

1 Molecular Investigation of the Multi-Phase Photochemistry of Fe(III)-
2 Citrate in Aqueous Solution

3
4

5 Christopher P. West¹, Ana C. Morales¹, Jackson Ryan¹, Maria V. Misovich¹, Anusha P. S.
6 Hettiyadura¹, Felipe Rivera-Adorno¹, Jay M. Tomlin¹, Andrew Darmody², Brittany N. Linn¹, Peng
7 Lin^{1,a}, Alexander Laskin^{1,3*}

8

9 ¹Department of Chemistry, Purdue University, West Lafayette, IN, USA

10 ²Department of Aeronautics and Aerospace Engineering, Purdue University, West Lafayette, IN

11 ³Department of Earth, Atmospheric & Planetary Sciences, Purdue University, West Lafayette,
12 IN, USA

13

14 ^aNow at California Air Resource Board (CARB), 1001 I Street, Sacramento, CA, USA

15

16 *Correspondence to: alaskin@purdue.edu

17

18

19

20

21

22

Manuscript Submitted to:

23

24

Environmental Science: Processes & Impacts

25

26

27

28

29

April 8, 2022

30

31 **Abstract:** Iron (Fe) is ubiquitous in nature and found as Fe^{II} or Fe^{III} in minerals or as dissolved
32 ions Fe²⁺ or Fe³⁺ in aqueous systems. The interactions of soluble Fe have important implications
33 on fresh water and marine biogeochemical cycles, which have impacts on global terrestrial and
34 atmospheric environments. Upon dissolution of Fe^{III} into natural aquatic systems, organic
35 carboxylic acids efficiently chelate Fe^{III} to form [Fe^{III}-carboxylate]²⁺ complexes that undergo a
36 wide range of photochemistry-induced radical reactions. The chemical composition and
37 photochemical transformations of these mixtures are largely unknown, making it challenging to
38 estimate their environmental impact. To investigate photochemical processing of Fe^{III}-
39 carboxylates at molecular-level, we conduct comprehensive experimental study employing UV-
40 visible spectroscopy, liquid chromatography coupled to photodiode array and high-resolution mass
41 spectrometry, and oil immersion flow microscopy. In this study, aqueous solutions of Fe^{III}-citrate
42 were photolyzed under 365 nm light in an experimental setup with an apparent quantum yield of
43 (Φ) \sim 0.02, followed by chemical analyses of reacted mixtures withdrawn at increment time
44 intervals of the experiment. The apparent photochemical reaction kinetics of Fe³⁺-citrate (aq)
45 were expressed as two generalized consecutive reactions of *Reactants* (*R*) $\xrightarrow{j_1}$ *Intermediates* (*I*) $\xrightarrow{j_2}$
46 *Products* (*P*) with the experimental rate constants of $j_1 \sim 0.12 \text{ min}^{-1}$ and $j_2 \sim 0.05 \text{ min}^{-1}$,
47 respectively. Molecular characterization results indicate that *R* and *I* consist of both water-soluble
48 organic and Fe-organic species, while *P* compounds are a mixture of water-soluble and colloidal
49 materials. The latter were identified as Fe-carbonaceous colloids formed at long photolysis times.
50 The carbonaceous content of these colloids was identified as unsaturated organic species with low
51 oxygen content and carbon with reduced oxidation state, indicative of their plausible radical
52 recombination mechanism at oxygen-deprived conditions typical for the extensively photolyzed
53 mixtures. Based on the molecular characterization results, we discuss the comprehensive reaction
54 mechanism of Fe^{III}-citrate photochemistry and report on the formation of previously unexplored
55 colloidal reaction products, which may contribute to atmospheric and terrestrial light-absorbing
56 material in aquatic environments.

57

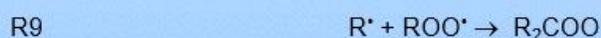
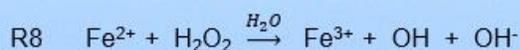
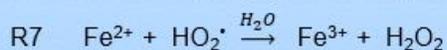
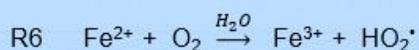
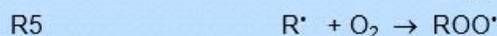
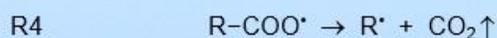
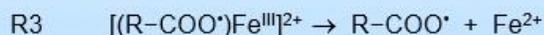
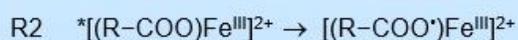
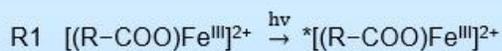
58 **Keywords:** water soluble Fe, Fe-carboxylate complexes, photolysis, reactive oxygen species
59 (ROS), semi-solid colloids, liquid chromatography (LC), UV-visible spectroscopy, high resolution
60 mass spectrometry (HRMS), flow microscopy

61 INTRODUCTION

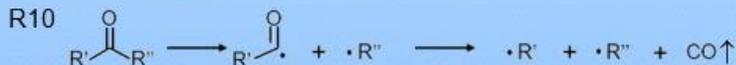
62 Iron (Fe) is the most common element on Earth by mass, and found in atmosphere, biosphere,
63 lithosphere, and hydrosphere.^{1,2} A majority of Fe is associated with the solid form of rocks of
64 crustal matter in oxygen-deficient settings,³ whereas soluble and colloidal Fe-containing species
65 are present in oceanic and terrestrial water environments.⁴⁻⁶ Atmospheric Fe is also naturally
66 abundant in wind-blown mineral dust and other components of particulate matter (PM).⁷⁻¹³ The
67 interactions, transport, and environmental fate of elemental Fe are vital in the Fe biogeochemical
68 cycle and necessary to sustain life and facilitate important environmental processes.^{14,15} The
69 soluble fractions of Fe from Fe-bearing materials¹⁶⁻¹⁸ are mobilized in aquatic environments as a
70 result of various chemical processes such as photolysis and reactions with inorganic acids. Upon
71 dissolution of Fe into aquatic aerobic systems (pH 4 – 8), available organic carboxylic acid
72 components of dissolved organic matter (DOM) efficiently chelate Fe³⁺ ions to form
73 photocatalytically active [Fe^{III}-carboxylate]²⁺ complexes that promote radical reactions in the
74 aquatic phase. Upon excitation by UV-visible light, the photochemically initiated radical reactions
75 modify the composition of DOM with subsequent effects on optical and physical properties of the
76 environmental aquatic systems.¹⁹

77 Molecular-level characterization of reactive components in the Fe-carboxylate photochemical
78 systems is still largely lacking and reported results remain ambiguous. Previous studies utilized
79 optical spectroscopy,²⁰⁻²⁴ electrochemistry,²¹ and liquid chromatography with optical detection²⁵
80 to infer reaction mechanisms and system properties solely from measurements of the bulk solution,
81 but not for the molecular-specific components. However, unravelling the molecular-specific
82 processes leading to the conversion of solid Fe^{III} to soluble Fe³⁺, Fe³⁺ complexation with carboxylic
83 acids, and the environmental reactions of [Fe^{III}-carboxylate]²⁺ complexes are still needed for
84 quantitative predictions of Fe^{III}-carboxylate photochemistry in aquatic systems.

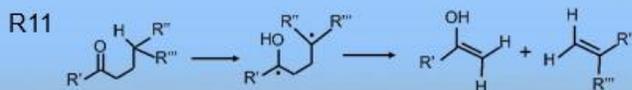
Photocatalytic cycle of Fe^{III}(Cit)



Norrish Reactions



Norrish Type I

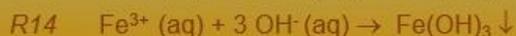


Norrish Type II

Radical Recombination



Fe^{III/II} sedimentation



85
86 **Scheme 1.** Reaction sequence of the photo-catalytic redox cycle of Fe^{III}-citrate complex in aquatic
87 environment. R in reactions R1-R5 and R9 corresponds to -(CH₂CO₂H)₂COH, 'R, "R, and ""R symbols
88 indicate various parts of organic molecules. Mononuclear monocitrate (or 1:1 complex, [(R-COO)Fe^{III}]²⁺)
89 is shown here as a representative complex in the reaction scheme. Herein, legends of Fe²⁺ and Fe³⁺ refer to
90 dissolved ions, whereas legends of Fe^{III} and Fe^{II} indicate complexation with ligands.

91
92 Photoreactive Fe^{III}-carboxylate complexes absorb light at wavelengths below 500 nm, which
93 catalyzes ligand to metal charge transfer (LMCT) reactions, resulting in the reduction of Fe^{III} to
94 Fe^{II} and oxidation of the carboxylate ligands, a process that represents an important sink of organic
95 acids in marine, terrestrial, and atmospheric water.^{22,26-31} Scheme 1 summarizes these and other

96 relevant aquatic reactions. The LMCT process occurs as a result of an electron transfer from the
97 innermost sp^3 orbital of the carboxylate ligand to the metal center, forming $[Fe^{III}-COO-R]^{2+*}$ -
98 excited state and the $[Fe^{III}-COO^{\bullet}-R]^{2+}$ radical complex, as shown in R1-R2.³² The equivalent
99 lifetime of the intermediate radical complex, determined in previous studies,^{24,33-35} was found to
100 be on the order of a few milliseconds, after which it decomposes into Fe^{2+} and $R-COO^{\bullet}$ as shown
101 in R3. Further reactions of Fe^{2+} and $R-COO^{\bullet}$ lead to loss of CO_2 as shown in R4, forming alkyl
102 radicals (R^{\bullet}).³⁶⁻³⁸ Organic peroxy radicals (RO_2^{\bullet}) are later formed in the reaction with oxygen
103 (O_2) as shown in R5.³⁹ Additional reactions of Fe^{2+} ions with dissolved O_2 lead to the formation of
104 reactive oxygen species (ROS)⁴⁰⁻⁴² such as HO_2^{\bullet} , hydrogen peroxide (H_2O_2), and hydroxyl radicals
105 (OH^{\bullet})⁴³⁻⁴⁵ while cycling Fe^{2+} back to Fe^{3+} through the Fenton reaction,⁴⁶ as shown in R6-8. The
106 ROS species react further with a variety of organic and inorganic compounds in aqueous
107 solution.⁴⁷⁻⁵⁰ Subsequently, organic peroxy radicals (ROO^{\bullet}) in solution may decompose into
108 smaller organic acid fragments such as oxygenated volatile organic compounds (OVOC), or
109 undergo radical recombination reactions as shown in R9, and R12 resulting in production of larger
110 oligomeric species.

111
112 Radical species with carbonyl groups formed after decarboxylation, may undergo Norrish type I
113 and II reactions as shown in R10-11,⁵¹⁻⁵⁴ triggering radical chain reactions with O_2 and various
114 organic compounds. Norrish reactions produce carbon monoxide (CO) and carbon-centered
115 radicals (CCRs), such as methyl ($^{\bullet}CH_3$), methyl ketones ($O=C-CH_3$), and
116 triphenylmethyl ($^{\bullet}C(C_6H_5)_3$), among other radicals.⁵⁵ As a result, the dissolved oxygen is
117 effectively consumed, while the CO_2 and CO products are degassed, resulting in oxygen-deficient
118 conditions after prolonged photolysis. As shown in R12, radical recombination forms compounds
119 with lower oxygen content and carbon in reduced oxidation state. These products have much lower
120 solubility than oxygenated organic carbon and therefore may contribute to the buildup of insoluble
121 photomineralization products.^{56,57} In addition to the reduced carbon colloids, precipitation of
122 insoluble iron hydroxides,¹⁹ $Fe(OH)_2$ and $Fe(OH)_3$, can take place as shown in R12-13,^{58,59} which
123 further contributes to the formation and growth of colloid particles. The insoluble colloids of
124 mixed $Fe(OH)_3$ /carbonaceous composition exhibit substantially larger light extinction properties
125 than the water-soluble components in the same mixtures.⁶⁰

126

127 Multiphase aquatic photochemistry of Fe^{III}-citrate has been studied^{20,25,41,55,61,62} as a laboratory
128 proxy for the environmental Fe^{III}-carboxylate systems. To date, the production and detection of
129 gas-phase OVOC upon photolysis of Fe^{III}-secondary organic aerosol proxies containing carboxylic
130 acids in flow reactors⁶³ have been investigated. Other studies reported changes in the
131 microphysical properties, chemical composition, and ROS capacity of viscous Fe-citrate particles
132 levitated in electrodynamic balance and probed by spectromicroscopy after exposure to UV
133 light.^{41,55} These studies^{41,55} demonstrated significant photochemical degradation after 24 hours of
134 irradiation, resulting in ~ 80% mass loss of particles due to degassing of CO₂ and OVOC, and slow
135 uptake/molecular diffusion of O₂ in the particle phase, consistent with reactions R1-8 of Scheme
136 1. The combination of these studies provides surface and physical property characterization
137 results. However, molecular-level composition, optical properties, and extent of the photochemical
138 reactions occurring in the condensed phase are still insufficiently investigated. This knowledge
139 gap precludes quantitative predictions of the photocatalytic processing for Fe-carboxylate systems
140 and their consequences on environmental systems.

141
142 In this work, the chemical composition and optical properties of aqueous Fe^{III}-citrate
143 photocatalytic components were investigated using high-performance liquid chromatography
144 coupled to a photodiode array and high-resolution mass spectrometry (HPLC-PDA-HRMS)
145 detectors,⁶⁴⁻⁷¹ and were further evaluated with direct infusion (DI) HRMS chemical analysis.^{66,70,72}
146 Colloidal products produced during photolysis reactions were imaged with *in situ* oil immersion
147 flow microscopy. Multi-modal datasets from these complementary techniques provide a unique
148 experimental description of various stages of Fe^{III}-citrate photochemistry, elucidate individual
149 components of this reacting system, determine mechanistic insights, and quantify environmental
150 parameters affecting the Fe^{III}-carboxylate photochemistry.

151
152

153 **EXPERIMENTAL METHODS**

154
155 **Sample Preparation** 10mM stock solutions of ferric citrate (C₆H₅O₇Fe, 16.5 – 18.5 % Fe basis,
156 BioReagent, CAS: 3522-50-7; Sigma Aldrich Inc.) and of citric acid (C₆H₈O₇·H₂O; 99% purity,
157 Mallinckrodt Inc.) analytical blank for HPLC analysis were prepared in 100mL borosilicate
158 volumetric flasks using 18.1 MΩ cm⁻¹ Milli-Q ultrapure water. The stock solutions were wrapped

159 in aluminum foil (Fisher Inc.) and stored in a refrigerator at ~ 5 °C. The ferric citrate stock solution
160 were further diluted to $\sim 9.0 \times 10^{-5}$ M (90 μ M) in the ultrapure water and were set to equilibrate
161 for ~ 2 -24 h in the dark prior to chemical characterization, as recommended in the literature.^{73,74}

162
163 **Photochemical Experiments and Optical Spectroscopy measurements.** A custom-built
164 photolysis reactor setup employing a constant wavelength light emitting diode (LED) that emits
165 blue light at $\sim 365 \pm 9$ nm (Model: M365LP1; Thor Labs Inc.) was placed ~ 20 cm above the
166 sample compartment (qpod2e, Quantum Northwest Inc.). The sample was placed in a 10 mm
167 quartz cuvette (Vernier Inc.). The mounted UV- LED optics and adjusted collimator lens in these
168 experiments resulted in an illuminated cross-sectional area of 0.78 ± 0.05 cm². All experiments
169 were conducted at room temperature (~ 24 -26 °C), and temperature was monitored throughout the
170 reaction. Additional details of the experimental methods and instrumentation parameters are
171 described in Figure S1 (Appendix A of the SI file). UV-Visible absorption spectra of the
172 photolyzed solutions were acquired using a USB 2000 UV-Vis-NIR (Ocean Optics Inc.) fiber
173 optic spectrophotometer. The samples were analyzed over the 200 – 900 nm wavelength range
174 employing 20 ms integration time, 6 scans to average, and boxcar width of 4.

175
176 The spectral flux of the LED radiation (F_{LED}) experienced by the samples was experimentally
177 determined by the chemical actinometry method using a 1 mM solution of potassium ferrioxalate
178 ($K_3[Fe(C_2O_4)_3]$; Sigma-Aldrich) in 0.05M sulfuric acid (Honeywell Inc.), described by Lehóczyki
179 et al., 2013.⁷⁵ The decrease in absorption of ferrioxalate ($\epsilon_{Fe-Ox} \sim 312 \pm 2$ L mol⁻¹ cm⁻¹ at 390 nm)⁷⁵
180 as a function of irradiation time was measured and converted to the $F_{LED} \sim 3.85 \times 10^{15}$ photons
181 cm⁻² s⁻¹ at 365 nm, with a quantum yield, $\Phi \sim 1.26$. Description and results of flux calculations
182 is provided in Appendix B and Figure S2 of the SI file.

183
184 **Quantitation of Dissolved Fe²⁺** Quantification of Fe²⁺ concentrations in irradiated solutions
185 employed colorimetric tests with 1,10-phenanthroline (C₁₂H₈N₂; $\geq 99\%$ purity, Sigma-Aldrich
186 Inc.), whereby dissolved Fe²⁺ ions and 1,10-phenanthroline form a red-orange complex that
187 absorbs visible light at ~ 510 nm.^{76,77} For analysis, 0.5 mL of analyte sampled at 0, 5, 20, 40, 80,
188 and 120 min irradiation time were mixed with 1.5 mL of 5mM 1,10-phenanthroline and filled to
189 3mL with 1mL of 18.1 MΩ cm⁻¹ Milli-Q ultrapure water. Quantitative [Fe²⁺] measurements were

190 performed in triplicate to establish reproducibility. For calibration, chemical standards of Fe^{II}
191 perchlorate hydrate (Fe(ClO₄)₂·H₂O; 98% purity, Sigma Aldrich Inc.), ranging from 1-90 μM
192 concentrations, were mixed with 5 mM 1,10- phenanthroline, and their respective absorbances
193 were measured and recorded via a seven-point calibration curve (Figure S3 in SI file). The limit
194 of detection (LOD) and molar absorptivity ($\epsilon_{510\text{nm}}$) obtained from calibration measurements in this
195 study are ~ 2.3 μM and ~ 7400 L mol⁻¹ cm⁻¹, respectively.

