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## Urbanization can accelerate climate change by increasing soil N<sub>2</sub>O emission while reducing CH<sub>4</sub> uptake

Running title: Urbanization effect on soil non-CO<sub>2</sub> GHG flux

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## Abstract

Urban land use change has the potential to affect local to global biogeochemical carbon (C) and nitrogen (N) cycles and associated greenhouse gas (GHG) fluxes. We conducted a meta-analysis to 1) assess the effects of urbanization-induced land-use conversion on soil nitrous oxide ( $N_2O$ ) and methane ( $CH_4$ ) fluxes, 2) quantify direct  $N_2O$  emission factors ( $EF_d$ ) of fertilized urban soils used e.g., as lawns or forests, and 3) identify the key drivers leading to flux changes associated with urbanization. On average, urbanization increases soil  $N_2O$  emissions by 153%, to  $3.0\text{ kg N ha}^{-1}\text{ yr}^{-1}$ , while rates of soil  $CH_4$  uptake are reduced by 50%, to  $2.0\text{ kg C ha}^{-1}\text{ yr}^{-1}$ . The global mean annual  $N_2O$   $EF_d$  of fertilized lawns and urban forests is 1.4%, suggesting that urban soils can be regional hotspots of  $N_2O$  emissions. On a global basis, conversion of land to urban greenspaces has increased soil  $N_2O$  emission by  $0.46\text{ Tg N}_2O\text{-N yr}^{-1}$  and decreased soil  $CH_4$  uptake by  $0.58\text{ Tg CH}_4\text{-C yr}^{-1}$ . Urbanization driven changes in soil  $N_2O$  emission and  $CH_4$  uptake are associated with changes in soil properties (bulk density, pH, total N content and C/N ratio), increased temperature, and management practices, especially fertilizer use. Overall, our meta-analysis shows that urbanization increases soil  $N_2O$  emissions and reduces the role of soils as a sink for atmospheric  $CH_4$ . These effects can be mitigated by avoiding soil compaction, reducing fertilization of lawns, and by restoring native ecosystems in urban landscapes.

**KEY WORDS:** urbanization; nitrous oxide; methane; climate change; greenspace; lawn; urban forest; emission factor

## 1 INTRODUCTION

Increasing populations and the search for social and economic opportunities are driving people to move from rural to urban areas (Montgomery, 2008; Wang et al., 2021a; Zhang et al., 2022a). Approximately 4% of the global land area is urbanized and half of the world's population lives in urban areas (United Nations, 2018; Zhang et al., 2021). The trend towards urbanization is driving conversion of natural (e.g., grassland and forest) and managed (e.g., arable land and grazed pasture) ecosystems into urban landscapes dominated by residential areas with interspersed greenspaces (e.g., lawns and urban forests) (IPCC, 2013; van Delden et al., 2018; Liu et al., 2019, 2020). These conversions affect ecosystem functions and services including biodiversity support, food security, and climate regulation by carbon sequestration (Zhou et al., 2004; Seto et al., 2012; Wang et al., 2021a; Yu et al., 2022; Zhang et al., 2022a).

Urbanization and associated increases in energy consumption are drivers of global climate change (e.g., Kaye et al., 2006; Grimm et al., 2008; Hopkins et al., 2016; Pan et al., 2020). Urban building, transport and industrial activities produce significant amounts of greenhouse gas (GHG) emissions (Hoornweg et al., 2011; Sari and Bayram, 2014; Ward et al., 2015). Cities also have significant greenspaces for recreation or improvement of air quality and urban climate (Gregg et al., 2003; Law and Patton, 2017; Nowak et al., 2018; Liu et al., 2021; Yu et al., 2022). These urban green areas can act as sources and/or sinks for GHGs, including carbon dioxide (CO<sub>2</sub>), which can be sequestered in urban forest ecosystems (Escobedo et al., 2011), as well as the non-CO<sub>2</sub> GHG methane (CH<sub>4</sub>) and nitrous oxide (N<sub>2</sub>O), which have global warming potentials 28 and 273 times higher than CO<sub>2</sub> over a 100-year horizon, respectively (IPCC, 2021). CH<sub>4</sub> and N<sub>2</sub>O fluxes at the soil-atmosphere interface are the result of simultaneously occurring microbial aerobic or anaerobic production and consumption processes, which

are regulated by a complex suite of environmental factors such as soil moisture and temperature as well as nutrient availability (Conrad, 1996; Butterbach-Bahl et al., 2013; van Delden et al., 2018; Ni and Groffman, 2018). Urbanization significantly alters these factors due to heat island effects, increased atmospheric N deposition as a result of industrial and traffic emissions, and soil compaction and redistribution in the course of urban landscaping. This alteration affects C and N biogeochemical cycles in urban soils and associated production and consumption of CH<sub>4</sub> and N<sub>2</sub>O (Kaye et al., 2006; Groffman et al., 2009; Ni and Groffman, 2018; Xu et al., 2022). Previous studies have reported low or full inhibition of soil atmospheric CH<sub>4</sub> uptake in urban forests and lawns in China and the USA due to increases in soil bulk density, which constrains diffusion of atmospheric CH<sub>4</sub> to sites of microbial CH<sub>4</sub> uptake by methanotrophs, or increases in soil N availability, which may directly inhibit CH<sub>4</sub> uptake (Kaye et al., 2004; Groffman and Pouyat, 2009; Zhang et al., 2014). Zhang et al. (2021) found that with increasing intensity of urbanization soil CH<sub>4</sub> uptake decreased, while the methanotrophic community was not affected.

With regard to soil N<sub>2</sub>O emissions from urban greenspaces, contradictory results have been reported. While some studies have reported low N<sub>2</sub>O emissions from urban soils (e.g., Groffman et al., 2009), other studies have reported high fluxes, especially if soils were fertilized (e.g., lawns or golf courses) (Kaye et al., 2004; van Delden et al., 2018; Xu et al., 2022). Dutt and Tanwar (2020) conducted a meta-analysis to evaluate the effect of fertilizer application on lawn N<sub>2</sub>O emissions, and found that lawn N<sub>2</sub>O emissions from fertilized plots were 41% (or 0.29 g N m<sup>-2</sup> yr<sup>-1</sup>) higher than those from non-fertilized controls. A literature review by Braun and Bremer (2018) found that annual N<sub>2</sub>O emissions from fertilized urban lawns ranged from 1.0 to 7.6 kg N ha<sup>-1</sup> yr<sup>-1</sup>, comparable to N<sub>2</sub>O fluxes from intensively fertilized agricultural soils. A study

conducted in Fort Collins, Colorado USA showed that even though fertilized urban lawns occupied only 6.4% of the land area, they contributed 30% of regional soil to atmosphere N<sub>2</sub>O emissions (Kaye et al., 2004). Fertilization rates of lawns can exceed 200 kg N ha<sup>-1</sup> yr<sup>-1</sup> (Braun and Bremer, 2018), and as soil N<sub>2</sub>O emissions can increase nonlinearly in response to fertilization (Shcherbak et al., 2014), the currently used N<sub>2</sub>O emission factor (EF<sub>d</sub>) for fertilized soils of 1% (IPCC, 2006) may not accurately estimate the regional source strength of soils in urban areas (Kaye et al., 2006). There is a clear need to test and develop N<sub>2</sub>O EF<sub>d</sub> for fertilized urban soils.

In this study, we performed a global synthesis of studies of soil CH<sub>4</sub> and N<sub>2</sub>O fluxes from urban greenspaces. Our objectives were (1) to provide an overall assessment of how urbanization-induced land-use conversion affects soil CH<sub>4</sub> and N<sub>2</sub>O fluxes, (2) to calculate a general global N<sub>2</sub>O EF<sub>d</sub> for fertilized soils of urban greenspaces, and (3) to explore how climate conditions, soil properties or land-use modulates the response of soil non-CO<sub>2</sub> GHGs fluxes to urbanization.

