

1 **Impact assessment of coastal marine range shifts to support proactive management**

2 Running head: Impact assessment of marine range shifts

5 Amy K. Henry¹ and Cascade J. B. Sorte¹

6 ¹Department of Ecology and Evolutionary Biology, University of California, Irvine, CA, 92697-

7 2525 USA

10 Corresponding author: Cascade J. B. Sorte

11 phone: 949-824-6971

12 email: csorte@uci.edu

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20 **Abstract**

21 Climate change is reshuffling Earth's biota as species ranges shift to track increasing habitat
22 temperatures. While redistribution may be necessary for species persistence, there can also be
23 impacts on existing communities upon arrival of novel, range-shifting species. Anticipating the
24 beneficial versus deleterious impacts of range-shifting species is essential for determining
25 whether active management is needed, which could include employing strategies from
26 facilitation (eg managed relocation) to suppression (eg prevention/control). We employ an
27 impact assessment protocol developed for invasive species to evaluate potential consequences
28 of range shifts in coastal marine ecosystems of North America. Our review demonstrates how
29 invasion impact assessment combined with species vulnerability assessment could support
30 decisions about management of range shifts. We found that ~50% of these shifting coastal
31 species have had negative impacts in their expanded range. The importance of proactive
32 management is likely to increase as the number and extent of range shifts accelerates.

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35 **In a nutshell:**

36 • Novel species can arrive in locations due to introduction by humans or shifts in species'
37 native ranges as habitat temperatures increase.

38 • There is no "one size fits all" approach to managing novel species, particularly for native
39 range shifts: range shifts can be necessary for native species to cope with climate change, and
40 our results show that in some cases, they lead to relatively peaceful coexistence, whereas in
41 other cases, range-shifting species may disrupt communities.

42 • Our review suggests that impact assessments developed for invasive species can be
43 used to anticipate the consequences (both negative and positive) of native species range shifts.

44 **Introduction**

45 Climate change is causing shifts in species' ranges in taxa and systems worldwide (Sorte *et al.*
46 2010, Pecl *et al.* 2017, Lenoir *et al.* 2020). As conditions become more hospitable at cold-
47 temperature range boundaries, species can expand into areas spatially contiguous with
48 documented native ranges. Importantly, range shifts occur piecemeal, with changes in species
49 distributions reshuffling ecosystems. Understanding the impacts of native species range shifts is
50 critical for anticipating effects of climate change and designing management plans for
51 preserving biodiversity. Here, we focus on the impacts of species additions to communities at
52 expanding range limits. We employed an impact assessment approach developed for invasive
53 species to evaluate the consequences of native species range shifts (ie expansion of range
54 limits) and, by so doing, explore the potential for this approach to inform management decisions.

55 Range shifts are both a response and driver of global change impacts: while
56 redistribution is necessary for some species to persist, range-shifting species can have impacts
57 in their expanded ranges that threaten current inhabitants. When climatic conditions become
58 unsuitable, species unable to adapt in place must move or perish (Berg *et al.* 2010).

59 Redistribution is, therefore, crucial for preventing extinctions. Across the Western Hemisphere,
60 Lawler *et al.* (2009) predicted 10-20% (and up to 90%) species turnover in the next century
61 based on analysis of 2,954 bird, mammal, and amphibian species. For species included in
62 climate change vulnerability analyses (primarily terrestrial animals and plants), the proportion
63 doomed to extinction approximately doubles in models without redistribution (Thomas *et al.*
64 2004, Urban 2015). Many species likely cannot shift their ranges fast enough to keep pace with
65 climate change (Urban 2015), begging the question of whether humans should intervene on
66 their behalf via deliberate, managed relocation (McLachlan *et al.* 2006, Hoegh-Guldberg *et al.*
67 2008). A main concern about managed relocation is that deliberate range shifts, as with those
68 occurring naturally, can lead to unintended consequences.

69 The consequences of both range shifts and deliberate translocation of native species

70 can include negative impacts on communities in the expanded range (ie decreases in native
71 populations or human activities), even to the point of causing extinctions. For example, in
72 western North America, the barred owl *Strix varia* has displaced a threatened species, the
73 northern spotted owl *Strix occidentalis caurina* (Kelly *et al.* 2003, Long & Wolfe 2019). Barred
74 owl removals are underway, and although eradication from its expanded range is not feasible,
75 suppression of barred owl populations may be necessary to prevent extinction of the spotted
76 owl (Long & Wolfe 2019). Similarly, management (via tree removal to slow spread) of the
77 shifting southern pine beetle *Dendroctonus frontalis* appears necessary for preventing
78 extirpation of pitch pines *Pinus rigida* in the Eastern US (Heuss *et al.* 2019). Not all range shift
79 impacts are negative: shifting species can play beneficial roles in their expanded ranges (ie
80 increase native populations or human activities), particularly when they increase resources. As
81 an example, the tropical seaweed *Turbinaria ornata* provides habitat and enhances food supply
82 for herbivorous fish in the South Pacific (Bittick *et al.* 2019), and it has chemical properties that
83 make it potentially useful in pharmaceuticals (Ananthi *et al.* 2010). However, this seaweed also
84 negatively impacts corals in some areas (Brown & Carpenter 2015), highlighting the importance
85 of considering both positive and negative interactions for evaluating management options for
86 range shifts. Given that climate change is accelerating (Cheng *et al.* 2019), there will likely be
87 an increase in the number and rate of range shifts and the immediacy of considering whether
88 range shifts should be actively managed. An essential step in evaluating management
89 scenarios is range shift impact assessment (McLachlan *et al.* 2006, Hoegh-Guldberg *et al.*
90 2008).

