

RESEARCH ARTICLE

Nitrogen addition promotes terrestrial plants to allocate more biomass to aboveground organs: A global meta-analysis

Huili Feng^{1,2}  | Jiahuan Guo^{1,2}  | Changhui Peng^{3,4}  | Daniel Kneeshaw³  |
Gabrielle Roberge³  | Chang Pan²  | Xuehong Ma²  | Dan Zhou²  |
Weifeng Wang² 

¹Key Laboratory of Ministry of Education for Genetics and Germplasm Innovation of Tropical Special Trees and Ornamental Plants/Hainan Biological Key Laboratory for Germplasm Resources of Tropical Special Ornamental Plants, College of Forestry, Hainan University, Haikou, China

²Co-Innovation Center for Sustainable Forestry in Southern China, College of Biology and the Environment, Nanjing Forestry University, Nanjing, China

³Department of Biological Sciences, University of Quebec at Montreal, Montreal, Quebec, Canada

⁴College of Geographic Science, Hunan Normal University, Changsha, China

Correspondence

Weifeng Wang, Co-Innovation Center for Sustainable Forestry in Southern China, College of Biology and the Environment, Nanjing Forestry University, Nanjing, Jiangsu 210037, China.

Email: wang.weifeng@njfu.edu.cn

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Abstract

A significant increase in reactive nitrogen (N) added to terrestrial ecosystems through agricultural fertilization or atmospheric deposition is considered to be one of the most widespread drivers of global change. Modifying biomass allocation is one primary strategy for maximizing plant growth rate, survival, and adaptability to various biotic and abiotic stresses. However, there is much uncertainty as to whether and how plant biomass allocation strategies change in response to increased N inputs in terrestrial ecosystems. Here, we synthesized 3516 paired observations of plant biomass and their components related to N additions across terrestrial ecosystems worldwide. Our meta-analysis reveals that N addition (ranging from 1.08 to 113.81 g m⁻² year⁻¹) increased terrestrial plant biomass by 55.6% on average. N addition has increased plant stem mass fraction, shoot mass fraction, and leaf mass fraction by 13.8%, 12.9%, and 13.4%, respectively, but with an associated decrease in plant reproductive mass (including flower and fruit biomass) fraction by 3.4%. We further documented a reduction in plant root-shoot ratio and root mass fraction by 27% (21.8%–32.1%) and 14.7% (11.6%–17.8%), respectively, in response to N addition. Meta-regression results showed that N addition effects on plant biomass were positively correlated with mean annual temperature, soil available phosphorus, soil total potassium, specific leaf area, and leaf area per plant. Nevertheless, they were negatively correlated with soil total N, leaf carbon/N ratio, leaf carbon and N content per leaf area, as well as the amount and duration of N addition. In summary, our meta-analysis suggests that N addition may alter terrestrial plant biomass allocation strategies, leading to more biomass being allocated to aboveground organs than belowground organs and growth versus reproductive trade-offs. At the global scale, leaf functional traits may dictate how plant species change their biomass allocation pattern in response to N addition.

KEYWORDS

biomass partitioning, leaf traits, plant biomass, reproductive trade-offs, terrestrial ecosystem

1 | INTRODUCTION

Plant growth and development depend on environmental variables, such as temperature, light intensity, and the availability of carbon, water, and mineral nutrients (Hermans et al., 2006). As a plant captures carbon and nutrients, the newly formed biomass is allocated to leaves, stems (including bole, branch, and bark), roots, or reproductive parts (Reich et al., 2014). To sustain the required physiological functions and grow normally, a plant must balance the distribution of biomass among its leaves, branches, stems, and roots (Shipley & Meziane, 2002; Yin et al., 2019). Modifying biomass allocation is one of the most fundamental ways to maximize plant growth rate, survival, and adaptability under various biotic and abiotic stresses (Poorter et al., 2015; Puglielli et al., 2021). Consequently, altered biomass allocation has profound implications for plant growth.