196
197 **Chemical Characterization** 300 uL aliquots of the initial (non-photolyzed) and irradiated
198 mixtures were withdrawn from the cuvette containing 3 mL of experimental solutions at 0, 5, 40,
199 80, and 120 min. Components of the Fe-citrate mixtures were characterized using a VanquishTM
200 high performance liquid chromatography (HPLC) system equipped with an autosampler, a
201 photodiode array (PDA) detector, and a Q ExactiveTM HF-X Orbitrap mass spectrometer
202 interfaced with an IonMAX source which housed a heated electrospray ionization (HESI) probe
203 operated in negative ion mode (all from Thermo Scientific Inc.). Chemical constituents were
204 separated on a zwitterionic hydrophilic interaction liquid chromatography (ZIC-HILIC) column
205 (SynchronisTM, 2 mm × 150 mm, 5 μm particles, 100 Å pore size, Thermo Inc.). ZIC-HILIC
206 column can retain ionic and highly polar, hydrophilic compounds, through liquid-liquid
207 partitioning, electrostatic interactions, and hydrogen bonding,⁷⁸ which cannot be separated on
208 conventional reversed-phase LC columns.⁷⁹ The incorporation of a zwitterionic HILIC stationary
209 phase is particularly applicable for retention of highly polar species such as charged metal-organic
210 complexes (i.e. Fe-citrate, Fe-malate, and Fe-siderophores).⁸⁰⁻⁸⁵ A multistep gradient elution was
211 performed at a flow rate of 0.2 mL min⁻¹ using a binary mobile phase-buffer system: (A) LC/MS-
212 grade water (OptimaTM, Sigma Aldrich Inc.) with 10 mM ammonium acetate buffer (v/v) and (B)
213 LC/MS-grade methanol (OptimaTM, Sigma Aldrich Inc.) with 10 mM ammonium acetate buffer
214 (Sigma Aldrich Inc.). The buffer pH for all LC experiments was held within 5.2 - 6.8 to preserve
215 the metal-ligand equilibria.⁸⁴ Elution proceeded as follows: 0–3 min held at 95% B, 3–7 min linear
216 gradient to 20% B, 7–15 min held at 20% B, 15-19 min linear gradient to 95% B, and re-
217 equilibration period at 19-57 min held at 95% B to prepare for the next injection in the sequence.
218 The autosampler compartment temperature where sample vials were stored was held at 15°C. The
219 column temperature was held at 35°C and an injection volume of 15 μL (~258 ng of injected
220 organic analyte) was used. No additional sample preparation was performed, and samples were

221 directly withdrawn and analyzed from the irradiated solutions. The UV–visible absorption spectra
222 of LC-separated components were measured using the PDA detector equipped with a deuterium
223 light-source and 1.0 cm fused-silica LightPipe™ flow cell, with a wavelength range of 200–680
224 nm (3D Field spectra), a scan rate of 20 Hz, and $\lambda \pm 4$ nm spectral resolution. The following
225 conditions were used for the HESI source operation in the HPLC experiments: 40 °C probe heater
226 temperature, 100 °C capillary temperature, 30 units of sheath gas flow, 10 units of auxiliary gas
227 flow, 0 units of sweep gas flow, 3.5 kV spray potential, and a funnel RF level of 30. A method
228 blank was used for all LC-PDA-HRMS experiments, corresponding to 90 μ M citric acid standard.
229 Analyzed samples were ionized in the negative ionization mode and HPLC-PDA-ESI(-)/HRMS
230 datasets were acquired using Xcalibur software (Thermo Inc.). The HRMS was operated at $m/\Delta m$
231 240,000 resolving power at 200 m/z , at a scan rate of 1.5 Hz in the full MS scan mode of 100 –
232 1300 m/z range. Custom mass calibration was performed using commercial calibration solutions
233 (Thermo Scientific, PI-88324) ionized in the negative ESI mode.

234
235 Additionally, two sets of HRMS experiments were performed by (1) directly infusing (DI-HRMS)
236 Fe^{III} -citrate solutions photolyzed at 0, 5, 40, 80, and 120 minutes (Exp. 1) in the negative mode to
237 screen for overall water soluble Fe^{III} -citrate components and their photochemical products, and
238 (Exp. 2) to specifically probe components of the unreacted and 120 min irradiated sample for
239 colloidal/carbonaceous products dissolved in 1 mL of a mixture of organic solvents (2:2:2:1 v/v%,
240 acetonitrile (ACN)/dichloromethane (DCM)/ hexanes/ toluene), the ‘*org-mix*’ hereafter, in the
241 positive ion mode. After this step, colloidal mixture underwent dissolution and intentional
242 disruption assisted by ultrasonication for ~ 40 min, followed by filtering using 0.45 μ m PTFE
243 syringe filter cartridges (Thermo Inc.), and solvent evaporation on the TurboVap (Biotage Inc.) at
244 1.5 L min^{-1} flowrate to ~ 1 mL level of remaining water solvent. 20% methanol was added before
245 DI-HRMS as recommended in the literature.^{73,74} DI-HRMS experiments were conducted using
246 similar MS tune conditions described above at 5 μ L min^{-1} injection flow rate for the water-soluble
247 fraction and slightly modified tune parameters for analysis of the colloid components dissolved in
248 the *org-mix*. For the analysis of colloid components dissolved in *org-mix*, the sheath, auxiliary gas
249 flow, funnel RF-level, capillary, and source heater temperature were adjusted to 30, 10 arbitrary
250 units, 80, 250 °C and 100 °C, respectively to allow for efficient droplet desolvation, ionization and
251 detection of larger molecular weight species. In Exp. 1 and Exp. 2 mass spectra were acquired in

252 negative and positive modes, respectively. Targeted MS² experiments were performed for
253 structural characterization of selected components of the Fe-citrate mixtures in the negative mode.
254 Additional details of DI-HRMS and targeted MS² experiments are described in Appendix C of the
255 SI file.

256

257 **Data Processing & Analysis.** Raw Xcalibur data files were background-subtracted prior to data
258 processing and analysis to remove any signal attributed to the method blank. The HPLC-
259 ESI/HRMS datasets were batch-processed with customized java script in the open-source
260 software, MZmine 2 (v. 2.51, <http://mzmine.github.io/>), to perform data preprocessing,
261 chromatogram construction, peak deconvolution, peak smoothing, peak alignment, peak
262 annotation, and analyte identification with the CAMERA R-package (Bioconductor R-
263 repository).^{86,87} The ADAP chromatogram builder⁸⁸ was used to reconstruct extracted ion
264 chromatograms (EIC) and obtain the output chromatographic feature list, using a minimum peak
265 height of 10³, LC peak duration of 0.6 min, and mass tolerance of 0.001 m/z. The combined
266 software algorithm identifies the analyte species, assuming the annotated ions in the range 100 –
267 1300 m/z were formed by the loss of a proton in the ESI(–) mode. The detected peaks identified
268 by MZmine 2 were cross evaluated using the Xcalibur software (Thermo Scientific Inc.), thus
269 eliminating erroneous formula assignments occasionally made by the algorithm. In addition, the
270 overall screenings of the Fe-containing peaks detected in the bulk samples were performed in DI-
271 ESI(–/+)/HRMS experiments. Obtained data sets were extracted with the DeconTools
272 AutoProcessor software developed at the Pacific Northwest National Laboratory (v. 1.0;
273 <http://omics.pnl.gov/software/>)⁸⁹ and processed using custom Excel macros (Microsoft Inc.)
274 developed for mass alignment, background subtraction, and ¹³C isotope filtering. Formula
275 assignments were assisted with grouping of the homologous species assigned based on first-order
276 (CH₂) and second-order (CH₂, H₂) Kendrick mass defects, followed by the group-representative
277 formula assignments⁹⁰ using the MIDAS molecular formula calculator (v 1.1;
278 <http://nationalmaglab.org/user-facilities/icr/icr-software>). Isotope distributions of Fe-containing
279 peaks were compared with results obtained from mass spectral isotope distribution simulator
280 software, IsoPro (v. 3.0; <https://sites.google.com/site/isoproms/>), as well as isotope peak scanner
281 function in MZmine 2 to search for m/z and intensity distribution of a calculated isotope pattern
282 within the feature list. The following constraints were applied for all formula assignments: C₁₋₆₀,

283 H₁₋₁₀₀, O₀₋₂₂, Fe₀₋₅, charge ≤ 2, and mass tolerance of ± 3 ppm. MS peaks with a signal to noise
284 (S/N) ratio ≥ 5 were only considered for this analysis. Double-bond equivalent (DBE) values of
285 the neutral assigned species were calculated using the following equation⁹¹:

286

$$287 \quad \text{DBE} = C - H/2 + N/2 + 1. \quad (1)$$

288

289 Throughout the manuscript, all molecular formulas discussed in the text below correspond to
290 neutral molecules in case of C_xH_yO_z composition and to ions in case Fe-containing organic
291 compounds, superscripted by their corresponding charges.

292

293 **Calculation of Mass Absorption Coefficient ($MAC_{\lambda,OM}$)**

294

295 **Optical Measurements of Bulk Solutions.** The absorption spectra presented in this work are in the
296 units of wavelength-dependent mass absorption coefficient ($MAC(\lambda)_{bulk}$ eq. 2),⁹² which is
297 calculated from log base-10 absorbance ($A_{10}^{solution}$) of the aqueous solution with organic mass
298 (OM) concentration (C_{mass} , g m⁻³), and optical path length ($b = 0.01$ m).

299

$$300 \quad MAC(\lambda)_{bulk} (\text{m}^2 \text{g}^{-1}) = \frac{A_{10}^{solution}(\lambda) \times \ln(10)}{b \times C_{mass}} \quad (2)$$

301

302 Whereas C_{mass} is calculated as mass concentration of the organic component (i.e., citrate, C₆H₅O₇³⁻
303). Therefore, C_{mass} in the case of 90 μM Fe-citrate (C₆H₅FeO₇) is ~ 17.02 g m⁻³. The calculation
304 assumes that total OM value does not change in the reacted Fe-citrate mixtures. This assumption,
305 however, underestimates $MAC(\lambda)_{bulk}$ values as decarboxylation reactions produce CO₂ and small
306 oxygenated VOC products, which can volatilize from the solution, therefore reducing OM
307 concentration in the reacted solutions. The open headspace in the experimental setup leads to
308 evaporation and degassing of VOC products from photoreacted solutions. Thus, the computed
309 $MAC(\lambda)_{bulk}$ values reported here should be viewed as the lower estimates. The wavelength
310 dependence of $MAC(\lambda)_{bulk}$ expressed through absorption Ångström exponent (AAE) was assumed
311 using the following formula:

312

313
$$MAC(\lambda)_{bulk} = k \times \lambda^{-AAE} \quad (3)$$

314
 315 *AAE* values were derived by computing the slope of the linear dependence of $MAC(\lambda)_{bulk}$ versus
 316 wavelength plotted on a ln-ln scale, while k is the imaginary part of the refractive index.⁹³ For the
 317 analytes of this study, we calculate *AAE* over 250-450 nm spectral range.

318
 319 **Photodiode Array (PDA) Optical Measurements of Solution Components.** The $MAC(\lambda)^{PDA}$ values
 320 derived from integrated UV-vis records obtained in LC-PDA measurements and corresponding to
 321 the total light absorption in each of the aliquot samples were computed using eq. 4.^{94,95}

322
 323
$$MAC(\lambda)_{\Sigma}^{PDA} (m^2 g^{-1}) = \frac{A(\lambda)_{\Sigma}^{PDA} (\mu AU) \times \Delta t \times F \times \ln(10)}{b (cm) \times m_{inj} \times 10} \quad (4)$$

324
 325 In this analytical expression, $A(\lambda)_{\Sigma}^{PDA}$ is the optical absorbance from PDA detector recorded at
 326 wavelength λ over the period of LC separation ($\Sigma = 0 - 12$ min), including the unretained fraction
 327 eluting at 0 – 3 min. Δt is the elution time range (12 min); F is the LC flow rate (0.2 mL min⁻¹); b
 328 is the optical pathlength of the PDA flow cell (1 cm), and m_{inj} is the injected mass (258 ng) of OM
 329 analyte. The coefficient 10 accounts for the combined conversion of μAU into AU, cm³ into m³,
 330 cm into m, and ng into g.^{94,95}

331
 332 The relative fractional contributions of $MAC(\lambda)_i^{PDA}$ attributable to each of the individual LC-
 333 separated features (i) in the analyzed samples were computed using equation 5.

334
 335
$$MAC(\lambda)_i^{PDA} = MAC(\lambda)_{\Sigma}^{PDA} \times \left(\frac{A(\lambda)_i^{PDA} \times \Delta t_i}{A(\lambda)_{\Sigma}^{PDA} \times \Delta t_{\Sigma}} \right) \quad (5)$$

336
 337 where $A(\lambda)_i^{PDA}$ (μAU) is the averaged UV-visible absorbance of the individual feature i and Δt_i
 338 (min) is its time duration. All UV-visible absorption data presented here were processed and
 339 exported using custom MATLAB (R2018a, MathWorks Inc., USA) scripts developed in our
 340 group. In our study, we report on the molecular composition of the Fe-Citrate reacting mixture and
 341 specify the individual $MAC(\lambda)_i^{PDA}$ contributions as a function of reaction time. From there, we

342 incorporate a ‘bottom up’ approach to interpret optical transformations of the bulk Fe-citrate
343 material in terms of a practical concept of simplified sequential reactions where first-order reaction
344 kinetics are quantified based on computed $MAC(\lambda)_{bulk}$ values.

345
346 ***In Situ* Flow Microscopy Detection of Colloids.** 300 μL of the unreacted, 5 min, 40 min, 80 min,
347 and 120 min irradiated solutions sampled from same cuvette were analyzed *in situ* using a
348 FlowCam[®] Nano instrument (Yokogawa Fluid Imaging Technologies Inc.) operated with the
349 VisualSpreadsheet 5.6.26 software (Fluid Imaging Technologies). The FlowCam[®] Nano
350 instrument uses oil immersion flow microscopy^{96,97} to focus and image particles as they pass
351 through a flow cell. Prior to imaging experiments, the instrument was autofocused with NIST
352 Traceable Polymer Microspheres standard (Lot No: 247697) with known diameter (D_p) of 0.702
353 ± 0.006 μm . The background intensity in all experiments were measured within the 173-178 A.U.
354 range. The instrument parameters are as follows: x40 objective lens, 0.1 μm distance to nearest
355 neighbor, threshold dark/light 18/18, 1 close hole (iterations), 60 μm flow cell depth, 500 μm flow
356 cell width, 0.3 mL sample volume, 0.025 mL min^{-1} sampling flow rate, auto image rate of 137
357 frames s^{-1} , sample run time 12.00 min, particle size range $0.3 - 60$ μm when operated in relative
358 count mode. Particle number concentrations were determined using count calibrated mode with
359 the instrument default settings of 0.1 μm distance to nearest neighbor, threshold dark/light 20/20,
360 1 close hole (iterations), 60 μm flow cell depth, 500 μm flow cell width, 0.3 mL sample volume,
361 0.025 mL min^{-1} sampling flow rate, auto image rate of 137 frames s^{-1} , sample run time 12.00 min,
362 and a particle size range $0.3 - 60$ μm . After each measurement, the microfluidic system was rinsed
363 with 0.20 μm filtered Windex[®] or 0.1% Liquinox surfactant soap in water (Optima[™], Sigma
364 Aldrich Inc.) followed with a rinse of pure water. Diameters reported from this measurement are
365 equivalent size diameter.

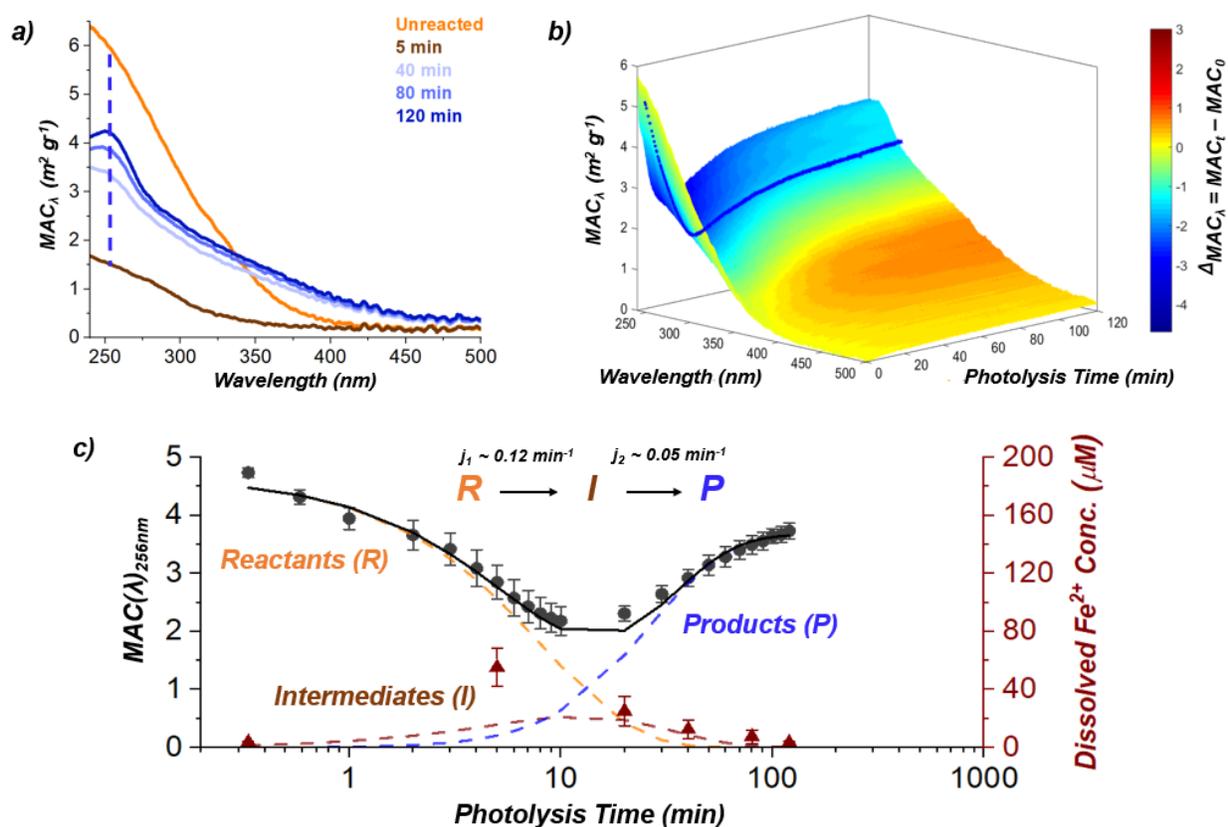
366

367 RESULTS AND DISCUSSION

368

369 **Photolysis of the Fe^{III}-Citrate aqueous system.** Figure 1a shows the $MAC(\lambda)_{bulk}$ absorption
370 spectra of the Fe^{III} citrate system measured at selected time intervals during the photolysis
371 experiment. Figure 1b further illustrates evolution of the same $MAC(\lambda)_{bulk}$ spectra but shown as a
372 continuing progression recorded at 5 s intervals using transient optical spectroscopy. The plots
373 show very rapid decrease of the MAC values over the first ~ 5 min of the irradiation, followed by

374 gradual increase over next ~100 min of the experiment. From the beginning of the reaction and
 375 during the initial decrease, $MAC(\lambda)_{bulk}$ spectra show a significant drop in their AAE values (from
 376 6.86 to 4.21) but remain featureless. In contrast, over the follow-up increase period, $MAC(\lambda)_{bulk}$
 377 spectra consistently show buildup of an absorption feature at 256 nm and AAE values of the spectra
 378 show only moderate change (from 4.21 to 3.62). Figure 1c shows a time-resolved record of
 379 MAC_{256nm} values over the entire irradiation time, indicating a reaction trend consistent with the
 380 formal kinetics of two first-order consecutive reactions of *Reactants* (R) $\xrightarrow{j_1}$ *Intermediates* (I) $\xrightarrow{j_2}$
 381 *Products* (P). Figure 1c shows kinetic modeling results using this apparent first-order kinetic
 382 scheme, featuring relative concentration profiles of R , I , and P components. Details of the kinetic
 383 model are described in Appendix D of the SI file. The pH of the initial ferric citrate solution was
 384 measured to be 6.7 and changes slightly to 6.52 in the 120 min photoreacted sample. Therefore all
 385 photolysis reactions were reproduced within the narrow pH range of this study.



