Frequency-Dependent Hybridization Contributes to Habitat Segregation in Monkeyflowers

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ABSTRACT: Spatial segregation of closely related species is usually attributed to differences in stress tolerance and competitive ability. For both animals and plants, reproductive interactions between close relatives can impose a fitness cost that is more detrimental to the rarer species. Frequency-dependent mating interactions may thus prevent the establishment of immigrants within heterospecific populations, maintaining spatial segregation of species. Despite strong spatial segregation in natural populations, two sympatric California monkeyflowers (*Mimulus nudatus* and *M. guttatus*) survive and reproduce in the other's habitat when transplanted reciprocally. We hypothesized that a frequency-dependent mating disadvantage maintains spatial segregation of these monkeyflowers during natural immigration. To evaluate this hypothesis, we performed two field experiments. First, we experimentally added immigrants in varying numbers to sites dominated by heterospecifics. Second, we reciprocally transplanted arrays of varying resident and immigrant frequencies. Immigrant seed viability decreased with conspecific rarity for *M. guttatus* but not for *M. nudatus*. We observed immigrant minority disadvantage for both species, but it was driven by different factors—frequency-dependent hybridization for *M. guttatus* and competition for resources and/or pollinators for *M. nudatus*. Overall, our results suggest a major role for reproductive interference in spatial segregation that should be evaluated along with stress tolerance and competitive ability.

Keywords: frequency dependence, hybridization, reproductive interference, habitat segregation, *Mimulus*, monkeyflower.

Introduction

Patterns of habitat segregation are common in closely related species and have been long mechanistically attributed to trade-offs between stress tolerance and competitive ability (Baker 1909; Connell 1961; Grace and Wetzel 1981; Yost et al. 2012; Peterson et al. 2013; Chen and Schemske 2015; Ferris and Willis 2018). Competition is well established to be an important mechanism driving zonation among closely related species with a high potential for niche overlap (Burns and Strauss 2011). However, mating between closely related species may also contribute to their spatial segregation (Anacker and Strauss 2014). The fitness costs of interspecific mating (reproductive interference) can drive patterns of segregation or coexistence through demographic displacement (Groning and Hochkirch 2008) or the evolution of increased prezygotic isolation in sympatry (Servedio and Noor 2003). The ecological and evolutionary outcomes of costly interspecific mating—spatial segregation (Groning and Hochkirch 2008) or character displacement (Brown and Wilson 1956)—depend on the relative rate of evolutionary change versus the rate of demographic decline in sympatry (Kyogoku and Wheatcroft 2020). While the initial establishment of habitat segregation could result from exclusion in sympatry or evolve directly in response to reproductive interference (Kyogoku and Kokko 2020), the long-term maintenance of zonation patterns requires mechanisms that result in the demographic decline of rare immigrants.

Plant species that co-occur and overlap in flowering often compete for pollinators and often suffer consequent fitness reductions (Mitchell et al. 2009). Competition for pollination can reduce fitness through reductions in visitation or through interspecific pollen transfer (Waser 1978*a*, 1978*b*). Both mechanisms of competition for pollination can affect fitness in a frequency-dependent manner. Rarer species may receive fewer visits and are expected to receive less conspecific pollen and more heterospecific pollen than common species. Models of reproductive interference thus predict fitness declines with increasing rarity (Levin and Anderson 1970; Kuno 1992). Rare-species disadvantage caused by reproductive interference is a major factor limiting

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persistence of polyploid species, which arise in populations of their diploid progenitors (minority cytotype exclusion; Levin 1975). Thus, considerable focus has been placed on these dynamics in co-occurring diploid and polyploid species (Husband 2000; Baack 2005; Buggs and Pannell 2006). However, the conditions experienced by rare polyploids are similar to those of parapatric species that meet in secondary contact (Lewis 1961; Ribeiro and Spielman 1986; Thum 2007), immigrants of species that are microspatially segregated within regions of sympatry (Singer 1990; Friberg et al. 2013; Christie and Strauss 2020), or nonnative species interacting with native congeners (Takakura et al. 2009; Takakura and Fujii 2010; Takakura 2013; Takakura and Fujii 2015).

An extreme cost to reproductive interactions is the production of inviable seeds from interspecific pollen transfer. This hybrid seed inviability is a common reproductive isolating barrier in plants and particularly among species in the *Mimulus guttatus* species complex (Vickery 1978; Garner et al. 2016; Oneal et al. 2016; Coughlan et al. 2020; Sandstedt et al. 2021). Inviable hybrid seeds produced between species in the *M. guttatus* species complex have a characteristic flat or shriveled shape, which is caused by arrested endosperm development. These inviable seeds can be distinguished by eye from round viable seeds with fully formed endosperm produced by conspecific pollinations.

Populations of the widespread *M. guttatus* have repeatedly evolved tolerance to harsh serpentine soils (Selby and Willis 2018) and co-occur with the geographically restricted serpentine soil endemic *M. nudatus* in the Northern Coast Range of California. *Mimulus guttatus* typically grows in streams, seeps, or meadows, whereas *M. nudatus* typically grows in washes, on rock outcroppings, or in drier rocky areas adjacent to the streams inhabited by *M. guttatus*. *Mimulus nudatus* typically grows in barer habitats with fewer co-occurring con- and heterospecifics than *M. guttatus*, a common pattern in serpentine soil endemics in contrast to species that occur both on and off serpentine soils (Sianta and Kay 2019). Since habitats occupied by *M. guttatus* and *M. nudatus* are often in close proximity (within meters) and immigrant individuals are regularly found, dispersal limitation does not explain spatial segregation in these species. When transplanted reciprocally at equal frequencies, *M. guttatus* has higher survival and produces more seed than *M. nudatus* in both habitats, demonstrating that stress tolerance or competition are not sufficient to explain patterns of spatial segregation between these species (Toll and Willis 2018). In those reciprocal transplants, both species produced many inviable F_1 hybrid seeds as immigrants, demonstrating the occurrence of reproductive interference (Toll and Willis 2018). While a combination of factors limits the persistence of *M. nudatus* immigrants in habitats occupied by *M. guttatus*, including a lack of toler-

ance to flooding in *M. nudatus* (Toll and Willis 2018) and inbreeding depression (Toll et al. 2021), these factors do not sufficiently explain the absence of *M. guttatus* in the drier habitats occupied by *M. nudatus*. We therefore hypothesize that frequency-dependent reproductive interference may maintain this spatial segregation pattern by imposing a strong fitness cost on natural immigrants, which usually occur alone or in very small numbers.

Using observational and experimental data collected across two field seasons, we asked three questions in this study. First, does hybridization reduce fecundity when *M. guttatus* and *M. nudatus* naturally occur in close proximity? If hybridization contributes to the maintenance of spatial segregation, we expect that seed viability will increase with increasing conspecific frequency. Second, when species immigrate to sites occupied by heterospecifics, do components of fitness in immigrants or residents (flowers per plant, seeds per flower, and seed viability) increase with their relative frequency? If resource competition contributes to the maintenance of habitat segregation, we expect that flowers per plant will increase with increasing conspecific frequency. If a lack of conspecific mates or pollinator competition contributes to the maintenance of habitat segregation, we expect that seeds per flower will increase with increasing conspecific frequency. If hybridization contributes to the maintenance of habitat segregation, we expect that seed viability will increase with increasing conspecific frequency. Third, does lifetime fecundity (measured as viable seeds per plant) in immigrants or residents increase with increasing conspecific frequency in experimental sympatry? If minority disadvantage contributes to the maintenance of habitat segregation, we would expect that viable seeds per immigrant or resident plant to increase with their increasing frequency, changing the relative performance of immigrants relative to residents.

Material and Methods

Reproductive Interference in Natural Populations: An Observational Transect

To characterize the existence and possible magnitude of reproductive interference in co-occurring populations of *Mimulus guttatus* and *M. nudatus*, we surveyed plants along a transect spanning an exposed serpentine outcrop that was inhabited by a pure stand of *M. nudatus*, a transition zone where both species were present at lower density, terminating in a serpentine seep with a pure stand of *M. guttatus* (38°51.528′N, 122°24.614′W). The entire length of the transect, spanning the two pure stands through a small zone of overlap, was 6 m. We counted all of the fruit produced by each species within eight 42 × 66-cm gridded quadrats centered on the transect line, with the long side

parallel to the transect line and evenly spaced along the length of the transect. We counted all fruit present within the quadrat at 7-cm intervals along the long side of the quadrat grid. At approximately 21-cm intervals along the long side of the quadrat, we collected one plant if only one species was present ($n = 13$ *M. nudatus* and $n = 9$ *M. guttatus*) or two plants if both species were present ($n = 10$ *M. nudatus* and $n = 10$ *M. guttatus*) at the midpoint of the short side of the quadrat. We counted fruit numbers instead of species abundances because flower number, not plant number, is what pollinators see in dense overlapping stands (Gardner and MacNair 2000). We counted all seeds produced by the plants collected at 21-cm intervals and categorized them as round, viable (conspecific), or flat, inviable F_1 hybrid seeds on the basis of morphology (Oneal et al. 2016). We calculated seed viability by dividing the number of round viable conspecific seeds by the total number of seeds produced by each plant. We estimated Pearson correlation coefficients for the relationship between the proportion of conspecific fruit and seed viability per sampled plant in R version 4.1.2 (R Core Team 2021).