## 2 MATERIALS AND METHODS

### 2.1 Literature search and data extraction

In order to study the effects of urbanization on non-CO<sub>2</sub> GHG (CH<sub>4</sub> and N<sub>2</sub>O) fluxes in soils of urban greenspaces, we carried out a detailed review of the peer-reviewed literature published before April 2022, using ISI-Web of Science and Google Scholar. Specific search terms were: urbanization; urban greenspace; turfgrass; lawn; urban forest; N<sub>2</sub>O; nitrous oxide; CH<sub>4</sub>; methane; GHG; greenhouse gas. Articles were retained for use in our meta-analysis if: (a) the study was field-based (i.e., excluding e.g., pot and laboratory experiments or model simulations); (b) the observation period covered at least a growing season (which usually refers to the main growing period, i.e., during the summer and/or wet season); (c) CH<sub>4</sub> and/or N<sub>2</sub>O fluxes were reported for soils of

urban greenspaces (lawns or urban forests) as well as for comparable land-uses in a non-urban setting; (d) the reported experiment included treatments of different N type or rate of fertilizer application in the urban greenspaces. The raw data were either extracted directly from tables and texts or retrieved digitally from graphs using GetData Graph Digitizer (version 2.25.0.32, <http://www.getdata-graph-digitizer.com/>).

We identified a total of 32 peer-reviewed publications reporting on 197 side-by-side comparisons of urban lawns and forests with natural (grassland and forest) or non-urban managed (arable land and grazed pasture) ecosystems, as well as 119 observational studies of soil N<sub>2</sub>O emissions in urban greenspaces (Dataset S1). The majority of these observations were made in North America, with a few studies in Europe, Australia or Asia. The 197 paired observations were further grouped into three categories: urbanization effects on (1) soil N<sub>2</sub>O emission (n = 78), (2) soil CH<sub>4</sub> uptake (n = 72) and (3) sum of non-CO<sub>2</sub> GHG (CH<sub>4</sub>+N<sub>2</sub>O) flux (n = 47). The 119 experimental observations were used to investigate N fertilization effects on soil N<sub>2</sub>O emission from urban forests and lawns.

For soil CH<sub>4</sub> fluxes, all observations covered at least an entire year, allowing us to directly retrieve or calculate annual cumulative uptake rates (in kg C ha<sup>-1</sup> yr<sup>-1</sup>). For soil N<sub>2</sub>O fluxes, not all studies report on measurements covering an entire observational year (n = 12 out of 28 studies). For cases reporting only growing season soil N<sub>2</sub>O fluxes (these studies were mainly located on the temperate zone), we gap-filled the data by assuming that non-growing time fluxes account for on average 28% of the total annual budget as was found by Zhan et al. (2021) for a temperate urban site in China. Cumulative non-CO<sub>2</sub> GHG fluxes, i.e., the sum of soil CH<sub>4</sub> and N<sub>2</sub>O fluxes, were only calculated for studies which simultaneously measured both GHG fluxes (n = 6 out of 15 studies). For this calculation we used the conversion factors of 28 and 273 for the

global warming potential of CH<sub>4</sub> and N<sub>2</sub>O, respectively, for a 100-year time horizon (IPCC, 2021), and expressed the non-CO<sub>2</sub> GHG flux as CO<sub>2</sub> equivalents (in kg CO<sub>2</sub>-eq ha<sup>-1</sup> yr<sup>-1</sup>). In addition, other key information including location of experimental site, rate and type of N fertilizer application, soil texture, soil bulk density (BD), soil pH, soil organic C (SOC), total N (TN) and C/N ratio was extracted and included in our database (Dataset S1).

## 2.2 Data and statistical analysis

An effect size of the natural logarithm of the response ratio (ln RR) was employed to assess the effect of urbanization-induced land-use conversion on soil N<sub>2</sub>O emission, soil CH<sub>4</sub> uptake and soil physicochemical properties (e.g., SOC, TN, C/N ratio, BD, pH and clay content) (Hedges et al., 1999):

$$\ln \text{RR} = \ln \left( \frac{\bar{X}_t}{\bar{X}_c} \right) = \ln(\bar{X}_t) - \ln(\bar{X}_c)$$

where  $\bar{X}_t$  and  $\bar{X}_c$  are the mean values of the selected annual N<sub>2</sub>O emissions, CH<sub>4</sub> uptake rates or soil property parameters for urban and non-urban land-use, respectively. The meta-analysis results were reported as percentage changes (i.e., %change= (exp (ln RR) - 1) × 100%). Positive percentage changes indicate an increase, and negative values denote a decrease in the variable as an effect of urbanization.

The values of non-CO<sub>2</sub> GHG (CH<sub>4</sub>+N<sub>2</sub>O) fluxes can be both positive and negative, which makes the abovementioned ln RR equation problematic. As suggested by previous studies (e.g., van Groenigen et al., 2011; Xia et al., 2017), therefore, the response of a variable (RR') was assessed by the formula:

$$\text{RR}' = \bar{X}_t - \bar{X}_c$$

For these analyses, effect sizes were weighted by the inverse of the pooled variance (var) of each individual observation, calculated by the following equation (Hedges et al., 1999):

$$\text{var} = \frac{s_t^2}{n_t \bar{X}_t^2} + \frac{s_c^2}{n_c \bar{X}_c^2}$$

where  $s_t$  and  $s_c$  are the standard deviations (SD) of  $\bar{X}_t$  and  $\bar{X}_c$ , respectively;  $n_t$  and  $n_c$  are the number of replicates per treatment for  $\bar{X}_t$  and  $\bar{X}_c$ , respectively. For studies in which the standard error (SE) rather than SD was reported, we recalculated the SD by:

$$SD = SE \times \sqrt{n}$$

where  $n$  is the number of replicates. When the SD or SE data were not reported, we estimated the SD values based on the average coefficient of variation for the known data (De Stefano and Jacobson, 2018).

To assess N fertilization effects on soil  $N_2O$  emissions in urban greenspaces, we performed a pair-wise meta-analysis using  $EF_d$  as the effect size, which was calculated by the equation (Wang et al., 2020):

$$EF_d(\%) = \frac{E_f - E_0}{N} \times 100$$

where  $N$  refers to the rate of N fertilizer application (in  $\text{kg N ha}^{-1} \text{yr}^{-1}$ ), and  $E_f$  and  $E_0$  represent annual cumulative  $N_2O$  emissions (in  $\text{kg N ha}^{-1} \text{yr}^{-1}$ ) for fertilized urban soils and non-fertilized controls, respectively. For sites where no background  $N_2O$  emissions from a non-fertilized control ( $E_0$ ) were reported, we estimated background emissions from studies conducted in the same region under comparable climate and soil conditions. More than half of the  $EF_d$  data lacked information on the pooled variance, so the meta-analysis of  $EF_d$  was unweighted (Chen et al., 2013; Qasim et al., 2021).

Means of effect sizes and their 95% confidence intervals (CIs) were calculated using a bootstrapping procedure with 999 iterations generated by MetaWin 2.1 (Rosenberg et al., 2000). If 95% CI values of the  $\ln RR$  for a variable did not overlap with zero, the effect size was considered to be significant at  $P < 0.05$ . According to Wang et al. (2021b), the frequency distribution of response ratios (i.e.,  $\ln RR$ ) for each variable

were plotted by a Gaussian function, which was generally normally distributed (Figure S1).

In this meta-analysis, several categorical variables were used to evaluate their effects on land-use conversion-induced changes in soil CH<sub>4</sub> and N<sub>2</sub>O fluxes as well as non-CO<sub>2</sub> GHG (CH<sub>4</sub>+N<sub>2</sub>O) fluxes, including land-use change type and climate conditions (mean annual temperature (MAT) and precipitation (MAP), Table S1). The subgroup classification of these variables was conducted in terms of aiming for maximal in-group homogenization and exploration of site-specific variation (Jeffery et al., 2011). For MAT and MAP, therefore, subgroups were split at MAT > 13°C and MAP > 1060 mm. In addition, the effects of N application rate, N type and soil properties (e.g., soil texture, pH, BD, SOC, TN and C/N ratio) on the N<sub>2</sub>O EF<sub>d</sub> were also examined (Table S1). Between-group heterogeneity (Q<sub>b</sub>) tests were used to examine the significance of each sub-group category (Table S2). A model selection analysis in the R package “*glmulti*” was used to identify the most important predictors of the response of soil N<sub>2</sub>O emission and CH<sub>4</sub> uptake to urbanization as well as the N<sub>2</sub>O EF<sub>d</sub> of urban soils (Calcagno and de Mazancourt, 2010). The relative importance of a particular predictor was expressed as the sum of Akaike weights for all multiple linear models that included this predictor, and a cutoff of 0.8 was set to differentiate between important and non-essential predictors (Terrer et al., 2016). We also used non-linear or linear regression analysis to analyze the relationships between environmental factors and soil N<sub>2</sub>O emission, CH<sub>4</sub> uptake or N<sub>2</sub>O EF<sub>d</sub> under urbanization. The regression analysis was conducted with SPSS version 23.0 (SPSS China, Beijing, China) and the graphs were drawn with the Origin software (Origin 2018 for Windows).