Nitrogen (N) is a necessary nutrient for plant growth and ecosystem production in both terrestrial and marine habitats (Elser et al., 2007). N deficit inhibits plant growth and development, lowers photosynthesis and leaf area, accelerates plant senescence, and eventually diminishes plant production (Mu & Chen, 2021). Therefore, N fertilization has been widely used to promote plant growth and production (Frink et al., 1999; Xia & Wan, 2008). Increased reactive N emissions through agricultural fertilization and fossil fuel combustion have resulted in a large amount of atmospheric inorganic N deposition at a global average rate of 105 Tg N year⁻¹ (Galloway et al., 2008; Lu et al., 2011). Global N deposition rates are expected to increase 2.5-fold by the end of the century (Janssens et al., 2010; Lamarque et al., 2005). As a result, N addition to terrestrial ecosystems is now considered one of the most widespread drivers of global change (Galloway et al., 2008; Humbert et al., 2016).

In most terrestrial ecosystems, plant growth is typically restricted by soil N availability (LeBauer & Treseder, 2008; Lu et al., 2011). Some previous studies have shown a strong covariation between soil N availability and plant biomass allocation in natural ecosystems (Cambui et al., 2011; Hermans et al., 2006). Therefore, plants, as a fundamental component of the terrestrial ecosystem, may be sensitive to enhanced N deposition in their growth (Bobbink et al., 2010; Phoenix et al., 2012). However, there is still much uncertainty about whether or how plant biomass allocation will change in response to increased N inputs in terrestrial ecosystems. A previous study has shown that plants adapt to environmental changes by changing biomass distribution across different plant organs (Hermans et al., 2006). Besides, leaf functional traits such as leaf carbon and N content (LCC and LNC), specific leaf area (SLA), and leaf area per plant (PLA) have a significant impact on the efficiency of assimilation of photosynthetic active radiation, and hence, plant photosynthetic capacity (Marron et al., 2005; Scott Green et al., 2003). Some research has suggested that leaf functional traits can also be used as potential covariates for understanding biomass allocation (Li et al., 2021; Yin et al., 2019). However, it is unclear whether leaf functional traits and environmental factors jointly drive the response of plant biomass allocation to N addition. Therefore, a quantitative synthesis across multiple studies is urgently needed to quantify the

effects of reactive N addition on terrestrial plant biomass and its composition and identify the key drivers accurately.

Previous manipulation experiments have improved our understanding of N addition effects on the allocation of plant biomass (Högberg et al., 2006; Holub & Tůma, 2010; Palmqvist & Dahlman, 2006). However, studies that rely on individual cases or are constrained to specific habitats make it difficult to get a complete picture of the effects of N addition on terrestrial plants. Moreover, the role of global change in promoting the feedback of terrestrial ecosystems and atmospheric N deposition has not been fully elucidated. Previous research has reported the effects of N deposition on plant photosynthesis (Liang et al., 2020), plant diversity (Bobbink et al., 2010), and plant species richness and abundance (Midolo et al., 2019); however, due to a lack of data, a synthesis was not produced. To date, although several studies have attempted to quantify N addition effect on plant biomass (Li et al., 2020; Yue et al., 2020), none of these studies considered the effects of environmental variables (e.g., climate, soil, and N deposition background values) and plant leaf functional traits (such as SLA, PLA, and leaf carbon/nitrogen ratio [LCN]) on the response of terrestrial plants to N addition. Since N availability affects carbon sequestration across terrestrial ecosystems (Gruber & Galloway, 2008), a better understanding of how and to what extent N addition affects biomass allocation in terrestrial plants is essential for accurate quantification of carbon sequestration.

In this study, we compiled a global dataset of 3516 paired sets of experimental field observations from 255 studies covering the majority of the dominant plant species in terrestrial ecosystems. The objectives of this study were to (1) clarify the effects of N addition on plant biomass in terrestrial ecosystems and (2) identify the key drivers and global patterns of biomass accumulation and allocation in response to reactive N addition. We hypothesized that (i) N addition promotes plants to allocate more biomass to vegetative (i.e., roots, stems, and leaves) rather than reproductive (i.e., flowers and fruits) organs and (ii) environmental factors and leaf functional traits jointly drive the response of plant biomass to N addition.