386
 387 **Figure 1.** a) $MAC(\lambda)$ spectra of 90 μM Fe^{III}-citrate (1:1 M:L) solution undergoing photochemical
 388 transformations recorded at $t = 0, 5, 40, 80, 120$ min of irradiation time. Dashed line indicates absorption
 389 at 256 nm b) 3D $MAC(\lambda)$ plot illustrating the same transformations recorded at high time resolution (5s) in

390 a separate experiment. The x, y, and z-axis represent the absorbing wavelength (nm), MAC_λ , and photolysis
391 time (min), respectively. Colormap is scaled by the difference in MAC units ($\Delta MAC = MAC_t - MAC_0$). 2D
392 contour line indicating transient absorption at 256 nm. c) Single-wavelength $MAC_{256\text{nm}}$ experimental data
393 of the Fe^{III} -citrate photolytic time-resolved transformations (black circles, error bars represent standard
394 deviation of 3 replicated measurements). The reaction extent is modeled (dashed lines) using Apparent First
395 Order Kinetics for a formal mechanism of two consecutive irreversible reactions: *Reactants (R)* \rightarrow
396 *Intermediates (I)* \rightarrow *Products (P)*. Right Y axis of figure 1c shows the overall concentrations of Fe^{2+} ions
397 (brown triangles) measured in the solutions at 0, 5, 20, 40, 80, 120 min of irradiation time.

398 The $MAC(\lambda)_{\text{bulk}}$ spectrum of the original Fe^{III} -citrate solution (Fig 1a, orange trace) exhibits
399 characteristic tailing that extends to the visible region of the spectrum. These absorption
400 characteristics are very consistent with previous literature reports,^{21,22,61,98,99} and are similar to
401 optical characteristics exhibited by other Fe^{III} – carboxylate complexes.^{34,100–102} Spectral
402 characteristics of Fe^{II} -ligand species presented here and in the later text may be attributed to the
403 λ_{max} at ~ 256 nm. Previous study¹⁰³ report the UV-vis spectra of Fe^{II} - tannic acid complexes at
404 acidic pH absorbing closer to 260 nm, while Fe^{II} -cationic complexes in aprotic solutions exhibit a
405 unique and narrow λ_{max} at 262 nm.¹⁰⁴ The photochemical mechanism of the Fe^{III} -citrate system
406 (scheme I) explains the evolution of the $MAC(\lambda)_{\text{bulk}}$ spectra observed in our experiment.
407 Specifically, the initial photolysis of Fe^{III} -citrate ($R \rightarrow I$ reactions) proceeds through rapid ligand to
408 metal charge transfer (LMCT) excitation,¹⁰⁵ reducing Fe^{3+} to Fe^{2+} and dissociating and oxidizing
409 the citrate-radical pair.³¹ Reduced Fe^{2+} products exist as intermediate species; they are oxidized
410 back to Fe^{3+} ($I \rightarrow P$ reactions) at the longer photolysis times in the presence of ROS species and
411 dissolved oxygen. The formal kinetics and modeled curves describing relative fractions of R , I ,
412 and P components in the photoreacting system shown in Figure 1c indicate a net apparent effect
413 of reactions similar to R1-9 (Scheme 1). To validate our assumption that I products are indeed
414 associated with Fe^{II} intermediates, we measured dissolved $[Fe^{2+}]$ concentrations at selected
415 irradiation times. The Fe^{2+} concentration is low in the unreacted sample (3.8 μM), which is close
416 to the limit of detection (~ 2.3 μM). As the photolysis progresses to 5 min, the averaged
417 concentration of dissolved Fe^{2+} increases to 55.3 μM , then gradually decreases to 25-12 μM at 20-
418 40 min and to 7-3 μM at 80-120 min, respectively. Figure 1c shows that overall concentrations of
419 Fe^{2+} ions in the solutions (brown triangles) agree well with the modeled time-resolved profile of
420 the intermediate products I (brown dashed lines). The processes influencing $[Fe^{2+}]$ production are

421 related to (1) dissociation of citrate ligand from Fe^{III} center and reduction following LMCT as
422 previously mentioned.^{31,105} (2) The reactions and rate of reactions of available ROS (i.e. H₂O₂, O₂,
423 and HO₂) in the solution with dissolved Fe²⁺ and (3) speciation of Fe(II) in the aqueous solution at
424 the higher pH (~6.5 - 6.7) ranges where peroxides react with photoreduced Fe²⁺ and re-oxidize
425 back to Fe³⁺, significantly faster than at the lower pH conditions in separate studies.^{43,62,106}

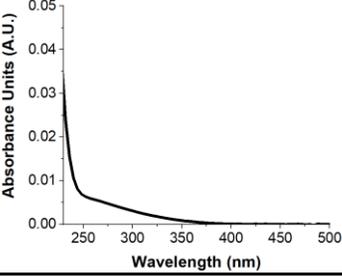
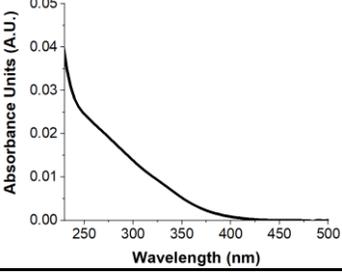
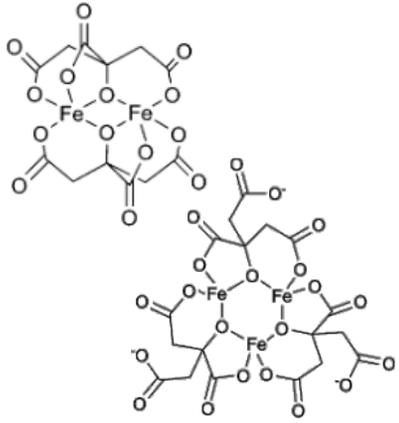
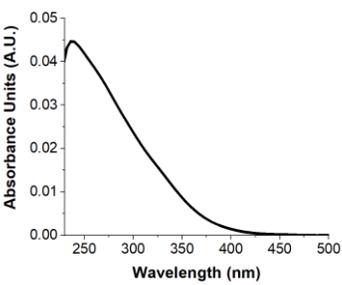
426 The apparent quantum yield, Φ of the generalized photochemical decomposition of R (Fe^{III}-citrate
427 reactant) was calculated based on the observed number of reacted Fe^{III}-citrate molecules divided
428 by the number of photons to which the experimental solutions were exposed. The latter was
429 calculated based on the photon flux measured in the actinometry experiments (Appendix B in SI
430 file). The calculated apparent Φ is ~ 0.02, which is only ~10% of the intrinsic Φ values reported
431 in the range of ~ 0.16 – 0.30 for Fe^{III}-carboxylates.^{25,61} Lower values of the apparent Φ are a
432 combined result of additional absorbance and scattering by other components in the reaction
433 mixture, which evolve over the reaction time. Therefore, apparent first-order reaction rates of $j_1=$
434 0.12 min^{-1} and $j_2= 0.05 \text{ min}^{-1}$ obtained from the formal kinetic modeling need to be considered as
435 specific to our experimental conditions. Details of the kinetic model are described in Appendix E
436 of the SI file. For the extrapolations of kinetic data to experiments external to our study, the
437 apparent rates and the apparent Φ need to be considered together, scalable to the relevant Φ values
438 determined in the corresponding experiments.

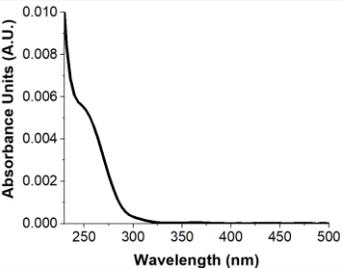
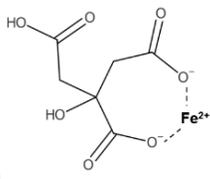
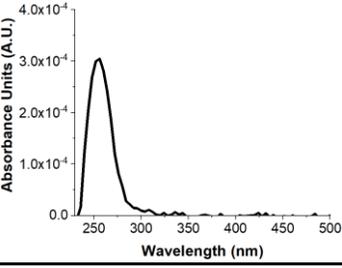
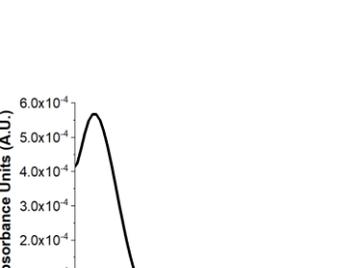
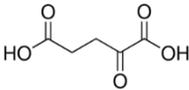
439 **Molecular Characterization of the Irradiated Samples.** Figure 2a illustrates HPLC-PDA
440 experimental results showing separated components of Fe^{III}-citrate solutions irradiated for 0, 5, 40,
441 80, and 120 min. For our analysis, the major separated absorbing components eluted between 0 –
442 3 min RT unretained period and 3 – 12 min RT for weakly to strongly retained species. Therefore,
443 we discuss those identified within these two LC ranges. The individual features are labeled
444 accordingly in each of the respective panels as reactants (R_i), intermediates (I_i), and products (P_i),
445 with their respective coloring schemes from Figure 1c describing the $R \rightarrow I \rightarrow P$ progression. In
446 total, 31 individual HPLC-PDA features (~ 95% of all features) are detected and identified in 5
447 irradiated samples based on the correlative assessment of HPLC-PDA and HPLC-ESI(-)/HRMS
448 records. Table 1 includes a list of separated features and proposed structures commonly grouped
449 with R_x , I_y , and P_z components, whereas the full list of separated features with corresponding
450 retention times, UV-vis spectra, and HRMS information are reported in Table S1 of the SI file.

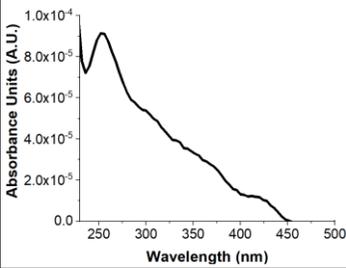
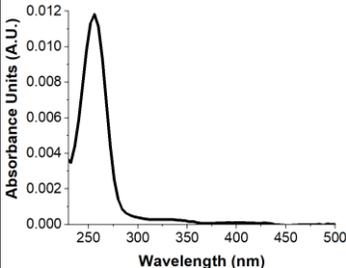
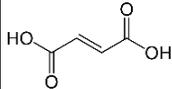
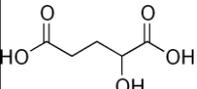
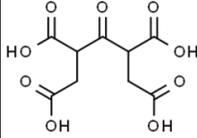
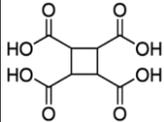
466 isotopes in the integrated MS spectra explicitly confirmed molecular assignments of the Fe-
467 containing ions. Extracted ion chromatograms of the Fe^{III}-citrate complexes overlaid with the LC-
468 PDA chromatographic peaks are illustrated in Figure S4 of the SI file. Details of DI-ESI(-)/HRMS
469 and ESI-MS² characterization of all Fe-containing peaks in the unreacted sample are described in
470 Figures S5, S6 and Table S2 of the SI file. After 5 min of photolysis, all *R* features decrease
471 substantially, indicating their lower concentrations. We estimate the summed concentration of
472 Fe^{III}-citrate complexes (*R*₂ and *R*₃) from the HPLC-PDA measurement based on the literature
473 reported molar extinction coefficient at $\lambda = 365$ nm ($\epsilon = 900$ L mol⁻¹ cm⁻¹)¹⁰⁷ to be ~ 78.4 and 1.4
474 μ M for 0 and 5 min samples, respectively. Considering that this concentration of the Fe^{III}-citrate
475 complexes in the unreacted mixture is lower than concentration of the dissolved Fe^{III}-citrate salt
476 ($\sim 90\mu$ M), the remaining 11.6 μ M likely corresponds to un-complexed/un-dissolved [Fe]³⁺,
477 [citrate]³⁻ ions, and surface-bound Fe^{III}-citrate complexes on the high surface area iron oxide (i.e.
478 FeOOH) colloids in the solution.¹⁰⁸ At the relevant Fe concentrations used in this study, formation
479 of insoluble iron oxide colloids precipitated in the aqueous solution at the higher pH setting
480 form.¹⁰⁹ For mechanistic interpretation of Fe(III)-citrate photochemical degradation discussed
481 here, special consideration of multi-phase photoreductive dissolution processes of colloidal
482 surface-bound Fe^{III}-citrate and other Fe-organic complexes in the aqueous solution should be
483 accounted for.^{108,110-112} The photodegradation of oxide bound Fe^{III}-citrate complexes lead to
484 dissolution of Fe^{III}-citrate, followed by Fe(III) reduction and immediate release of soluble Fe²⁺ and
485 citrate from colloid to the aqueous solution.¹⁰⁸ Overall, photodegradation of the soluble and surface
486 bound Fe^{III}-citrate complexes initiated through LMCT reactions³¹ R1-R4 (Scheme 1) is very fast,
487 resulting in rapid reduction of Fe^{III} to Fe^{II} over the first 5 min of photolysis. *R* components were
488 not detected in the samples collected at longer irradiation times.

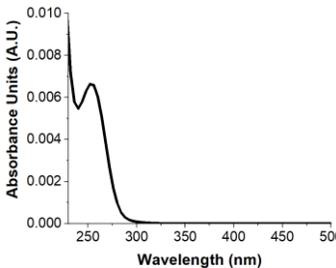
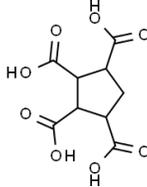
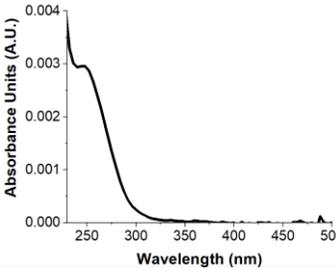
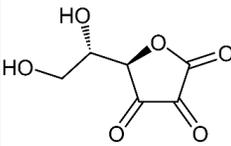
489

490 **Table 1:** Summary table of identified components in LC-PDA-HRMS datasets.

<i>ID</i>	<i>RT (min)</i>	<i>UV-Vis Spectrum</i>	<i>Exp. m/z</i>	<i>Chemical Formula</i>	<i>Mass Error (ppm)</i>	<i>Proposed Structure & Description</i>
			(-ESI)			
<i>Reactants (R_i)</i>						
<i>R₁</i>	2.3		159.972	$\text{Fe}_2^{\text{III}}\text{C}_{11}\text{H}_{12}\text{O}_4^{-2}$	1.875	
<i>R₂</i>	2.6		366.401 375.406 243.931 488.870 271.899	$\text{Fe}_3^{\text{III}}\text{C}_{18}\text{H}_{13}\text{O}_{21}^{-2}$ $\text{Fe}_3^{\text{III}}\text{C}_{18}\text{H}_{15}\text{O}_{22}^{-2}$ $\text{Fe}_2^{\text{III}}\text{C}_{12}\text{H}_8\text{O}_{14}^{-2}$ $\text{Fe}_2^{\text{III}}\text{C}_{12}\text{H}_9\text{O}_{14}^{-1}$ $\text{Fe}_3^{\text{III}}\text{C}_{12}\text{H}_{10}\text{O}_{14}^{-2}$	0.689 0.533 0.152 0.466 0.345	 <i>Iron(III)-citrate (Fe: Cit) complexes based on previous literature, Fe isotope distribution, and optical signals/UV-vis spectra.^{74,113}</i>
<i>R₃</i>	8.4		366.400 375.406 243.931 488.870 271.898	$\text{Fe}_3^{\text{III}}\text{C}_{18}\text{H}_{13}\text{O}_{21}^{-2}$ $\text{Fe}_3^{\text{III}}\text{C}_{18}\text{H}_{15}\text{O}_{22}^{-2}$ $\text{Fe}_2^{\text{III}}\text{C}_{12}\text{H}_8\text{O}_{14}^{-2}$ $\text{Fe}_2^{\text{III}}\text{C}_{12}\text{H}_9\text{O}_{14}^{-1}$ $\text{Fe}_3^{\text{III}}\text{C}_{12}\text{H}_{10}\text{O}_{14}^{-2}$	0.307 0.134 0.152 0.998 1.530	
<i>Intermediates (I_i)</i>						

I_1	2.4		244.939	$\text{Fe}^{\text{II}}\text{C}_6\text{H}_5\text{O}_7^{-1}$	1.48	 <p><i>Iron(II) Citrate based on spectral database match and isotope distribution.</i>¹¹⁴</p>
			I_2	3.1		211.028
197.012	$\text{Fe}^{\text{II}}\text{C}_4\text{H}_{13}\text{O}_5^{-1}$	3.2				
I_3	6.2		255.018	$\text{Fe}^{\text{II}}\text{C}_6\text{H}_{15}\text{O}_7^{-1}$	2.01	
			211.028	$\text{Fe}^{\text{II}}\text{C}_5\text{H}_{15}\text{O}_5^{-1}$	3.2	
			145.014	$\text{C}_5\text{H}_6\text{O}_5$	0.37	 <p><i>Alpha-ketoglutaric acid</i></p>
I_4	8.8		230.923	$\text{Fe}^{\text{II}}\text{C}_5\text{H}_3\text{O}_7^{-1}$	0.23	
			190.965	$\text{Fe}^{\text{II}}\text{C}_4\text{H}_7\text{O}_5^{-1}$	0.58	

			176.949	$\text{Fe}^{\text{II}}\text{C}_3\text{H}_5\text{O}_5^{-1}$	0.79	
<i>Products (P_i)</i>						
<i>P₁</i>	2.2		115.004	$\text{C}_4\text{H}_4\text{O}_4$	0.853	 <i>Fumaric acid</i>
			147.029	$\text{C}_5\text{H}_8\text{O}_5$	0.200	 <i>2-hydroxyglutaric acid</i>
			261.025	$\text{C}_9\text{H}_{10}\text{O}_9$	0.594	 <i>3-oxo-1,2,4,5-pentanetetracarboxylic acid</i>
<i>P₂</i>	2.3		231.015	$\text{C}_8\text{H}_8\text{O}_8$	0.431	 <i>cyclobutanetetracarboxylic acid</i>