91 Impact assessment protocols developed for invasive species could be effective tools for
92 anticipating outcomes of native species range shifts. These assessments involve compiling
93 published data or expert opinions to categorize species based on likely impacts. The
94 transferability of impact assessments depends on whether a species' impacts in one (ie
95 previously studied) location is representative of its impacts elsewhere (ie in areas where it has

96 not yet arrived or been studied). For invasions, the impacts of a particular invasive species tend
97 to be consistently positive or negative; however, the magnitude of these impacts can vary
98 across locations with different environments and community members (Kulhanek *et al.* 2011,
99 Kumschick *et al.* 2015). These caveats of using existing data from one location/community to
100 anticipate impacts in a different location/community likely also apply to range-shift impact
101 assessments. Impact assessments are, thus, best used to “flag species with high potential
102 impacts” (Blackburn *et al.* 2014), and identify those that might pose high risk within groups of
103 spreading species (eg Rockwell-Postel *et al.* 2020).

104 The goals and characteristics of impact assessments are the same whether novel
105 species are invasive or native in adjacent locations: to evaluate novel species’ impacts in a way
106 that is comparable across taxa and locations and transparently incorporates the best available
107 data with moderate effort (Blackburn *et al.* 2014, Hawkins *et al.* 2015, Eisenmenger *et al.* 2016,
108 Turbé *et al.* 2017). At least 29 protocols exist for invasive species impact assessment, some of
109 which identify maximum potential impacts while others predict likelihood of introduction and
110 spread (Roy *et al.* 2017). Invasion impact assessments have yielded both species-specific
111 information (eg prioritizing weed species for management in the northeastern US; Rockwell-
112 Postel *et al.* 2020) as well as an understanding of the most impactful taxa and impact
113 mechanisms (eg greater impacts of invasive mammals than birds in Europe, particularly via
114 feeding habits; Kumschick *et al.* 2011). However, it is unclear whether impact assessments
115 developed for invasive species will be useful for anticipating impacts of range shifts, particularly
116 if data availability is low for native species that have not been seen as problematic or targeted
117 for study.

118 The objective of this review is to assess the potential impacts of range-shifting native
119 species on populations of interacting species in the expanded range. We applied an impact
120 assessment modified from the Environmental Impact Classification of Alien Taxa (EICAT;
121 Blackburn *et al.* 2014, Hawkins *et al.* 2015). The EICAT protocol was chosen because it was

122 recently adopted for use by the IUCN (International Union for the Conservation of Nature), the
123 body that manages the Global Invasive Species Database (www.iucngisd.org/) and Red List of
124 Threatened Species (www.iucnredlist.org/). We also applied a modified version of the Socio-
125 economic Impact Classification of Alien Taxa (SEICAT) protocol (Bacher *et al.* 2017), which
126 uses the same approach and yields scores on the same scale as EICAT.

127 We used EICAT and SEICAT to evaluate both detrimental and beneficial impacts of
128 range shifts in coastal marine ecosystems of North America (Table 1). We collated impacts data
129 from both expanded and native ranges of shifting species, defining impacts based on the
130 relationship between a range shifter's presence/abundance and the robustness of an interacting
131 species' population or human activity. In addition to evaluating the sign and magnitude of these
132 impacts, we tested the hypothesis that impacts increase outside of species' native ranges (as
133 shown for many invasive species; Cure *et al.* 2012). Our study is the first to demonstrate the
134 effectiveness of this impact assessment approach as a tool for evaluating outcomes of native
135 species range shifts that could be incorporated into management plans.

136

137 **Methods**

138

139 **Identification of study species**

140 We identified 39 marine species whose poleward range limits were documented as shifting
141 northward along the coastline (<15 km from shore) of North America, including plants,
142 invertebrates, fish, a protist, and a bird (WebTable 1). Of these, 26 species were compiled by
143 Sorte *et al.* (2010), and we added 13 species from an updated literature review. We searched
144 Google Scholar (on 08/20/2019) using this search string: marine "range expansion" species
145 "range shift". We reviewed titles and, when appropriate, text of the first 600 results, identifying
146 11 additional species from 14 papers (WebTable 1). We added two species from our literature
147 files (WebTable 1).