2 | MATERIALS AND METHODS

2.1 | Data collection

We systematically searched the peer-reviewed published literature from the ISI Web of Science and Google Scholar databases for specific search terms (Table S1; Figure S1). Each paper was then screened to determine whether it met the following criteria: (1) the climatic, soil, and vegetation variables (or conditions) were the same between the control and the N addition treatment sites (i.e., studies along gradients of N deposition were not considered); (2) the research only included species that naturally exist in terrestrial habitats; (3) means, standard deviations (SDs), or errors (SEs), and the sampling sizes for both the control and treatment groups could be directly obtained or could be calculated; (4) the statistical assumption of independence

across observations for the long-term N addition study were met and, only the most recent results were included; (5) only data from control and N addition plots were collected when global change treatments (e.g., added CO₂ or increased temperature) other than N addition were tested; and (6) data from transgenic plants or plants that had been handled with herbicides, hormones, and/or heavy metals were eliminated since they might have presented a different pattern of biomass distribution. Subsequently, if a publication included several experiments in different locations, these observations were considered independent cases. As supplements, articles used in previous relevant meta-analyses (Liang et al., 2020; Xia & Wan, 2008; Yue et al., 2020) were also included and rescreened according to the above criteria.

For the literature that meets our criteria, we acquired the mean values, SD, and sample size directly from the table, text, and supplementary files or extracted these data indirectly from figures by using GetData software (version 2.26). We also collected geographic variables (i.e., latitude, longitude, and elevation), environmental variables (i.e., MAT, mean annual temperature; MAP, mean annual precipitation), and other reported variables including experimental duration (years, the number of years with N applied repeatedly), the magnitude of N applied (quantity of N addition per year, ranging from 0.3 to 112 g m⁻² year⁻¹), and the form of N applied (the type or types of N fertilizer). If N addition varied during the N addition experiment, we estimated the total applied dosage based on the amount of N fertilizer used and the duration of the treatment (total N application/duration; Xu et al., 2021). Data on atmospheric inorganic N (ranging from 0.02 to 2.52 g m⁻² year⁻¹) were acquired from global N deposition maps (Ackerman et al., 2018). Other climatic data such as precipitation in the warmest quarter (WQP) were obtained from WorldClim2 at 1 km spatial resolution (Fick & Hijmans, 2017).

For each species, plants were classified to the family and genus level, and grouped according to their photosynthesis pathway (e.g., C3, C4, or CAM), functional types (i.e., tree, shrub, grass, forb, fern, lichen, and moss), plant N fixation capacity (N-fixing plant or non-N-fixing plant), and life cycles (i.e., annual, biennial, or perennial). As categorical predictors, seven commonly measured plant leaf functional traits were obtained from the TRY plant trait database for each species (Kattge et al., 2020): SLA (mm² mg⁻¹), PLA (m²), LCN, leaf nitrogen/phosphorus (N/P) ratio (LNP), leaf photosynthesis rate (LPR, μmol m⁻² s⁻¹), LCC (g m⁻²), and LNC (g m⁻²). We averaged every functional trait observed for each species after excluding duplicate functional trait measurements and outliers (values larger than triple standard deviations from the trait mean for each species; Adler et al., 2014). As continuous predictors, four soil variables including soil total nitrogen (TN, %), soil total potassium (TK, %), soil available phosphorus (AP, mg kg⁻¹), soil pH, and aridity index (AI) were extracted from Global Soil Datasets in Earth System Models (GSDE; Shangguan et al., 2014) and the CGIAR Consortium for Spatial Information (CGIAR-CSI) dataset (Antonio & Robert, 2019) at a 30 arc-second (approximately 1 km²) spatial resolution. Finally, we compiled a database from 255 published articles, containing a total of 3516 paired observations of 401 terrestrial plant species (see Figure 1; Dataset S1).

2.2 | Data analysis

We used the log response ratio (RR) to conduct a meta-analysis to assess how plant biomass and biomass allocation responded to N addition in terrestrial ecosystems (Hedges et al., 1999). The following is how the RR was calculated:

$$RR = \ln(\bar{X}_t / \bar{X}_c), \quad (1)$$

where $\bar{X}_t$ and $\bar{X}_c$ were the means of each observation for the treatment (i.e., elevated N) and the control (i.e., ambient N) groups, respectively. The sampling variation (v) for each RR was calculated as:

$$v = \frac{S_t^2}{N_t \bar{X}_t^2} + \frac{S_c^2}{N_c \bar{X}_c^2}, \quad (2)$$

where N_t and N_c represent the sample size of the treatment and control groups, S_t and S_c were the standard deviation of the treatment and control groups.