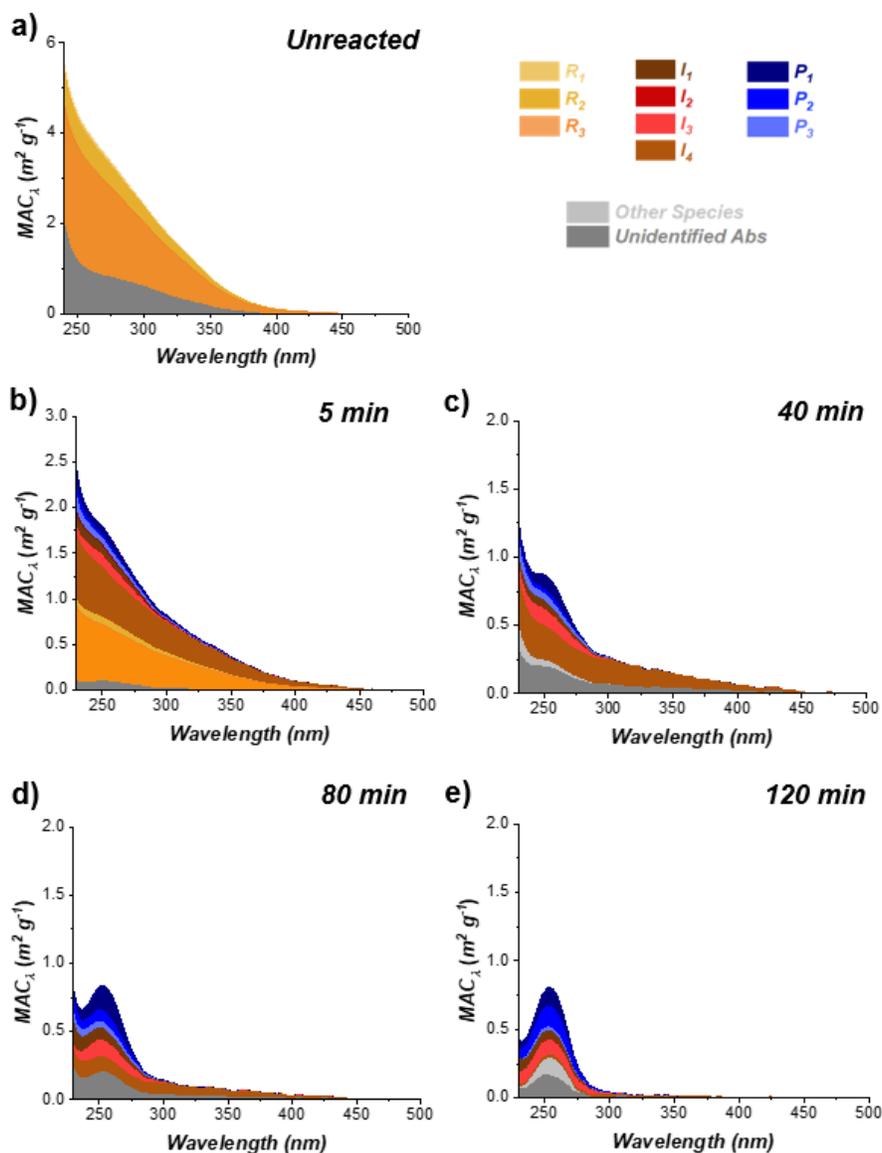
		 <p>Absorbance Units (A.U.) vs Wavelength (nm) plot for cyclopentanecarboxylic acid. The x-axis ranges from 250 to 500 nm, and the y-axis ranges from 0.000 to 0.010. A prominent peak is observed at 245.030 nm with an absorbance of approximately 0.0065.</p>	245.030	$C_9H_{10}O_8$	0.451	 <p><i>cyclopentanecarboxylic acid</i></p>
P_3	2.6	 <p>Absorbance Units (A.U.) vs Wavelength (nm) plot for dehydroascorbic acid. The x-axis ranges from 250 to 500 nm, and the y-axis ranges from 0.000 to 0.004. A peak is observed at 173.009 nm with an absorbance of approximately 0.003.</p>	173.009	$C_6H_6O_6$	0.122	 <p><i>Dehydroascorbic acid</i></p>

492 The identified HPLC-PDA features eluting at RTs of 2.4, 3.1, 6.2, and 8.8 min in the samples from
493 5 and 40 min irradiation times correspond to Fe^{II}-organic intermediates *I*₁, *I*₂, *I*₃, and *I*₄,
494 respectively. *I*₁ is plausibly identified as the reduced mono-charged Fe^{II}-citrate complex,
495 corresponding to ⁵⁶Fe^{II}C₆H₅O₇¹⁻, which is chemically different than the Fe^{III}-citrate complex(es)
496 observed as *R* species. Fe^{II}-citrate was detected in all of the photoreacted samples. Extracted ion
497 chromatograms for both ⁵⁶Fe^{II}Cit and ⁵⁶Fe₂^{III}Cit₂ complexes eluting at different retention times
498 along with corresponding UV-visible spectra recorded by PDA detector are described in Figure
499 S7 of the SI file. The complexation of Fe²⁺ to citric acid occurs at neutral and alkaline pH
500 conditions (within LC buffer ranges), as the hydroxyl groups in citrate molecule become ionized
501 in solution forming the monomeric Fe^{II}-citrate complex.¹¹⁵ Our results show that Fe^{II}-citrate and
502 Fe^{III}-citrate are uniquely identifiable and distinguishable in these LC experiments. However, those
503 species are considered unstable and relatively weak complexes in the aqueous solution¹¹⁶;
504 therefore it is difficult to quantify Fe^{II}-citrate (aq) in these experiments. Nevertheless, to the best
505 of our knowledge, separated and distinguished Fe^{III}- and Fe^{II}-citrate complexes are shown here for
506 the first time. Detection of these Fe^{II}-citrate intermediates implies that a fraction of these species
507 did not degrade immediately following LMCT conversion of both soluble and oxide bound¹⁰⁸ Fe^{III}-
508 citrate complexes to Fe²⁺ and citrate ions. *I*₂ corresponds to two coeluting monocharged Fe-
509 organic complexes of ⁵⁶Fe^{II}C₅H₁₅O₅¹⁻ and ⁵⁶Fe^{II}C₄H₁₅O₅¹⁻ while *I*₃ corresponds to a mixture of
510 two ions: ⁵⁶Fe^{II}C₆H₁₅O₇¹⁻ monocharged complex and α-ketoglutaric di-carboxylic acid (C₅H₆O₅),
511 which was reported in a previous study²⁵ as an abundant organic intermediate in the
512 photodecomposition of Fe^{III}-citrate. Finally, strongly absorbing *I*₄ feature corresponds to several
513 co-eluting mono-charged ions: ⁵⁶Fe^{II}C₃H₅O₅¹⁻, ⁵⁶Fe^{II}C₄H₇O₅¹⁻, and ⁵⁶Fe^{II}C₅H₃O₇¹⁻, which exhibit
514 maximum absorbance at ~ 250 – 256 nm and tailing in the 350 – 400 nm ranges (see Table 1 and
515 Table S1 of SI file). Consistently with the observed *R*→*I*→*P* formal kinetics, all *I* features
516 identified here degrade substantially at longer irradiation times and are therefore poorly seen in
517 the 80 and 120 min samples. The degradation of Fe(II) intermediate species here is likely due to
518 the oxidation of Fe²⁺ back to Fe³⁺ in the presence of citrate and available oxygen at the relevant
519 pH conditions similar to previous work.¹¹⁷ The rate of Fe(II)-citrate oxidation is enhanced in the
520 presence of high concentrations of citrate in the solution,¹¹⁸ therefore we should expect the rate of
521 Fe(II) oxidation to be proportional to the concentration of citrate ions in the solution.

522 Features P_1 , P_2 , and P_3 , eluting at 2.2, 2.3, and 2.6 min RT were detected in each of the irradiated
523 mixtures. However, they become most abundant in the 80 and 120 min samples. The plausible
524 identification of P_1 is a mixture of co-eluting organic compounds such as fumaric acid ($C_4H_4O_4$),
525 2-hydroxyglutaric acid ($C_5H_8O_5$), and 3-oxo-1,2,4,5-pentanetetracarboxylic acid ($C_9H_{10}O_9$), while
526 P_2 feature is associated with cyclobutanetetracarboxylic ($C_8H_8O_8$) and cyclopentane carboxylic
527 ($C_9H_{10}O_8$) acids. Detection of small organic acid byproducts such as fumaric (C_4) and hydroxy
528 glutaric acid (C_5) in the irradiated mixtures results from decarboxylation of citrate radical in the
529 presence of oxygen (R4 in Scheme 1). In addition, plausible Norrish I and II photodegradation
530 pathways (R10-11 in Scheme 1) are the source of small organic acids with C=C double bonds and
531 other radical recombinant species (R-R). Photolysis of citric acid by OH^\bullet mediated heterogenous
532 oxidation in a separate study¹¹⁹ results in extensive fragmentation of C-C bonds through acid-base,
533 free-radical chemical reactions, as well as decarboxylation and formation of carboxyl and
534 carbonyl-groups. Therefore, fumaric acid, 2-hydroxy glutaric acid and other small molecules are
535 expected byproducts of the decarboxylation and they are also common enzymatic metabolites in
536 the citric acid ‘Krebs’ cycle.¹²⁰ Alternatively, detection of highly oxygenated multi-carboxylic
537 acid and hydroxy dimers in this photolyzed system, such as $C_7H_8O_8$, $C_8H_{10}O_9$, and $C_{10}H_{12}O_{10}$, are
538 consistent with a previous study investigating the heterogenous photooxidation of dicarboxylic
539 acids catalyzed on the surface of TiO_2/Fe_2O_3 dust particles.¹²¹ Carboxylic dimer products can arise
540 from radical-radical recombination reactions (R9-10 in Scheme 1) of two carboxylic monomer
541 radicals in the irradiated solution. Lastly, HPLC feature P_3 eluting at 2.6 min RT corresponds
542 plausibly to dehydroascorbic acid ($C_6H_6O_6$), a cyclic molecule with multiple carbonyl (C=O)
543 groups and two hydroxyl groups. The characteristic absorption profile of this molecule is similar
544 to that reported in previous study¹²² investigating similar species and ascorbic acid via HPLC-
545 PDA. However, this molecule may also undergo consequent Norrish photochemistry or
546 reactions¹²³ with H_2O_2 to yield other oxidized, ring-opening products such as $C_6H_8O_8$ detected in
547 the irradiated samples. Full list of identified CHO-containing species is included in Table S3 of
548 the SI file. Overall, cyclic Fe redox reactions in the solution catalyzes the oxidation of citrate to
549 several of the intermediate and product species representing structurally similar species.

550

551 **Optical Properties of the Irradiated Samples** Figure 3 illustrates contributions of individual *R*,
552 *I* and *P* components to $MAC(\lambda)_{\Sigma}^{PDA}$ spectra recorded by HPLC-PDA for each of the five analyzed
553 samples. Overall, the systematic and quantitative changes in the absolute values of $MAC(\lambda)_{\Sigma}^{PDA}$
554 and the relative fractions of *R*, *I* and *P* contributions can be observed as the photolysis reactions
555 progress. Components *R* dominate absorbance in the unreacted sample. Their contribution is
556 halved in the sample irradiated for 5 min, where contributions by the *R* and *I* components become
557 nearly equal, and influence of the *P* components also becomes detectable. At the later irradiation
558 times (40 – 120 min), contributions from *P* continue to grow, while contributions from *I* decline
559 and contributions from *R* are undetectable. Notably, absolute values of $MAC(\lambda)_{\Sigma}^{PDA}$ decline
560 steadily as the photolysis reactions proceed and the spectral shape changes from featureless to
561 more defined with a well-distinguished peak at 256 nm. At 80 – 120 min irradiation time, the
562 $MAC(\lambda)_{\Sigma}^{PDA}$ values decrease by a factor of ~ 5 compared to those measured for the unreacted
563 sample. These observations are in striking contrast to the $MAC(\lambda)_{bulk}$ measurements using a UV-
564 vis spectrophotometer shown in Figures 1a and 1b. While both measurements show the same trend
565 of the *MAC* declining during the photolysis, absorption detected by the UV-vis spectroscopy of
566 bulk samples is much higher than that inferred from the HPLC-PDA measurements of the same
567 samples. The plausible reason for this discrepancy is a difference in the analyte material probed in
568 two measurements. Specifically, the $MAC(\lambda)_{bulk}$ spectra reflect absorption by total organic carbon
569 (TOC) which includes dissolved water-soluble organic carbon (WSOC) and colloids, while the
570 $MAC(\lambda)_{\Sigma}^{PDA}$ spectra correspond to the light absorption by WSOC only. Therefore $MAC(\lambda)_{\Sigma}^{PDA} \stackrel{\text{def}}{=}$
571 $MAC(\lambda)_{\Sigma}^{WSOC}$.

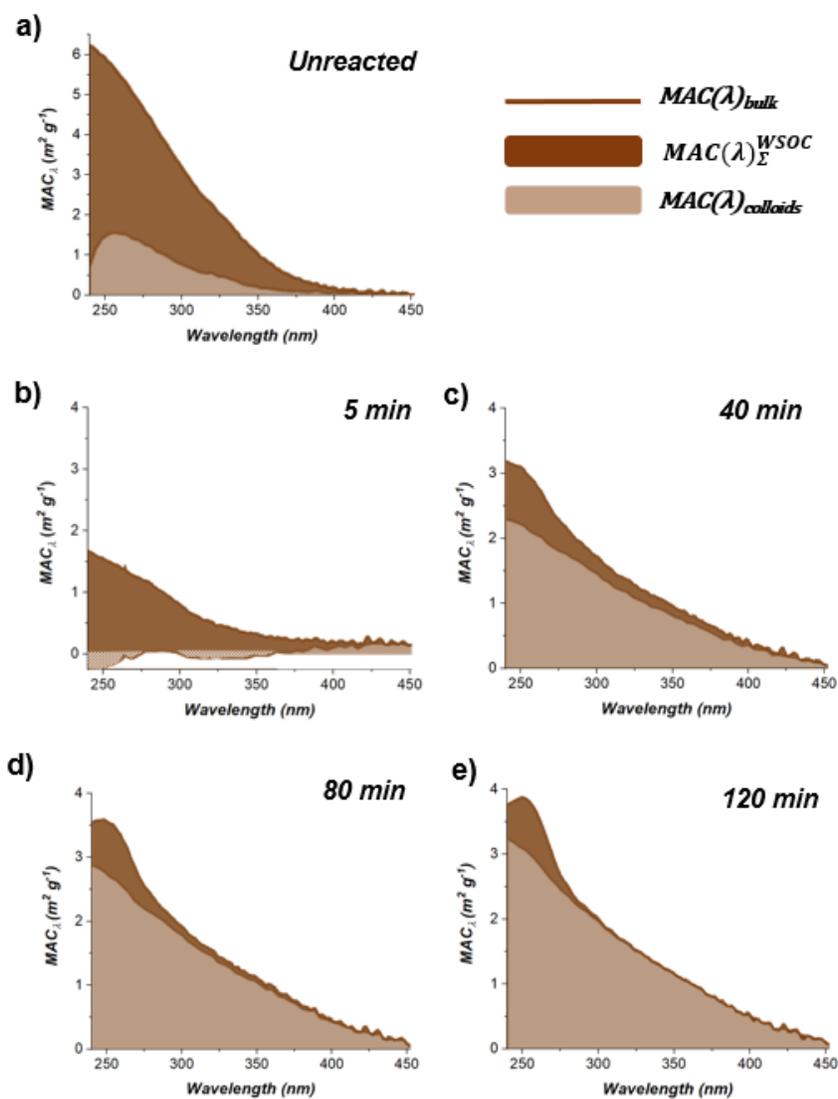


572

573 **Figure 3:** Fractions of $MAC(\lambda)_{\Sigma}^{PDA}$ attributed to water-soluble individual components of the Fe^{III} -citrate
 574 irradiated mixtures. Orange, brown, and blue colors indicate contributions from reactants (R), intermediates
 575 (I), and products (P) defined in Figure 2.

576 Figure 4 further illustrates the relationship between $MAC(\lambda)_{bulk}$ and $MAC(\lambda)_{\Sigma}^{WSOC}$ values
 577 determined for all 5 samples studied here. Individual panels illustrate relative fractions of
 578 $MAC(\lambda)_{\Sigma}^{WSOC}$ and $MAC(\lambda)_{colloids}$ out of $MAC(\lambda)_{bulk}$ spectra measured for the unreacted and
 579 irradiated samples. The $MAC(\lambda)_{\Sigma}^{WSOC}$ is the major fraction of $MAC(\lambda)_{bulk}$ for the unreacted and 5
 580 min irradiated mixtures, suggesting only minor effects of colloids (likely $Fe(OH)_3$) at the
 581 beginning of reaction. At later irradiation times (40-120 min), the buildup of colloidal products is

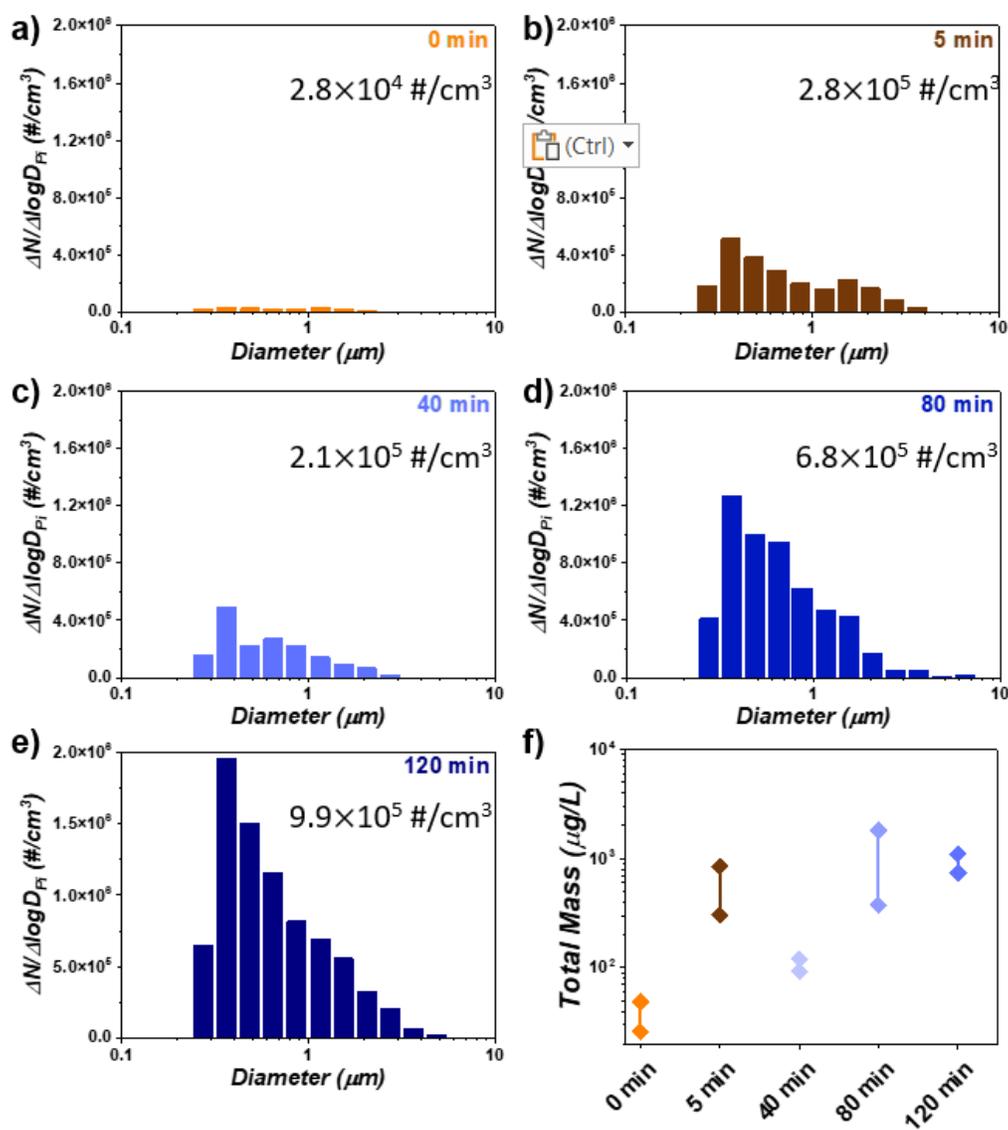
582 evident from increased values of $MAC(\lambda)_{colloids} = MAC(\lambda)_{bulk} - MAC(\lambda)_{\Sigma}^{WSOC}$ illustrated in the
 583 plots. The $MAC(\lambda)_{colloids}$ absorbing fraction in these mixtures is presumably related to formation of
 584 insoluble $Fe(OH)_3$ and poorly soluble reduced carbon products of reactions R12 (Scheme 1) that
 585 contribute to formation and growth of colloids. To test this assumption, we conducted a focused
 586 study to provide insights on the sizes, concentration, and chemical composition of these colloidal
 587 products.



