decreases in resources while poor nutrient competitors experience decreases in T_{opt} and T_{max} that make them more sensitive to warming and reduce their relative abundance in the community. An additional source of complexity that we do not discuss here is that resource competitive abilities are also a function of temperature (Tilman et al. 1981, Lewington-Pearce et al. 2019). These feedbacks between temperature and nutrients have not yet been adequately incorporated into our predictions of the global change effects on organisms, populations and communities.

- 5) Within species, populations located in low-resource regions today – such as the oceanic gyres or drylands – may be a valuable source of genetic diversity. Being adapted to low resource levels, they may be better able to tolerate high temperatures under high-nutrient conditions than populations presently living in high-resource regions. They could therefore form a reservoir of (relative) heat tolerance. Heat waves in adjacent high-resource regions may provide opportunities for immigrants from low-resource environments by removing competitors adapted to high-resource conditions. These preadapted genotypes can either disperse into novel environments on their own or be transplanted deliberately to rescue declining populations (Bay et al. 2017).
- 6) Just as species are expected to migrate towards cooler regions, species from hot environments that also experience other stresses at present (low resource availability or low pH, for example) may survive by migrating towards high-resource or moderate pH environments. Such migrations may favour the persistence of otherwise vulnerable taxa. This complicates predictions of extirpation and extinction based solely on thermal limits. It can also lead to more complex spatial and temporal patterns of community reorganization than presently envisioned. A species that persists by migrating towards high-resource or moderate pH environments necessarily competes with resident taxa, possibly causing extirpations. This complex outcome of environmental warming will be hard to predict or model, but properly accounting for interactions is a necessary step towards achieving this.
- 7) The interacting effects of temperature and resources also cascade through food webs. If prey species decline due to warming, this may trigger a similar temperature-food interaction problem in their predators. They may become more sensitive to high temperatures due to resource (food) limitation, and this may amplify the negative effects of warming on consumers. While some studies are starting to address the indirect effects of temperature on food webs (Gibert 2019), we know very little about how resource limitation will shape the TPCs of different trophic levels. We need to incorporate such interactions for multiple trophic levels into food web models to realistically model multiple driver effects on food webs.
- 8) Phenological shifts can also change resource availability for different trophic levels (Visser and Both 2005, Nord and Lynch 2009) and may therefore increase vulnerability to high temperatures. Flowering plants in peak summer may be an especially important resource for local pollinator communities and their predators. Shifts in flowering times leading to lower resource availability for pollinators (Solga et al. 2014) may make pollinators – especially the specialists – more vulnerable to high temperatures, including heat waves. Changes in fruiting times may also have important effects on consumer species' heat tolerances.
- 9) Selection on temperature tolerance is likely much stronger in nature than anticipated from lab studies, because

of the increased heat stress associated with periods of low food and other stresses. Evolutionary adaptation to high temperature may itself be affected by resource availability and other environmental drivers. Under the suboptimal levels of other environmental factors, adaptation to rising temperatures may either be faster due to stronger selection or be slowed down or arrested, if there are trade-offs between temperature tolerances and resource requirements (Aranguren-Gassis et al. 2019), in addition to the simpler reason that population sizes may be reduced. Evolution experiments under different combinations of environmental drivers and temperature would help determine how driver interactions affect thermal adaptation.

- 10) Fertilization practices in agriculture are likely to be especially important to consider as the climate warms. Although excess fertilization is a major environmental concern because of the consequent greenhouse gas emissions (Tian et al. 2020) and aquatic eutrophication (Conley et al. 2009), preventing periods of nutrient limitation in plants could provide protection against heat waves. Nutrient supply has been proposed as a crop cultivation strategy to offset the negative effects of high temperatures (Waraich et al. 2012).
- 11) Because some mutualistic interactions appear to increase high-temperature tolerance (at least in plants), developing and using crop mutualists may alleviate future increases in heat stress. Mutualists from hot environments may improve high-temperature performance of existing crops, a phenomenon known as 'habitat-adapted symbiosis' (Rodriguez et al. 2008). Conversely, disease will likely reduce tolerance for high temperatures. Optimizing crops for future heat waves may require approaches that increase mutualist abundance and resource availability while decreasing disease prevalence. These are likely steps that would be useful even in the absence of warming, but protection from heat stress makes them even more valuable. Possible trade-offs that reduce the possibility of accomplishing these goals are worth investigating to improve crop performance, such as between receptiveness to mutualism and resistance against infection.