### 3 RESULTS

#### 3.1 Effects of urbanization-induced land-use conversion on soil non-CO<sub>2</sub> GHG (CH<sub>4</sub> and N<sub>2</sub>O) fluxes

Overall, urbanization-induced land-use conversion (hereafter referred to as urbanization) significantly increased soil N<sub>2</sub>O emissions by 153% (CIs: 107% to 214%). The increase was 102% (CIs: 19% to 339%) if arable land and grazed pasture (hereafter referred to as managed ecosystems) are changed into fertilized lawns, 225%-1259% if natural grassland and forest (hereafter referred to as natural ecosystems) are converted to fertilized lawns, and 126% if natural forest is converted to fertilized urban forests (Figure 1a). In contrast, the increase tended to be relatively lower (ranging from 111% to 266%) if natural ecosystems are converted to non-fertilized urban lawns and forests.

On average, annual soil N<sub>2</sub>O emissions increased from  $1.2 \pm 0.2$  (mean  $\pm$  SE) kg N ha<sup>-1</sup> yr<sup>-1</sup> (median: 0.5 kg N ha<sup>-1</sup> yr<sup>-1</sup>) in non-urban environments to  $3.0 \pm 0.4$  kg N ha<sup>-1</sup> yr<sup>-1</sup> (median: 1.8 kg N ha<sup>-1</sup> yr<sup>-1</sup>) in urban greenspaces (Figure 2a).

The response of soil N<sub>2</sub>O emission to urbanization varied with climate conditions (Figures. 1 and S3). For example, the positive effects of urbanization on soil N<sub>2</sub>O emission tended to be higher at sites with a MAT  $>13$  °C or if the MAP exceeded 1060 mm, but their trends were not statistically significant.

Our analysis shows that urbanization also affects soil properties such as BD (+10.4%), TN (+34.9%), pH (+ 9.7%), SOC (-19.2%), C/N ratio (-21.6%) and clay content (-26.7%) (Figure S2). Regression analysis indicated that the ln RR of soil N<sub>2</sub>O emission was positively correlated with the ln RR of SOC and clay content (Figure S3). These findings were further supported by assessing the performance of multiple linear models using Akaike weights, showing that MAT, clay content and C/N ratio were key environmental factors explaining the variance of soil N<sub>2</sub>O emissions in response to

urbanization (Figure 3a).

On average, urbanization significantly decreased soil CH<sub>4</sub> uptake by 50% (CIs: 45% to 55%) (Figure 1b). However, the effects of urbanization on soil CH<sub>4</sub> uptake depended greatly on the type of land-use change. Urbanization significantly reduced soil CH<sub>4</sub> uptake by 48%-98% for changing natural ecosystems into urban greenspaces regardless of N fertilization, but showed the tendency of increased CH<sub>4</sub> uptake (39%, CIs: -4% to 106%) if managed ecosystems are converted to lawns (Figure 1b). The mean annual CH<sub>4</sub> uptake by soils of urban greenspaces was  $2.0 \pm 0.2 \text{ kg C ha}^{-1} \text{ yr}^{-1}$  (median:  $2.1 \text{ kg C ha}^{-1} \text{ yr}^{-1}$ ), lower than uptake by non-urban soils (mean:  $4.3 \pm 0.3 \text{ kg C ha}^{-1} \text{ yr}^{-1}$ ; median:  $4.4 \text{ kg C ha}^{-1} \text{ yr}^{-1}$ ) (Figure 2b).

Similar to soil N<sub>2</sub>O emission, the response of soil CH<sub>4</sub> uptake to urbanization was affected by soil properties. The ln RR of soil CH<sub>4</sub> uptake was negatively correlated with the ln RR of BD, TN and pH, and there was a quadratic relationship with clay content (Figure S3). Our multiple linear model selection analysis confirmed the importance of land-use type and TN on the response of soil CH<sub>4</sub> uptake to urbanization (Figure 3b).

Across all studies reporting simultaneous measurements of soil N<sub>2</sub>O and CH<sub>4</sub> fluxes, urbanization significantly increased the fluxes of non-CO<sub>2</sub> GHG (CH<sub>4</sub>+N<sub>2</sub>O) by 674 kg CO<sub>2</sub>-eq ha<sup>-1</sup> yr<sup>-1</sup> (CIs: 497 to 842 kg CO<sub>2</sub>-eq ha<sup>-1</sup> yr<sup>-1</sup>) (Figure 1c). Increases for changing natural ecosystems into urban forests and lawns were with 661-1000 kg CO<sub>2</sub>-eq ha<sup>-1</sup> yr<sup>-1</sup>, higher than the effect of converting managed ecosystems to lawns (381 kg CO<sub>2</sub>-eq ha<sup>-1</sup> yr<sup>-1</sup>, CIs: 82 to 767 kg CO<sub>2</sub>-eq ha<sup>-1</sup> yr<sup>-1</sup>).

### 3.2 Direct N<sub>2</sub>O emission factors (EF<sub>d</sub>) for fertilized soils in urban greenspaces

The mean annual N<sub>2</sub>O emission from fertilized urban soils was  $3.5 \pm 0.4 \text{ kg N ha}^{-1} \text{ yr}^{-1}$ , significantly higher than emissions from non-fertilized soils ( $2.0 \pm 0.4 \text{ kg N ha}^{-1} \text{ yr}^{-1}$ ) (Figure 4). Annual N<sub>2</sub>O emissions from urban soils increased linearly with

increasing N fertilization, but this correlation was weak, as the determination coefficient  $R^2$  was only 0.04 (Figure S4).

Overall, the global mean annual  $N_2O$  EF<sub>d</sub> of fertilized urban soils was 1.4% (CIs: 1.1% to 1.8%), but this varied with the type of urban greenspace (Figure 5a). On average, the annual  $N_2O$  EF<sub>d</sub> was 3.7% (CIs: 2.6% to 4.9%) for fertilized urban forests. For fertilized urban lawns, the annual  $N_2O$  EF<sub>d</sub> decreased in the following order: athletic lawn (5.4%) > golf course (2.3%) > residential lawn (1.2%) > institutional lawn (1.0%) > other lawns (public park and ornamental landscape) (0.8%), with an overall mean value of 1.3%. Further analyses showed that the annual  $N_2O$  EF<sub>d</sub> was not affected by climate or type of N fertilization (Figure 5). However, a higher  $N_2O$  EF<sub>d</sub> was calculated for soils in urban greenspaces with  $BD > 1.3 \text{ g cm}^{-3}$  (Figures. 5g and S5).

Our model selection analysis revealed that soil BD and pH were the most important variables explaining variations of the  $N_2O$  EF<sub>d</sub> (Figure 3c).

## 4 DISCUSSION

### 4.1 Effects of urbanization on soil $N_2O$ and $CH_4$ fluxes and its drivers

Urbanization is driving local to global alterations of biogeochemical cycles and these changes also affect soil GHG emissions (Kaye et al., 2006; Grimm et al., 2008). While the effect of urbanization on regional  $CO_2$  fluxes is well documented, showing that urban areas are strong net sources (while terrestrial ecosystems, specifically natural and forested landscapes may function as carbon sinks) (e.g., Seto et al., 2012; Chien and Krumins, 2022), little information is available on effects of urbanization on soil non- $CO_2$  GHG ( $CH_4$  and  $N_2O$ ) fluxes. In this study, our synthesis shows that non- $CO_2$  GHG emissions from soils of urban greenspaces are significantly higher than emissions from non-urban soils. This difference was due to increases in soil  $N_2O$  emission (+153%), while uptake of atmospheric  $CH_4$  by soils was decreased by about 50% in urban

greenspaces (Figures. 1 and 2). Urban land conversion had a positive effect on N<sub>2</sub>O emissions even if the conversion does not involve fertilizer. However, relatively higher rates of soil CH<sub>4</sub> uptake were observed if managed ecosystems were converted to urban lawns (+39%).