588
 589 **Figure 4.** The relative fractions of $MAC(\lambda)_{\Sigma}^{WSOC}$ and $MAC(\lambda)_{colloids}$ out of $MAC(\lambda)_{bulk}$ spectra measured
 590 for the samples analyzed in this study.

591

592 ***In Situ* Flow Microscopy Detection of Colloids** Colloidal components of the unreacted and
593 irradiated samples at 5, 40, 80, and 120 min were measured using the FlowCam® Nano oil
594 immersion flow microscope that is capable to image and count individual colloidal particles larger
595 than 0.3 μm . Figures 5a-5e shows the 8-bin per decade number size distributions ($\Delta N/\Delta \log D_p$;
596 particles cm^{-3}) for the unreacted, 5 min, 40, 80 min, and 120 min photoreacted samples. In the
597 unreacted sample, the particle number concentration is low. The representative particle images
598 (included in Figure S8 of the SI file) reveal relatively smaller and round colloidal particles in the
599 unreacted system, which are likely commonly occurring $\text{Fe}(\text{OH})_3$ colloids and the presumed oxide
600 bound Fe^{III} -citrate particles. As the photolysis progresses from 5 – 120 min, the number
601 concentrations and size distributions of the particles detected grow in the reacted solutions, which
602 reveal significantly larger and more fractional colloids, as shown in Figure S8 of the SI file. The
603 estimated mass loadings shown in Figure 5f indicate progressive increase from $<50 \mu\text{g/L}$ in the
604 unreacted sample to $>1000 \mu\text{g/L}$ in the irradiated samples. Ranges of lower and upper mass limits
605 are shown to reduce bias of larger particles ($> 3 \mu\text{m}$) that are sparsely present. Details of the data
606 analysis are described in Appendix F of the SI file. With respect to the unreacted sample, the mass-
607 size distributions are shifted to larger particle sizes, suggesting that reduced carbon components
608 formed after prolonged photolysis contribute to the buildup of colloids in the irradiated samples
609 (see Figure S9 of the SI file). The mass loading of colloids in the 80 and 120 min irradiated sample
610 corresponds to $\sim 8\%$ and 5% of the original $\text{Fe}(\text{III})$ -Cit reactant mass. However, particles within
611 the $0.3 - 0.5 \mu\text{m}$ size ranges contribute most to the observed scattering of UV-visible light in the
612 near visible regions of the absorption spectrum shown in Figure 1a. Nevertheless, these results
613 show the formation of new colloids that effectively increase in concentration under prolonged
614 irradiation to UV-Vis light, representing an important pathway to nanoparticle products formation
615 in photochemical Fe^{III} -carboxylate systems.



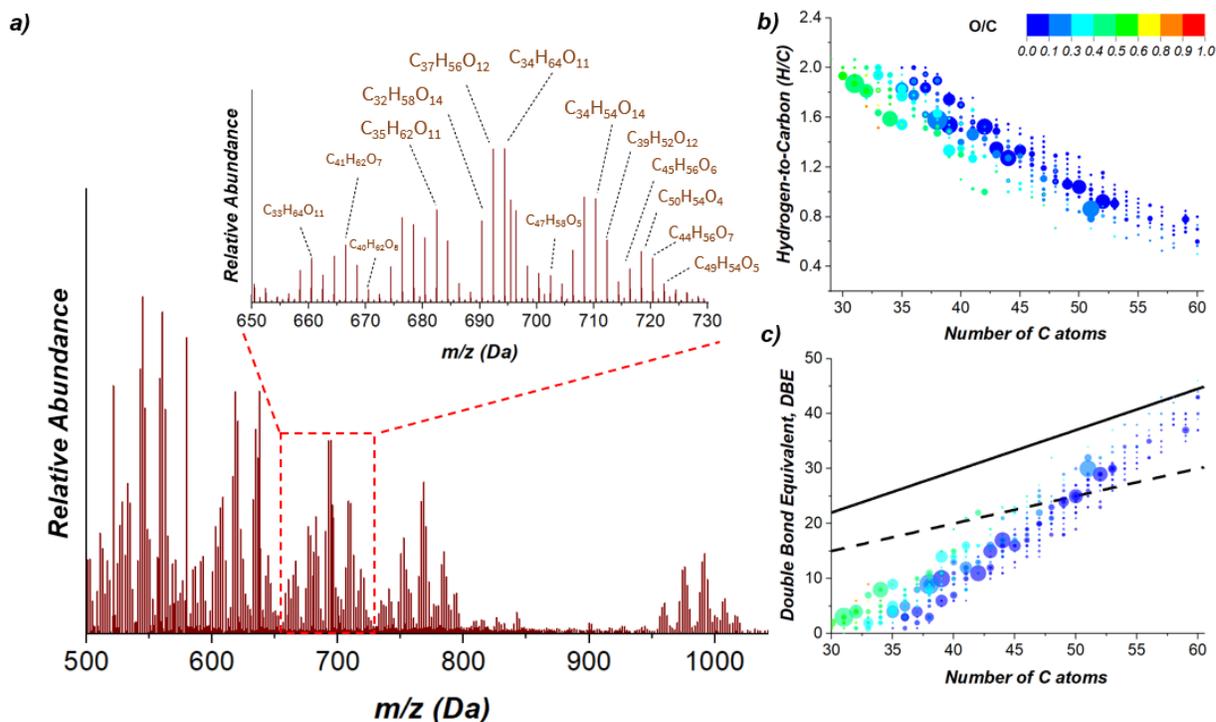
616

617 **Figure 5.** Particle number size distribution of colloids detected in the unreacted (orange) and
 618 photoreacted samples at 5 (brown), 40, 80, and 120 min (light blue - dark blue). Diameter reported
 619 by the flow microscope is the equivalent size diameter (ESD) in μm. Calculated mass loadings for
 620 each of the cases are reported as upper and lower limit values in panel f (see Fig S9 for additional
 621 details).

622

623 **Chemical Characterization of Colloidal Material** Colloidal material found in the 120 min
 624 irradiated sample was used for chemical characterization. Extracted colloidal material was
 625 dissolved in the 'org-mix' solvent and the obtained solutions were analyzed in DI-ESI(+)-HRMS

626 experiments for detection of high molecular weight components formed from radical
627 recombination (R12 of Scheme 1). Figure 6a illustrates results of the ESI(+)-HRMS analysis,
628 indicating MS features identified and assigned in the 500-1100 Da mass range. The representative
629 inset plot in Figure 6a illustrates selected CHO peaks and their formula assignments in the 650-
630 730 m/z range. Figure 6b and 6c show the hydrogen-to-carbon (H/C) elemental ratios and DBE of
631 species found in colloidal components, plotted as a function of number of carbon atoms (C) and
632 color-coded by their oxygen-to-carbon (O/C) ratio (Figure 6b). Reference lines in Figure 6c
633 correspond to cata-PAH's ($DBE = 0.75 \times C - 0.5$)¹²⁴ and linear conjugated polyenes C_xH_x+2 (DBE
634 $= 0.5 \times C$).¹²⁵ The molecular species representative of dissolved colloids in this HRMS dataset
635 span 30 – 60 carbon atoms with varying degree of oxygenation and unsaturation due to double
636 bonds in the organic structures. The adsorbed components of the colloidal material is composed
637 of moderately saturated aliphatic and unsaturated hydrocarbon species with variable DBE and low
638 oxygen content. Majority of the assigned features with higher HRMS signal intensity fall below
639 the linear conjugated polyene boundary line, while some of the lower intensity features fall closer
640 to the condensed aromatic reference line. Few selected species such as: $C_{34}H_{64}O_{11}$ ($DBE = 3$; O/C
641 $= 0.32$), $C_{37}H_{56}O_{12}$ ($DBE = 10$; $O/C = 0.32$), $C_{40}H_{62}O_8$ ($DBE = 10$; $O/C = 0.2$), $C_{47}H_{58}O_5$ ($DBE =$
642 19 ; $O/C = 0.11$), and $C_{49}H_{54}O_5$ ($DBE = 23$; $O/C = 0.1$) represent less oxygenated/ more reduced
643 carbon fragments that were adsorbed on colloid surfaces.



644
 645 **Figure 6.** a) (+)ESI-HRMS spectrum representative of the ‘*org-mix*’ digested colloidal components
 646 identified in the Fe^{III}-citrate mixture irradiated for 120 min. Inset plot shows zoomed in HRMS spectrum
 647 in the range 650-730 Da, b) H/C ratio vs number of carbon atoms (C) plot and c) DBE vs C plot scaled to
 648 O/C ratio. Reference lines represent DBE values of linear polyenes (dotted line; $0.5 \times C$)¹²⁵ and condensed
 649 aromatics (dashed line; $0.75 \times C - 0.5$).¹²⁴ The size of all individual symbols is scaled to the square root of the
 650 MS peak intensity.

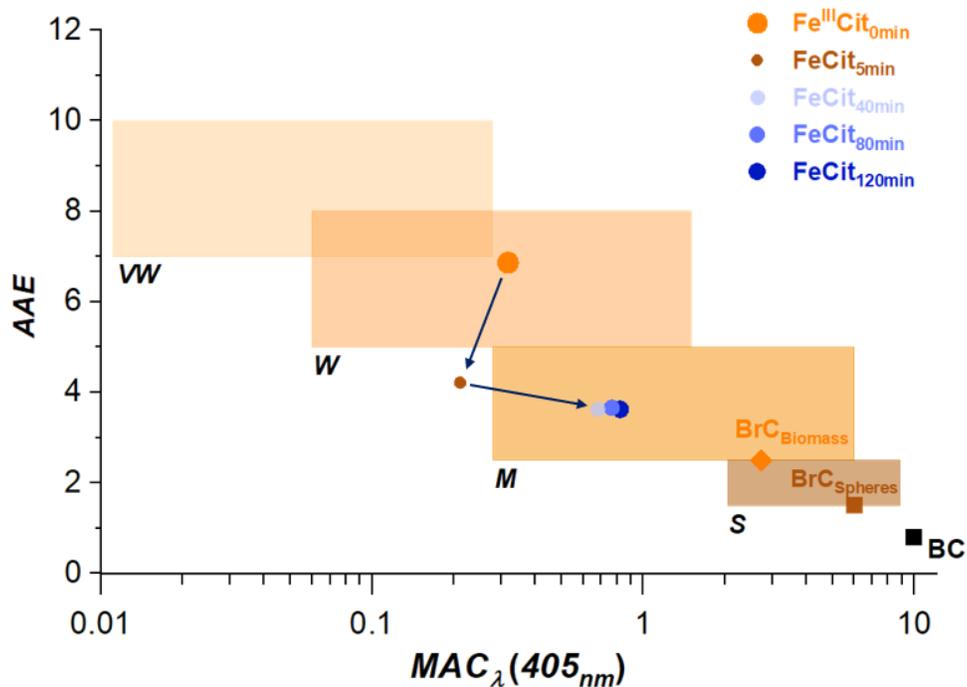
651
 652 Overall, we observe unique low and high molecular weight features in the mixture comprising
 653 dissolved colloid fragments composed of *C*, *H* and *O* immediately following dissolution in ‘*org-*
 654 *mix*’ followed by ultrasonication. Of note, those features were only observed in the HRMS spectra
 655 of sonicated samples, confirming their apportionment to the degraded colloids (see Figure S9 of
 656 SI file for more details). Furthermore, those features were also absent in MS spectra acquired from
 657 the unreacted sample prepared following the same dissolution procedures (see Figure S10 of the
 658 SI file). Therefore, adsorbed colloidal analytes probed at the longer photolysis time are composed
 659 of less oxygenated (more reduced) carbon material, resulting from the radical recombination
 660 reactions in R12 of Scheme 1.

661

662 CONCLUSIONS

663 This study investigated multi-phase photochemistry of Fe-citrate and reported on the formation of
664 unexplored colloidal products under prolonged light exposure. Apparent reaction rates of the $R \xrightarrow{j_1}$
665 $I \xrightarrow{j_2} P$ formal kinetics derived from our experiments are $j_1 \sim 0.12 \text{ min}^{-1}$ and $j_2 \sim 0.05 \text{ min}^{-1}$, which
666 correspond to the laboratory half-lifetimes of $\tau_{lab} \sim 8.3$ and 20 min, respectively (Appendix D).
667 Scaling these half -lifetimes with respect to intensity of the solar radiation at summer solstice in
668 Southern California and Mediterranean region suggests environmental half-lifetimes of ~ 32 min
669 and ~ 1.3 hr, respectively. (Appendixes G and H). Therefore, we conclude that photochemical
670 formation of colloidal products with carbon in a relatively reduced oxidation state may take place
671 at the time scale relevant to the real-world environment of the Fe-carboxylate systems.

672 The light-absorbing properties of the photochemically reacting Fe-citrate mixtures are very much
673 dynamic and overall comparable to those reported for atmospheric brown carbon (BrC). Figure 7
674 illustrates the $MAC(\lambda)_{bulk}$ values for the Fe-citrate samples of our study shown with respect to the
675 optical-based AAE versus MAC_{405nm} classification framework.^{70,95,126,127} The light-absorbing
676 properties of the unreacted and photoreacted Fe-citrate mixtures are skewed between weak (W)
677 and moderate (M) light-absorbing classes, and they are less-absorbing than BrC from biomass
678 burning^{93,128} Over the course photochemical aging, BrC characteristics of the Fe-citrate mixture
679 follows the $W \rightarrow M$ transition, resulting in stronger absorption by the photoreacted mixtures. This
680 darkening is largely attributed to the carbonaceous colloidal material formed after prolonged
681 photolysis under oxygen-deprived conditions. To date, no study has observed these systematic
682 changes of Fe^{III} -carboxylates leading to production of insoluble carbonaceous colloids in aqueous
683 photochemical reactions. To that extent, we therefore presume that formation of condensed
684 reduced carbon species in this system is most likely similar to other extensively photolyzed brown
685 carbon proxies,¹²⁹ and other Fe^{III} -carboxylate systems.^{20,24,34}



686

687 **Figure 7.** Optical properties of Fe-citrate photolysis samples mapped in the $AAE-MAC_{405\text{ nm}}$ space
 688 introduced by Saleh.¹²⁶ The shaded regions represent “optical bins” for “very weak (VW)”, “weak (W)”,
 689 “moderate (M)”, and “strong (S)” absorbing BrC classes. Literature-reported MAC values at 405 nm for
 690 BrC related to biomass burning,⁹³ spherical tar-balls,¹²⁸ and black carbon (BC)¹³⁰ are shown.

691 Overall, the results of this work demonstrate a multi-phase mechanism of the Fe^{III} -citrate
 692 photochemistry in aquatic systems of environmental relevance. The photochemical products
 693 identified in this representative Fe^{III} -carboxylate system are water-soluble components (R , I , and
 694 P components) and carbonaceous colloids formed under oxygen-deprived conditions in the
 695 extensively photolyzed mixtures. The detection of colloidal products in this system may contribute
 696 to yet unrecognized atmospheric and terrestrial light-absorbing material in aquatic environments
 697 with direct effects on the radiative forcing of climate. Further investigation of additional
 698 photoreactive Fe^{III} -carboxylate systems need to consider and account for formation of colloidal
 699 products that remain poorly characterized, overlooked in previous studies, and received little
 700 attention in the literature.^{62,131} In particular, ferric oxalate in natural aquatic environments may also
 701 behave similarly in the context of our work. Aqueous photolysis of ferric oxalate complex as
 702 shown in Figure S11 of the SI file follows similar trend, where degradation of $Fe(III)$ -oxalate
 703 complex occurs, followed by formation of short-lived $Fe(II)$ -intermediate species, and build-up of

704 water-soluble and insoluble colloidal reaction products where the light extinction extends beyond
705 500nm. Additionally, the aqueous photolysis of Fe-carboxylate complexes in natural waters are
706 important for Fe^{III/II} catalytic cycling, controlling the bioavailability of Fe(II) to microorganisms,¹³²
707 photo-mineralization of DOM, and oxidation of organic carbon in irradiated environmental waters
708 and in atmospheric aerosol particles.^{62,133–135}

709 Photoreactive Fe-carboxylates may also play an important role in initiating photocatalytic
710 reactions in snowpack. In particular, snow covered areas of inland central Asia^{127,136–138} located
711 near Fe^{III} dust sources may serve as an important medium for photocatalytic cycling of Fe^{III}-
712 carboxylates and the subsequent photooxidation of other water-soluble species (i.e. brown carbon,
713 carboxylic acids, ketones), halogenated species, and other inorganics in the snowpack. The
714 interactions of externally mixed Fe-mineral dust¹³⁷ and subsequent photochemical transformations
715 due to Fe^{III}-carboxylates can significantly enhance light absorption and modify the resulting snow
716 and surface albedo, with implications to indirect radiative forcing changes on the snow/surface
717 energy budgets. Specifically, carbonaceous colloid products resulted from Fe-carboxylate
718 photochemistry may accelerate melting of the snowpack, similar to BC and BrC deposits,¹³⁶
719 leading to reduction of the snow albedo.

720

721 SUPPORTING INFORMATION

722 Appendix A and Figure S1 describe transient optical and photochemical experimental setup of this
723 study; Appendix B and Figure S2 describes chemical actinometry experiments and relevant
724 calculations; Appendix C describes detailed direct infusion-HRMS parameters; Appendix D, E, F,
725 G, and H describes the formal kinetic modelling, quantum yields, particle size distributions,
726 atmospheric scaling, and TUV model calculations and selected parameters, respectively. Figure
727 S3 illustrates experimental quantitation of dissolved Fe²⁺ and calibration curve of Fe^{II}-
728 phenanthroline complexes; Figure S4 illustrate selected ion chromatograms (SIC`s) of Fe^{III}-citrate
729 complex ions; Figure S5 and S6 describe DI-HRMS screening and MS² experiments of Fe^{III}-citrate
730 complexes for identification and structural characterization. Figure S7 illustrates SIC`s for Fe^{III/II}-
731 citrate complexes detected in HPLC-HRMS. Figure S8 show *in situ* FlowCam Nano images of
732 colloidal particles. Figure S9 shows mass size distributions of colloids detected by FlowCam Nano.