Conclusions

Across different organisms and ecosystems, a variety of abiotic and biotic drivers modify organisms' ability to tolerate high temperatures. Because these effects appear so widespread, we need to explicitly consider how temperature interacts with other environmental factors, including global change stressors, to develop better predictions of how warming will affect species and communities. So far, most research on environmental driver interactions with temperature has focused on nutrients and water availability, but the effects of many other environmental factors – especially biotic ones – on TPCs remains underexplored. A focused research agenda to investigate systematically the effects of multiple interacting stressors on species' TPCs from a wide range of habitats in oceanic, freshwater and terrestrial ecosystems (including agricultural

systems), would align well with the ongoing efforts to implement the multiple driver/multistressor framework in global change research (Boyd et al. 2019, Wake 2019). Among the key topics to address are: how universal the negative effects of other stressors on high temperature tolerances are, the magnitudes and the mechanisms of the effects and whether adding more than one or two stressors exacerbates thermal sensitivity even further. New research would help to better assess the effects of global warming on species growth, future geographic ranges, productivity and biodiversity. Moreover, it is essential for developing predictive models for conservation, agriculture, fisheries and climate change mitigation.

Acknowledgements – This is Kellogg Biological Station contribution 2325. Open Access funding enabled and organized by Projekt DEAL.
Funding – The work presented in this article results, in part, from a grant to SCOR from the US National Science Foundation (OCE-1840868) to the Changing Oceans Biological Systems project and the NSF grants OCE-1638958 and DEB-1754250 to EL and C. Klausmeier.

Author contributions

Elena Litchman: Conceptualization (equal); Writing – original draft (equal); Writing – review and editing (equal).
Mridul K. Thomas: Conceptualization (equal); Writing – original draft (equal); Writing – review and editing (equal).

Data availability statement

Data sharing is not applicable to this article as no new data were created or analyzed in this study.