Experimental Transplants

To test whether frequency dependence contributes to the maintenance of habitat segregation, we performed two reciprocal transplants over two years. In the first field season (2018), we transplanted immigrants in varying abundances into natural populations of resident heterospecifics. This allowed us to simulate a situation where immigrants arrive at low abundances to sites already established by heterospecifics. However, since we did not transplant the naturally occurring resident species with the immigrants, we lacked an appropriate comparison for immigrant performance relative to residents. In the second field season (2019), we transplanted immigrants and residents at varying frequencies reciprocally, allowing us to compare the overall fecundities of experimental immigrants and residents.
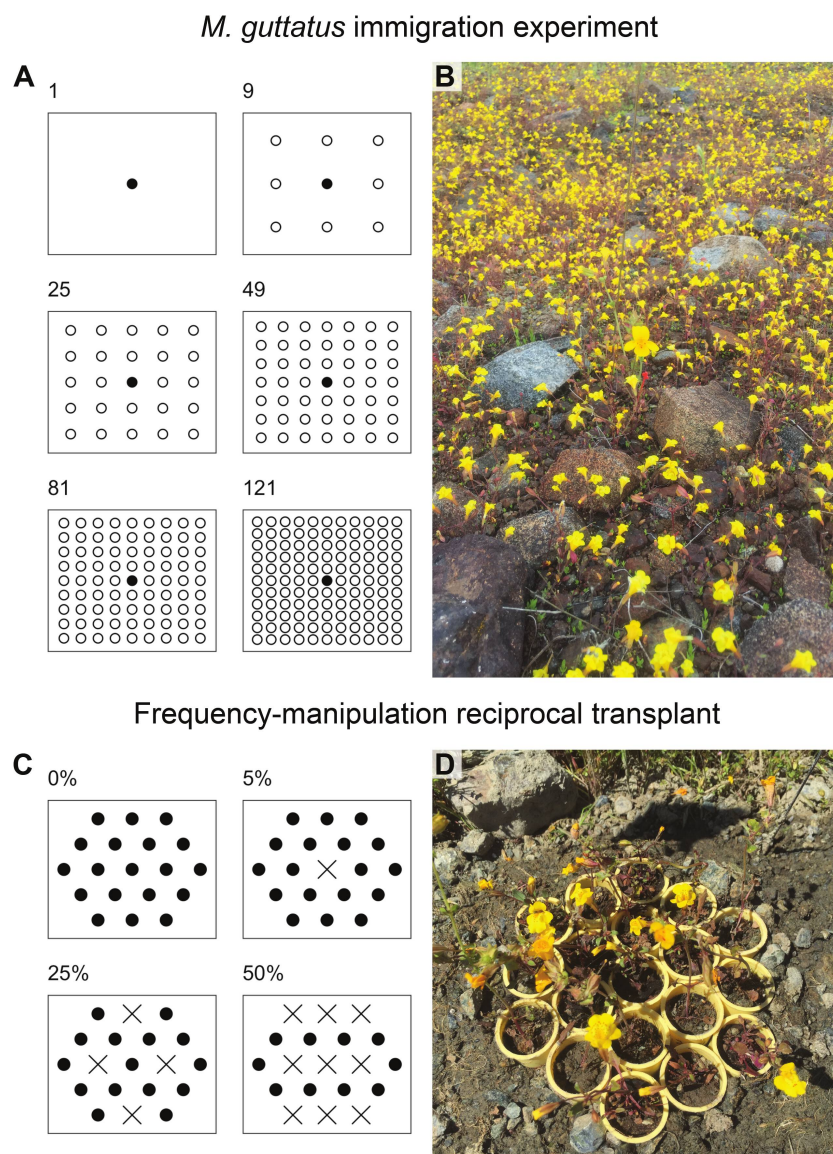
Mimulus guttatus Immigration Experiment. In February 2018, we transplanted variable numbers of immigrants into naturally dense patches of native seedlings (>100 per half meter). While we performed this experiment reciprocally, all but nine immigrant *M. nudatus* transplants died before flowering. We lacked sufficient seedlings to replace them, and they were not analyzed further. Therefore, this experiment only allowed us to examine the immigration dynamics of *M. guttatus* into *M. nudatus* habitat (the direction less well explained by our previous studies).

We collected seeds from a local population of *M. guttatus* at the University of California McLaughlin Reserve (38°51.528′N, 122°24.614′W) in 2013. We grew seeds from field-collected maternal families in a greenhouse and crossed a single plant from each family to a single plant from a different family to produce outbred maternal families in 2013 (described in more detail in Toll et al. 2021). To minimize effects of inbreeding on fecundity (see Toll et al. 2021), we germinated six outbred maternal families of *M. guttatus* to use as focal plants. These plants were initially grown in the shade house at the University of California McLaughlin Reserve in 2018. We did not have sufficient seed to use these same outbred lines as neighbors, so we pooled equal numbers of seed from 20 field-collected maternal families of *M. guttatus* to germinate in the shade house to use as neighbors. We planted 72 blocks of focal *M. guttatus* plants in random order by maternal family and then randomly assigned each family to one of the six levels of the initial conspecific neighbor treatments (0, 8, 24, 48, 80, and 120 neighbors; fig. 1*A*). We transplanted a range of *M. guttatus* seedlings that were ecologically relevant if a single fruit dispersed into *M. nudatus* habitat, given that each fruit produces an average of 200 seeds (Toll and Willis 2018). *Mimulus guttatus* seedlings were planted within naturally dense patches of *M. nudatus* (>100 native *M. nudatus* within a half meter of each focal *M. guttatus* plant; fig. 1*B*). We planted *M. guttatus* neighbors within a half meter of each focal plant in a regularly spaced array, and we planted these blocks more than 3 m away from one another.

We replaced *M. guttatus* focal transplants as they died of transplant shock, resulting in 62 focal plants (of the original 72 plus 28 replacements) that survived to flower by the end of the season. Neighbor transplant mortality rates were also high; 78% of *M. guttatus* transplants did not survive to flowering, resulting in a range of flowering neighbors from zero to 103, with a median of one neighbor surviving to flowering. In subsequent analyses of the effect of conspecific neighbors on components of fitness, we included only neighbors that successfully survived to flower, as only they could potentially reduce reproductive interference. Mortality rates from this transplant experiment are consistent with previous estimates at this site: 73%, 33%, and 82% of *M. guttatus* seedlings died at this site in 2015, 2016, and 2017, respectively (site QV in Toll and Willis 2018). Thus, we are confident that our transplants represent an ecologically realistic range of potential surviving immigrant *M. guttatus* seedlings into *M. nudatus* habitat. We counted all seeds produced by the 62 focal plants that produced fruit and categorized them as viable round conspecific-pollinated or inviable flat heterospecific-pollinated (hybrid) seeds on the basis of morphology.

Statistical analysis. To examine how the number of conspecific neighbors affected focal plant fitness components (flowers per plant, total seed per flower, and seed viability per focal *M. guttatus* transplant), we used generalized linear mixed models in the R package glmmTMB

M. guttatus immigration experiment

Frequency-manipulation reciprocal transplant

Figure 1: *A*, Schematic of the 2018 *Mimulus guttatus* immigration experiment showing positions of focal immigrant *M. guttatus* plants (filled circles) and conspecific neighbors (open circles) within each block. *B*, Photograph of an *M. guttatus* transplant with no conspecific neighbors, surrounded by native *Mimulus nudatus*. *C*, Schematic of the 2019 frequency-manipulation reciprocal transplant experiment showing the positions of immigrant (crosses) and resident (filled circles) plants within each immigrant frequency treatment. Figure adapted from Nagy (1997). Five replicates per treatment were planted at each site, resulting in 20 blocks containing residents and 15 blocks containing immigrants at each site. *D*, Photograph of an example 50% (9N:10G) immigrant plot at the seep site. The diameter of each pot was 2.54 cm, and the largest diameter of each array was 12.7 cm. Native *Mimulus* plants within 0.25 m of each array were removed prior to transplanting.

(Brooks et al. 2017). We fit models with each fitness component as the dependent variable, the number of surviving conspecific neighbors as a fixed effect, and maternal family ($n = 6$) of the focal plant as a random effect. Models of total seeds per flower had an additional nested random effect term for individual plant ($n = 62$), since each focal plant produced multiple flowers.

We fit each count model first with a Poisson error distribution and identified the best-fitting error distribution by evaluating model diagnostics with the R package DHARMa (Hartig 2021). The effect of conspecific neighbors on flowers per focal transplant was estimated using a negative binomial model because Poisson models were significantly overdispersed. The effect of conspecific neighbors

on seeds per flower was estimated using zero-inflated negative binomial models with a constant intercept because Poisson and negative binomial models were significantly overdispersed and zero inflated. The effect of conspecific neighbors on seed viability was estimated using a beta-binomial model because the binomial model was significantly overdispersed.

We assessed the significance of the fixed effect (the number of conspecific neighbors) with the best-fitting model for each fitness component using Wald type II χ^2 tests (R package car; Fox and Weisberg 2019). We predicted marginal effects and bootstrapped ($n = 500$ iterations) 95% confidence intervals with the R package ggeffects (Lüdecke 2018). We plotted raw data and predictions with ggplot2 (Wickham 2016) and combined plots with patchwork (Pederson 2020).