733 Figure S10, S11 illustrate DI-HRMS results of dissolved colloidal products under specified
734 conditions. Figure S12 UV-visible spectrum of ferric oxalate photolysis. List of all molecular
735 assignments in HPLC-PDA-HRMS and DI-HRMS are reported in Table S1, S2, and S3.

736 **Data Availability** The dataset used for this work is available for download as a .zip file from
737 <https://doi.org/10.4231/xxx-yyyy> (West et al., 2022).

738

739 AUTHOR INFORMATION

740 Corresponding Author

741 **Alexander Laskin** – Department of Chemistry, Purdue University, Indiana 47906, United States;
742 orcid.org/0000-0002-7836-8417; Email: alaskin@purdue.edu

743 Authors

744 **Christopher P. West** – Department of Chemistry, Purdue University, Indiana 47906, United
745 States; orcid.org/0000-0001-9337-2820

746 **Jackson T. Ryan** – Department of Chemistry, Purdue University, Indiana 47906, United States

747 **Ana C. Morales** – Department of Chemistry, Purdue University, Indiana 47906, United States;
748 orcid.org/0000-0001-6969-2883

749 **Maria V. Misovich** – Department of Chemistry, Purdue University, Indiana 47906, United States;
750 orcid.org/0000-0001-6748-049X

751 **Anusha P. S. Hettiyadura** – Department of Chemistry, Purdue University, Indiana 47906, United
752 States; orcid.org/0000-0002-5757-9784

753 **Felipe Rivera-Adorno** – Department of Chemistry, Purdue University, Indiana 47906, United
754 States; orcid.org/0000-0002-7355-7999

755 **Jay M. Tomlin** – Department of Chemistry, Purdue University, Indiana 47906, United States;
756 orcid.org/0000-0002-3081-1512

757 **Andrew Darmody** – Department of Chemistry, Purdue University, Indiana 47906, United States;
758 orcid.org/0000-0002-7749-397X

759 **Brittany N. Linn** – Department of Chemistry, Purdue University, Indiana 47906, United States

760 **Peng Lin** – Department of Chemistry, Purdue University, Indiana 47906, United States

761

762 **Author Contributions.** C.P.W and A.L designed the overall project framework and experiments.
763 C.P.W, J.R, M.M, and B.N.L performed the photolysis and chemical actinometry experiments.
764 C.P.W, M.M, A.P.S.H, and P.L performed the LC separations and method development protocols.
765 C.P.W performed the molecular characterization experiments and processed, analyzed, and
766 interpreted the hyphenated LC-PDA-HRMS and direct infusion HRMS data with the assistance of
767 J.R, A.D., and B.N.L. F.R and J.M.T performed microscopy imaging. C.P.W and A.C.M
768 performed FlowCam measurements and data analysis. A.C.M designed custom MATLAB scripts
769 used for processing and analysis of all UV-visible spectroscopy data in this work. C.P.W and A.L
770 wrote the manuscript with contributions with all co-authors.

771

772 **Competing Interests.** The authors declare that they have no conflict of interest.

773

774 **Acknowledgments.** We acknowledge support from the startup funds allocated to A. Laskin by the
775 Department of Chemistry at Purdue University and partial support from the National Science
776 Foundation award AGS-2039985 at the time of this manuscript preparation.

777

778

779 **REFERENCES**

- 780 (1) Huang, J.; Jones, A.; Waite, T. D.; Chen, Y.; Huang, X.; Rosso, K. M.; Kappler, A.; Mansor, M.;
781 Tratnyek, P. G.; Zhang, H. Fe(II) Redox Chemistry in the Environment. *Chem. Rev.* **2021**.
782 <https://doi.org/10.1021/acs.chemrev.0c01286>.
- 783 (2) Taylor, K. G.; Konhauser, K. O. Iron in Earth Surface Systems: A Major Player in Chemical and
784 Biological Processes. *Elements* **2011**, *7* (2), 83–88. <https://doi.org/10.2113/gselements.7.2.83>.
- 785 (3) Taylor, S. R. Abundance of Chemical Elements in the Continental Crust: A New Table. *Geochim.*
786 *Cosmochim. Acta* **1964**, *28* (8), 1273–1285. [https://doi.org/10.1016/0016-7037\(64\)90129-2](https://doi.org/10.1016/0016-7037(64)90129-2).
- 787 (4) Hutchins, D. A.; Boyd, P. W. Marine Phytoplankton and the Changing Ocean Iron Cycle. *Nat. Clim.*
788 *Change* **2016**, *6* (12), 1072–1079. <https://doi.org/10.1038/nclimate3147>.
- 789 (5) Lis, H.; Shaked, Y.; Kranzler, C.; Keren, N.; Morel, F. M. M. Iron Bioavailability to Phytoplankton: An
790 Empirical Approach. *ISME J.* **2015**, *9* (4), 1003–1013. <https://doi.org/10.1038/ismej.2014.199>.
- 791 (6) Boyd, P. W.; Jickells, T.; Law, C. S.; Blain, S.; Boyle, E. A.; Buesseler, K. O.; Coale, K. H.; Cullen, J. J.;
792 Baar, H. J. W. de; Follows, M.; Harvey, M.; Lancelot, C.; Levasseur, M.; Owens, N. P. J.; Pollard, R.;
793 Rivkin, R. B.; Sarmiento, J.; Schoemann, V.; Smetacek, V.; Takeda, S.; Tsuda, A.; Turner, S.; Watson,
794 A. J. Mesoscale Iron Enrichment Experiments 1993–2005: Synthesis and Future Directions. *Science*
795 **2007**, *315* (5812), 612–617. <https://doi.org/10.1126/science.1131669>.
- 796 (7) Jickells, T. D.; An, Z. S.; Andersen, K. K.; Baker, A. R.; Bergametti, G.; Brooks, N.; Cao, J. J.; Boyd, P.
797 W.; Duce, R. A.; Hunter, K. A.; Kawahata, H.; Kubilay, N.; laRoche, J.; Liss, P. S.; Mahowald, N.;
798 Prospero, J. M.; Ridgwell, A. J.; Tegen, I.; Torres, R. Global Iron Connections Between Desert Dust,
799 Ocean Biogeochemistry, and Climate. *Science* **2005**, *308* (5718), 67–71.
800 <https://doi.org/10.1126/science.1105959>.
- 801 (8) Sedwick, P. N.; Sholkovitz, E. R.; Church, T. M. Impact of Anthropogenic Combustion Emissions on
802 the Fractional Solubility of Aerosol Iron: Evidence from the Sargasso Sea. *Geochem. Geophys.*
803 *Geosystems* **2007**, *8* (10). <https://doi.org/10.1029/2007GC001586>.
- 804 (9) Gonet, T.; Maher, B. A. Airborne, Vehicle-Derived Fe-Bearing Nanoparticles in the Urban
805 Environment: A Review. *Environ. Sci. Technol.* **2019**, *53* (17), 9970–9991.
806 <https://doi.org/10.1021/acs.est.9b01505>.
- 807 (10) Deguillaume, L.; Leriche, M.; Desboeufs, K.; Mailhot, G.; George, C.; Chaumerliac, N. Transition
808 Metals in Atmospheric Liquid Phases: Sources, Reactivity, and Sensitive Parameters. *Chem. Rev.*
809 **2005**, *105* (9), 3388–3431. <https://doi.org/10.1021/cr040649c>.
- 810 (11) Tomlin, J. M.; Jankowski, K. A.; Rivera-Adorno, F. A.; Fraund, M.; China, S.; Stirm, B. H.; Kaeser, R.;
811 Eakins, G. S.; Moffet, R. C.; Shepson, P. B.; Laskin, A. Chemical Imaging of Fine Mode Atmospheric
812 Particles Collected from a Research Aircraft over Agricultural Fields. *ACS Earth Space Chem.* **2020**, *4*
813 (11), 2171–2184. <https://doi.org/10.1021/acsearthspacechem.0c00172>.
- 814 (12) Cwiertny, D. M.; Baltrusaitis, J.; Hunter, G. J.; Laskin, A.; Scherer, M. M.; Grassian, V. H.
815 Characterization and Acid-Mobilization Study of Iron-Containing Mineral Dust Source Materials. *J.*
816 *Geophys. Res. Atmospheres* **2008**, *113* (D5). <https://doi.org/10.1029/2007JD009332>.
- 817 (13) Al-Abadleh, H. A. Aging of Atmospheric Aerosols and the Role of Iron in Catalyzing Brown Carbon
818 Formation. *Environ. Sci. Atmospheres* **2021**, *1* (6), 297–345. <https://doi.org/10.1039/D1EA00038A>.
- 819 (14) Tagliabue, A.; Bowie, A. R.; Boyd, P. W.; Buck, K. N.; Johnson, K. S.; Saito, M. A. The Integral Role of
820 Iron in Ocean Biogeochemistry. *Nature* **2017**, *543* (7643), 51–59.
821 <https://doi.org/10.1038/nature21058>.
- 822 (15) Kappler, A.; Bryce, C.; Mansor, M.; Lueder, U.; Byrne, J. M.; Swanner, E. D. An Evolving View on
823 Biogeochemical Cycling of Iron. *Nat. Rev. Microbiol.* **2021**, *19* (6), 360–374.
824 <https://doi.org/10.1038/s41579-020-00502-7>.

- 825 (16) Lovley, D. R.; Holmes, D. E.; Nevin, K. P. Dissimilatory Fe(III) and Mn(IV) Reduction. In *Advances in*
826 *Microbial Physiology*; Academic Press, 2004; Vol. 49, pp 219–286. <https://doi.org/10.1016/S0065->
827 [2911\(04\)49005-5](https://doi.org/10.1016/S0065-2911(04)49005-5).
- 828 (17) Bosch, J.; Heister, K.; Hofmann, T.; Meckenstock, R. U. Nanosized Iron Oxide Colloids Strongly
829 Enhance Microbial Iron Reduction. *Appl. Environ. Microbiol.* **2010**, *76* (1), 184–189.
830 <https://doi.org/10.1128/AEM.00417-09>.
- 831 (18) Roden, E. E. Geochemical and Microbiological Controls on Dissimilatory Iron Reduction. *Comptes*
832 *Rendus Geosci.* **2006**, *338* (6), 456–467. <https://doi.org/10.1016/j.crte.2006.04.009>.
- 833 (19) Lueder, U.; Barker Jørgensen, B.; Kappler, A.; Schmidt, C. Photochemistry of Iron in Aquatic
834 Environments. *Environ. Sci. Process. Impacts* **2020**, *22* (1), 12–24.
835 <https://doi.org/10.1039/C9EM00415G>.
- 836 (20) Pozdnyakov, I. P.; Kolomeets, A. V.; Plyusnin, V. F.; Melnikov, A. A.; Kompanets, V. O.; Chekalin, S.
837 V.; Tkachenko, N.; Lemmetyinen, H. Photophysics of Fe(III)–Tartrate and Fe(III)–Citrate Complexes
838 in Aqueous Solutions. *Chem. Phys. Lett.* **2012**, *530*, 45–48.
839 <https://doi.org/10.1016/j.cplett.2012.01.051>.
- 840 (21) Vukosav, P.; Mlakar, M.; Tomišić, V. Revision of Iron(III)–Citrate Speciation in Aqueous Solution.
841 Voltammetric and Spectrophotometric Studies. *Anal. Chim. Acta* **2012**, *745*, 85–91.
842 <https://doi.org/10.1016/j.aca.2012.07.036>.
- 843 (22) Weller, C.; Horn, S.; Herrmann, H. Photolysis of Fe(III) Carboxylate Complexes: Fe(II) Quantum
844 Yields and Reaction Mechanisms. *J. Photochem. Photobiol. Chem.* **2013**, *268*, 24–36.
845 <https://doi.org/10.1016/j.jphotochem.2013.06.022>.
- 846 (23) Pozdnyakov, I. P.; Melnikov, A. A.; Tkachenko, N.; Chekalin, S. V.; Lemmetyinen, H.; Plyusnin, V. F.
847 Ultrafast Photophysical Processes for Fe(III)–Carboxylates. *Dalton Trans.* **2014**, *43* (47), 17590–
848 17595. <https://doi.org/10.1039/C4DT01419G>.
- 849 (24) Glebov, E. M.; Pozdnyakov, I. P.; Grivin, V. P.; Plyusnin, V. F.; Zhang, X.; Wu, F.; Deng, N.
850 Intermediates in Photochemistry of Fe(III) Complexes with Carboxylic Acids in Aqueous Solutions.
851 *Photochem. Photobiol. Sci.* **2011**, *10* (3), 425–430. <https://doi.org/10.1039/C0PP00151A>.
- 852 (25) Abrahamson, H. B.; Rezvani, A. B.; Brushmiller, J. G. Photochemical and Spectroscopic Studies of
853 Complexes, of Iron(III) with Citric Acid and Other Carboxylic Acids. *Inorganica Chim. Acta* **1994**, *226*
854 (1), 117–127. [https://doi.org/10.1016/0020-1693\(94\)04077-X](https://doi.org/10.1016/0020-1693(94)04077-X).
- 855 (26) Weller, C.; Tilgner, A.; Bräuer, P.; Herrmann, H. Modeling the Impact of Iron–Carboxylate
856 Photochemistry on Radical Budget and Carboxylate Degradation in Cloud Droplets and Particles.
857 *Environ. Sci. Technol.* **2014**, *48* (10), 5652–5659. <https://doi.org/10.1021/es4056643>.
- 858 (27) George, C.; D’Anna, B.; Herrmann, H.; Weller, C.; Vaida, V.; Donaldson, D. J.; Bartels-Rausch, T.;
859 Ammann, M. Emerging Areas in Atmospheric Photochemistry. In *Atmospheric and Aerosol*
860 *Chemistry*; McNeill, V. F., Ariya, P. A., Eds.; Topics in Current Chemistry; Springer: Berlin,
861 Heidelberg, 2014; pp 1–53. https://doi.org/10.1007/128_2012_393.
- 862 (28) Mangiante, David. M.; Schaller, R. D.; Zarzycki, P.; Banfield, J. F.; Gilbert, B. Mechanism of Ferric
863 Oxalate Photolysis. *ACS Earth Space Chem.* **2017**, *1* (5), 270–276.
864 <https://doi.org/10.1021/acsearthspacechem.7b00026>.
- 865 (29) Borer, P.; Hug, S. J. Photo-Redox Reactions of Dicarboxylates and α -Hydroxydicarboxylates at the
866 Surface of Fe(III)(Hydr)Oxides Followed with in Situ ATR-FTIR Spectroscopy. *J. Colloid Interface Sci.*
867 **2014**, *416*, 44–53. <https://doi.org/10.1016/j.jcis.2013.10.030>.
- 868 (30) Herrmann, H.; Schaefer, T.; Tilgner, A.; Styler, S. A.; Weller, C.; Teich, M.; Otto, T. Tropospheric
869 Aqueous-Phase Chemistry: Kinetics, Mechanisms, and Its Coupling to a Changing Gas Phase. *Chem.*
870 *Rev.* **2015**, *115* (10), 4259–4334. <https://doi.org/10.1021/cr500447k>.

- 871 (31) Wasgestian, F. V. Balzani Und V. Carassiti: Photochemistry of Coordination Compounds. Academic
872 Press, London, New York 1970. 432 Seiten. Preis: 150 s.; 1971.
873 <https://doi.org/10.1002/BBPC.19710750220>.
- 874 (32) Cieřla, P.; Kocot, P.; Mytych, P.; Stasicka, Z. Homogeneous Photocatalysis by Transition Metal
875 Complexes in the Environment. *J. Mol. Catal. Chem.* **2004**, *224* (1), 17–33.
876 <https://doi.org/10.1016/j.molcata.2004.08.043>.
- 877 (33) Feng, W.; Nansheng, D.; Glebov, E. M.; Pozdnyakov, I. P.; Grivin, V. P.; Plyusnin, V. F.; Bazhin, N. M.
878 Kinetics and Mechanism of Photolysis of the Iron(III) Complex with Tartaric Acid. *Russ. Chem. Bull.*
879 **2007**, *56* (5), 900–903. <https://doi.org/10.1007/s11172-007-0136-7>.
- 880 (34) Pozdnyakov, I. P.; Kel, O. V.; Plyusnin, V. F.; Grivin, V. P.; Bazhin, N. M. New Insight into
881 Photochemistry of Ferrioxalate. *J. Phys. Chem. A* **2008**, *112* (36), 8316–8322.
882 <https://doi.org/10.1021/jp8040583>.
- 883 (35) Zhang, X.; Gong, Y.; Wu, F.; Deng, N.; Pozdnyakov, I. P.; Glebov, E. M.; Grivin, V. P.; Plyusnin, V. F.;
884 Bazhinb, N. M. Photochemistry of the Iron(III) Complex with Pyruvic Acid in Aqueous Solutions.
885 *Russ. Chem. Bull.* **2009**, *58* (9), 1828–1836. <https://doi.org/10.1007/s11172-009-0249-2>.
- 886 (36) Abel, B.; Assmann, J.; Buback, M.; Grimm, C.; Kling, M.; Schmatz, S.; Schroeder, J.; Witte, T.
887 Ultrafast Decarboxylation of Carbonyloxy Radicals: Influence of Molecular Structure. *J. Phys.*
888 *Chem. A* **2003**, *107* (45), 9499–9510. <https://doi.org/10.1021/jp0350823>.
- 889 (37) Bockman, T. M.; Hubig, S. M.; Kochi, J. K. Direct Observation of Ultrafast Decarboxylation of
890 Acyloxy Radicals via Photoinduced Electron Transfer in Carboxylate Ion Pairs. *J. Org. Chem.* **1997**,
891 *62* (7), 2210–2221. <https://doi.org/10.1021/jo9617833>.
- 892 (38) Hilborn, J. W.; Pincock, J. A. Rates of Decarboxylation of Acyloxy Radicals Formed in the
893 Photocleavage of Substituted 1-Naphthylmethyl Alkanoates. *J. Am. Chem. Soc.* **1991**, *113* (7),
894 2683–2686. <https://doi.org/10.1021/ja00007a049>.
- 895 (39) Sonntag, C. von; Schuchmann, H.-P. The Elucidation of Peroxyl Radical Reactions in Aqueous
896 Solution with the Help of Radiation-Chemical Methods. *Angew. Chem. Int. Ed. Engl.* **1991**, *30* (10),
897 1229–1253. <https://doi.org/10.1002/anie.199112291>.
- 898 (40) George, C.; Ammann, M.; D’Anna, B.; Donaldson, D. J.; Nizkorodov, S. A. Heterogeneous
899 Photochemistry in the Atmosphere. *Chem. Rev.* **2015**, *115* (10), 4218–4258.
900 <https://doi.org/10.1021/cr500648z>.
- 901 (41) Dou, J.; Alpert, P. A.; Corral Arroyo, P.; Luo, B.; Schneider, F.; Xto, J.; Huthwelker, T.; Borca, C. N.;
902 Henzler, K. D.; Raabe, J.; Watts, B.; Herrmann, H.; Peter, T.; Ammann, M.; Krieger, U. K.
903 Photochemical Degradation of Iron(III) Citrate/Citric Acid Aerosol Quantified with the Combination
904 of Three Complementary Experimental Techniques and a Kinetic Process Model. *Atmospheric*
905 *Chem. Phys.* **2021**, *21* (1), 315–338. <https://doi.org/10.5194/acp-21-315-2021>.
- 906 (42) Zuo, Y.; Hoigne, J. Formation of Hydrogen Peroxide and Depletion of Oxalic Acid in Atmospheric
907 Water by Photolysis of Iron(III)-Oxalato Complexes. *Environ. Sci. Technol.* **1992**, *26* (5), 1014–1022.
908 <https://doi.org/10.1021/es00029a022>.
- 909 (43) Zepp, R. G.; Faust, B. C.; Hoigne, J. Hydroxyl Radical Formation in Aqueous Reactions (PH 3-8) of
910 Iron(II) with Hydrogen Peroxide: The Photo-Fenton Reaction. *Environ. Sci. Technol.* **1992**, *26* (2),
911 313–319. <https://doi.org/10.1021/es00026a011>.
- 912 (44) Balmer, M. E.; Sulzberger, B. Atrazine Degradation in Irradiated Iron/Oxalate Systems: Effects of
913 PH and Oxalate. *Environ. Sci. Technol.* **1999**, *33* (14), 2418–2424.
914 <https://doi.org/10.1021/es9808705>.
- 915 (45) Southworth, B. A.; Voelker, B. M. Hydroxyl Radical Production via the Photo-Fenton Reaction in the
916 Presence of Fulvic Acid. *Environ. Sci. Technol.* **2003**, *37* (6), 1130–1136.
917 <https://doi.org/10.1021/es020757l>.