148

149 **Review of published impacts**

150 Evidence of species' impacts was compiled from online database searches and literature
151 review. We conducted individual Web of Science searches for the 39 shifting species using
152 each species' scientific name (and synonyms). Papers reporting species impacts were identified
153 by reviewing titles and abstracts. For species with >800 Web of Science results, the first 400
154 results were reviewed and remaining results were filtered using this search string: "ecology" OR
155 "invas*" OR "impact". For species with <100 Web of Science results, we also performed Google
156 Scholar searches, and relevant papers were identified from the first 400 results. Additional
157 impact studies were added opportunistically from citations within papers found in database
158 searches. In total, we reviewed 11,508 papers for this impact assessment of 39 range-shifting
159 species.

160

161 **Impact assessment**

162 We evaluated environmental and socioeconomic impacts using modified versions of the
163 Environmental Impact Classification of Alien Taxa (EICAT; Hawkins *et al.* 2015) and Socio-
164 economic Impact Classification of Alien Taxa (SEICAT; Bacher *et al.* 2017) protocols. The
165 EICAT and SEICAT protocols focus on impacts on native, non-human populations and human
166 activities, respectively. Primary modifications were the inclusion of beneficial (rather than only
167 detrimental) impacts and use of studies in species' native and expanded ranges to estimate
168 impacts (rather than only non-native ranges). These modifications were intended to minimize
169 the influence of study/publication bias, although we acknowledge that researchers historically
170 focused on negative over positive interactions (Bertness & Callaway 1994) and are more likely
171 to study/publish results of strong over weak interactions (Gurevitch & Hedges 1999).

172 Impacts were classified by mechanism. We identified the following mechanisms as
173 responsible for negative impacts by shifting species on native (non-human) species:

174 competition, predation, herbivory, disease transmission, interaction with other invaders, physical
175 disturbance, poisoning/toxicity, and “other” negative impacts (including those with unknown
176 mechanisms). We also found evidence of positive ecological impacts by the following
177 mechanisms: food provisioning, habitat provisioning, and “other” positive impacts. Our SEICAT
178 analysis revealed socioeconomic impacts associated with alterations in health; material and
179 immaterial assets; and social, spiritual, or cultural relations.

180 We assigned levels of impacts based on categories described in the EICAT and SEICAT
181 protocols (Hawkins *et al.* 2015, Bacher *et al.* 2017). Impacts range across a semi-quantitative
182 gradient from 1 (lowest) to 5 (highest). For each published study, we scored impacts of shifting
183 species based on the highest level response from the categories shown in Table 1. Impact
184 scores, thus, represent the maximum impact that has been observed. Both EICAT and SEICAT
185 protocols were modified to incorporate positive impacts, essentially switching the direction or
186 sign of negative impacts (Table 1). Species for which we found no published papers on impacts
187 were categorized as “data deficient”.

188 For both EICAT and SEICAT assessments, we collected additional information about the
189 shifting species and study. These characteristics included taxonomic classifications, study
190 location, and whether the study was conducted in the shifting species’ “native” or non-native,
191 “expanded” range. Ranges were defined as “native” or “expanded” based primarily on
192 documentation within the source reporting the range shift (WebTable 1). “Expanded” ranges
193 were designated as such conservatively, acknowledging potential lack of benchmark data for
194 species ranges, with most range shifts documented after 1985 (Sorte *et al.* 2010). We evaluated
195 the relationship between average EICAT impact levels in the native versus expanded range for
196 the 7 species that were studied in both range types. This analysis was performed using a linear
197 mixed effect model (lmer; lmerTest R package, Kunetsova 2017) in the statistical computing
198 language R (version 4.0.2, R Core Team 2020) with range (native or expanded) as a fixed factor
199 and species as a random effect. Visual inspection of Pearson residuals indicated no deviation

200 from linearity or normality and no major outliers. There was also no deviation from
201 homoscedasticity (Levene's test, $F = 0.0002$, $p > 0.05$). We validated the fit of this model against
202 a model without random effects using AICc (nlme R package, Pinheiro *et al.* 2020).

203

204 **Results & Discussion**

205 The effectiveness of this assessment approach for anticipating impacts of range shifts depends
206 partly on data availability, and we found that environmental impact was studied for a similar
207 proportion of these 39 range shifters as for invasive species. Environmental impacts were
208 documented for 32 (82%) of the 39 shifting species while 7 (18%) of the species were data
209 deficient (WebTable 2). In reviews of invasive species impacts, the proportion of species that
210 were data deficient ranged from 4% (2 of 50 alien mammal and bird species in Europe;
211 Kumschick *et al.* 2011) and 18% (18 of 100 invasive plant species in the northeastern US;
212 Rockwell-Postel *et al.* 2020) in regional studies to 71% (296 of 415 bird species; Evans *et al.*
213 2016) for global studies of all invasive species within a taxonomic group. Our results suggest
214 that data availability does not preclude using assessments such as EICAT to anticipate impacts
215 of range shifts.