Frequency-Manipulation Reciprocal Transplant. To test whether frequency dependence contributes to the maintenance of habitat segregation, we reciprocally transplanted *M. guttatus* and *M. nudatus* at different frequencies but with the same number of individuals in experimental patches. These mixed species patches were transplanted into a seep dominated by *M. guttatus* and a wash dominated by *M. nudatus*. The plants used in this experiment were all collected directly from natural populations at each transplant site in the year (2019) of the experiment. In April 2019, we dug up 380 seedlings per species from each site and repotted them in yellow 49-mL (2.54 cm in diameter) Cone-tainers filled with potting soil (Stuewe and Sons, Tangent, OR). We maintained seedlings in a shade house at the University of California McLaughlin Reserve for a week prior to redistributing seedlings in Cone-tainers at both sites. We sunk Cone-tainers into the ground so that the top was flush with the soil surface at each site in 19 plant hexagonal arrays (fig. 1C). Arrays were planted within natural populations of each species, but we cleared native *Mimulus* plants within 0.25 m of each array. We designed this experiment following the frequency manipulation arrays from Nagy (1997): 0% immigrants (0 immigrants, 19 residents); 5% immigrants (1 immigrant, 18 residents); ~25% immigrants (4 immigrants, 15 residents); and ~50% immigrants (9 immigrants, 10 residents; fig. 1D). Higher levels of immigrants than natives would have been unrealistic in this system. For clarity, we abbreviate frequency treatments as the ratio of *M. guttatus* (G) and *M. nudatus* (N) transplants. We assigned plants to four frequency treatments at the seep site (19G, 18G∶1N, 15G∶4N, and 10G∶9N), which is naturally occupied by *M. guttatus*, and four frequency treatments at the wash site (19N, 18N∶1G, 15N∶4G, and 10N∶9G), which is naturally occupied by *M. nudatus*. We planted five replicates per frequency treatment

per site, for a total of 20 blocks per site and 40 blocks overall. Blocks were planted at least 3 m away from each other. We added supplemental water to blocks for a month after transplanting and replaced seedlings that died prior to flowering to ensure that the same number of plants flowered in each frequency treatment. We left containers in the field until all plants terminated flowering and ripened their fruit. We collected fruit from each transplant in the field prior to digging up containers. We counted and categorized all seeds produced by experimental plants and estimated the hybridization rate using the same methods as the observational transect. We estimated lifetime fecundity (viable seeds per plant) from three fitness components: the total number of flowers produced per plant, the total number of seeds produced per flower, and seed viability (= viable seeds/total seeds produced).

Statistical analysis. To examine how the immigrant frequency treatments affected plant fitness components (flowers per plant, total seed per flower, and seed viability) and lifetime fecundity (viable seeds per plant), we used generalized linear mixed models in the R package glmmTMB (Brooks et al. 2017). We tested whether fitness components differed among frequency treatments for each species at each site separately because no immigrants were transplanted in the 0% immigrant frequency treatment (fig. 1D). We fit models with each fitness component as the dependent variable, immigrant frequency as a fixed effect, and block as a random effect (block $n = 15$ for immigrants, $n = 20$ for residents). We also tested whether lifetime fecundity differed between species, but only within the treatments where both immigrants and residents were present (5%, 25%, and 50% immigrant frequency treatments) at each site separately. We fit models with lifetime fecundity as the dependent variable, immigrant frequency, species, and their interaction as fixed effects and block as a random effect (block $n = 15$).

Poisson models of count fitness components and lifetime fecundity were significantly overdispersed, and models of seeds per flower and viable seeds per plant were significantly zero inflated. Thus, the effect of immigrant frequency on flowers per plant was estimated using a negative binomial distribution, while total seeds per flower and viable seeds per plant were estimated using a zero-inflated negative binomial distribution with a constant zero-inflation intercept applied to all observations. The effect of immigrant frequency on seed viability was analyzed using a binomial model for immigrants and a beta-binomial model for residents because binomial models for residents were significantly overdispersed.

We assessed the significance of the fixed effect (immigrant frequency treatment) with the best-fitting model for each fitness component using Wald type II χ^2 tests and assessed the significance of the fixed effects (immigrant

frequency treatment, species, and their interaction) for lifetime fecundity using Wald type III χ^2 tests (R package car; Fox and Weisberg 2019). We tested whether fitness components differed among treatments within species and whether species differed in lifetime fecundity within each frequency treatment with Tukey post hoc contrasts (R package emmeans; Lenth 2022). We estimated marginal means and bootstrapped ($n = 500$ iterations) 95% confidence intervals with the R package ggeffects (Lüdecke 2018). We plotted marginal means with ggplot2 (Wickham 2016) and combined plots with patchwork (Pederson 2020).

Results

Reproductive Interference in Natural Populations: An Observational Transect

Individuals collected along a transect produced a high proportion of viable seeds in zones where only conspecifics were present (average: 0.92 for *Mimulus guttatus*, 0.84 for *M. nudatus*), even though heterospecifics were in very close proximity (always within 3 m). In contrast, individuals collected in the 3-m overlap zone of the two species pro-

duced far lower proportions of viable seeds (average: 0.53 for *M. guttatus*, 0.66 for *M. nudatus*; fig. 2). Seed viability was positively correlated with the local frequency of conspecific fruit for *M. guttatus* plant sampled along the transect (Pearson correlation coefficient $= 0.66$, 95% CI $= 0.29$ to 0.86, $t = 3.61$, df $= 17$, $P = .002$), meaning that greater numbers of adjacent conspecific flowers and fruits were associated with greater proportions of viable seeds. However, seed viability was not significantly correlated with the local frequency of conspecific fruit for *M. nudatus* plants sampled along the transect (Pearson correlation coefficient $= 0.18$, 95% CI $= -0.28$ to 0.57, $t = 0.79$, df $= 19$, $P = .44$). These data from a natural transition zone suggest that reproductive interference is more likely to be frequency dependent in *M. guttatus* immigrants than in *M. nudatus* immigrants.

Experimental Transplants

Mimulus guttatus Immigration Experiment. As expected from the observational transect, *M. guttatus* immigrant transplants hybridized less and produced a greater proportion of viable conspecific seeds when surrounded by more
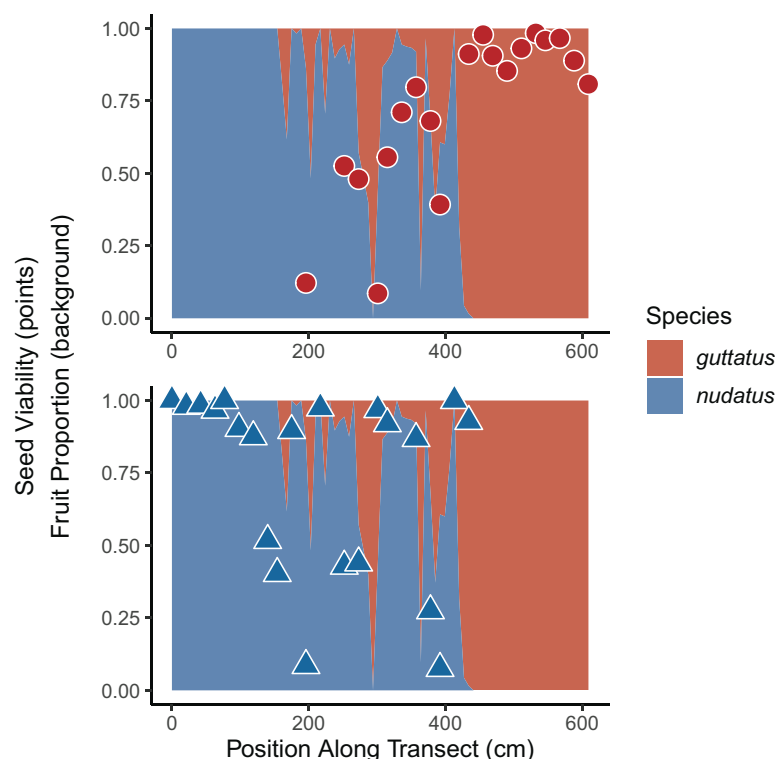

Figure 2: Transect of co-occurring *Mimulus nudatus* and *M. guttatus*. Background colors depict the frequency of *M. nudatus* fruit (blue) versus *M. guttatus* fruit (red). Points indicate the proportion of inviable F_1 hybrid seeds produced by individual *M. nudatus* (triangles) and *M. guttatus* (circles) plants collected every 21 cm along the transect.

flowering conspecific neighbor transplants (fig. 3*C*), as would occur if an intact fruit or multiple seeds colonized a habitat patch occupied by heterospecifics. Seed viability increased with increasing numbers of conspecific neighbors (Wald type II χ^2 test: $\chi^2 = 7.49$, df $= 1$, $P = .006$; coefficient estimate $= 0.023$, SE $= 0.009$, $z = 2.74$, $P = .006$). The model-predicted seed viability for rare immigrants increased from 0.24 for immigrants without conspecific neighbors to 0.78 for the plant with 103 neighbors (fig. 3*C*).

Flowers per plant and total seeds per flower for focal immigrant *M. guttatus* transplants were not significantly associated with the number of flowering conspecific neighbor transplants (fig. 3*A*, 3*B*; Wald type II χ^2 tests: flowers per plant $\chi^2 = 0.30$, df $= 1$, $P = .58$; seeds per flower $\chi^2 = 0.52$, df $= 1$, $P = .47$).