For each study, the weighting factor (w) was calculated as:

$$w = \frac{1}{v + \tau^2}, \quad (3)$$

where τ^2 was the estimated between-study variance of estimates. Following that, the overall effects ($\overline{RR}$) for all observations was estimated as:

$$\overline{RR} = \frac{\sum_{i=1}^m \sum_{j=1}^n w_{ij} \ln RR_{ij}}{\sum_{i=1}^m \sum_{j=1}^n w_{ij}}, \quad (4)$$

where m ($i = 1, 2, \dots, m$) and n ($j = 1, 2, \dots, n$) indicate the number of groups (e.g., functional type) and the number of observations in the i th group, respectively. For a better interpretation, the mean effect size was transformed back to the percentage change induced by the experimental N addition treatment (Bai et al., 2013):

$$\text{Percentage (\%)} = [1 - \exp(RR)] \times 100. \quad (5)$$

Then, the multilevel mixed-effect meta-analysis was conducted with the `rma.mv` function in the “metafor” R package (Viechtbauer, 2010) to determine whether N addition had a substantial impact on the biomass accumulation and biomass allocation of terrestrial plants. The effects of the treatment were considered significantly different from the control if their 95% confidence intervals did not include zero (Hedges & Olkin, 1985). The heterogeneity of effect sizes was tested with the Q statistic to determine whether the variability of the observed effect sizes was greater than that anticipated by chance (Cauvy-Fraunié & Dangles, 2019). As projected, a significant residual heterogeneity was found in the mixed effect meta-analysis for the N treatment on plant biomass ($Q_t = 160,533.5$, $p < .0001$) and biomass allocation ($Q_t = 18,879.8$,