References

- Addo-Bediako, A. et al. 2000. Thermal tolerance, climatic variability and latitude. – *Proc. R. Soc. B* 267: 739–745.
- Aldea-Sánchez, P. et al. 2021. Heat tolerance, energetics and thermal treatments of honeybees parasitized with *Varroa*. – *Front. Ecol. Evol.* 9: 656504.
- Anderson, C. J. et al. 2015. Placing bounds on extreme temperature response of maize. – *Environ. Res. Lett.* 10: 124001.
- Aranguren-Gassis, M. et al. 2019. Nitrogen limitation inhibits marine diatom adaptation to high temperatures. – *Ecol. Lett.* 22: 1860–1869.
- Bay, R. A. et al. 2017. Genomic models predict successful coral adaptation if future ocean warming rates are reduced. – *Sci. Adv.* 3: e1701413.
- Berkelmans, R. and van Oppen, M. J. 2006. The role of zooxanthellae in the thermal tolerance of corals: a ‘nugget of hope’ for coral reefs in an era of climate change. – *Proc. R. Soc. B* 273: 2305–2312.
- Bestion, E. et al. 2018. Nutrient limitation constrains thermal tolerance in freshwater phytoplankton. – *Limnol. Oceanogr. Lett.* 3: 436–443.
- Bill, B. D. et al. 2016. Effects of temperature and salinity on the growth of *Alexandrium* (Dinophyceae) isolates from the Salish Sea. – *J. Phycol.* 52: 230–238.
- Boyd, P. W. 2019. Physiology and iron modulate diverse responses of diatoms to a warming Southern Ocean. – *Nat. Clim. Change* 9: 148–152.
- Boyd, P. W. et al. 2019. SCOR WG149 handbook to support the SCOR best practice guide for ‘multiple drivers’ marine research. – Univ. of Tasmania for Scientific Committee on Oceanic Research (SCOR).
- Brett, J. R. 1971. Energetic responses of salmon to temperature. A study of some thermal relations in the physiology and freshwater ecology of sockeye salmon (*Oncorhynchus nerka*). – *Am. Zool.* 11: 99–113.
- Brown, J. H. et al. 2004. Toward a metabolic theory of ecology. – *Ecology* 85: 1771–1789.
- Busseni, G. et al. 2020. Large scale patterns of marine diatom richness: drivers and trends in a changing ocean. – *Global Ecol. Biogeogr.* 29: 1915–1928.
- Carter, E. K. et al. 2016. Separating heat stress from moisture stress: analyzing yield response to high temperature in irrigated maize. – *Environ. Res. Lett.* 11: 094012.
- Conley, D. J. et al. 2009. Controlling eutrophication: nitrogen and phosphorus. – *Science* 323: 1014–1015.
- Deutsch, C. A. et al. 2008. Impacts of climate warming on terrestrial ectotherms across latitude. – *Proc. Natl Acad. Sci.* 105: 6668–6672.
- Du, E. et al. 2020. Global patterns of terrestrial nitrogen and phosphorus limitation. – *Nat. Geosci.* 13: 221–226.
- Edwards, K. F. et al. 2012. Allometric scaling and taxonomic variation in nutrient utilization traits and growth rates of marine and freshwater phytoplankton. – *Limnol. Oceanogr.* 57: 554–566.
- Edwards, K. F. et al. 2016. Phytoplankton growth and the interaction of light and temperature: a synthesis at the species and community level. – *Limnol. Oceanogr.* 61: 1232–1244.
- El Haddad, N. et al. 2021. High-temperature and drought stress effects on growth, yield and nutritional quality with transpiration response to vapor pressure deficit in lentil. – *Plants* 11: 95.
- Ern, R. et al. 2016. Oxygen dependence of upper thermal limits in fishes. – *J. Exp. Biol.* 219: 3376–3383.
- Fahad, S. et al. 2017. Crop production under drought and heat stress: plant responses and management options. – *Front. Plant Sci.* 8: 1147.
- Fernández, P. A. et al. 2020. Nitrogen sufficiency enhances thermal tolerance in habitat-forming kelp: implications for acclimation under thermal stress. – *Sci. Rep.* 10: 3186.
- Fontaine, S. S. et al. 2022. Experimental manipulation of microbiota reduces host thermal tolerance and fitness under heat stress in a vertebrate ectotherm. – *Nat. Ecol. Evol.* 6: 405–417.
- García, H. et al. 2018. World Ocean Database 2018. Users manual (prerelease). – NOAA Atlas NESDIS 81.
- Gerecht, A. C. et al. 2018. Phosphorus limitation and heat stress decrease calcification in *Emiliana huxleyi*. – *Biogeosciences* 15: 833–845.
- Gibert, J. P. 2019. Temperature directly and indirectly influences food web structure. – *Sci. Rep.* 9: 5312.
- Giri, A. et al. 2017. Heat stress decreases levels of nutrient-uptake and -assimilation proteins in tomato roots. – *Plants* 6: 6.
- Greenspan, S. E. et al. 2017. Infection increases vulnerability to climate change via effects on host thermal tolerance. – *Sci. Rep.* 7: 9349.
- Grover, J. P. 1997. Resource competition. – Chapman and Hall.
- Hayashida, H. et al. 2020. Background nutrient concentration determines phytoplankton bloom response to marine heatwaves. – *Global Change Biol.* 26: 4800–4811.
- Hector, T. E. et al. 2019. Pathogen exposure disrupts an organism’s ability to cope with thermal stress. – *Global Change Biol.* 25: 3893–3905.
- Hector, T. E. et al. 2021. Thermal limits in the face of infectious disease: how important are pathogens? – *Global Change Biol.* 27: 4469–4480.