Previous studies have shown that natural forest and grassland soils are a major sink for atmospheric CH<sub>4</sub> via microbial consumption by methanotrophic bacteria (e.g., Kirschke et al., 2013; Ni and Groffman, 2018). Generally, farming practices such as tillage, N fertilization or grazing decrease soil methanotrophic populations and activities, and thus, soil CH<sub>4</sub> uptake (e.g., Ding et al., 2004; Chen et al., 2011). Conversion of such agricultural soils into lawns reduces soil disturbance, which may explain why uptake of atmospheric CH<sub>4</sub> was increased by conversion to lawns (van Delden et al., 2018).

In our study, the estimated N<sub>2</sub>O EF<sub>d</sub> for urban forests (3.7%) was higher than the mean EF<sub>d</sub> of 1.43% for global fertilized forest soils (Liu et al., 2017); while the mean N<sub>2</sub>O EF<sub>d</sub> for urban lawns (1.3%) was in the middle of the reported range for fertilized grassland soils (0.30%-2.49%) at regional and global scales (Freibauer and Kutschmitt, 2003; Aguilera et al., 2013; van der Weerden et al., 2016; Liu et al., 2017). The overall mean N<sub>2</sub>O EF<sub>d</sub> (1.4%) for soils in urban greenspaces was higher than the estimated EF<sub>d</sub> of 1.2% for tropical and subtropical agricultural systems (Albanito et al., 2017), or the IPCC default value of 1% for global croplands (IPCC, 2006). All these comparisons indicate that urban soils are regional hotspots for N<sub>2</sub>O emissions.

While there are no reliable published global estimates of the area occupied by urban greenspaces, several studies have suggested that urban ecosystems currently comprise at least 2% of the global terrestrial land area (Kaye et al., 2004; Potere and Schneider, 2007; Hall et al., 2008). Also, multiple studies have shown that for urban greenspaces approximately 50% of lawns and 10% of forests are fertilized (Law et al., 2004; Fraser

et al., 2013; Polsky et al., 2014; Groffman et al., 2016; Locke et al., 2018, 2019). Therefore, we roughly estimate that the total area of individual greenspaces is 149 Mha for fertilized lawns, 30 Mha for fertilized urban forests and 119 Mha for non-fertilized lawns and forests (Table 1). To estimate the global effect of urbanization on soil non-GHG emissions, we multiplied the differences in soil  $N_2O$  emission and  $CH_4$  uptake due to urbanization-induced land-use conversion (both expressed as area-scaled metrics) by the corresponding total greenspace areas. We estimate that conversion from non-urban landscapes to urban greenspaces increases  $N_2O$  emission by 0.46 (CIs: 0.30-0.64) Tg  $N_2O$ -N  $yr^{-1}$  and decreases  $CH_4$  uptake by 0.58 (CIs: 0.38-0.80) Tg  $CH_4$ -C  $yr^{-1}$ , which is equivalent to 6.3% of global anthropogenic  $N_2O$  emissions (i.e., 7.3 Tg  $N_2O$ -N  $yr^{-1}$ ) and 2.6% of global soil  $CH_4$  uptake (i.e., 22.5 Tg  $CH_4$ -C  $yr^{-1}$ ) (IPCC, 2021). Overall, the urban land conversion-induced increase in soil non- $CO_2$  GHG emissions is 218 (CIs: 158-286) Tg  $CO_2$ -eq  $yr^{-1}$ , which is equivalent to 3.5% of global agriculture non- $CO_2$  GHG emissions (i.e., 6.2 Pg  $CO_2$ -eq  $yr^{-1}$ ) (IPCC, 2021).

It is well known that soil  $N_2O$  and  $CH_4$  fluxes are strongly affected by many abiotic, biotic and anthropogenic factors, including climate conditions, soil properties and management practices (e.g., Groffman and Pouyat, 2009; Butterbach-Bahl et al., 2013). With regard to climate conditions, the best-documented feature of urbanization is the urban heat island effect, which elevates air temperatures in urban environments by 1-3 °C compared to rural sites (Santamouris, 2015). At increased soil temperatures, if soil moisture does not become limiting, decomposition of organic matter and nutrient mineralization is stimulated, which contribute to the increase in  $N_2O$  emission from urban soils (Bijoor et al., 2008; Zhan et al., 2021). Likewise, the urban heat island effect may also promote soil methanogenic and methanotrophic activities (i.e., the rates of  $CH_4$  production and oxidation) (Butterbach-Bahl and Papen, 2002; Zhang et al., 2021).

Besides increased temperatures, urban areas also often receive higher precipitation, since higher concentrations of condensation nuclei and surface roughness in urban environments have been found to stimulate precipitation (Gregg et al., 2003; Liu and Niyogi, 2019; Sun et al., 2021). This change in precipitation affects soil moisture, soil aeration and redox potential, and may support the development of soil anaerobiosis which promotes the activity of methanogens and stimulates microbial denitrification processes (Hartmann et al., 2011; van Delden et al., 2018). These changes decrease net uptake of CH<sub>4</sub> and increase N<sub>2</sub>O emissions. Consistent with this reasoning, our results show that N<sub>2</sub>O emissions from urban soils tended to be higher at sites with MAP > 1060 mm, while the response of soil CH<sub>4</sub> uptake tended to decrease with increasing precipitation (Figure 1).

In addition to climate conditions, greater atmospheric N deposition (due to fossil fuel combustion and industrial processes) as well as anthropogenic management (e.g., N fertilization and irrigation) in urban greenspaces can result in increased abundances and activity of ammonia oxidizer and denitrifier populations, and higher soil inorganic N concentrations and soil TN content (as shown in this and previous (Dutt and Tanwar, 2020) studies). These changes in soil nutrient concentrations and the soil microbiome in urban soils support greater rates of nitrification and denitrification and subsequent N<sub>2</sub>O emissions (Rao et al., 2014; Xu et al., 2022; Zhang et al., 2022b). Higher rates of N cycling in urban soils usually leads to a reduction in CH<sub>4</sub> uptake by suppressing methanotroph populations and activities (Kravchenko et al., 2002; Costa and Groffman, 2013; Bodelier, 2011; Tate, 2015; Wu et al., 2022).

Soil properties can also be altered by human disturbances associated with urbanization, such as removal, compaction, burial or construction activities (Byrne, 2007). Our meta-analysis shows that urbanization significantly increases soil BD, TN and pH and

decreases soil C/N ratio, SOC and clay content (Figure S2). The urbanization-induced changes in these soil parameters likely contribute to the observed changes in soil CH<sub>4</sub> and N<sub>2</sub>O fluxes. The higher BD associated with urban soil compaction reduces porosity and gas diffusivity, decreasing the flow of CH<sub>4</sub> from the atmosphere to oxidizing populations. Increases in BD can increase N<sub>2</sub>O production by increasing the fraction of anaerobic volume in soils (De Neve and Hofman, 2000; Byrne, 2007; Pulido-Moncada et al., 2022). This effect is supported by our regression analysis showing that the ln RR of soil CH<sub>4</sub> uptake was negatively correlated with the ln RR of BD (Figure S3). The higher TN and lower C/N ratio in urban soils likely support higher rates of N transformation (particularly mineralization and nitrification) and associated N<sub>2</sub>O production, while hampering methanotrophic activities and CH<sub>4</sub> uptake (Zhu and Carreiro, 2004; Zhang et al., 2021). This is supported by the negative correlation between the ln RR of CH<sub>4</sub> uptake and the ln RR of soil TN (Figure S3). In addition, the alteration of soil pH associated with urbanization may affect soil N<sub>2</sub>O and CH<sub>4</sub> fluxes. For example, increasing soil pH has been observed to accelerate soil N<sub>2</sub>O emissions, due to stimulation of nitrification and N<sub>2</sub>O production by nitrifiers (Baggs et al., 2010; Wang et al., 2021b). On the other hand, increases in soil pH may negatively affect CH<sub>4</sub> uptake or even result in net emissions of CH<sub>4</sub> from arable and forest soils due to effects on soil inorganic N and labile C availability (Butterbach-Bahl and Papen, 2002; Wang et al., 2021b). However, our study also shows that urbanization is associated with decreases in SOC and soil clay content. This is most likely due to the addition of materials from a variety of sources (e.g., the presence of dust and materials from the buildings, external soils and other materials with low differentiation of clay), the change of arrangement or sequence of horizons in urban soil profile during the urbanization processes, or the lower coefficient of clay formation values in the urban soils

(Doichinova et al., 2006; Sun et al., 2013; Herrmann et al., 2018). This indicates that the availability of C substrates and the tendency of soils to become anaerobic is lower in urban soils as compared to soils in rural areas. These changes would tend to decrease soil N<sub>2</sub>O emissions and increase CH<sub>4</sub> uptake (Li et al., 2005; Yao et al., 2019). The net effect of the interacting soil factors controlling N<sub>2</sub>O and CH<sub>4</sub> fluxes appears to be changes in environmental conditions and soil physicochemical properties that stimulate urban soil N<sub>2</sub>O emissions, while reducing soil CH<sub>4</sub> uptake in urban environments (Figure 6).