- 918 (46) Fenton, H. J. H. LXXIII.—Oxidation of Tartaric Acid in Presence of Iron. *J. Chem. Soc. Trans.* **1894**, 65
919 (0), 899–910. <https://doi.org/10.1039/CT8946500899>.
- 920 (47) Zhang, C.; Wang, L.; Wu, F.; Deng, N. Quantitation of Hydroxyl Radicals from Photolysis of Fe(III)-
921 Citrate Complexes in Aerobic Water (5 Pp). *Environ. Sci. Pollut. Res.* **2006**, *13* (3), 156–160.
922 <https://doi.org/10.1065/espr2005.10.287>.
- 923 (48) Kim, S.-M.; Vogelpohl, A. Degradation of Organic Pollutants by the Photo-Fenton-Process. *Chem.*
924 *Eng. Technol.* **1998**, *21* (2), 187–191. [https://doi.org/10.1002/\(SICI\)1521-
925 4125\(199802\)21:2<187::AID-CEAT187>3.0.CO;2-H](https://doi.org/10.1002/(SICI)1521-4125(199802)21:2<187::AID-CEAT187>3.0.CO;2-H).
- 926 (49) Sun, Yunfu.; Pignatello, J. J. Activation of Hydrogen Peroxide by Iron(III) Chelates for Abiotic
927 Degradation of Herbicides and Insecticides in Water. *J. Agric. Food Chem.* **1993**, *41* (2), 308–312.
928 <https://doi.org/10.1021/jf00026a034>.
- 929 (50) Fukushima, M.; Tatsumi, K. Degradation Pathways of Pentachlorophenol by Photo-Fenton Systems
930 in the Presence of Iron(III), Humic Acid, and Hydrogen Peroxide. *Environ. Sci. Technol.* **2001**, *35* (9),
931 1771–1778. <https://doi.org/10.1021/es001088j>.
- 932 (51) Marchetti, B.; Karsili, T. N. V.; Ashfold, M. N. R. Exploring Norrish Type I and Type II Reactions: An
933 Ab Initio Mechanistic Study Highlighting Singlet-State Mediated Chemistry. *Phys. Chem. Chem.*
934 *Phys.* **2019**, *21* (26), 14418–14428. <https://doi.org/10.1039/C8CP07292B>.
- 935 (52) Norrish, R. G. W.; Appleyard, M. E. S. 191. Primary Photochemical Reactions. Part IV.
936 Decomposition of Methyl Ethyl Ketone and Methyl Butyl Ketone. *J. Chem. Soc. Resumed* **1934**, No.
937 0, 874–880. <https://doi.org/10.1039/JR9340000874>.
- 938 (53) Pitts, J. N.; Blacet, F. E. METHYL ETHYL KETONE PHOTOCHEMICAL PROCESSES. *J. Am. Chem. Soc.*
939 **1950**, *72* (6), 2810–2811. <https://doi.org/10.1021/ja01162a544>.
- 940 (54) Norrish, R. G. W.; Bamford, C. H. Photo-Decomposition of Aldehydes and Ketones. *Nature* **1937**,
941 *140* (3535), 195–196. <https://doi.org/10.1038/140195b0>.
- 942 (55) Alpert, P. A.; Dou, J.; Corral Arroyo, P.; Schneider, F.; Xto, J.; Luo, B.; Peter, T.; Huthwelker, T.;
943 Borca, C. N.; Henzler, K. D.; Schaefer, T.; Herrmann, H.; Raabe, J.; Watts, B.; Krieger, U. K.;
944 Ammann, M. Photolytic Radical Persistence Due to Anoxia in Viscous Aerosol Particles. *Nat.*
945 *Commun.* **2021**, *12* (1), 1769. <https://doi.org/10.1038/s41467-021-21913-x>.
- 946 (56) Zhang, Y.; Xie, H. Photomineralization and Photomethanification of Dissolved Organic Matter in
947 Saguenay River Surface Water. *Biogeosciences* **2015**, *12* (22), 6823–6836.
948 <https://doi.org/10.5194/bg-12-6823-2015>.
- 949 (57) Borduas-Dedekind, N.; Ossola, R.; David, R. O.; Boynton, L. S.; Weichlinger, V.; Kanji, Z. A.; McNeill,
950 K. Photomineralization Mechanism Changes the Ability of Dissolved Organic Matter to Activate
951 Cloud Droplets and to Nucleate Ice Crystals. *Atmospheric Chem. Phys.* **2019**, *19* (19), 12397–12412.
952 <https://doi.org/10.5194/acp-19-12397-2019>.
- 953 (58) Barbeau, K.; Moffett, J. W. Laboratory and Field Studies of Colloidal Iron Oxide Dissolution as
954 Mediated by Phagotrophy and Photolysis. *Limnol. Oceanogr.* **2000**, *45* (4), 827–835.
955 <https://doi.org/10.4319/lo.2000.45.4.0827>.
- 956 (59) Wells, M. L.; Mayer, L. M.; Donard, O. F. X.; de Souza Sierra, M. M.; Ackelson, S. G. The Photolysis
957 of Colloidal Iron in the Oceans. *Nature* **1991**, *353* (6341), 248–250.
958 <https://doi.org/10.1038/353248a0>.
- 959 (60) Chen, X.; Ye, X.; Chu, W.; Olk, D. C.; Cao, X.; Schmidt-Rohr, K.; Zhang, L.; Thompson, M. L.; Mao, J.;
960 Gao, H. Formation of Char-Like, Fused-Ring Aromatic Structures from a Nonpyrogenic Pathway
961 during Decomposition of Wheat Straw. *J. Agric. Food Chem.* **2020**, *68* (9), 2607–2614.
962 <https://doi.org/10.1021/acs.jafc.9b06037>.
- 963 (61) Abida, O.; Kolar, M.; Jirkovsky, J.; Mailhot, G. Degradation of 4-Chlorophenol in Aqueous Solution
964 Photoinduced by Fe(III)–Citrate Complex. *Photochem. Photobiol. Sci.* **2012**, *11* (5), 794–802.
965 <https://doi.org/10.1039/C2PP05358F>.

- 966 (62) Faust, B. C.; Zepp, R. G. Photochemistry of Aqueous Iron(III)-Polycarboxylate Complexes: Roles in
967 the Chemistry of Atmospheric and Surface Waters. *Environ. Sci. Technol.* **1993**, *27* (12), 2517–2522.
968 <https://doi.org/10.1021/es00048a032>.
- 969 (63) Arroyo, P. C.; Malecha, K. T.; Ammann, M.; Nizkorodov, S. A. Influence of Humidity and Iron(III) on
970 Photodegradation of Atmospheric Secondary Organic Aerosol Particles. *Phys. Chem. Chem. Phys.*
971 **2018**, *20* (47), 30021–30031. <https://doi.org/10.1039/C8CP03981J>.
- 972 (64) Aiona, P. K.; Lee, H. J.; Leslie, R.; Lin, P.; Laskin, A.; Laskin, J.; Nizkorodov, S. A. Photochemistry of
973 Products of the Aqueous Reaction of Methylglyoxal with Ammonium Sulfate. *ACS Earth Space*
974 *Chem.* **2017**, *1* (8), 522–532. <https://doi.org/10.1021/acsearthspacechem.7b00075>.
- 975 (65) Fleming, L. T.; Lin, P.; Roberts, J. M.; Selimovic, V.; Yokelson, R.; Laskin, J.; Laskin, A.; Nizkorodov, S.
976 A. Molecular Composition and Photochemical Lifetimes of Brown Carbon Chromophores in
977 Biomass Burning Organic Aerosol. *Atmospheric Chem. Phys.* **2020**, *20* (2), 1105–1129.
978 <https://doi.org/10.5194/acp-20-1105-2020>.
- 979 (66) Lin, P.; Fleming, L. T.; Nizkorodov, S. A.; Laskin, J.; Laskin, A. Comprehensive Molecular
980 Characterization of Atmospheric Brown Carbon by High Resolution Mass Spectrometry with
981 Electrospray and Atmospheric Pressure Photoionization. *Anal. Chem.* **2018**, *90* (21), 12493–12502.
982 <https://doi.org/10.1021/acs.analchem.8b02177>.
- 983 (67) Lavi, A.; Lin, P.; Bhaduri, B.; Carmieli, R.; Laskin, A.; Rudich, Y. Characterization of Light-Absorbing
984 Oligomers from Reactions of Phenolic Compounds and Fe(III). *ACS Earth Space Chem.* **2017**, *1* (10),
985 637–646. <https://doi.org/10.1021/acsearthspacechem.7b00099>.
- 986 (68) Budisulistiorini, S. H.; Riva, M.; Williams, M.; Chen, J.; Itoh, M.; Surratt, J. D.; Kuwata, M. Light-
987 Absorbing Brown Carbon Aerosol Constituents from Combustion of Indonesian Peat and Biomass.
988 *Environ. Sci. Technol.* **2017**, *51* (8), 4415–4423. <https://doi.org/10.1021/acs.est.7b00397>.
- 989 (69) Lin, Y.-H.; Budisulistiorini, S. H.; Chu, K.; Siejack, R. A.; Zhang, H.; Riva, M.; Zhang, Z.; Gold, A.;
990 Kautzman, K. E.; Surratt, J. D. Light-Absorbing Oligomer Formation in Secondary Organic Aerosol
991 from Reactive Uptake of Isoprene Epoxydiols. *Environ. Sci. Technol.* **2014**, *48* (20), 12012–12021.
992 <https://doi.org/10.1021/es503142b>.
- 993 (70) West, C. P.; Hettiyadura, A. P. S.; Darmody, A.; Mahamuni, G.; Davis, J.; Novosselov, I.; Laskin, A.
994 Molecular Composition and the Optical Properties of Brown Carbon Generated by the Ethane
995 Flame. *ACS Earth Space Chem.* **2020**, *4* (7), 1090–1103.
996 <https://doi.org/10.1021/acsearthspacechem.0c00095>.
- 997 (71) Siemens, K.; Morales, A.; He, Q.; Li, C.; Hettiyadura, A. P. S.; Rudich, Y.; Laskin, A. Molecular
998 Analysis of Secondary Brown Carbon Produced from the Photooxidation of Naphthalene. *Environ.*
999 *Sci. Technol.* **2022**, *56* (6), 3340–3353. <https://doi.org/10.1021/acs.est.1c03135>.
- 1000 (72) Lee, H. J. (Julie); Aiona, P. K.; Laskin, A.; Laskin, J.; Nizkorodov, S. A. Effect of Solar Radiation on the
1001 Optical Properties and Molecular Composition of Laboratory Proxies of Atmospheric Brown
1002 Carbon. *Environ. Sci. Technol.* **2014**, *48* (17), 10217–10226. <https://doi.org/10.1021/es502515r>.
- 1003 (73) Silva, A. M. N.; Kong, X.; Parkin, M. C.; Cammack, R.; Hider, R. C. Iron(III) Citrate Speciation in
1004 Aqueous Solution. *Dalton Trans.* **2009**, No. 40, 8616. <https://doi.org/10.1039/b910970f>.
- 1005 (74) Gautier-Luneau, I.; Merle, C.; Phanon, D.; Lebrun, C.; Biaso, F.; Serratrice, G.; Pierre, J.-L. New
1006 Trends in the Chemistry of Iron(III) Citrate Complexes: Correlations between X-Ray Structures and
1007 Solution Species Probed by Electrospray Mass Spectrometry and Kinetics of Iron Uptake from
1008 Citrate by Iron Chelators. *Chem. – Eur. J.* **2005**, *11* (7), 2207–2219.
1009 <https://doi.org/10.1002/chem.200401087>.
- 1010 (75) Lehoczki, T.; Józsa, É.; Ósz, K. Ferrioxalate Actinometry with Online Spectrophotometric Detection.
1011 *J. Photochem. Photobiol. Chem.* **2013**, *251*, 63–68.
1012 <https://doi.org/10.1016/j.jphotochem.2012.10.005>.

- 1013 (76) Stucki, J. W. The Quantitative Assay of Minerals for Fe²⁺ and Fe³⁺ Using 1,10-Phenanthroline: II. A
1014 Photochemical Method. *Soil Sci. Soc. Am. J.* **1981**, *45* (3), 638–641.
1015 <https://doi.org/10.2136/sssaj1981.03615995004500030040x>.
- 1016 (77) Stucki, J. W.; Anderson, W. L. The Quantitative Assay of Minerals for Fe²⁺ and Fe³⁺ Using 1,10-
1017 Phenanthroline: I. Sources of Variability. *Soil Sci. Soc. Am. J.* **1981**, *45* (3), 633–637.
1018 <https://doi.org/10.2136/sssaj1981.03615995004500030039x>.
- 1019 (78) Greco, G.; Letzel, T. Main Interactions and Influences of the Chromatographic Parameters in HILIC
1020 Separations. *J. Chromatogr. Sci.* **2013**, *51* (7), 684–693. <https://doi.org/10.1093/chromsci/bmt015>.
- 1021 (79) Buszewski, B.; Noga, S. Hydrophilic Interaction Liquid Chromatography (HILIC)—a Powerful
1022 Separation Technique. *Anal. Bioanal. Chem.* **2012**, *402* (1), 231–247.
1023 <https://doi.org/10.1007/s00216-011-5308-5>.
- 1024 (80) AlChoubassi, G.; Aszyk, J.; Pisarek, P.; Bierla, K.; Ouerdane, L.; Szpunar, J.; Lobinski, R. Advances in
1025 Mass Spectrometry for Iron Speciation in Plants. *TrAC Trends Anal. Chem.* **2018**, *104*, 77–86.
1026 <https://doi.org/10.1016/j.trac.2017.11.006>.
- 1027 (81) Spagou, K.; Tsoukali, H.; Raikos, N.; Gika, H.; Wilson, I. D.; Theodoridis, G. Hydrophilic Interaction
1028 Chromatography Coupled to MS for Metabonomic/Metabolomic Studies. *J. Sep. Sci.* **2010**, *33* (6–
1029 7), 716–727. <https://doi.org/10.1002/jssc.200900803>.
- 1030 (82) Dell'mour, M.; Schenkeveld, W.; Oburger, E.; Fischer, L.; Kraemer, S.; Puschenreiter, M.;
1031 Lämmerhofer, M.; Koellensperger, G.; Hann, S. Analysis of Iron-Phytosiderophore Complexes in
1032 Soil Related Samples: LC-ESI-MS/MS versus CE-MS. *ELECTROPHORESIS* **2012**, *33* (4), 726–733.
1033 <https://doi.org/10.1002/elps.201100466>.
- 1034 (83) Tsednee, M.; Mak, Y.-W.; Chen, Y.-R.; Yeh, K.-C. A Sensitive LC-ESI-Q-TOF-MS Method Reveals
1035 Novel Phytosiderophores and Phytosiderophore–Iron Complexes in Barley. *New Phytol.* **2012**, *195*
1036 (4), 951–961. <https://doi.org/10.1111/j.1469-8137.2012.04206.x>.
- 1037 (84) Rellán-Álvarez, R.; Giner-Martínez-Sierra, J.; Orduna, J.; Orera, I.; Rodríguez-Castrillón, J. Á.; García-
1038 Alonso, J. I.; Abadía, J.; Álvarez-Fernández, A. Identification of a Tri-Iron(III), Tri-Citrate Complex in
1039 the Xylem Sap of Iron-Deficient Tomato Resupplied with Iron: New Insights into Plant Iron Long-
1040 Distance Transport. *Plant Cell Physiol.* **2010**, *51* (1), 91–102. <https://doi.org/10.1093/pcp/pcp170>.
- 1041 (85) Xuan, Y.; Scheuermann, E. B.; Meda, A. R.; Hayen, H.; von Wirén, N.; Weber, G. Separation and
1042 Identification of Phytosiderophores and Their Metal Complexes in Plants by Zwitterionic
1043 Hydrophilic Interaction Liquid Chromatography Coupled to Electrospray Ionization Mass
1044 Spectrometry. *J. Chromatogr. A* **2006**, *1136* (1), 73–81.
1045 <https://doi.org/10.1016/j.chroma.2006.09.060>.
- 1046 (86) Pluskal, T.; Castillo, S.; Villar-Briones, A.; Orešič, M. MZmine 2: Modular Framework for Processing,
1047 Visualizing, and Analyzing Mass Spectrometry-Based Molecular Profile Data. *BMC Bioinformatics*
1048 **2010**, *11* (1), 395. <https://doi.org/10.1186/1471-2105-11-395>.
- 1049 (87) Kuhl, C.; Tautenhahn, R.; Böttcher, C.; Larson, T. R.; Neumann, S. CAMERA: An Integrated Strategy
1050 for Compound Spectra Extraction and Annotation of LC/MS Data Sets. *Anal. Chem.* **2012**, *84* (1),
1051 283–289. <https://doi.org/10.1021/ac202450g>.
- 1052 (88) Myers, O. D.; Sumner, S. J.; Li, S.; Barnes, S.; Du, X. One Step Forward for Reducing False Positive
1053 and False Negative Compound Identifications from Mass Spectrometry Metabolomics Data: New
1054 Algorithms for Constructing Extracted Ion Chromatograms and Detecting Chromatographic Peaks.
1055 *Anal. Chem.* **2017**, *89* (17), 8696–8703. <https://doi.org/10.1021/acs.analchem.7b00947>.
- 1056 (89) Jaitly, N.; Mayampurath, A.; Littlefield, K.; Adkins, J. N.; Anderson, G. A.; Smith, R. D. Decon2LS: An
1057 Open-Source Software Package for Automated Processing and Visualization of High Resolution
1058 Mass Spectrometry Data. *BMC Bioinformatics* **2009**, *10* (1), 87. <https://doi.org/10.1186/1471-2105-10-87>.
- 1059