216 Fewer data were available to assess impacts of range shifts on socioeconomic systems,
217 for which 72% of species were unstudied and categorized as data deficient (WebTable 2). In
218 comparison, only 26% (78 of 300) of invasive species in Europe (comprising mammals, birds,
219 fish, insects, and plants) were data deficient for socioeconomic impacts (Kumschick *et al.* 2015).
220 Kumschick *et al.* (2015) showed that environmental and socioeconomic impacts were highly
221 correlated, both within and across taxonomic groups. Although this relationship supports use of
222 an impact assessment approach, more studies of range shift impacts on human systems are
223 needed.

224 In total, environmental and/or socioeconomic impacts were documented for 34 (87%) of
225 39 species. Our assessments were based on 184 papers, 154 papers reporting environmental

226 impacts and 30 papers about socioeconomic impacts (~6 and ~3 papers per studied species,
227 respectively) (Figure 1).

228 Published impacts of range shifts were more often negative than positive, although half
229 (51%) of species were documented as having both beneficial and detrimental impacts across
230 environmental and socioeconomic systems (S/EICAT score of 2+). Overall, only negative
231 impacts were reported for 26% of species and only positive impacts were reported for 10% of
232 species (Figures 2, 3, WebTable 2). Environmental impacts on interacting species were
233 observed for 51% of range-shifting species. Of these, 30% were documented as having
234 primarily positive impacts, including habitat-forming seagrass and coral, and fishery species of
235 crab. Half (50%) of the shifting species had documented impacts that were primarily negative,
236 most due to consumption (herbivory or predation) of native species by range-shifting gastropods
237 and fish, as well as two shifters acting as competitors and one disease-causing protistan
238 parasite. The remaining 20% of species had recorded impacts that were both negative and
239 positive. These species (including sponge, coral, mangrove plant, crab, and fish species)
240 provided food or habitat while also negatively impacting native species via consumption,
241 competition, or physical disturbance.

242 The maximum impacts reported for these shifting species were Major impacts (score of
243 4), meaning a native species was lost or gained in a community because of the range shift, but
244 not permanently so. In most cases, including 7 of 9 studies and 4 of 6 species, Major impacts
245 were related to habitat availability. Creation of habitat was often beneficial but could also be
246 detrimental. For example, habitat created by shifting mangrove species supported a native
247 parrotfish and increased commercial fishery yields (Mumby *et al.* 2004) but also altered
248 community structure and increased invasive species (Demopoulos & Smith 2010). Two shifting
249 species had Major impacts via consumptive effects: the predatory sea slug *Phidiana hiltoni*
250 decreased native sea slugs (Goddard *et al.* 2011), and *Lottia picta*, an herbivorous gastropod
251 (limpet), was associated with catastrophic declines in seagrass meadows (Zimmerman *et al.*

252 1996). We did not find evidence of global extinctions (Massive impacts; score of 5) caused by
253 coastal marine range shifts in North America. This is perhaps not surprising given impacted
254 species are generally characterized by high fecundity, little to no parental care, and broad
255 dispersal capacity, which could allow population replenishment from few surviving individuals
256 (McCauley *et al.* 2015, Le Pape *et al.* 2017). Since redistribution is likely necessary for global
257 persistence of many species, range shifts may contribute more to biodiversity preservation than
258 to biodiversity loss.

259 Most impact studies synthesized here were conducted in shifting species' native ranges,
260 including 69% of studies on environmental impacts and 96% of studies on socioeconomic
261 impacts. If impact assessment for range shifts only included impacts measured in species' non-
262 native ranges, as for invasion impact assessments underway, then the number of data deficient
263 species would increase from 13% to 63%. This pattern highlights the need for more studies in
264 expanded ranges of shifting species as well as the importance of understanding whether
265 impacts in the native range are indicative of impacts in the expanded range of shifting species.

266 We compared impacts between native and expanded ranges for 7 species that were
267 studied in both (Figure 4, WebFigure 1). We found that species with stronger negative impacts
268 documented in the native range were also shown to be more detrimental when shifting into new
269 communities. However, impacts in expanded ranges tended to be more negative than impacts
270 in native ranges, with impact increasing by more than one level between the native and
271 expanded range (fixed effect estimate 1.78 [95% CI: 0.89, 2.59]). For 6 of 7 species, impacts
272 were more negative in the expanded than native range, while mean impacts were the same in
273 both ranges for 1 species (Figure 4). Species' impacts were never documented as more positive
274 after range shifts (Figure 4). Furthermore, average impacts were negative for 4 of 7 species in
275 native ranges and 6 of 7 species in expanded ranges (Figure 4). Thus, impacts reported in
276 expanded ranges of shifting species were indicated by – yet often more negative than – impacts
277 in native ranges.