One focal immigrant *M. guttatus* transplant had more than twice as many surviving neighbors as any other focal transplant. To examine whether including this individual
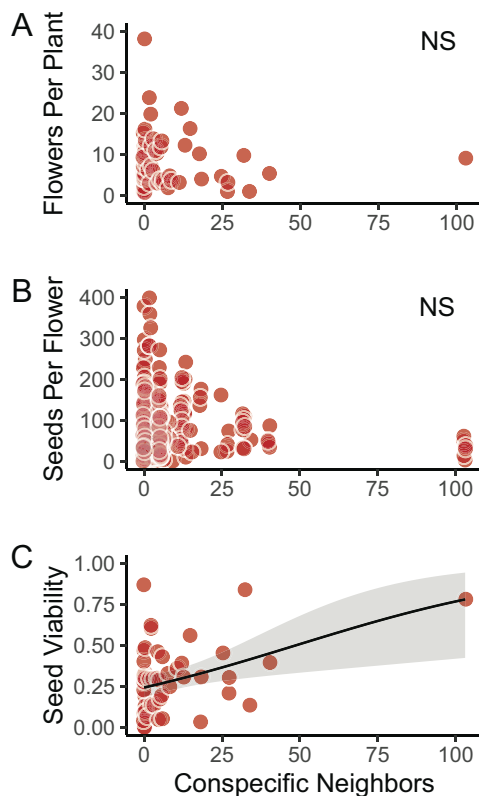
Figure 3: Components of fitness from the 2018 *Mimulus guttatus* immigration experiment. Fitness components: flowers per plant (*A*), total seeds per flower (*B*), and seed viability (*C*). Each red circle represents a single immigrant focal *M. guttatus* transplant in a wash dominated by *Mimulus nudatus*. Conspecific neighbor abundance was significantly associated with focal plant hybridization rate. NS = not significant.

transplant influenced the relationship between fitness components and the number of conspecific neighbors, we reanalyzed the data in two ways: by log transforming the number of neighbors and removing the outlier (figs. S1, S2, available online). Neither of these changes qualitatively changed the analysis presented with the raw data.

Frequency-Manipulation Reciprocal Transplant. Fitness components and lifetime fecundity were analyzed in the wash and seep transplant sites.

Fitness components in wash. For immigrant *M. guttatus* transplants in the wash, seed viability was significantly associated with the frequency treatment (fig. 4*E*; Wald type II χ^2 test: $\chi^2 = 10.602$, df $= 2$, $P = .005$). In contrast, flowers per plant (fig. 4*A*; Wald type II χ^2 test: $\chi^2 = 1.75$, df $= 2$, $P = .42$) and seeds per flower (fig. 4*C*; Wald type II χ^2 test: $\chi^2 = 0.44$, df $= 2$, $P = .80$) were not associated with the frequency treatment. Seed viability in immigrant *M. guttatus* transplants increased from 0.14 to 0.48 between the 5% (1G:18N) and 50% (9G:10N) immigrant frequency treatments (Tukey post hoc contrast: $P = .006$; table S1; tables S1–S6 are available online).

For resident *M. nudatus* transplants in the wash, seed viability was significantly associated with the frequency treatment (fig. 4*F*; Wald type II χ^2 test: $\chi^2 = 11.40$, df $= 3$, $P = .01$), but flowers per plant (fig. 4*B*; Wald type II χ^2 test: $\chi^2 = 3.18$, df $= 3$, $P = .37$) and seeds per flower (fig. 4*D*; Wald type II χ^2 test: $\chi^2 = 6.47$, df $= 3$, $P = .09$) were not. Seed viability in resident *M. nudatus* transplants decreased from 0.83 to 0.68 between the 0% (1G:18N) and 25% (9G:10N) immigrant frequency treatments (Tukey post hoc contrast: $P = .005$; table S2).

Fitness components in seep. For immigrant *M. nudatus* transplants in the seep, flowers per plant (fig. 5*A*; Wald type II χ^2 test: $\chi^2 = 11.74$, df $= 2$, $P = .003$) and seeds per flower (fig. 5*C*; Wald type II χ^2 test: $\chi^2 = 12.21$, df $= 2$, $P = .002$) were significantly associated with frequency treatment, but seed viability was not (fig. 5*E*; Wald type II χ^2 test: $\chi^2 = 1.52$, df $= 2$, $P = .47$). Flowers per plant in immigrant *M. nudatus* transplants decreased from approximately six to approximately two flowers per plant between the 25% (4N:15G) and 50% (9N:10G) immigrant frequency treatments (Tukey post hoc contrast: $P = .005$; table S3). This decline resulted in a lower proportion of *M. nudatus* flowers than expected in the highest immigrant frequency treatment (proportion immigrant *M. nudatus* flowers by immigrant frequency treatment: 5% [1N:18G] treatment $= 0.06$, 25% [4N:15G] treatment $= 0.33$, 50% [9N:10G] treatment $= 0.29$). Seeds per flower in immigrant *M. nudatus* transplants increased from 27 to 64 seeds per flower
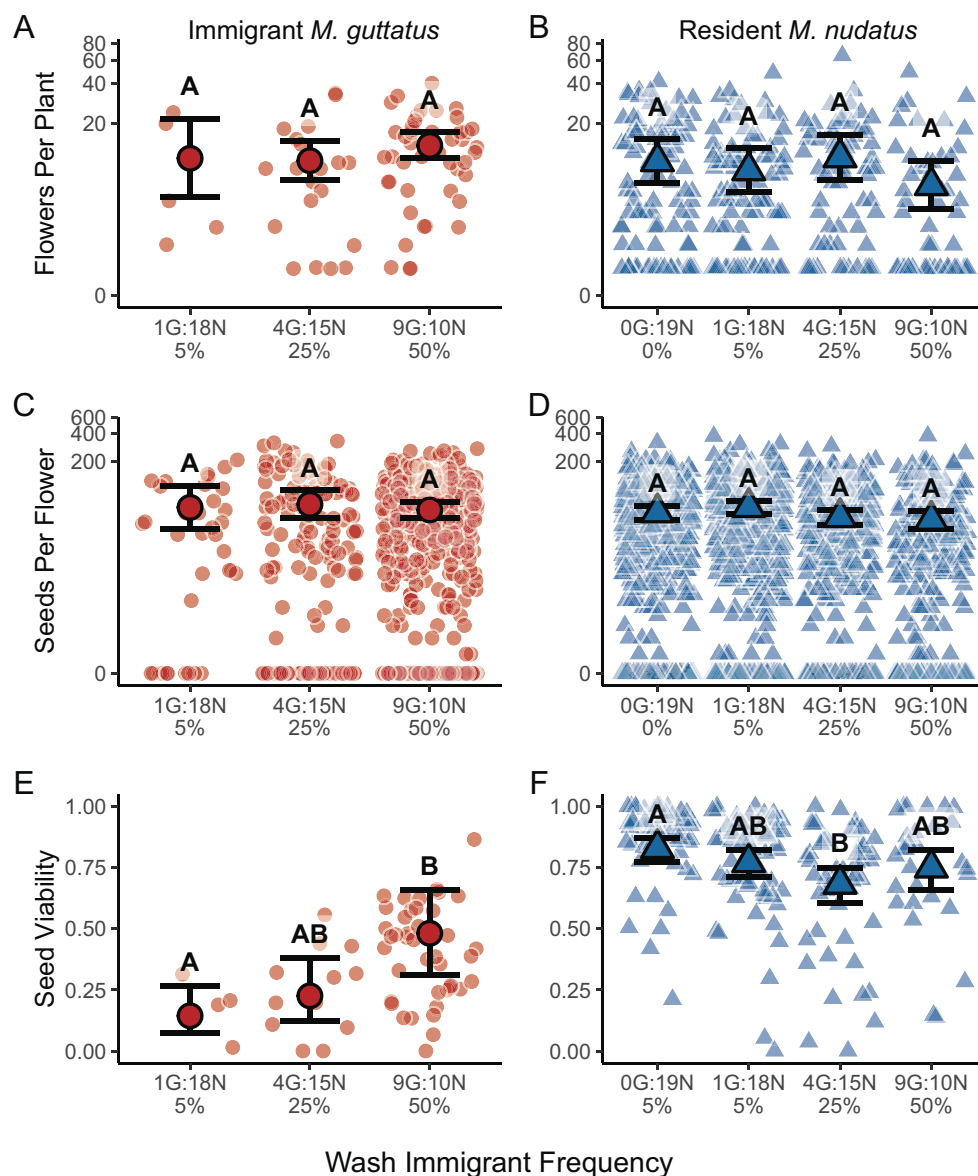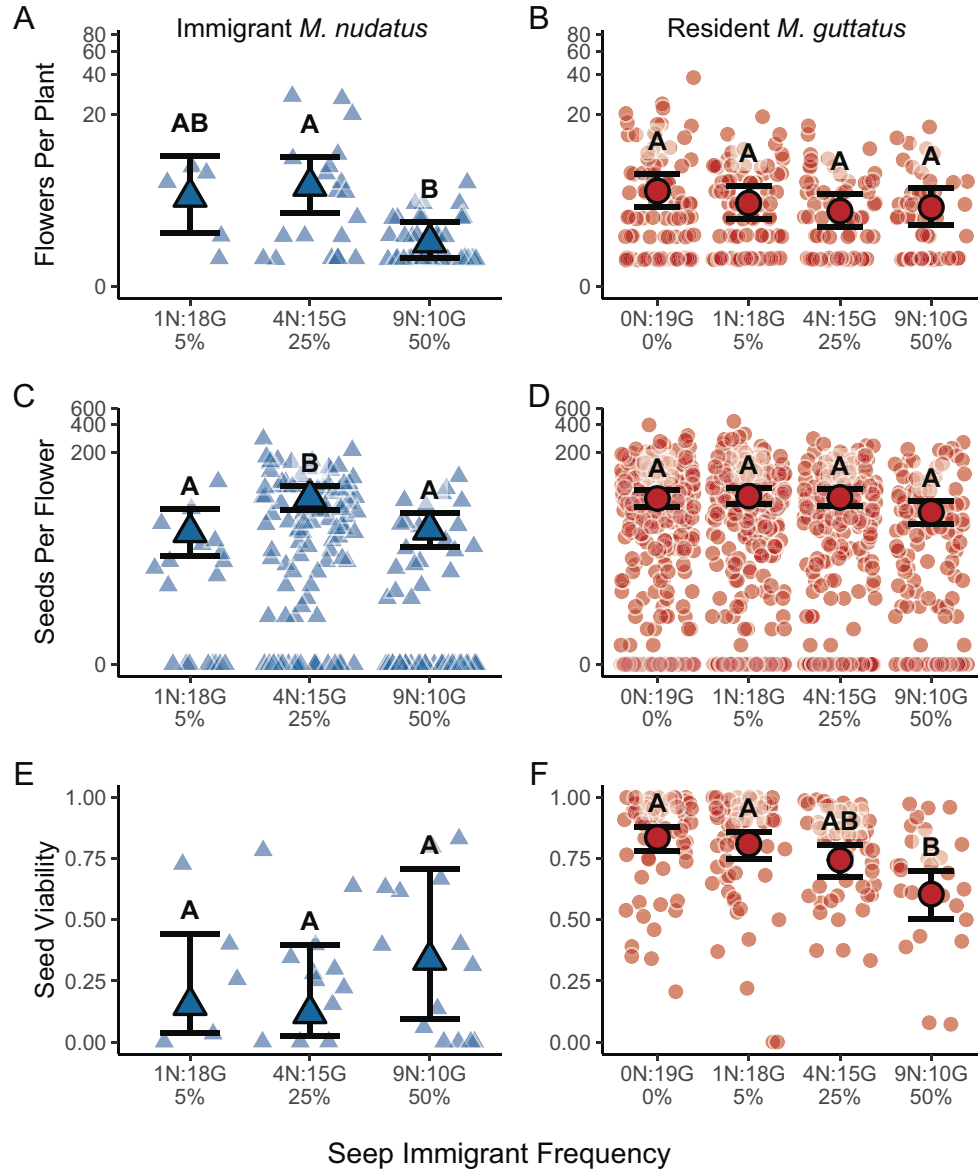
Figure 4: Predicted marginal means and 95% confidence intervals for components of fitness in the wash transplant site from the 2019 reciprocal transplant experiment. Models: fitness component = immigrant frequency + (1|block). Fitness components: flowers per plant (*A*, *B*), total seeds per flower (*C*, *D*), and seed viability (*E*, *F*). Immigrant *Mimulus guttatus* = red circles (*A*, *C*, *E*); resident *Mimulus nudatus* = blue triangles (*B*, *D*, *F*). Letters indicate groupings from Tukey post hoc contrasts.