- Hou, E. et al. 2021. Latitudinal patterns of terrestrial phosphorus limitation over the globe. – *Ecol. Lett.* 24: 1420–1431.
- Hubbard, M. et al. 2014. Fungal endophytes enhance wheat heat and drought tolerance in terms of grain yield and second-generation seed viability. – *J. Appl. Microbiol.* 116: 109–122.
- Huey, R. B. and Kingsolver, J. G. 2019. Climate warming, resource availability and the metabolic meltdown of ectotherms. – *Am. Nat.* 194: E140–E150.
- Huxley, P. J. et al. 2021. The effect of resource limitation on the temperature dependence of mosquito population fitness. – *Proc. R. Soc. B* 288: 20203217.
- IPCC 2021. Climate Change 2021: The physical science basis. Contribution of Working Group I to the Sixth Assessment Report of the Intergovernmental Panel on Climate Change (Masson-Delmotte, V., P. Zhai, A. Pirani, S.L. Connors, C. Péan, S. Berger, N. Caud, Y. Chen, L. Goldfarb, M.I. Gomis, M. Huang, K. Leitzell, E. Lonnoy, J.B.R. Matthews, T.K. Maycock, T. Waterfield, O. Yelekçi, R. Yu, and B. Zhou (eds.)). – Cambridge Univ. Press.
- Jaramillo, A. and Castañeda, L. E. 2021. Gut microbiota of *Drosophila subobscura* contributes to its heat tolerance and is sensitive to transient thermal stress. – *Front. Microbiol.* 12: 654108.
- Kendrick, G. A. et al. 2019. A systematic review of how multiple stressors from an extreme event drove ecosystem-wide loss of resilience in an iconic seagrass community. – *Front. Mar. Sci.* 6: 455.
- Kimmance, S. et al. 2006. Do temperature–food interactions matter? Responses of production and its components in the model heterotrophic flagellate *Oxyrrhis marina*. – *Aquatic Microb. Ecol.* 42: 63–73.
- Kovács, A. W. et al. 2016. Thermal-dependent growth characteristics for *Cylindrospermopsis raciborskii* (Cyanoprokaryota) at different light availabilities: methodological considerations. – *Aquatic Ecol.* 50: 623–638.
- Laidlaw, T. et al. 2020. Pathogen exposure reduces sexual dimorphism in a host's upper thermal limits. – *Ecol. Evol.* 10: 12851–12859.
- Lewington-Pearce, L. et al. 2019. Temperature-dependence of minimum resource requirements alters competitive hierarchies in phytoplankton. – *Oikos* 128: 1194–1205.
- Locarnini, M. et al. 2019. Temperature, World Ocean Atlas 2018. – NOAA Atlas NESDIS 81.
- Márquez, L. M. et al. 2007. A virus in a fungus in a plant: three-way symbiosis required for thermal tolerance. – *Science* 315: 513–515.
- McLeod, I. M. et al. 2013. Climate change and the performance of larval coral reef fishes: the interaction between temperature and food availability. – *Conserv. Physiol.* 1: cot024.
- Nord, E. A. and Lynch, J. P. 2009. Plant phenology: a critical controller of soil resource acquisition. – *J. Exp. Bot.* 60: 1927–1937.
- O'Donnell, D. R. et al. 2018. Rapid thermal adaptation in a marine diatom reveals constraints and trade-offs. – *Global Change Biol.* 24: 4554–4565.
- Payne, N. L. et al. 2016. Temperature dependence of fish performance in the wild: links with species biogeography and physiological thermal tolerance. – *Funct. Ecol.* 30: 903–912.
- Pelosi, J. et al. 2021. Thermally tolerant symbionts may explain Caribbean octocoral resilience to heat stress. – *Coral Reefs* 40: 1113–1125.
- Redman, R. S. et al. 2002. Thermotolerance generated by plant/fungal symbiosis. – *Science* 298: 1581.
- Ritchie, M. E. 2018. Reaction and diffusion thermodynamics explain optimal temperatures of biochemical reactions. – *Sci. Rep.* 8: 11105.