278 Including studies from the native range not only increased the proportion of species for
279 which impact scores could be assigned, it also led to a more balanced assessment of both
280 detrimental and beneficial impacts of range shifts. While non-native species invasions are
281 decreasing global biodiversity (Doherty *et al.* 2016), range shifts are becoming increasingly
282 necessary for maintaining biodiversity (Thomas *et al.* 2004, Urban 2015) despite sometimes
283 causing negative impacts locally (see examples above). Therefore, while a focus on negative
284 impacts may be appropriate for invasive species management, decisions about range shifts will
285 need to consider both negative and positive impacts. Negative impacts may be more often
286 reported in expanded ranges where shifting species are more likely to be seen as detrimental,
287 while positive impacts may be more often studied in native ranges where species are deemed
288 beneficial. Therefore, while similarities between range shifts and invasions allow their impacts to
289 be assessed using a common protocol, range shifts are unique in their potential benefits for
290 global biodiversity. A more balanced impact assessment for range shifts, which helps to
291 minimize the influence of study and publication bias, would ideally include positive impacts and
292 studies conducted in native ranges.

293

294 **Conclusions**

295 Whereas managing non-native species invasions focuses on suppression (eg Hulme 2006),
296 management of range shifts is likely to require considering a broader scope of options, including
297 facilitation. Impact assessments developed for invasive species could be used as indicators of
298 potential consequences of range shifts, whether they occur with or without direct human
299 intervention. Our study of 39 range-shifting coastal marine species showed that data were
300 available to assess environmental impacts of >80% of species, similar to the proportion of
301 invasive species that can be assessed using the EICAT protocol. Given that this approach relies
302 on previously published studies, and in light of likely biases in the available literature, we
303 advocate for incorporating both negative and positive impacts studied across the native and

304 expanded ranges of shifting species. We note that the EICAT approach is largely precautionary
305 as it focuses on maximum recorded impacts. For well-studied species, it might be useful to
306 consider average and most commonly reported impacts. Still, given our finding that impacts
307 were more negative in expanded than native ranges, we should not be complacent about the
308 potential for impacts to be more detrimental than previously recorded. Since socioeconomic
309 impacts are rarely reported (<30% of study species), expert opinion could be solicited to fill this
310 data gap.

311 Impact assessments for range shifts would ideally be paired with vulnerability
312 assessments, both for species impacted in the expanded range and for the shifting species
313 themselves. Both shifting species and impacted species are candidates for management,
314 depending on their vulnerability (to changing climate or range-shift impacts), perceived value
315 (eg to biodiversity or economy), and cost/feasibility of interventions. Managers are probably
316 already aware of species in their jurisdiction that are endangered or of conservation concern,
317 and impacts on these species will likely be common justification for suppressing range shifts (as
318 with removals underway for shifting barred owls and pine beetles; Long & Wolfe 2019, Heuss *et*
319 *al.* 2019). In contrast, facilitation of range shifts (and even managed relocation) might be
320 considered when the potential range shifter is endangered. A first step in vulnerability
321 assessment would be to determine whether the shifting species or impacted species are
322 included on the IUCN Red List (www.iucnredlist.org/) of >32,000 species threatened with global
323 extinction (Van der Colff *et al.* 2020). Second, for species not on the IUCN Red List, vulnerability
324 could be assessed using an established protocol, such as the IUCN Red List extinction risk
325 assessment protocol (eg Short *et al.* 2011). The approach proposed here, combining impact
326 assessment with vulnerability assessment, minimizes cost because (1) both assessments rely
327 on previously published data, and (2) by starting with the impact assessment of an identified
328 range-shifting species, vulnerability assessments can target the shifting species itself and a
329 subset of species in the expanded range that are likely to be impacted. Management

330 alternatives can then be compared following a structured decision framework, such as those
331 developed for managed relocations, which incorporate information on the risks and feasibility of
332 options for attaining management goals (McLachlan *et al.* 2006, Hoegh-Guldberg *et al.* 2008).
333 Even when management is not feasible, the results of these impact assessments could inform
334 adaptation strategies (eg governance of transboundary shifts in fisheries species; Lindegren &
335 Brander 2018, Pinsky *et al.* 2018).

336 The recommendations above are based on a species-specific approach to management,
337 which may not be feasible for species that are not well studied, particularly as range shifts
338 accelerate. Thus, future studies should seek to identify generalities in the consequences of
339 range shifts. For example, Bradley *et al.* (2019) demonstrated that for invasions, impacts accrue
340 more rapidly from species of higher trophic levels, highlighting the need for more proactive
341 management of invasive predators and herbivores. As with invasive species, impacts of range
342 shifts are likely to be greatest for species with highest population sizes and individual effects.
343 Therefore, strong impacts of range shifts might be indicated by characteristics such as life
344 history strategies and trophic levels of shifting species or community-level resistance to
345 disturbance in expanded ranges (Catford *et al.* 2009, Wallingford *et al.* 2020). However,
346 generalizing impacts between species groups is premature, as our study revealed high
347 variability between species, with strong impacts by species at both the top and bottom (eg
348 habitat-forming primary producers) of the food chain and few taxonomic patterns (WebTable 2).