between the 5% (1N : 18G) and 25% (4N : 15G) immigrant frequency treatments and decreased from 64 to 29 between the 25% (4N : 15G) and 50% (9N : 10G) immigrant frequency treatments (Tukey post hoc contrasts: $P < .05$; table S3).

For resident *M. guttatus* transplants in the seep, seed viability was significantly associated with frequency treatment (fig. 5*F*; Wald type II χ^2 test: $\chi^2 = 21.56$, df = 3, $P < .001$), but flowers per plant (fig. 5*B*; Wald type II χ^2 test: $\chi^2 = 3.14$, df = 3, $P = .37$) and seeds per flower

(fig. 5*D*; Wald type II χ^2 test: $\chi^2 = 5.80$, df = 3, $P = .122$) were not. Seed viability decreased from 0.84 to 0.60 between the 0% (0N : 19G) and 50% (9N : 10G) immigrant frequency treatments (Tukey post hoc contrast: $P < .001$; table S4) and decreased from 0.81 to 0.60 between the 5% (1N : 18G) and 50% (9N : 10G) immigrant frequency treatments (Tukey post hoc contrast: $P = .001$; table S4).

Lifetime fecundity in wash. Viable seeds per plant in the wash was significantly associated with immigrant

Figure 5: Predicted marginal means and 95% confidence intervals for components of fitness in the seep transplant site from the 2019 reciprocal transplant experiment. Models: fitness component = immigrant frequency + (1|block). Fitness components: flowers per plant (*A*, *B*), total seeds per flower (*C*, *D*), and seed viability (*E*, *F*). Immigrant *Mimulus nudatus* = blue triangles (*A*, *C*, *E*); resident *Mimulus guttatus* = red circles (*B*, *D*, *F*). Letters indicate groupings from Tukey post hoc contrasts.

frequency treatment, species, and an immigrant frequency × species interaction (fig. 6*A*; Wald type III χ^2 test: frequency, $\chi^2 = 6.83$, df = 2, $P = .03$; species, $\chi^2 = 12.14$, df = 1, $P < .001$; frequency × species, $\chi^2 = 14.12$, df = 2, $P < .001$). Immigrant *M. guttatus* transplants produced one-sixth as many viable seeds per plant as resident *M. nudatus* transplants in the 5% (1N : 18G) immigrant frequency treatment (60 vs. 347; Tukey post hoc contrast: $P = .008$; table S5).

Lifetime fecundity in seep. Viable seeds per plant in the seep was significantly associated with species and an immigrant frequency treatment × species interaction (fig. 6*B*; Wald type III χ^2 test: frequency, $\chi^2 = 2.14$, df = 2, $P = .34$; species, $\chi^2 = 18.85$, df = 1, $P < .001$; frequency × species, $\chi^2 = 15.21$, df = 2, $P < .001$). Immigrant *M. nudatus* transplants produced one-twelfth as many viable seeds per plant as resident *M. guttatus* transplants in the 5% (1N : 18G) immigrant frequency treatment (15 vs. 187;
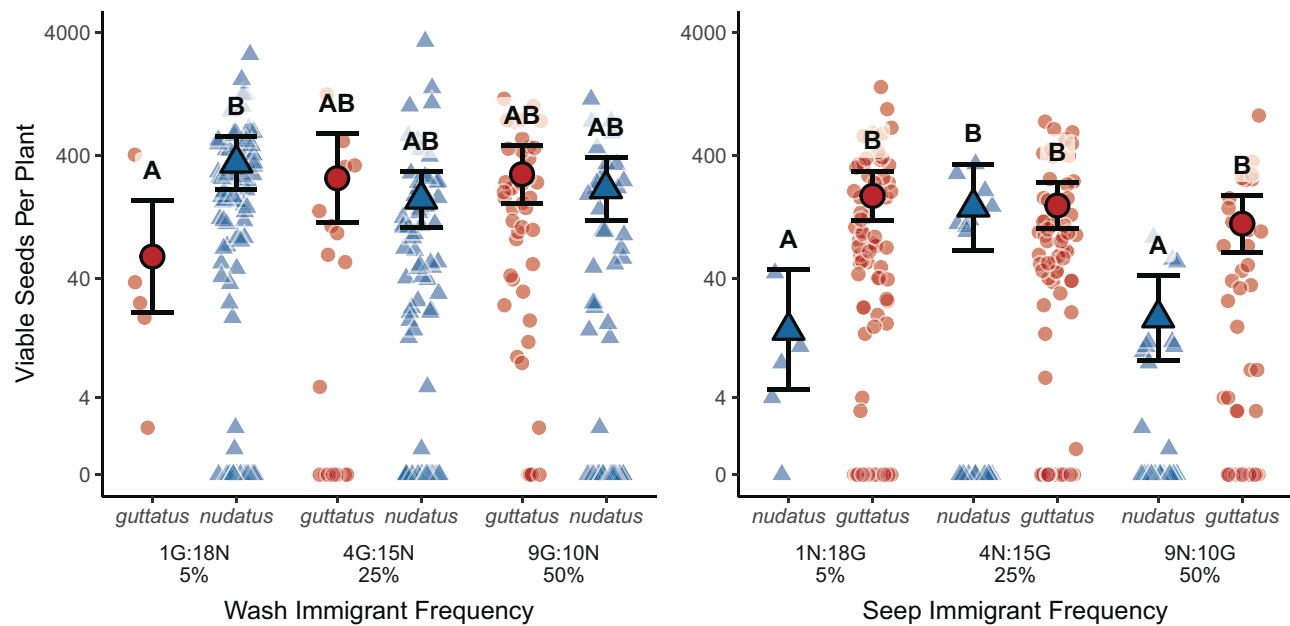
Figure 6: Predicted marginal means and 95% confidence intervals for lifetime fecundity in a wash (*A*) and seep (*B*) in the 2019 reciprocal transplant experiment. *Mimulus nudatus* = blue triangles; *Mimulus guttatus* = red circles. Marginal means were predicted from models using data from plots where both species were present at each site. Models: lifetime fecundity = species + immigrant frequency + species × immigrant frequency + (1|block). Letters indicate significant groupings from Tukey post hoc tests.

Tukey post hoc contrast: *P* < .001) and one-sixth as many in the 50% (9N:10G) treatment (19 vs. 111; Tukey post hoc contrast: *P* = .001; table S6).

Experimental Transplant Summary. Seed viability was positively associated with immigrant frequency for immigrant *M. guttatus* transplants but not for immigrant *M. nudatus* transplants, consistent with the findings from the observational transect. Resident transplants of both species also suffered from decreased seed viability as immigrant transplants increased in frequency. Unexpectedly, fitness components declined for immigrant *M. nudatus* transplants when they were most common, resulting in immigrant disadvantage in the highest frequency treatment. Immigrant transplants had lower lifetime fecundity, measured as viable seeds per plant, than resident transplants when immigrants were rarest. For immigrant *M. guttatus*, this rarity disadvantage could not be explained by fitness component differences prior to hybridization. In contrast, rarity disadvantage for immigrant *M. nudatus* was observed for fitness components prior to hybridization.