- Rodriguez, R. J. et al. 2008. Stress tolerance in plants via habitat-adapted symbiosis. – *ISME J.* 2: 404–416.
- Sarmiento, J. L. et al. 2004. Response of ocean ecosystems to climate warming. – *Global Biogeochem. Cycles* 18: GB3003.
- Sarrazin, J. and Sperfeld, E. 2022. Food quality mediates responses of *Daphnia magna* life history traits and heat tolerance to elevated temperature. – *Freshwater Biol.* 67: 1521–1531.
- Shi, H. et al. 2021. Terrestrial biodiversity threatened by increasing global aridity velocity under high-level warming. – *Proc. Natl Acad. Sci. USA* 118: e2015552118.
- Siebert, S. et al. 2017. Heat stress is overestimated in climate impact studies for irrigated agriculture. – *Environ. Res. Lett.* 12: 054023.
- Solga, M. J. et al. 2014. Timing is everything: an overview of phenological changes to plants and their pollinators. – *Nat. Areas J.* 34: 227–234, 228.
- Starke, C. W. E. et al. 2021. Interactive effects of water temperature and stoichiometric food quality on *Daphnia pulex*. – *Freshwater Biol.* 66: 256–265.
- Sunday, J. M. et al. 2011. Global analysis of thermal tolerance and latitude in ectotherms. – *Proc. R. Soc. B* 278: 1823–1830.
- Sunday, J. M. et al. 2012. Thermal tolerance and the global redistribution of animals. – *Nat. Climate Change* 2: 686–690.
- Taulavuori, K. et al. 2014. Truths or myths, fact or fiction, setting the record straight concerning nitrogen effects on levels of frost hardiness. – *Environ. Exp. Bot.* 106: 132–137.
- Thomas, M. K. et al. 2012. A global pattern of thermal adaptation in marine phytoplankton. – *Science* 338: 1085–1088.
- Thomas, M. K. et al. 2017. Temperature–nutrient interactions exacerbate sensitivity to warming in phytoplankton. – *Global Change Biol.* 23: 3269–3280.
- Tian, H. et al. 2020. A comprehensive quantification of global nitrous oxide sources and sinks. – *Nature* 586: 248–256.
- Tilman, D. et al. 1981. Competition and nutrient kinetics along a temperature gradient – an experimental test of a mechanistic approach to niche theory. – *Limnol. Oceanogr.* 26: 1020–1033.
- Toca, A. et al. 2017. Species ecology determines the role of nitrogen nutrition in the frost tolerance of pine seedlings. – *Tree Physiol.* 38: 96–108.
- Tomanek, L. 2008. The importance of physiological limits in determining biogeographical range shifts due to global climate change: the heat-shock response. – *Physiol. Biochem. Zool.* 81: 709–717.
- Tüzün, N. and Stoks, R. 2018. Evolution of geographic variation in thermal performance curves in the face of climate change and implications for biotic interactions. – *Curr. Opin. Insect Sci.* 29: 78–84.
- Vale, P. F. et al. 2008. Temperature-dependent costs of parasitism and maintenance of polymorphism under genotype-by-environment interactions. – *J. Evol. Biol.* 21: 1418–1427.
- Visser, M. E. and Both, C. 2005. Shifts in phenology due to global climate change: the need for a yardstick. – *Proc. R. Soc. B* 272: 2561–2569.
- Wake, B. 2019. Experimenting with multistressors. – *Nat. Clim. Change* 9: 357–357.
- Waraich, E. A. et al. 2012. Alleviation of temperature stress by nutrient management in crop plants: a review. – *J. Soil Sci. Plant Nutr.* 12: 221–244.
- Weisse, T. et al. 2002. Interactive effect of temperature and food concentration on growth rate: a test case using the small freshwater ciliate *Utricha farcta*. – *Limnol. Oceanogr.* 47: 1447–1455.