#### 4.2 Uncertainties and implications for future studies

There are several important uncertainties in our analysis. First, some bias may arise from uneven distribution of experimental sites, with a dominance of studies in North America and other areas in the temperate zone, and an unbalanced selection of ecosystem types. Specifically, there were no studies available for South America and Africa. Given the rapid urban sprawl in all parts of the world, particularly in emerging countries like China and India, more studies in these regions are needed to better understand the importance of urbanization on soil non-CO<sub>2</sub> GHG fluxes in a global context. Second, too few studies had multiple years of data. The majority of studies in our database lasted for 1 or 2 years, and the few observations available reporting on flux measurements of >2 years showed high interannual variations (Ni and Groffman, 2018; van Delden et al., 2018). There is a clear need for future field measurements spanning multiple years, particularly for underrepresented regions or climate zones (e.g., in tropical regions of South America) in order to more accurately quantify changes in non-CO<sub>2</sub> GHG fluxes under urbanization. Lastly, the complex interactions among various biotic and abiotic factors (i.e., climate conditions, soil properties and anthropogenic management activities) involved in the consumption, production and net

flux of CH<sub>4</sub> and N<sub>2</sub>O from urban soils results in unavoidable uncertainties. Future studies should focus on processes of CH<sub>4</sub> and N<sub>2</sub>O formation and consumption in urban soils to allow for comparative evaluation of biogeochemical changes associated with urban land use change.

Given ongoing trends in urbanization, the importance of urban soils as sources of non-CO<sub>2</sub> GHG at regional scales will increase. This highlights the urgent need for accurate mapping of dynamic land use change due to urbanization and to incorporate urbanization effects on soil GHG fluxes into Earth system models to allow better evaluation of the relationships between urbanization and climate. These assessments will also help to identify effective and efficient mitigation and adaptation strategies (e.g., avoiding urban soil compaction, reducing N input in urban environments and improving urban afforestation) to achieve sustainable and resilient urban development.

## 5 CONCLUSIONS

Our meta-analysis provides a first comprehensive assessment of the response of soil N<sub>2</sub>O emission and CH<sub>4</sub> uptake to urbanization. Overall, our results show that urbanization can accelerate global climate change by increasing soil N<sub>2</sub>O emissions while reducing rates of CH<sub>4</sub> uptake in soils in urban greenspaces. The global mean annual N<sub>2</sub>O EF<sub>a</sub> for urban forests and lawns are at least 50% higher than the current IPCC default value, highlighting that urban soils can be significant hotspots for regional N<sub>2</sub>O emissions. With a rapid proliferation of urban land use and cover worldwide, greater non-CO<sub>2</sub> GHG emissions from urban soils can be anticipated. However, the magnitude of these urbanization effects is largely dependent on environmental conditions, soil properties and land-use type. Overall, our study may help to better understand effects of urbanization on soil GHG fluxes as an underestimated driver of climate change, to stimulate more research in this field, and to consider urban soil

management as an overlooked tool for sustainable development.

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## CONFLICT OF INTEREST

The authors declare no conflict of interest.

## DATA AVAILABILITY STATEMENT

The data that supports the findings of this study are publicly available at Dryad via <https://doi.org/10.5061/dryad.v9s4mw714>.

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**TABLE 1** Global-scale estimates of urbanization-induced changes in soil nitrous oxide ( $\Delta N_2O$ ) emission and methane ( $\Delta CH_4$ ) uptake as well as the non-carbon dioxide ( $CO_2$ ) greenhouse ( $\Delta non-CO_2$  GHG) emission on basis of the total area of fertilized and non-fertilized urban greenspaces.

Greenspace type	Area <sup>a</sup> (Million ha)	The difference in soil $N_2O$ and $CH_4$ fluxes due to urbanization-induced land-use conversion <sup>b</sup>		Urbanization-induced changes at global scale <sup>c</sup>		
		Soil $N_2O$ emission (kg N ha <sup>-1</sup> yr <sup>-1</sup> )	Soil $CH_4$ uptake (kg C ha <sup>-1</sup> yr <sup>-1</sup> )	$\Delta N_2O$ emission (Tg $N_2O$ -N yr <sup>-1</sup> )	$\Delta CH_4$ uptake (Tg $CH_4$ -C yr <sup>-1</sup> )	$\Delta non-CO_2$ GHG (Tg $CO_2$ -eq yr <sup>-1</sup> )
Fertilized lawn	149	1.0 (0.5~1.6)	-2.0 (-3.2~-0.9)	0.15 (0.07~0.24)	-0.30 (-0.48~-0.13)	75 (48~108)
Fertilized urban forest	30	5.4 (3.9~6.5)	-0.6	0.16 (0.12~0.20)	-0.02	69 (52~87)
Non-fertilized soil	119	1.3 (0.9~1.7)	-2.2 (-2.5~-1.9)	0.15 (0.11~0.20)	-0.26 (-0.30~-0.23)	74 (58~94)
Total	298			0.46 (0.30~0.64)	-0.58 (-0.80~-0.38)	218 (158~286)

<sup>a</sup> The area of each urban greenspace type was calculated by combining information on the urban greenspace proportion of global land area (i.e., ca. 2% of global land area, Kaye et al., 2004; Potere and Schneider, 2007; Hall et al., 2008) with information on the distribution of fertilized and non-fertilized urban lawns and forests (i.e., ca. 50% of fertilized lawns, 10% of fertilized urban forests and 40% of non-fertilized soils, Law et al., 2004; Fraser et al., 2013; Polsky et al., 2014; Groffman et al., 2016; Locke et al., 2018, 2019).

<sup>b</sup> The difference in soil  $N_2O$  emission and  $CH_4$  uptake due to urban land conversions was calculated on basis of our dataset S1.

<sup>c</sup>  $\Delta non-CO_2$  GHG emission was estimated by  $\Delta N_2O + \Delta CH_4$  and applying IPCC global warming potential factors for  $CH_4$  (28) and  $N_2O$  (273), respectively, over a 100-year time horizon (IPCC, 2021).

Negative sign (-) indicates a reduction of soil  $CH_4$  uptake due to urbanization. Numbers in brackets represent 95% confidence intervals.

**FIGURE CAPTIONS:**

**FIGURE 1** Response of soil nitrous oxide ( $N_2O$ ) emission (a), methane ( $CH_4$ ) uptake (b) and non-carbon dioxide ( $CO_2$ ) greenhouse gas (GHG, i.e.,  $CH_4+N_2O$ ) fluxes (c) to urbanization-induced land-use conversion. Managed ecosystem refers to arable land and grazed pasture. The data in brackets indicates the number of paired observations in each sub-group. Error bars represent 95% confidence intervals (CIs). Variables were significant at  $P < 0.05$ , if 95% CIs did not overlap with zero. MAT = mean annual temperature, MAP = mean annual precipitation.

**FIGURE 2** Box-plots of annual nitrous oxide ( $N_2O$ ) emission (a) and methane ( $CH_4$ ) uptake (b) as well as non-carbon dioxide ( $CO_2$ ) greenhouse gas (GHG, i.e.,  $CH_4+N_2O$ ) fluxes (c) from urban greenspaces and non-urban soils. Solid and dashed lines inside the boxes represent medians and means, respectively. Box boundaries represent 75th and 25th percentiles, whisker caps represent 95th and 5th percentiles, and circles represent data points. Curves represent the normal distribution of data. Different letters (a or b) indicate significant differences between urban and non-urban soils ( $P < 0.05$ ).