349 Our findings serve as evidence that there is no “one size fits all” approach for managing
350 range shifts, which depend on the level and type of impacts combined with human interests and
351 options for intervention. Of range shifts reviewed here, ~50% led to observed negative impacts
352 on environmental and/or socioeconomic systems (Figures 2,3). At the same time, redistribution
353 is increasingly important for global persistence of these species, some of which have already
354 experienced contractions of low-latitude range boundaries (eg Fenberg *et al.* 2014, Timbs *et al.*
355 2019). While our review focused on impacts of species addition (through range shifts or

356 managed relocation), this approach could also be used to illuminate impacts of species loss in
357 areas of range contraction. In summary, impact assessments developed for invasive species
358 combined with vulnerability assessments is a promising approach for evaluating whether range
359 shifts are, on balance, detrimental or beneficial. The next step in proactive management
360 involves determining the level of negative impacts that we are willing to accept, particularly
361 given beneficial impacts and extinction risks for range-shifting species.

362

363 **Data Availability**

364 The final dataset is available on Dryad, DOI 10.7280/D1770W.

365

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536 **Figure captions**

537

538 **Figure 1.** Number of published studies on environmental and socioeconomic impacts of 39
539 coastal marine species that have undergone range shifts in North America.

540

541 **Figure 2.** Maximum negative (red) and positive (blue) scores for environmental impacts of
542 range-shifting species based on the EICAT protocol (Table 1).

543

544 **Figure 3.** Maximum negative (red) and positive (blue) scores for socioeconomic impacts of
545 range-shifting species based on the SEICAT protocol (Table 1).

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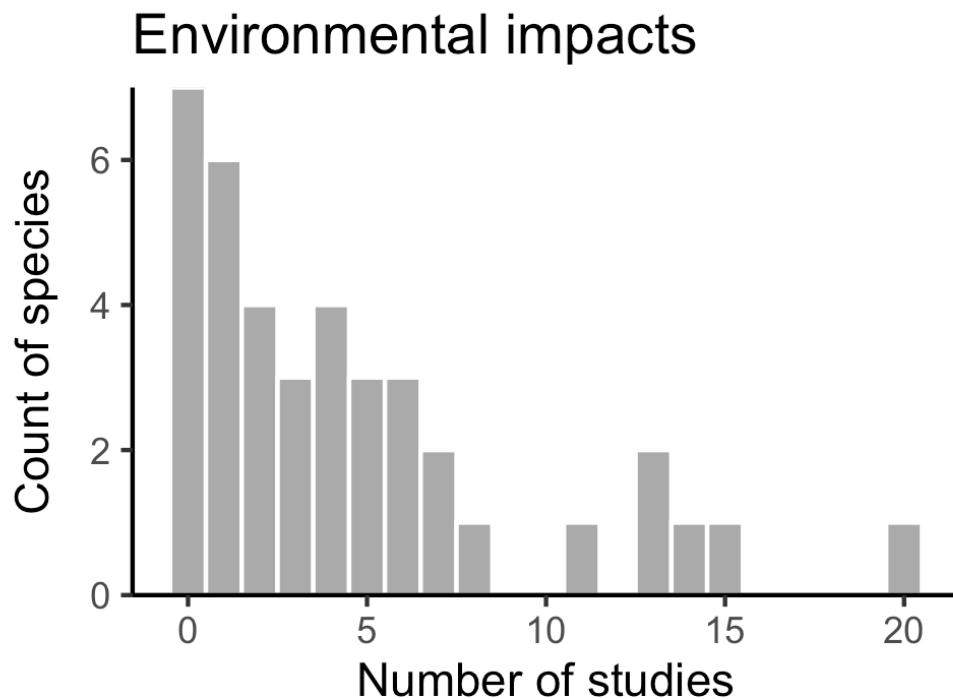
548 **Figure 4.** (a) Comparison of impact scores averaged across studies conducted in native versus
549 expanded ranges. Values range from -4 (major detrimental impact) to +4 (major beneficial
550 impact) for 8 species. The quadrants represent possible scenarios. For quadrants lying along
551 the diagonal 1:1 line, direction of impacts is the same in both ranges (negative at bottom left,
552 positive at top right). Alternately, impacts could switch from negative to positive (top left) or from
553 positive to negative (bottom right) during the range shift. Species in this analysis included (b)
554 black mangrove *Avicennia germinans*, (c) mangrove snapper *Lutjanus griseus*, and (d) dark
555 unicorn whelk *Mexacanthina lugubris*. Photo credits: (b) AR Hughes, (c) SA Bedgood, (d) DJ
556 Eernisse.