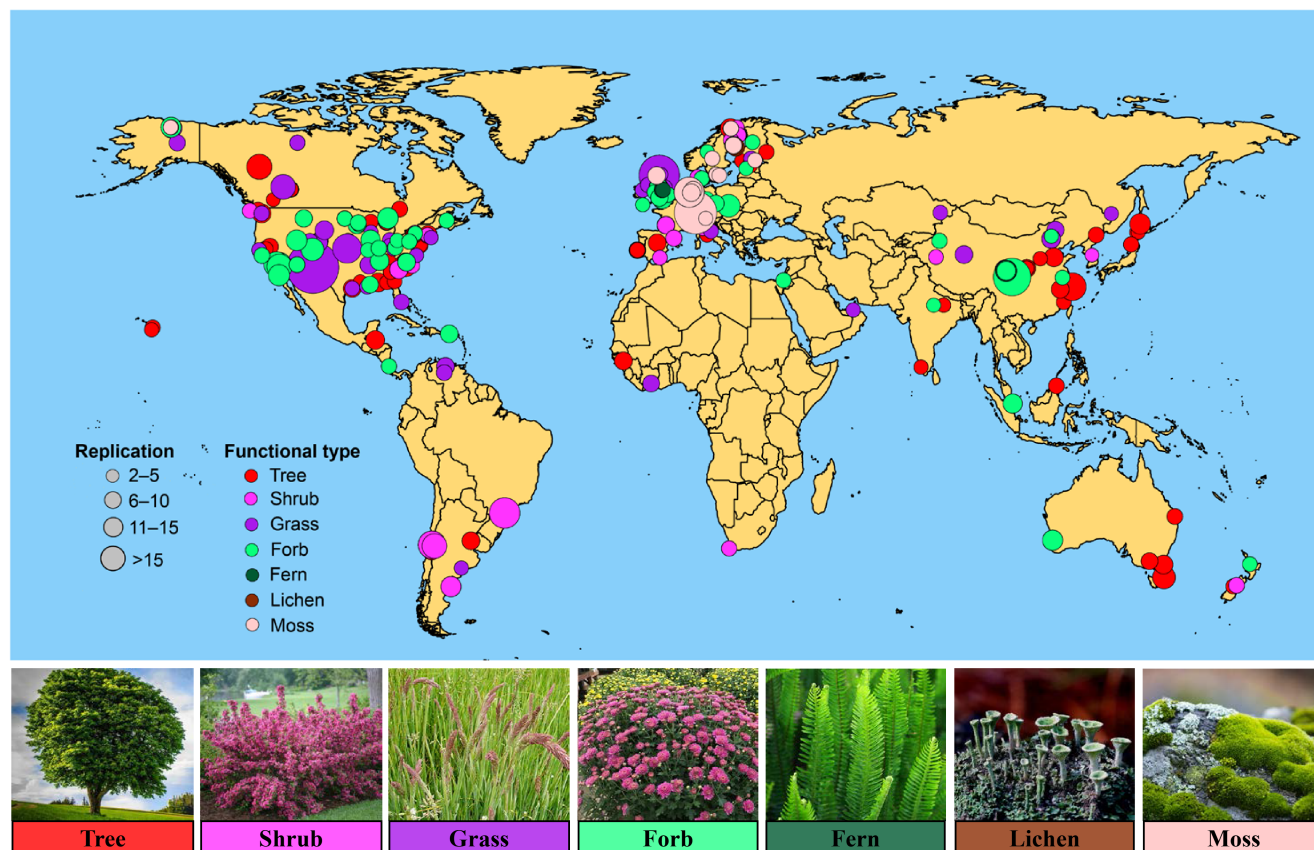


FIGURE 1 Geographical location of study sites in the meta-analysis. The point sizes represent the number of replications. Red, magenta, purple, lime, green, saddle brown, and pink points indicate functional types of tree, shrub, grass, forb, fern, lichen, and moss. Map lines delineate study areas and do not necessarily depict accepted national boundaries. [Colour figure can be viewed at [wileyonlinelibrary.com](https://onlinelibrary.wiley.com/doi/10.1111/gcb.16731)]

$p < .0001$), which we tried to interpret with various moderators (Table S2).

In addition, we used single mixed-effect meta-regression modeling to investigate the link across effect sizes with moderators. The explained heterogeneity Q statistic (Q_m) was developed to assess for significance in single covariance meta-regressions (Guo et al., 2023). A significant Q_m denotes a statistic in which moderators contribute to the heterogeneity in effect sizes (Cauvy-Fraunié & Dangles, 2019). All parameter estimates reported in this meta-analysis were from the best models fitted by a restricted maximum likelihood approach.

To avoid overfitting, we used the chart.Correlation function in the “PerformanceAnalytics” R package (Peterson & Carl, 2020) to perform Pearson correlation analysis of the predictors to test for multi-collinearity, and then exclude variables with correlation coefficients larger than .7 (Dormann et al., 2013). As a sensitivity test, we ran a model-selection procedure with the maximum likelihood method (Terrer et al., 2021). We used the glmulti function in the “glmulti” package (Calcagno & de Mazancour, 2010) to list all alternative models based on the corrected Akaike information criterion (AICc) values. All models with a $\Delta AICc$ value less than 2 were considered equivalent to the best-fitting model (Guo et al., 2023; Table S3).

The robustness of our findings to publication bias was evaluated using the funnel plots, Egger's test (Egger et al., 1997), and Rosenberg's fail-safe numbers (Rosenberg, 2005). The funnel plots found no

evidence for funnel asymmetry (Figure S2). Egger's test suggested that the overall effect sizes for N treatment on plant biomass (including biomass accumulation and biomass allocation) were robust (Egger's test: $z = .176$, $p = .860$). Rosenberg's fail-safe numbers were large enough (fail-safe N: 3.86×10^7 , $p < .0001$) to be convincing with respect to the estimate's robustness. As a consequence, we did not consider publication bias to be a problem in interpreting the findings. All analyses were carried out using the R 4.1.3 software (R Core Team, 2022).

3 | RESULTS

3.1 | Effects of nitrogen addition on plant biomass accumulation

For all of the terrestrial plants evaluated in this study, N addition increased biomass by 55.6% (back-transformed 95% CI: 51.4–59.8; Figure 2). The biomass of N-fixing plants increased by 53.6% (41.9%–65.4%), which was more variable but on average only slightly less than that of non-N-fixing plants 55.9% (51.4%–60.4%; Figure 3; Table S4). The greatest increase in biomass was for deciduous trees and broadleaved trees which increased by 74.2% (63.3%–85.2%) and 81.3% (70.8%–91.9%), respectively. This was considerably higher than evergreen trees and coniferous trees, which both increased by 57.1% (48%–66.2%) and 40.2%