Discussion

In this study, we found that frequency-dependent hybridization likely contributes to habitat segregation between two closely related monkeyflower species. Each species

suffered from hybridization when in close contact with heterospecifics in both field experiments (figs. 3*C*, 4*E*, 5*E*) as well as in natural populations (fig. 2). Immigrants of both species had lower lifetime fecundity than residents when rare (fig. 6). However, this minority disadvantage was caused by different factors for each species. Partitioning of the fitness components demonstrated that rarity disadvantage for *Mimulus guttatus* immigrants was most consistent with reproductive interference (fig. 4), whereas aboveground resource and/or pollinator competition together with reproductive interference all likely contributed to rarity disadvantage for *Mimulus nudatus* immigrants (fig. 5). *Mimulus nudatus* immigrants also suffered fitness reductions consistent with intraspecific competition when they were at their highest frequency (fig. 5*A*). This negative frequency dependence is predicted to facilitate immigrant invasion and break down spatial segregation (Grainger et al. 2019). However, since it was identified only for *M. nudatus*, it is unlikely to lead to breakdown. Our previous studies in this system identified flooding and inbreeding depression as major factors limiting the long-term establishment of *M. nudatus* in habitats occupied by *M. guttatus* (Toll and Willis 2018; Toll et al. 2021). The strong rarity disadvantage for *M. guttatus* immigrants is further evidence that multiple asymmetric processes contribute to the maintenance of strong spatial segregation between these closely related species.

In foundational studies of intertidal algae and barnacles, stress tolerance and competition differed between species, determining zonation patterns (Baker 1909; Connell 1961). The factors limiting each species are mechanistically different in our study, but the asymmetric nature of the limitation is fundamentally similar to the mechanisms of zonation found in those classic studies. For example, we found that a strong barrier limiting *M. nudatus* persistence in wetter sites, occupied by *M. guttatus*, was its inability to survive seasonal flooding (Toll and Willis 2018). Persistent flooding occurs during the winter rainy season in the seeps and meadows inhabited by *M. guttatus*. In contrast, flooding is uncommon in quickly draining washes and rocky outcrops inhabited by *M. nudatus*. An additional asymmetrical barrier limiting immigrant persistence in this system is inbreeding depression (Toll et al. 2021). Since immigrants have fewer opportunities to mate with conspecifics, they likely produce most viable seeds through selfing. While recurrent selfing reduces fitness for both species, the reduction is greater for *M. nudatus*, especially in foreign habitats. Selfing reduces the fitness of immigrant *M. nudatus* to zero in fewer generations than it does for immigrant *M. guttatus* (Toll et al. 2021). Combined, these prior studies provided a strong mechanistic rationale for why *M. nudatus* does not invade *M. guttatus* streams. However, why *M. guttatus* does not invade *M. nudatus* washes was still largely unexplained, and the experiments presented in this study suggest that frequency-dependent hybridization is a contributing factor.

Frequency Dependence of Hybridization

This study identifies frequency-dependent hybridization as a strong barrier limiting the persistence of *M. guttatus* in habitats occupied by *M. nudatus* (fig. 4*E*). This result is similar to that of reciprocal transplant experiments of *Limnanthus* species and *Gilia* subspecies, where frequency-dependent reductions in fecundity for one species (or subspecies) but not the other was attributed to asymmetric impacts of heterospecific pollen transfer on seed set (Nagy 1997; Runquist 2012; Runquist and Stanton 2012). In contrast, Christie and Strauss (2020) observed symmetric declines in seed viability of reciprocally transplanted *Streptanthus* species, even though experimental crosses reduced seed set asymmetrically (Christie and Strauss 2019). Since heterospecific pollen transfer between *M. nudatus* and *M. guttatus* reduces viable seed set symmetrically (Oneal et al. 2016), asymmetry of the frequency dependence of seed viability is surprising. However, this asymmetry can be explained when we consider how fecundity components prior to hybridization changed across immigrant frequency treatments for *M. nudatus* immigrants. *Mimulus nudatus*

immigrants produced the fewest flowers per plant in the highest immigrant frequency treatment, resulting in a lower proportion of *M. nudatus* gametes than suggested by the plant frequency treatment (~50% plants vs. ~29% flowers). *Mimulus nudatus* has fewer ovules and pollen grains per flower than *M. guttatus* (Ritland and Ritland 1989; Oneal et al. 2016). This difference in pollen production further amplified the difference in effective gamete frequency than would be expected solely by the manipulation of the frequency of *M. nudatus*. Thus, while other studies were able to predict asymmetries in frequency dependence based on laboratory crossing data (Nagy 1997; Runquist and Stanton 2012), the asymmetry we observe is driven by ecological interactions (the reduction in flower number with increasing conspecific frequency) and evolved life history differences (pollen production).

While naturally occurring *M. nudatus* also hybridizes with *M. guttatus*, seed viability was also not associated with conspecific fruit frequency in the observational transect (fig. 2). Our result differs from that of Gardner and Mac-Nair (2000), who found a significant negative correlation between fruit frequency and seed viability in *M. nudatus* co-occuring with *M. guttatus*. Our studies occurred in the same region but at qualitatively different sites. Gardner and MacNair (2000) only sampled *M. nudatus* plants with heterospecifics in close proximity, equivalent to the middle portion of our transect where both species co-occurred. Compared with *M. guttatus*, seed viability in the transition zone for *M. nudatus* was more variable (fig. 2), but since we did not record phenology or pollinator behavior over time, the factors contributing to this pattern remain open questions. While the onset of flowering differs between these species at some sites or under some experimental conditions (Wu et al. 2010; Selby et al. 2014; Sianta and Kay 2021), these species have highly overlapping flowering periods (Gardner and MacNair 2000; K. Toll, personal observation, 2013–2019). Furthermore, the relative order and even whether the onset of flowering differs between species is site and day length dependent (Friedman and Willis 2013; Toll 2017; K. Toll and J. H. Willis, unpublished data).

Our observations from the transect study suggest that fitness costs of hybridization are highly localized and decay within meters from the nearest heterospecific (fig. 2), which makes sense given the narrow boundaries that we see in most populations. In the reciprocal transplant experiment, resident transplants of both species suffered decreased seed viability when immigrant transplants were at high frequencies (figs. 4*F*, 5*F*), despite the high abundance of native conspecifics for resident transplants at each site. These results suggest that low pollinator constancy leads to high rates of interspecific pollen transfer between species at the local scale and that most pollination

is occurring at a very small scale. These localized and frequency-dependent fitness costs to residents are consistent with the observed stability of spatial segregation patterns. Immigration rates would have to be extremely high, and immigrants would have to be widespread to negatively affect resident populations.

Competition May Contribute Asymmetrically to Habitat Segregation

Competition could promote or limit coexistence depending on the degree of niche overlap between species and whether any species is competitively dominant (Chesson 2000). In the absence of strong intraspecific competition, if niche overlap between the species is high, immigrant fitness components would be positively associated with immigrant frequency, potentially resulting in competitive exclusion of the immigrant species (Gause 1934; Hardin 1960). It is possible that one species could always be competitively dominant over the other, or dominance could be context dependent and contribute asymmetrically to habitat segregation (Connell 1961; Grace and Wetzel 1981; Bertness 1991). However, since intraspecific competition is usually stronger than interspecific competition (Adler et al. 2018), rare immigrants may instead experience competitive release in habitats lacking conspecifics. If niche overlap between species is low, immigrant fitness components would be negatively associated with immigrant frequency, requiring greater fitness costs of hybridization to counterbalance the negative frequency dependence that would promote local coexistence and erode spatial segregation (Adler et al. 2007, 2018).

The *M. guttatus*/*M. nudatus* study system presented a unique opportunity to separate effects of hybridization on fitness from effects of competition and abiotic factors. While the combined effects of reproductive and ecological niche overlap could amplify positive frequency dependence and hasten exclusion (Schreiber et al. 2019), we instead observed declines in flowers per plant and seeds per flower for *M. nudatus* immigrants in their highest frequency treatment. The decline at high densities suggests that some form of competition was more important than release from reproductive interference for *M. nudatus* immigrants at high immigrant frequencies. As the transplants occurred in plastic Cone-tainers, preempting belowground competition, aboveground competition for space, light, and/or pollinator visits may have suppressed flower and seed production in these potted immigrants (fig. 1*D*). Because of the experimental design of our frequency arrays (fig. 1*C*), we were unable to estimate the strength of intra- versus interspecific competition (Inouye 2001).

Immigrant *M. nudatus* transplants produced the greatest number of seeds per flower in the intermediate immigrant frequency treatment (fig. 5*C*). We do not have a clear interpretation for this phenomenon. However, we hypothesize that pollinator behavior might have driven this U-shaped pattern in fecundity. Benadi and Pauw (2018) observed that South African Fynbos species received the highest visitation rates at intermediate flower abundances relative to rare and common species. Future studies of pollinator visitation rates are needed to test this hypothesis. *Mimulus guttatus* and *M. nudatus* are visited by many overlapping generalist bee species, but *M. nudatus* visitors tend to be small bodied, while *M. guttatus* visitors are small to large bodied (Gardner and MacNair 2000; Koski et al. 2015; K. Toll and D. B. Lowry, unpublished data).