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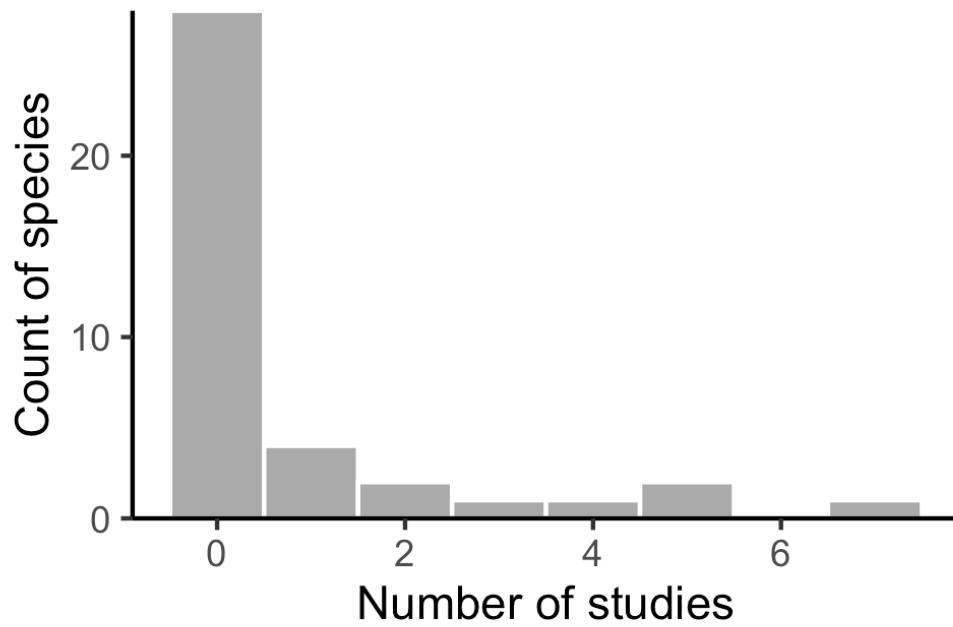
Table 1. Risk assessment impact levels

		EICAT (Environmental impacts)		SEICAT (Socioeconomic impacts)	
Impact Level		Negative	Positive	Negative	Positive
1	Minimal	Impacts possible (based on known interactions) but no change in native fitness observed	Impacts possible (based on known interactions) but no change in native fitness observed	Impacts possible (based on known uses) but no change in human activities observed	Impacts possible (based on known uses) but no change in human activities observed
2	Minor	Decreased fitness of a native species	Increased fitness of a native species	People continued to participate in an activity but with difficulty	People began to participate in an activity but with difficulty
3	Moderate	Decreased population size of a native species	Increased population size of a native species	Fewer people participated in an activity	More people participated in an activity
4	Major	Extirpation of a native population that could reestablish if the expander were removed	Establishment of a native population which would be lost if the expander were removed	An activity was suspended locally but would continue if the expander were removed	An activity commenced locally but would stop if the expander were removed
5	Massive	Extirpation of a native population which would not recover even if the expander were removed	Establishment of a native population which would persist even if the expander were removed	An activity was permanently lost in a location	An activity was permanently adopted in a location