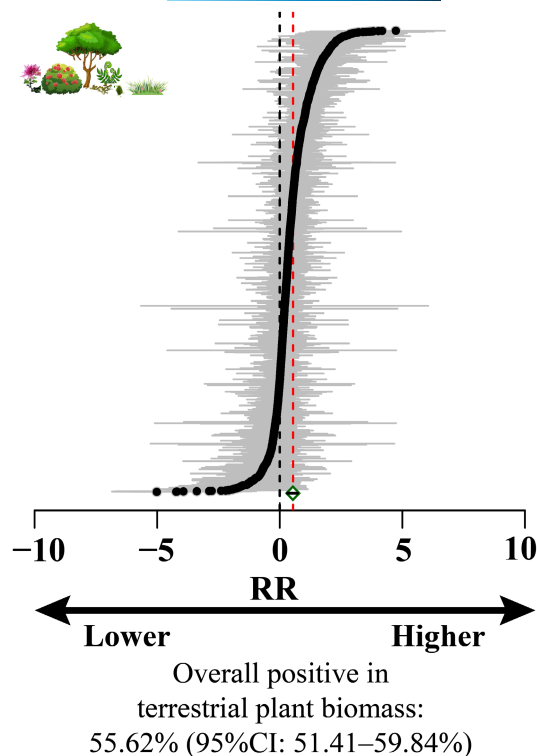


FIGURE 2 Overall effect size estimates of N addition on terrestrial plant biomass at a global scale. RR, log response ratios (effect sizes). RR=0, black dashed line; overall effect size, red dashed line, and green diamond (with 95% CI in black line). [Colour figure can be viewed at [wileyonlinelibrary.com](https://onlinelibrary.wiley.com/doi/10.1111/gcb.16731)]

(33.2%–47.2%). For different photosynthesis pathways, N addition increased biomass by 55.2% (50.8%–59.6%) in C3 plants and by 81.5% (56.7%–106.3%) in C4 plants. Our evaluation by life history traits showed that N addition increased the biomass of annual herbs and perennial herbs by 63.7% (50.0%–77.4%) and 55.5% (48.3%–62.6%), respectively.

By the growth forms of plants, we observed that N addition increased the biomass of herbaceous plants and woody plants by 60.1% (53.6%–66.6%) and 53.8% (48.2%–59.5%), respectively (Figure 4; Table S5). Our evaluation of biological realms of plants showed that N addition increased the biomass of seed plants by 56.6% (52.3%–60.9%) but had no effects on spore plants (Figure 3; Table S4). Furthermore, the effects of N addition differed among taxonomic groups, with biomass increasing by 65% for trees (57.9%–72%), 33% for shrubs (24%–42%), 71.4% for grasses (62.4%–80.4%), and 42.7% for forbs (33.6%–51.8%), but had no effects on lichens and mosses.

3.2 | Effects of nitrogen addition on plant biomass allocation

Overall, the meta-analysis demonstrated that N addition had different effects on different functional organs (or tissues) of plants (Figure 5). For the whole plant, N addition increased the biomass by 76.6% (Figure 5a; Table S6). Furthermore, more biomass appeared to be allocated to aboveground (61.7%) than belowground (i.e., root; 42.7%) biomass in