In greenhouse studies, Brigham (2003) failed to find deleterious effects of interspecific competition to either *M. guttatus* or *M. nudatus*. In the field, the effects of intraspecific competition for *M. nudatus* varied by year, site, and the timing of germination, resulting in conspecific neighbor removals having positive, negative, or no effect on *M. nudatus* survival (Brigham 2001). In our study, we did not observe reductions in *M. guttatus* fecundity prior to hybridization (i.e., in flower or fruit set). These findings accord well with the expected competitive ability of *M. guttatus*, which occur in streams that are full of both conspecifics as well as a high density of many other plant species. The rocky serpentine washes inhabited by *M. nudatus* are mostly sparsely vegetated, with only occasional areas of high density of monkeyflowers and very few other small forbs. This is consistent with the broader pattern of serpentine endemics occurring in barer (low-competition) habitats compared with closely related species (Sianta and Kay 2019). Therefore, it is perhaps not surprising that in our arrays, which intentionally created a density somewhat intermediate and appropriate for both habitats, *M. nudatus* would suffer from competition.

Reciprocal Transplant Experiments and the Asymmetry of Barriers

Reciprocal transplant experiments provide strong evidence of local adaptation, whether between populations, ecotypes, or closely related species (Clausen et al. 1940; Kawecki and Ebert 2004; Hereford 2009). In annual plants, the response variable in these studies is usually seed set. Disentangling the relative effects of interspecific mating from environmental differences between habitats can be achieved by varying the frequencies of each species reciprocally in native and foreign sites. This allows comparisons across frequencies of fitness components that are not affected by interspecific mating interactions (e.g., survival, flower number) to those that could be (e.g., seed

abortion, hybrid seed production; Nagy 1997; Runquist and Stanton 2013; Christie and Strauss 2020).

In reciprocal transplants, the fitness of local residents and foreign immigrants are compared at native and foreign sites and changes in the rank of fitness, where local residents outperform foreign immigrants at each site, is taken as evidence of adaptation to local conditions (Kawecki and Ebert 2004). Beyond differential habitat-based selection, transplanting individuals of closely related species reciprocally reduces the distance between heterospecific individuals. This in turn increases the chances of reproductive interference. While reciprocal transplant experiments are often used to explain habitat segregation between closely related species (e.g., Yost et al. 2012; Peterson et al. 2013; Chen and Schemske 2015; Ferris and Willis 2018; DiVittorio et al. 2020), reproductive interactions are rarely considered to be important drivers of these patterns (but see Toll and Willis 2018; Christie and Strauss 2020). Since heterospecific pollen transfer can reduce seed set in a pattern indistinguishable from divergent adaptation, deleterious effects of interspecific mating on fitness need to be accounted for in assessments of local adaptation.

As demonstrated here, both divergent habitat adaptation (i.e., differences in flooding tolerance) and reproductive interference contribute to the spatial separation of the two monkeyflower species. The underlying mechanisms of resident species advantage in reciprocal transplants commonly operate asymmetrically (Popovic and Lowry 2020), and thus a pattern consistent with divergent adaptation may arise as an emergent property of several asymmetric barriers to foreign habitat persistence. Therefore, interpretation of a single fitness metric (i.e., flower production, seed set) across species is difficult without knowledge of frequency- and density-dependent processes and the underlying mechanisms. In a previous study, *M. guttatus* and *M. nudatus* were transplanted reciprocally at equal frequencies (Toll and Willis 2018) and thus minority disadvantage could not be tested for, as we did here.

Conclusion

Patterns of spatial segregation are ubiquitous across animal and plant taxa and diverse environments (Wisheu 1998). In closely related species, negative reproductive interactions can contribute to these patterns if they occur in a frequency-dependent manner and are stronger than intraspecific competition (Weber and Strauss 2016; Schreiber et al. 2019). Using both experimental and observational approaches, we found that hybridization operated in a frequency-dependent manner for one species of monkeyflower when invading habitats dominated by its congener but was not frequency dependent when the other species

invaded. Despite this asymmetry, hybridization contributed to rare immigrant disadvantage for both species. Overall, this result, combined with other results from this system (Toll and Willis 2018; Toll et al. 2021), suggests that spatial segregation is an emergent property of many asymmetric barriers operating in complementary and contrasting ways. Therefore, it seems likely that evidence of divergent adaptation from reciprocal transplant experiments of closely related species have the potential to be driven by reproductive interference, in addition to the more commonly invoked differences in stress tolerance and competitive ability. Parsing out the specific effects of each mechanism across many systems will be necessary for a fuller understanding of why closely related species segregate into different habitats.

Statement of Authorship

K.T. and D.B.L. conceived of the study. K.T. designed and performed the field experiments and collected and analyzed the data. K.T. and D.B.L. wrote and edited the manuscript.

Data and Code Availability

Data and R code are accessible through Dryad (https://doi.org/10.5061/dryad.d51c5b04g; Toll and Lowry 2022) and Zenodo (https://doi.org/10.5281/zenodo.5829342).

Literature Cited

Adler, P. B., J. HilleRisLambers, and J. M. Levine. 2007. A niche for neutrality. Ecology Letters 10:95–104.

Adler, P. B., D. Smull, K. H. Beard, R. T. Choi, T. Furniss, A. Kulmatiski, J. M. Meiners, A. T. Tredennick, and K. E. Veblen. 2018. Competition and coexistence in plant communities: intraspecific competition is stronger than interspecific competition. Ecology Letters 21:1319–1329.

Anacker, B. L., and S. Y. Strauss. 2014. The geography and ecology of plant speciation: range overlap and niche divergence in sister species. Proceedings of the Royal Society B 281:20132980.

Baack, E. J. 2005. Ecological factors influencing tetraploid establishment in snow buttercups (*Ranunculus adoneus*, Ranunculaceae):

minority cytotype exclusion and barriers to triploid formation. American Journal of Botany 92:1827–1835.

Baker, S. M. 1909. On the causes of the zoning of brown seaweeds on the seashore. New Phytologist 8:196–202.

Benadi, G., and A. Pauw. 2018. Frequency dependence of pollinator visitation rates suggests that pollination niches can allow plant species coexistence. Journal of Ecology 106:1892–1901.

Bertness, M. D. 1991. Zonation of *Spartina patens* and *Spartina alterniflora* in New England salt marsh. Ecology 72:138–148.

Buggs, R. J. A., and J. R. Pannell. 2006. Rapid displacement of a monoecious plant lineage is due pollen swamping by a dioecious relative. Current Biology 16:996–1000.

Burns, J. H., and S. Y. Strauss. 2011. More closely related species are more ecologically similar in an experimental test. Proceedings of the National Academy of Sciences of the USA 108: 5302–5307.

Brigham, C. A. 2001. Maintenance of diversity: mechanisms of persistence in rare plants. PhD. diss. University of California, Davis.

———. 2003. Factors affecting persistence in formerly common and historically rare plants. Pages 59–97 in C. A. Brigham and M. W. Schwartz, eds. Population viability in plants: conservation, management, and modeling of rare plants. Springer Science and Business Media, New York.

Brooks, M. E., K. Kristensen, K. J. van Benthem, A. Magnusson, C. W. Berg, A. Nielsen, H. J. Skaug, M. Maechler, and B. M. Bolker. 2017. glmmTMB balances speed and flexibility among packages for zero-inflated generalized linear mixed modeling. R Journal 9:378–400.

Brown, W. L., and E. O. Wilson. 1956. Character displacement. Systematic Zoology 7:49–64.

Chen, G. F., and D. W. Schemske. 2015. Ecological differentiation and local adaptation in two sister species of Neotropical *Costus* (Costaceae). Ecology 96:440–449.

Chesson, P. 2000. Mechanisms of maintenance of species diversity. Annual Review of Ecology and Systematics 31:343–366.

Christie, K., and S. Y. Strauss. 2019. Reproductive isolation and the maintenance of species boundaries in two serpentine endemic jewelflowers. Evolution 73:1375–1391.

———. 2020. Frequency-dependent fitness and reproductive dynamics contribute to habitat segregation in sympatric jewelflowers. Proceedings of the Royal Society B 287:20200559.

Clausen, J., D. D. Keck, and H. M. Hiesey. 1940. Experimental studies on the nature of species. I. Effect of varied environment on western North American plants. Publication no. 520. Carnegie Institution of Washington, Washington, DC.

Connell, J. H. 1961. The influence of interspecific competition and other factors on the distribution of the barnacle *Chthamalus stellatus*. Ecology 42:710–723.

Coughlan, J. M., M. W. Brown, and J. H. Willis. 2020. Patterns of hybrid seed inviability in the *Mimulus guttatus* sp. complex reveal a potential role of parental conflict in reproductive isolation. Current Biology 30:83–93.

DiVittorio, C. T., S. Singhal, A. B. Roddy, F. Zapata, D. D. Ackerly, B. G. Baldwin, C. R. Brodersen, et al. 2020. Natural selection maintains species despite frequent hybridization in the desert shrub *Encelia*. Proceedings of the National Academy of Sciences of the USA 117:33373–33383.

Ferris, K. G., and J. H. Willis. 2018. Differential adaptation to a harsh granite outcrop habitat between sympatric *Mimulus* species. Evolution 72:1225–1241.

Fox, J., and S. Weisberg. 2019. An R companion to applied regression. 3rd ed. Sage, Thousand Oaks, CA.

Friberg, M., O. Leimar, and C. Wiklund. 2013. Heterospecific courtship, minority effects and niche separation between cryptic butterfly species. Journal of Evolutionary Biology 26:971–979.

Friedman, J., and J. H. Willis. 2013. Major QTLs for critical photoperiod and vernalization underlie extensive variation in flowering in the *Mimulus guttatus* species complex. New Phytologist 199:571–583.