- 1060 (90) Roach, P. J.; Laskin, J.; Laskin, A. Higher-Order Mass Defect Analysis for Mass Spectra of Complex
1061 Organic Mixtures. *Anal. Chem.* **2011**, *83* (12), 4924–4929. <https://doi.org/10.1021/ac200654j>.
- 1062 (91) Vetter, W. F. W. McLafferty, F. Turecek. Interpretation of Mass Spectra. Fourth Edition (1993).
1063 University Science Books, Mill Valley, California. *Biol. Mass Spectrom.* **1994**, *23* (6), 379–379.
1064 <https://doi.org/10.1002/bms.1200230614>.
- 1065 (92) Chen, Y.; Bond, T. C. Light Absorption by Organic Carbon from Wood Combustion. *Atmospheric*
1066 *Chem. Phys.* **2010**, *10* (4), 1773–1787. <https://doi.org/10.5194/acp-10-1773-2010>.
- 1067 (93) Kirchstetter, T. W.; Novakov, T.; Hobbs, P. V. Evidence That the Spectral Dependence of Light
1068 Absorption by Aerosols Is Affected by Organic Carbon. *J. Geophys. Res. Atmospheres* **2004**, *109*
1069 (D21). <https://doi.org/10.1029/2004JD004999>.
- 1070 (94) Hettiyadura, A. P. S.; Laskin, A. Quantitative Analysis of Polycyclic Aromatic Hydrocarbons Using
1071 High-Performance Liquid Chromatography-Photodiode Array-High-Resolution Mass Spectrometric
1072 Detection Platform Coupled to Electrospray and Atmospheric Pressure Photoionization Sources. *J.*
1073 *Mass Spectrom.* **2022**, *57* (2), e4804. <https://doi.org/10.1002/jms.4804>.
- 1074 (95) Hettiyadura, A. P. S.; Garcia, V.; Li, C.; West, C. P.; Tomlin, J.; He, Q.; Rudich, Y.; Laskin, A. Chemical
1075 Composition and Molecular-Specific Optical Properties of Atmospheric Brown Carbon Associated
1076 with Biomass Burning. *Environ. Sci. Technol.* **2021**, *55* (4), 2511–2521.
1077 <https://doi.org/10.1021/acs.est.0c05883>.
- 1078 (96) Krause, N.; Kuhn, S.; Frotscher, E.; Nikels, F.; Hawe, A.; Garidel, P.; Menzen, T. Oil-Immersion Flow
1079 Imaging Microscopy for Quantification and Morphological Characterization of Submicron Particles
1080 in Biopharmaceuticals. *AAPS J.* **2021**, *23* (1), 13. <https://doi.org/10.1208/s12248-020-00547-9>.
- 1081 (97) Hasberg, A. K. M.; Bijaksana, S.; Held, P.; Just, J.; Melles, M.; Morlock, M. A.; Opitz, S.; Russell, J. M.;
1082 Vogel, H.; Wennrich, V. Modern Sedimentation Processes in Lake Towuti, Indonesia, Revealed by
1083 the Composition of Surface Sediments. *Sedimentology* **2019**, *66* (2), 675–698.
1084 <https://doi.org/10.1111/sed.12503>.
- 1085 (98) Seraghni, N.; Belattar, S.; Mameri, Y.; Debbache, N.; Sehili, T. Fe(III)-Citrate-Complex-Induced
1086 Photooxidation of 3-Methylphenol in Aqueous Solution
1087 <https://www.hindawi.com/journals/ijp/2012/630425/> (accessed 2021 -03 -02).
1088 <https://doi.org/10.1155/2012/630425>.
- 1089 (99) Ruales-Lonfat, C.; Barona, J. F.; Sienkiewicz, A.; Vélez, J.; Benítez, L. N.; Pulgarín, C. Bacterial
1090 Inactivation with Iron Citrate Complex: A New Source of Dissolved Iron in Solar Photo-Fenton
1091 Process at near-Neutral and Alkaline PH. *Appl. Catal. B Environ.* **2016**, *180*, 379–390.
1092 <https://doi.org/10.1016/j.apcatb.2015.06.030>.
- 1093 (100) Ogi, Y.; Obara, Y.; Katayama, T.; Suzuki, Y.-I.; Liu, S. Y.; Bartlett, N. C.-M.; Kurahashi, N.;
1094 Karashima, S.; Togashi, T.; Inubushi, Y.; Ogawa, K.; Owada, S.; Rubešová, M.; Yabashi, M.; Misawa,
1095 K.; Slaviček, P.; Suzuki, T. Ultraviolet Photochemical Reaction of [Fe(III)(C₂O₄)₃]³⁻ in Aqueous
1096 Solutions Studied by Femtosecond Time-Resolved X-Ray Absorption Spectroscopy Using an X-Ray
1097 Free Electron Laser. *Struct. Dyn.* **2015**, *2* (3), 034901. <https://doi.org/10.1063/1.4918803>.
- 1098 (101) Dekkiche, B. A.; Seraghni, N.; Debbache, N.; Ghoul, I.; Sehili, T. Effect of Natural and Artificial
1099 Light on Fe(III) Organic Complexes Photolysis: Case of Fe (III)-Malonate and Fe(III)-Malate. *Int. J.*
1100 *Chem. React. Eng.* **2019**, *17* (2). <https://doi.org/10.1515/ijcre-2018-0106>.
- 1101 (102) Wang, L.; Zhang, C.; Mestankova, H.; Wu, F.; Deng, N.; Pan, G.; Bolte, M.; Mailhot, G.
1102 Photoinduced Degradation of 2,4-Dichlorophenol in Water: Influence of Various Fe(III)
1103 Carboxylates. *Photochem. Photobiol. Sci.* **2009**, *8* (7), 1059–1065.
1104 <https://doi.org/10.1039/B902607J>.
- 1105 (103) Fu, Z.; Chen, R. Study of Complexes of Tannic Acid with Fe(III) and Fe(II). *J. Anal. Methods Chem.*
1106 **2019**, *2019*, e3894571. <https://doi.org/10.1155/2019/3894571>.

- 1107 (104) Thorarinsdottir, A. E.; Gaudette, A. I.; Harris, T. D. Spin-Crossover and High-Spin Iron(II)
1108 Complexes as Chemical Shift ¹⁹F Magnetic Resonance Thermometers. *Chem. Sci.* **2017**, *8* (3),
1109 2448–2456. <https://doi.org/10.1039/C6SC04287B>.
- 1110 (105) Baxendale, J. H.; Magee, J. The Photochemical Oxidation of Benzene in Aqueous Solution by
1111 Ferric Ion. *Trans. Faraday Soc.* **1955**, *51* (0), 205–213. <https://doi.org/10.1039/TF9555100205>.
- 1112 (106) Moffett, J. W.; Zika, R. G. Reaction Kinetics of Hydrogen Peroxide with Copper and Iron in
1113 Seawater. *Environ. Sci. Technol.* **1987**, *21* (8), 804–810. <https://doi.org/10.1021/es00162a012>.
- 1114 (107) Quici, N.; Morgada, M. E.; Gettar, R. T.; Bolte, M.; Litter, M. I. Photocatalytic Degradation of
1115 Citric Acid under Different Conditions: TiO₂ Heterogeneous Photocatalysis against Homogeneous
1116 Photolytic Processes Promoted by Fe(III) and H₂O₂. *Appl. Catal. B Environ.* **2007**, *71* (3), 117–124.
1117 <https://doi.org/10.1016/j.apcatb.2006.09.001>.
- 1118 (108) Waite, T. D.; Morel, F. M. Photoreductive Dissolution of Colloidal Iron Oxides in Natural Waters.
1119 *Environ. Sci. Technol.* **1984**, *18* (11), 860–868. <https://doi.org/10.1021/es00129a010>.
- 1120 (109) Meng, X.; Ryu, J.; Kim, B.; Ko, S. Application of Iron Oxide as a PH-Dependent Indicator for
1121 Improving the Nutritional Quality. *Clin. Nutr. Res.* **2016**, *5* (3), 172–179.
1122 <https://doi.org/10.7762/cnr.2016.5.3.172>.
- 1123 (110) Li, F.; Koopal, L.; Tan, W. Roles of Different Types of Oxalate Surface Complexes in Dissolution
1124 Process of Ferrihydrite Aggregates. *Sci. Rep.* **2018**, *8* (1), 2060. <https://doi.org/10.1038/s41598-018-20401-5>.
- 1126 (111) Borer, P.; Hug, S. J.; Sulzberger, B.; Kraemer, S. M.; Kretzschmar, R. Photolysis of Citrate on the
1127 Surface of Lepidocrocite: An in Situ Attenuated Total Reflection Infrared Spectroscopy Study. *J.*
1128 *Phys. Chem. C* **2007**, *111* (28), 10560–10569. <https://doi.org/10.1021/jp0685941>.
- 1129 (112) Lv, Y.; Liu, J.; Zhu, R.; Zhu, J.; Chen, Q.; Liang, X.; He, H. Photoreductive Dissolution of Iron
1130 (Hydr)Oxides and Its Geochemical Significance. *ACS Earth Space Chem.* **2022**.
1131 <https://doi.org/10.1021/acsearthspacechem.1c00334>.
- 1132 (113) Silva, A. M. N.; Kong, X.; Parkin, M. C.; Cammack, R.; Hider, R. C. Iron(III) Citrate Speciation in
1133 Aqueous Solution. *Dalton Trans.* **2009**, No. 40, 8616–8625. <https://doi.org/10.1039/B910970F>.
- 1134 (114) National Center for Biotechnology Information. “PubChem Compound Summary for CID
1135 14298643, Ferrous Citrate” PubChem, [https://pubchem.ncbi.nlm.nih.gov/compound/Ferrous-](https://pubchem.ncbi.nlm.nih.gov/compound/Ferrous-Citrate)
1136 Citrate. Accessed 18 December, 2020.
- 1137 (115) Pierre, J. L.; Gautier-Luneau, I. Iron and Citric Acid: A Fuzzy Chemistry of Ubiquitous Biological
1138 Relevance. *Biometals* **2000**, *13* (1), 91–96. <https://doi.org/10.1023/A:1009225701332>.
- 1139 (116) Königsberger, L.-C.; Königsberger, E.; May, P. M.; Hefter, G. T. Complexation of Iron(III) and
1140 Iron(II) by Citrate. Implications for Iron Speciation in Blood Plasma. *J. Inorg. Biochem.* **2000**, *78* (3),
1141 175–184. [https://doi.org/10.1016/S0162-0134\(99\)00222-6](https://doi.org/10.1016/S0162-0134(99)00222-6).
- 1142 (117) Pham, A. N.; Waite, T. D. Oxygenation of Fe(II) in the Presence of Citrate in Aqueous Solutions at
1143 PH 6.0–8.0 and 25 °C: Interpretation from an Fe(II)/Citrate Speciation Perspective. *J. Phys. Chem. A*
1144 **2008**, *112* (4), 643–651. <https://doi.org/10.1021/jp077219l>.
- 1145 (118) Harris, D. C.; Aisen, P. Facilitation of Fe(II) Antioxidation by Fe(III) Complexing Agents. *Biochim.*
1146 *Biophys. Acta BBA - Gen. Subj.* **1973**, *329* (1), 156–158. [https://doi.org/10.1016/0304-4165\(73\)90019-6](https://doi.org/10.1016/0304-4165(73)90019-6).
- 1148 (119) Liu, M. J.; Wiegel, A. A.; Wilson, K. R.; Houle, F. A. Aerosol Fragmentation Driven by Coupling of
1149 Acid–Base and Free-Radical Chemistry in the Heterogeneous Oxidation of Aqueous Citric Acid by
1150 OH Radicals. *J. Phys. Chem. A* **2017**, *121* (31), 5856–5870.
1151 <https://doi.org/10.1021/acs.jpca.7b04892>.
- 1152 (120) Moharregh-Khiabani, D.; Linker, R. A.; Gold, R.; Stangel, M. Fumaric Acid and Its Esters: An
1153 Emerging Treatment for Multiple Sclerosis. *Curr. Neuropharmacol.* **2009**, *7* (1), 60–64.
1154 <https://doi.org/10.2174/157015909787602788>.

- 1155 (121) Ponczek, M.; Hayeck, N.; Emmelin, C.; George, C. Heterogeneous Photochemistry of Dicarboxylic
1156 Acids on Mineral Dust. *Atmos. Environ.* **2019**, *212*, 262–271.
1157 <https://doi.org/10.1016/j.atmosenv.2019.05.032>.
- 1158 (122) Mazurek, A.; Jamroz, J. Precision of Dehydroascorbic Acid Quantitation with the Use of the
1159 Subtraction Method – Validation of HPLC–DAD Method for Determination of Total Vitamin C in
1160 Food. *Food Chem.* **2015**, *173*, 543–550. <https://doi.org/10.1016/j.foodchem.2014.10.065>.
- 1161 (123) Dewhirst, R. A.; Fry, S. C. The Oxidation of Dehydroascorbic Acid and 2,3-Diketogulonate by
1162 Distinct Reactive Oxygen Species. *Biochem. J.* **2018**, *475* (21), 3451–3470.
1163 <https://doi.org/10.1042/BCJ20180688>.
- 1164 (124) Siegmann, K.; Sattler, K. Formation Mechanism for Polycyclic Aromatic Hydrocarbons in
1165 Methane Flames. *J. Chem. Phys.* **2000**, *112* (2), 698–709. <https://doi.org/10.1063/1.480648>.
- 1166 (125) Cain, J.; Laskin, A.; Kholghy, M. R.; Thomson, M. J.; Wang, H. Molecular Characterization of
1167 Organic Content of Soot along the Centerline of a Coflow Diffusion Flame. *Phys. Chem. Chem. Phys.*
1168 **2014**, *16* (47), 25862–25875. <https://doi.org/10.1039/C4CP03330B>.
- 1169 (126) Saleh, R. From Measurements to Models: Toward Accurate Representation of Brown Carbon in
1170 Climate Calculations. *Curr. Pollut. Rep.* **2020**, *6* (2), 90–104. [https://doi.org/10.1007/s40726-020-](https://doi.org/10.1007/s40726-020-00139-3)
1171 [00139-3](https://doi.org/10.1007/s40726-020-00139-3).
- 1172 (127) Zhou, Y.; West, C. P.; Hettiyadura, A. P. S.; Niu, X.; Wen, H.; Cui, J.; Shi, T.; Pu, W.; Wang, X.;
1173 Laskin, A. Measurement Report: Molecular Composition, Optical Properties, and Radiative Effects
1174 of Water-Soluble Organic Carbon in Snowpack Samples from Northern Xinjiang, China.
1175 *Atmospheric Chem. Phys.* **2021**, *21* (11), 8531–8555. <https://doi.org/10.5194/acp-21-8531-2021>.
- 1176 (128) Alexander, D. T. L.; Crozier, P. A.; Anderson, J. R. Brown Carbon Spheres in East Asian Outflow
1177 and Their Optical Properties. *Science* **2008**, *321* (5890), 833–836.
1178 <https://doi.org/10.1126/science.1155296>.
- 1179 (129) Misovich, M. V.; Hettiyadura, A. P. S.; Jiang, W.; Zhang, Q.; Laskin, A. Molecular-Level Study of
1180 the Photo-Oxidation of Aqueous-Phase Guaiacyl Acetone in the Presence of 3C*: Formation of
1181 Brown Carbon Products. *ACS Earth Space Chem.* **2021**, *5* (8), 1983–1996.
1182 <https://doi.org/10.1021/acsearthspacechem.1c00103>.
- 1183 (130) Zangmeister, C. D.; You, R.; Lunny, E. M.; Jacobson, A. E.; Okumura, M.; Zachariah, M. R.;
1184 Radney, J. G. Measured In-Situ Mass Absorption Spectra for Nine Forms of Highly-Absorbing
1185 Carbonaceous Aerosol. *Carbon* **2018**, *136*, 85–93. <https://doi.org/10.1016/j.carbon.2018.04.057>.
- 1186 (131) Miles, C. J.; Brezonik, P. L. Oxygen Consumption in Humic-Colored Waters by a Photochemical
1187 Ferrous-Ferric Catalytic Cycle. *Environ. Sci. Technol.* **1981**, *15* (9), 1089–1095.
1188 <https://doi.org/10.1021/es00091a010>.
- 1189 (132) Byrne, J. M.; Klueglein, N.; Pearce, C.; Rosso, K. M.; Appel, E.; Kappler, A. Redox Cycling of Fe(II)
1190 and Fe(III) in Magnetite by Fe-Metabolizing Bacteria. *Science* **2015**, *347* (6229), 1473–1476.
1191 <https://doi.org/10.1126/science.aaa4834>.
- 1192 (133) Emmenegger, L.; Schönenberger, R.; Sigg, L.; Sulzberger, B. Light-Induced Redox Cycling of Iron
1193 in Circumneutral Lakes. *Limnol. Oceanogr.* **2001**, *46* (1), 49–61.
1194 <https://doi.org/10.4319/lo.2001.46.1.0049>.
- 1195 (134) Cwiertny, D. M.; Young, M. A.; Grassian, V. H. Chemistry and Photochemistry of Mineral Dust
1196 Aerosol. *Annu. Rev. Phys. Chem.* **2008**, *59* (1), 27–51.
1197 <https://doi.org/10.1146/annurev.physchem.59.032607.093630>.
- 1198 (135) Siefert, R. L.; Pehkonen, S. O.; Erel, Y.; Hoffmann, M. R. Iron Photochemistry of Aqueous
1199 Suspensions of Ambient Aerosol with Added Organic Acids. *Geochim. Cosmochim. Acta* **1994**, *58*
1200 (15), 3271–3279. [https://doi.org/10.1016/0016-7037\(94\)90055-8](https://doi.org/10.1016/0016-7037(94)90055-8).

- 1201 (136) Pu, W.; Wang, X.; Wei, H.; Zhou, Y.; Shi, J.; Hu, Z.; Jin, H.; Chen, Q. Properties of Black Carbon and
1202 Other Insoluble Light-Absorbing Particles in Seasonal Snow of Northwestern China. *The Cryosphere*
1203 **2017**, *11* (3), 1213–1233. <https://doi.org/10.5194/tc-11-1213-2017>.
- 1204 (137) Shi, T.; Cui, J.; Chen, Y.; Zhou, Y.; Pu, W.; Xu, X.; Chen, Q.; Zhang, X.; Wang, X. Enhanced Light
1205 Absorption and Reduced Snow Albedo Due to Internally Mixed Mineral Dust in Grains of Snow.
1206 *Atmospheric Chem. Phys.* **2021**, *21* (8), 6035–6051. <https://doi.org/10.5194/acp-21-6035-2021>.
- 1207 (138) Zhou, Y.; West, C. P.; Hettiyadura, A. P. S.; Pu, W.; Shi, T.; Niu, X.; Wen, H.; Cui, J.; Wang, X.;
1208 Laskin, A. Molecular Characterization of Water-Soluble Brown Carbon Chromophores in Snowpack
1209 from Northern Xinjiang, China. *Environ. Sci. Technol.* **2022**, *56* (7), 4173–4186.
1210 <https://doi.org/10.1021/acs.est.1c07972>.
- 1211