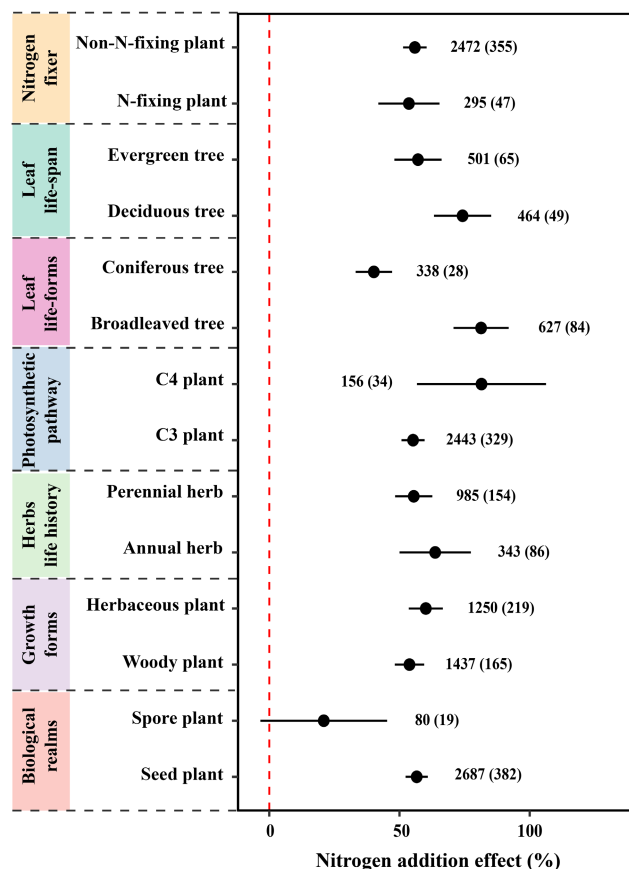


FIGURE 3 Effect size of N addition on terrestrial plant biomass, expressed on the global scale (RR [$\pm$ 95% confidence interval]). The vertical dashed red line denotes a null effect size (RR=0). The number of observations is beside each attribute without parentheses, and the number of plant species is in parentheses. RR, log response ratio. [Colour figure can be viewed at [wileyonlinelibrary.com](https://onlinelibrary.wiley.com/doi/10.1111/gcb.16731)]

response to N addition. Within the components of the aboveground plant structure, plant shoots showed the strongest positive response to N addition, with biomass increasing by 67.3%. Plant leaf and stem biomass, on the other hand, were increased by the addition of N, by 33.2% and 37.2%, respectively. The addition of N increased the mass fractions of stem, shoot, and leaf by 13.8%, 12.9%, and 13.4%, respectively (Figure 5b; Table S7). In contrast, plant root-shoot ratio and root mass fraction showed a negative response to N addition, which decreased by 27% and 14.7%, respectively. Although there were no clear directional effects on the reproductive (i.e., flowers and fruits) biomass of plants, more biomass from plants was allocated to vegetative organs than to reproductive organs. And we did note that plant reproductive mass fraction decreased by 3.4% in response to N addition.

3.3 | Moderators of nitrogen addition effects on plant biomass

The meta-regression results revealed that N addition effects on terrestrial plant biomass were positively correlated with MAT

($p < .0001$), AP ($p < .01$), and TK ($p < .01$; Table 1). On the contrary, N addition effects on plant biomass showed a significantly negative correlation with soil TN ($p < .0001$), the duration of the N addition

($p < .0001$), and the amount of N addition ($p < .01$). Furthermore, there was a negative relationship between N addition effects on plant biomass and latitude ($p < .05$). For plant functional traits, the effect size of N addition on plant biomass was positively correlated with SLA ($p < .0001$) and PLA ($p < .01$) and negatively correlated with LCN ($p < .01$), LCC ($p < .0001$), and LNC ($p < .0001$).

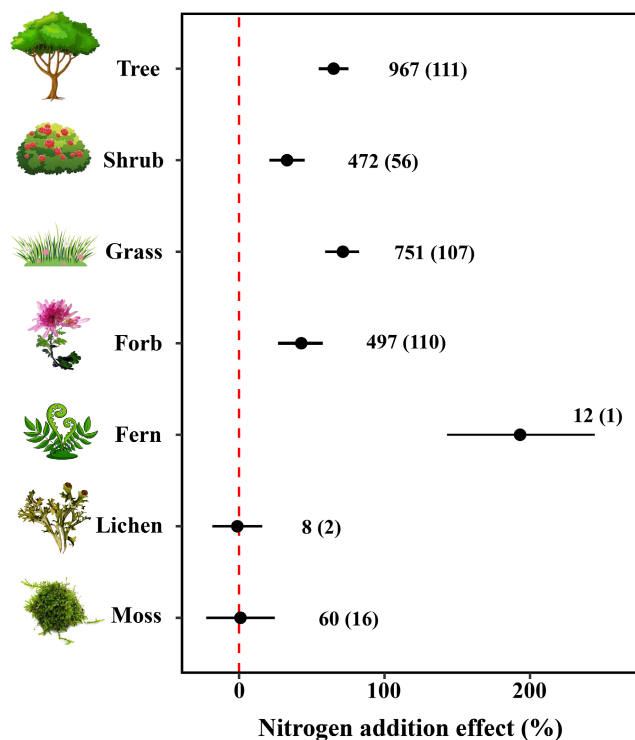


FIGURE 4 Effect size of N addition on terrestrial plant biomass of different growth forms, expressed on the global scale (RR [$\pm 95\%$ confidence interval]). The vertical dashed red line denotes a null effect size (RR=0). The number of observations is beside each attribute without parentheses, and the number of plant species is in parentheses. RR, log response ratio. [Colour figure can be viewed at [wileyonlinelibrary.com](https://onlinelibrary.wiley.com/doi/10.1111/gcb.16731)]