Gardner, M., and M. Macnair. 2000. Factors affecting the coexistence of the serpentine endemic *Mimulus nudatus* Curran and its presumed progenitor, *Mimulus guttatus* Fischer ex DC. Biological Journal of the Linnean Society 69:443–459.

Garner, A. G., A. M. Kenney, L. Fishman, and A. L. Sweigart. 2016. Genetic loci with parent-of-origin effects cause hybrid seed lethality in crosses between *Mimulus* species. New Phytologist 211:319–331.

Gause, G. F. 1934. The struggle for existence. Williams & Wilkins, Baltimore.

Grace, J. B., and R. G. Wetzel. 1981. Habitat partitioning and competitive displacement in cattails (*Typha*): experimental field studies. American Naturalist 118:463–474.

Grainger, T. N., J. M. Levine, and B. Gilbert. 2019. The invasion criterion: a common currency for ecological research. Trends in Ecology and Evolution 34:925–935.

Groning, J., and A. Hochkirch. 2008. Reproductive interference between animal species. Quarterly Review of Biology 83:257–282.

Hardin, G. 1960. The competitive exclusion principle. Science 131: 1292–1297.

Hartig, F. 2021. DHARMa: residual diagnostics for hierarchical (multi-level/mixed) regression models. R package version 0.4.4.

Hereford, J. 2009. A quantitative survey of local adaptation and fitness trade-offs. American Naturalist 173:579–588.

Husband, B. C. 2000. Constraints on polyploid evolution: a test of the minority cytotype exclusion principle. Proceedings of the Royal Society B 267:217–223.

Inouye, B. D. 2001. Response surface experimental designs for investigating interspecific competition. Ecology 82:2696–2706.

Kawecki, T. J., and D. Ebert. 2004. Conceptual issues in local adaptation. Ecology Letters 7:1225–1241.

Koski, M. H., G. A. Meindl, G. Arceo-Gomez, M. Wolowski, K. A. Lecroy, T.-L. Ashman. 2015. Plant-flower visitor networks in a serpentine metacommunity: assessing traits associated with keystone plant species. Arthropod-Plant Interactions 9:9–21.

Kuno, E. 1992. Competitive exclusion through reproductive interference. Researches on Population Ecology 34:275–284.

Kyogoku, D., and H. Kokko. 2020. Species coexist more easily if reinforcement is based on habitat preferences than on species recognition. Journal of Animal Ecology 89:2605–2616.

Kyogoku, D., and D. Wheatcroft. 2020. Heterospecific mating interactions as an interface between ecology and evolution. Journal of Evolutionary Biology 33:1330–1344.

Lenth, R. V. 2022. emmeans: estimated marginal means, aka least-squares means. R package version 1.7.2.

Levin, D. A. 1975. Minority cytotype exclusion in local plant populations. Taxon 24:35–43.

Levin, D. A., and W. W. Anderson. 1970. Competition for pollinators between simultaneously flowering species. American Naturalist 104:455–467.

Lewis, H. 1961. Experimental sympatric populations of *Clarkia*. American Naturalist 95:155–168.

Lüdecke, D. 2018. ggeffects: tidy data frames of marginal effects from regression models. Journal of Open Source Software 3:772.

Mitchell, R. J., R. J. Flanagan, B. J. Brown, N. M. Waser, and J. D. Karron. 2009. New frontiers in competition for pollination. Annals of Botany 103:1403–1413.

Nagy, E. S. 1997. Frequency-dependent seed production and hybridization rates: implications for gene flow between locally adapted plant populations. Evolution 51:703–714.

Oneal, E., J. H. Willis, and R. G. Franks. 2016. Disruption of endosperm development is a major cause of hybrid seed inviability between *Mimulus guttatus* and *Mimulus nudatus*. New Phytologist 210:1107–1120.

Pederson, T. L. 2020. patchwork: the composer of plots. R package version 1.1.1.

Peterson, M., K. J. Rice, and J. P. Sexton. 2013. Niche partitioning between close relatives suggests trade-offs between adaptation to local environments and competition. Ecology and Evolution 3:512–522.

Popovic, D., and D. B. Lowry. 2020. Contrasting environmental factors drive local adaptation at opposite ends of an environmental gradient in the yellow monkeyflower (*Mimulus guttatus*). American Journal of Botany 107:298–307.

R Core Team. 2021. R: a language and environment for statistical computing. R Foundation for Statistical Computing, Vienna.

Ribeiro, J. M. C., and A. Spielman. 1986. The Satyr effect: a model predicting parapatry and species extinction. American Naturalist 128:513–528.

Ritland, C., and K. Ritland. 1989. Variation of sex allocation among eight taxa of the *Mimulus guttatus* species complex (Scrophulariaceae). American Journal of Botany 1989:1731–1739.

Runquist, R. B. 2012. Pollinator-mediated competition between two congeners, *Limnanthes douglasii* subsp. *rosea* and *L. alba* (Limnanthaceae). American Journal of Botany 99:1125–1132.

Runquist, R. B., and M. L. Stanton. 2013. Asymmetric and frequency-dependent pollinator-mediated interactions may influence competitive displacement in two vernal pool plants. Ecology Letters 16:183–190.

Sandstedt, G. D., C. A. Wu, and A. L. Sweigart. 2021. Evolution of multiple postzygotic barriers between species of the *Mimulus tilingii* complex. Evolution 75:600–613.

Schreiber, S. J., M. Yamamichi, and S. Y. Strauss. 2019. When rarity has costs: coexistence under positive frequency-dependence and environmental stochasticity. Ecology 100:e02664.

Selby, J. P., A. L. Jeong , K. Toll, K. M. Wright, and D. B. Lowry . 2014. Methods and discoveries in the pursuit of understanding the genetic basis of adaptation to harsh environments in *Mimulus*. Pages 243–265 in N. Rajakaruna, ed. Plant ecology and evolution in harsh environments. NOVA Science, Hauppauge, NY.

Selby, J. P., and J. H. Willis. 2018. Major QTL controls adaptation to serpentine soils in *Mimulus guttatus*. Molecular Ecology 27:5073–5087.

Servedio, M. R., and M. A. F. Noor. 2003. The role of reinforcement in speciation: theory and data. Annual Review of Ecology, Evolution and Systematics 34:339–364.

Sianta, S. A., and K. M. Kay. 2019. Adaptation and divergence in edaphic specialists and generalists: serpentine soil endemics in the California flora occur in barer serpentine habitats with lower soil calcium levels than serpentine tolerators. American Journal of Botany 106:690–703.

———. 2021. Parallel evolution of phenological isolation across the speciation continuum in serpentine-adapted annual wildflowers. Proceedings of the Royal Society B 288:20203076.

Singer, F. 1990. Reproductive costs arising from incomplete habitat segregation among three species of *Leucorrhinia* dragonflies. Behaviour 115:188–201.

Takakura, K. I. 2013. Two-way but asymmetrical reproductive interference between an invasive *Veronica* species and a native congener. American Journal of Plant Sciences 4:535–542.

Takakura, K. I., and S. Fujii. 2010. Reproductive interference and salinity tolerance differentiate habitat use between two alien cockleburs: *Xanthium occidentale* and *X. italicum* (Compositae). Plant Ecology 206:309–319.

———. 2015. Island biogeography as a test of reproductive interference. Population Ecology 57:307–319.

Takakura, K. I., T. Nishida, T. Matsumoto, and S. Nishida. 2009. Alien dandelion reduces the seed-set of a native congener through frequency-dependent and one-sided effects. Biological Invasions 11:973–981.

Thum, R. A. 2007. Reproductive interference, priority effects and the maintenance of parapatry in *Skistodiaptomus* copepods. Oikos 116:759–768.

Toll, K. 2017. Mechanisms of habitat segregation between co-occurring monkeyflower species. PhD diss. Duke University, Durham, NC.

Toll, K., E. F. LoPresti, and D. B. Lowry. 2021. Inbreeding depression contributes to the maintenance of habitat segregation between closely related monkeyflower species. Evolution 75:832–846.

Toll, K., and D. B. Lowry. 2022. Data from: Frequency-dependent hybridization contributes to habitat segregation in monkeyflowers. American Naturalist, Dryad Digital Repository, https://doi.org/10.5061/dryad.d51c5b04g.

Toll, K., and J. H. Willis. 2018. Hybrid inviability and differential submergence tolerance drive habitat segregation between two congeneric monkeyflowers. Ecology 99:2776–2786.

Vickery, R. K. 1978. Case studies in the evolution of species complexes in *Mimulus*. Evolutionary Biology 11:405–507.

Waser, N. M. 1978*a*. Competition for hummingbird pollination and sequential flowering in two Colorado wildflowers. Ecology 59:934–944.

———. 1978*b*. Interspecific pollen transfer and competition between co-occurring plant species. Oecologia 36:223–236.

Weber, M. G., and S. Y. Strauss. 2016. Coexistence in close relatives: beyond competition and reproductive isolation in sister taxa. Annual Review of Ecology, Evolution, and Systematics 47:359–381.

Wickham, H. 2016. ggplot2: elegant graphics for data analysis. Springer, New York.

Wisheu, I. C. 1998. How organisms partition habitats: different types of community organization can produce identical patterns. Oikos 83:246–258.

Wu, C. A., D. B. Lowry, L. I. Nutter, and J. H. Willis. 2010. Natural variation for drought-response traits in the *Mimulus guttatus* species complex. Oecologia 162:23–33.

Yost, J. M., T. Barry, K. M. Kay, and N. Rajakaruna. 2012. Edaphic adaptation maintains the coexistence of two cryptic species on serpentine soils. American Journal of Botany 99:890–897.

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