561 Figure 1



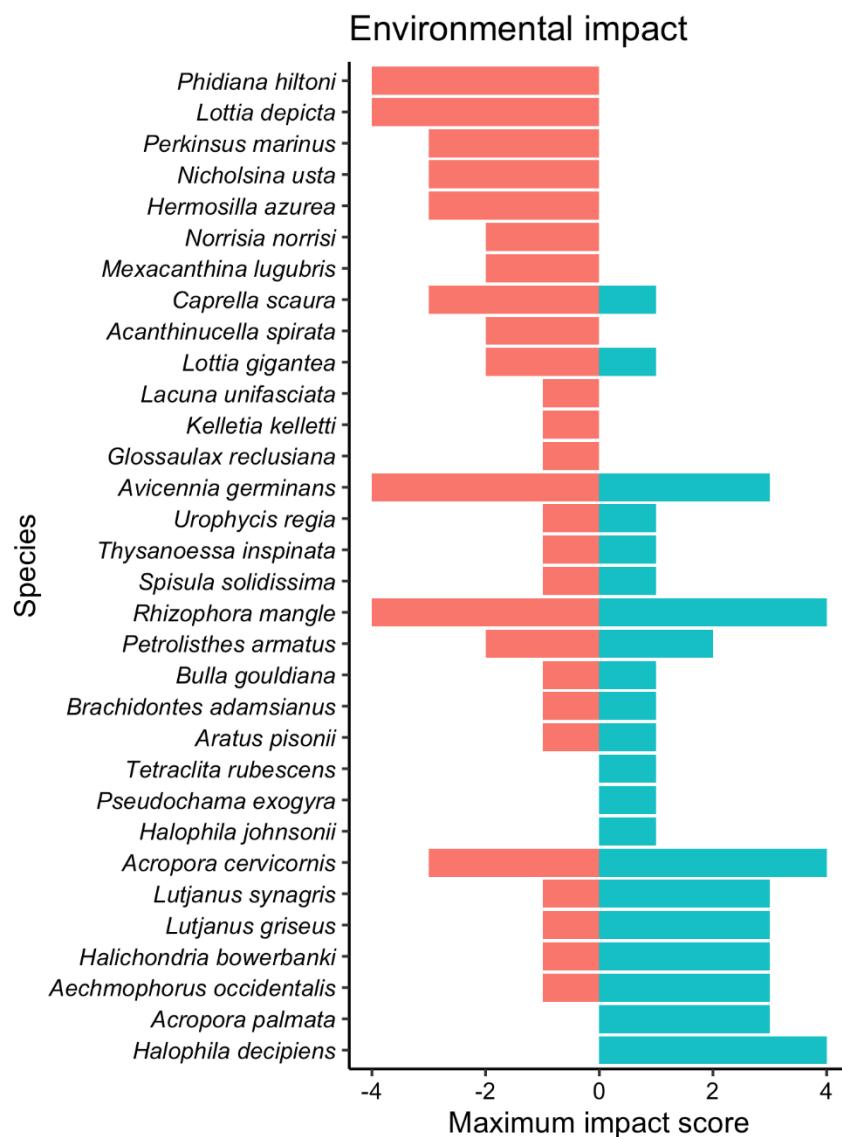
Socioeconomic impacts



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564 Figure 2

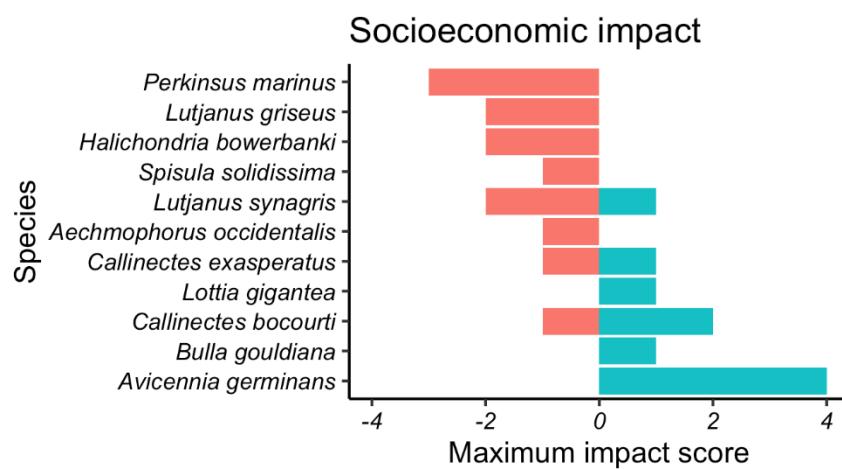


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567 Figure 3

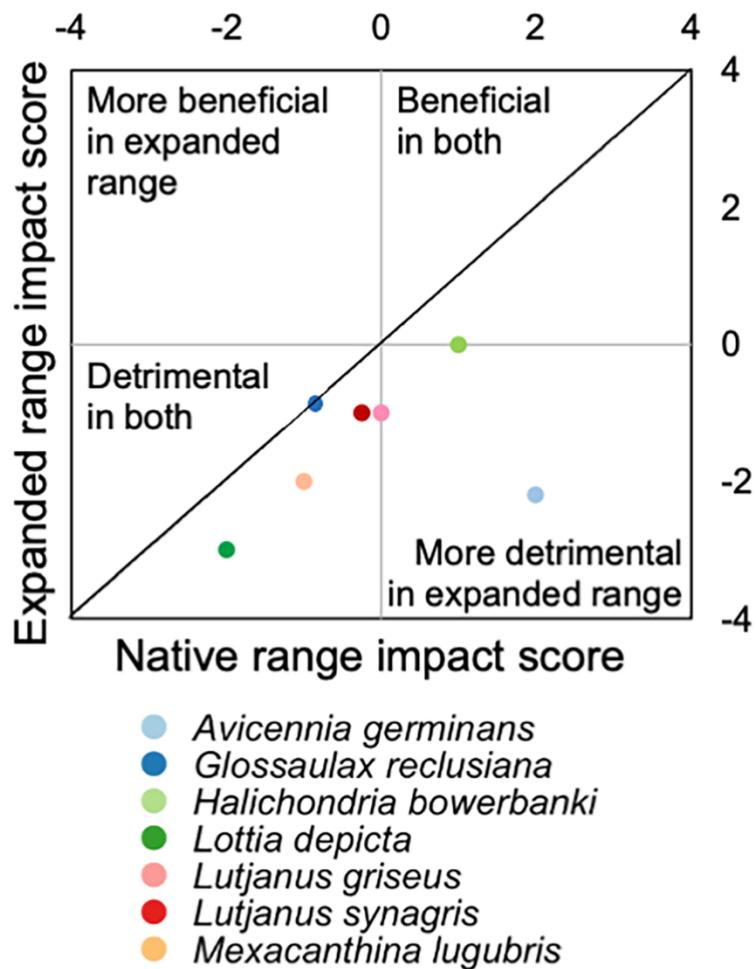
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571 Figure 4a



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Figure 4b



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577 Figure 4c



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580 Figure 4d



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