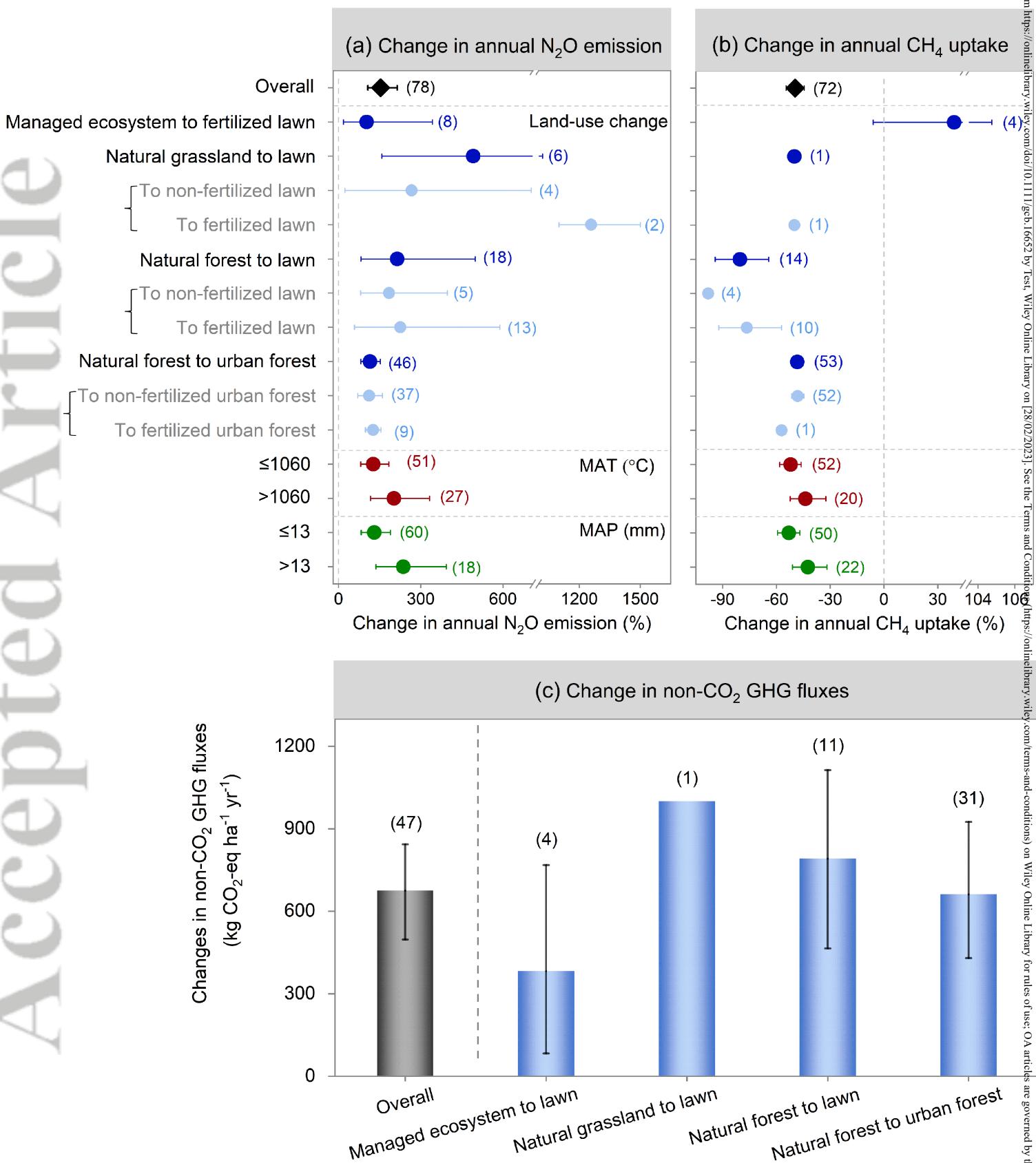
**FIGURE 3** Model-averaged importance of predictors of the response of soil nitrous oxide ( $N_2O$ ) emission (a) and methane ( $CH_4$ ) uptake (b) to urbanization-induced land-use conversion, as well as the annual direct  $N_2O$  emission factor ( $EF_d$ ) for soils of urban greenspaces (c). The relative importance is based on the sum of Akaike weights derived from the model selection using corrected Akaike's Information Criteria. The cutoff set at 0.8 (dashed line) differentiates between essential ( $> 0.8$ ) and non-essential ( $< 0.8$ )

model predictors. MAT = mean annual temperature, MAP = mean annual precipitation, BD = bulk density, SOC = soil organic C, TN = total N.

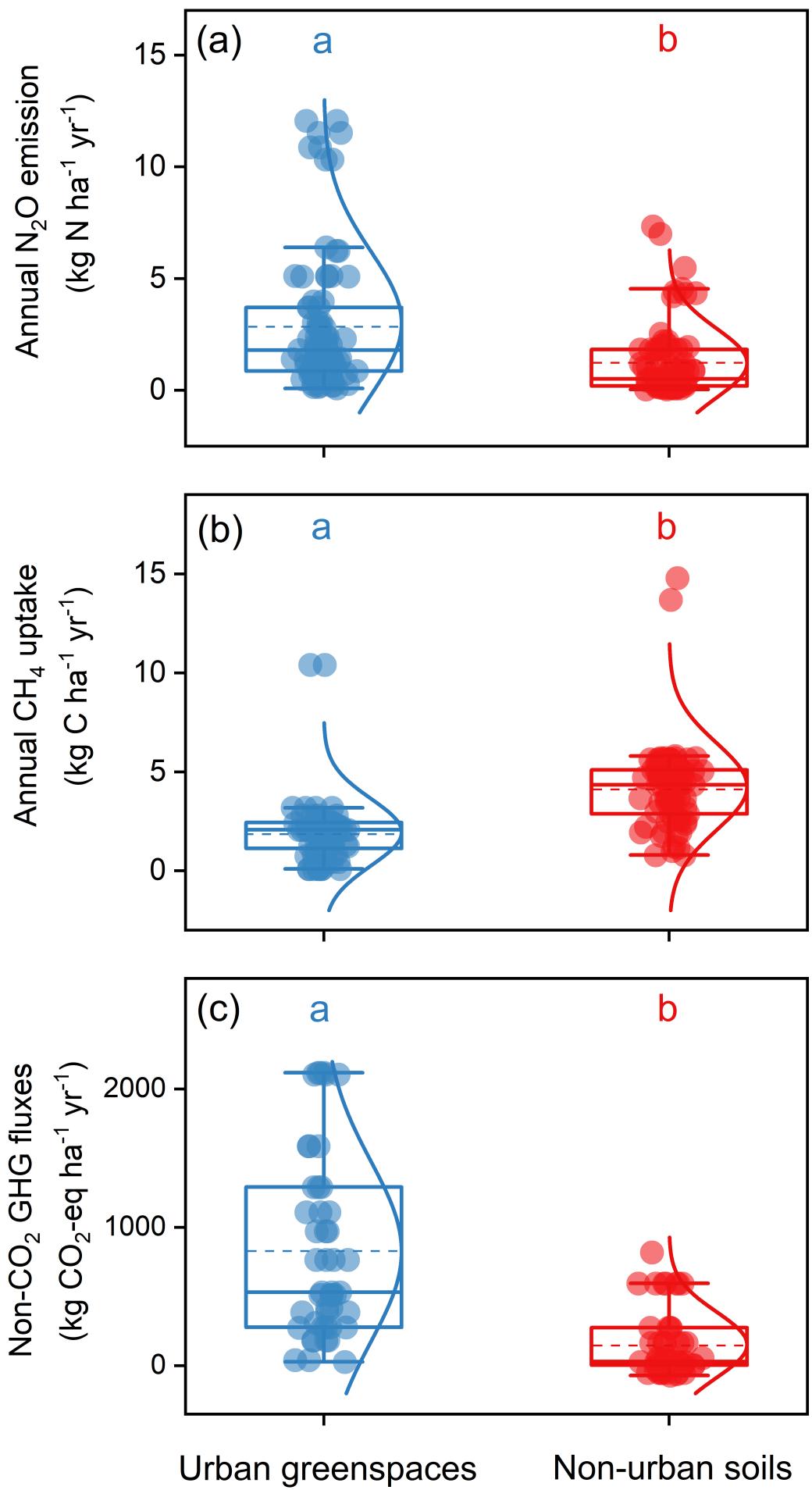
**FIGURE 4** Box-plots of annual nitrous oxide ( $N_2O$ ) emissions from fertilized urban soils and non-fertilized controls. Solid and dashed lines inside the boxes represent medians and means, respectively. Box boundaries represent 75th and 25th percentiles, whisker caps represent 95th and 5th percentiles, and circles represent data points. Curves represent the normal distribution of data. Different letters (a or b) indicate significant differences between fertilized and non-fertilized soils ( $P < 0.05$ ).

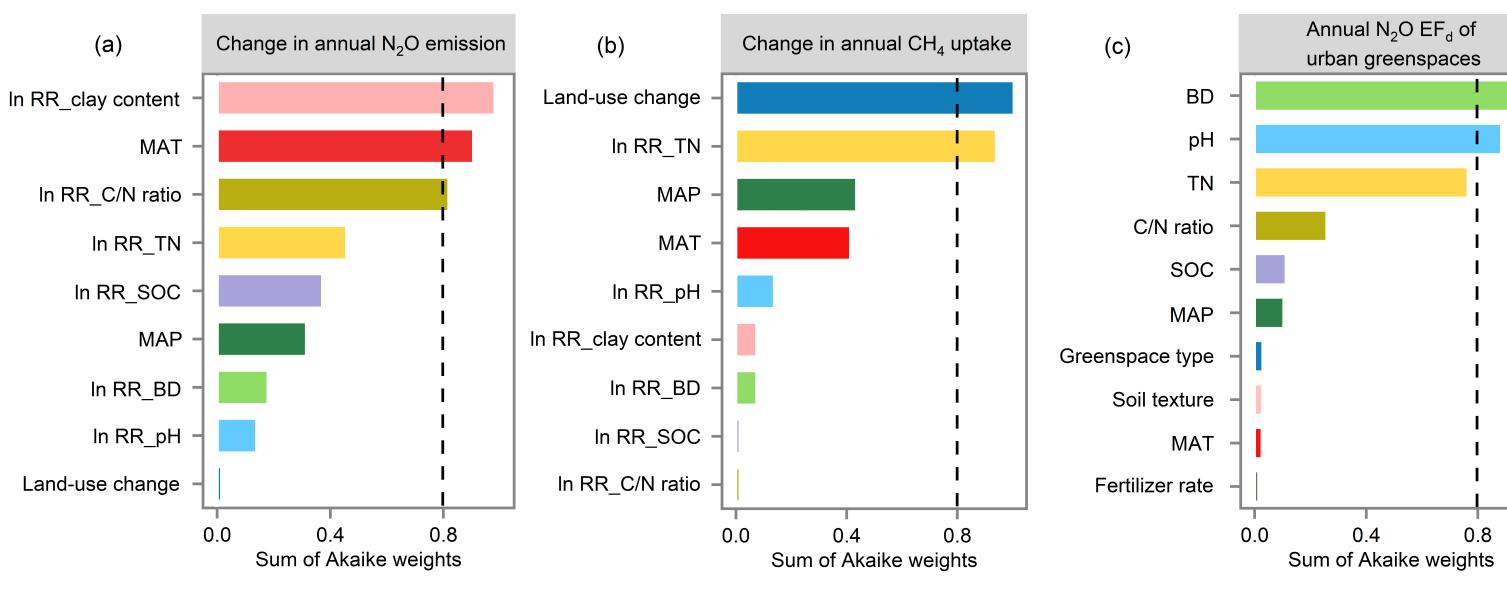
**FIGURE 5** Effects of type of greenspace (a), mean annual temperature (MAT) (b), mean annual precipitation (MAP) (c), N fertilization rate (d), N fertilizer type (e), soil texture (f), bulk density (BD) (g), soil pH (h), soil organic C (SOC) (i), total N (TN) (j) and C/N ratio (k) on annual direct emission factors ( $EF_d$ ) of nitrous oxide ( $N_2O$ ) in urban soils. The values in brackets indicate the number of observations in each subgroup. Error bars represent 95% confidence intervals. “Other lawns” refer to public park and ornamental landscapes. “Other synthetic” refers to synthetic N fertilizer other than urea and slow-release fertilizer, e.g., calcium nitrate and ammonium nitrate. Acid, neutral and alkaline mean  $pH \leq 6.5$ ,  $6.6 \leq pH < 7.3$  and  $pH > 7.3$ , respectively.

**FIGURE 6** Conceptual diagram illustrating urbanization effects on soil nitrous oxide ( $N_2O$ ) emission and methane ( $CH_4$ ) uptake as well as the associated soil processes. Non- $CO_2$  GHG emissions = non-carbon dioxide ( $CO_2$ ) greenhouse gases (GHG, i.e.,  $CH_4+N_2O$ ).  $EF_d$  = direct  $N_2O$  emission factor.

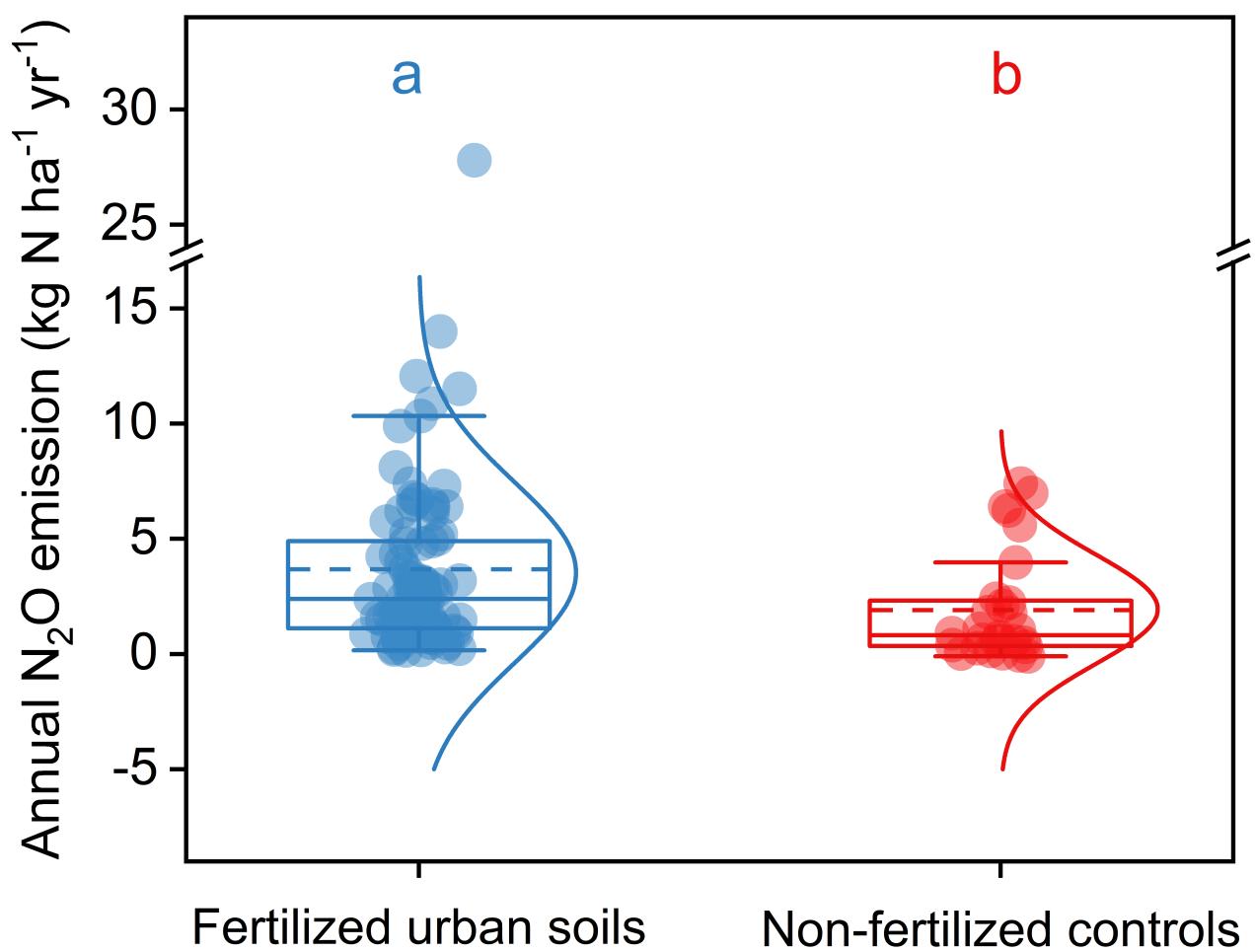


GCB\_16652 FIGURE1.png

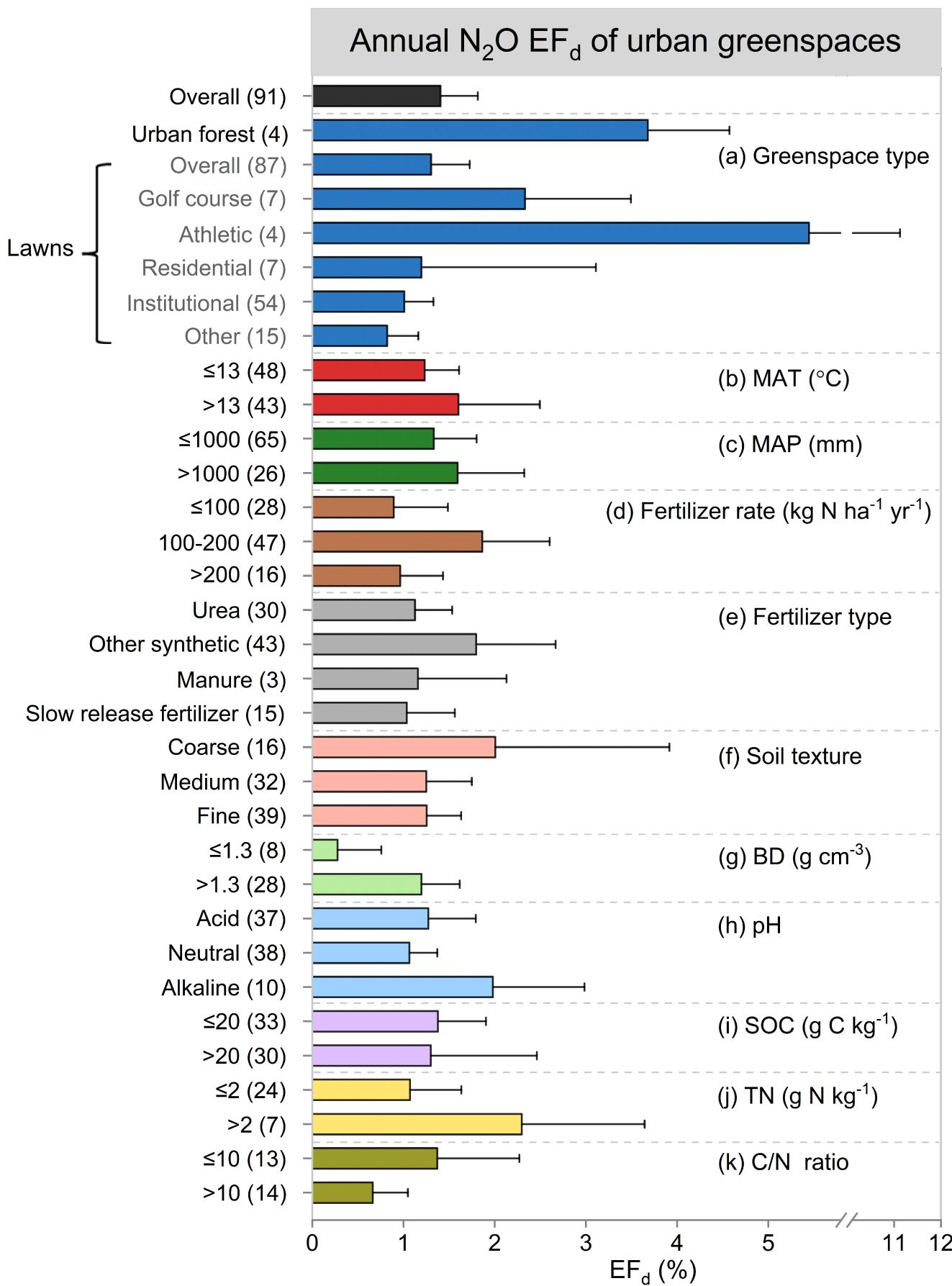


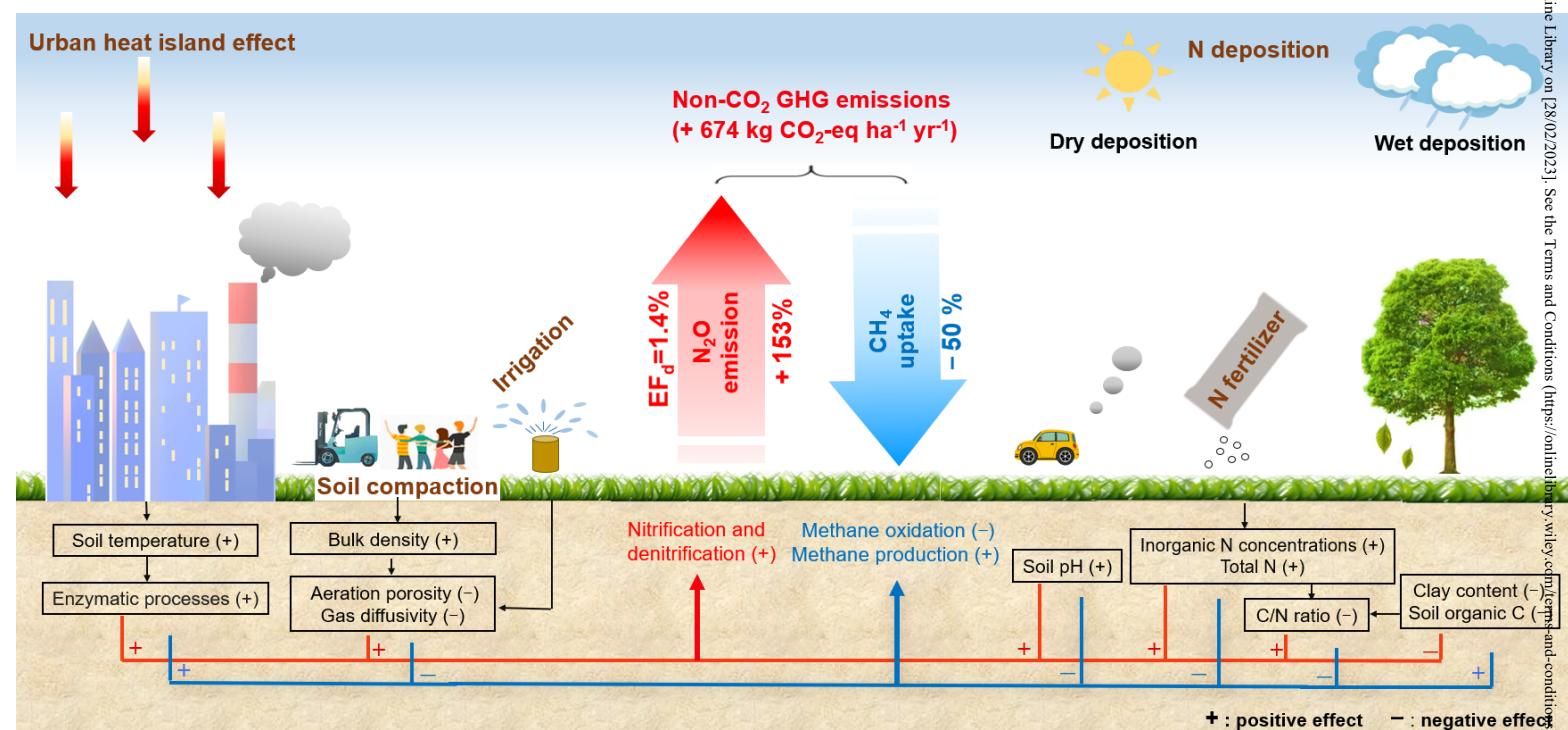


GCB\_16652 FIGURE3.png



GCB\_16652 FIGURE4.png





GCB\_16652 FIGURE6.png