4 | DISCUSSION

4.1 | Mechanisms of plant biomass allocation in response to nitrogen addition

For most parts of the plants, N resources are limited. Reduced photosynthesis occurs in N-deficient plants due to sugar accumulation, as sugar exerts metabolite feedback and influences many genes responsible for photosynthesis (Bläsing et al., 2005). N addition can therefore reduce sugar accumulation in leaves (Hermans et al., 2006), increase N allocation to photosynthetic apparatus (Li et al., 2019), and promote biomass allocation to vegetative organs (Yan et al., 2019). Consistent with the regulatory mechanisms described above and confirming our hypothesis, our meta-analysis shows that N addition results in a lower biomass fraction of roots and reproductive organs and a decrease in the root-shoot ratio. This indicates that for terrestrial plants, N deficiency may constrain photosynthesis more strongly than reproduction.

Previous research has demonstrated that C4 plants have a higher photosynthetic potential compared to C3 plants. (Zhu et al., 2008). Our research shows that C4 plants also respond more strongly than C3 plants in terms of biomass gain with N addition. Compared with herbaceous plants, the carbon and nutrient storage capacity of woody plants is expected to be much larger (Eyles et al., 2009).

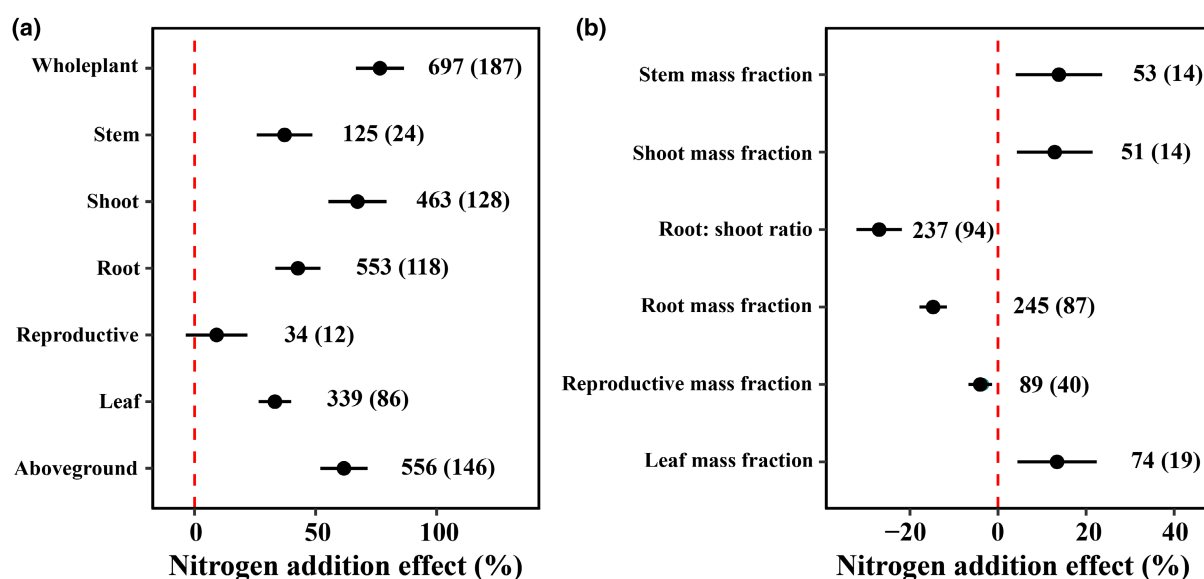


FIGURE 5 Effect size of N addition on terrestrial plant biomass for different tissue types (a) and mass fractions and root-shoot ratio (b), expressed on the global scale (RR [$\pm 95\%$ confidence interval]). The vertical dashed red line denotes a null effect size (RR=0). The number of observations is beside each attribute without parentheses, and the number of plant species is in the parentheses. RR, log response ratio. [Colour figure can be viewed at [wileyonlinelibrary.com](https://onlinelibrary.wiley.com/doi/10.1111/gcb.16731)]

TABLE 1 Effect sizes and regression coefficient of the factors affecting terrestrial plant biomass response to N addition on the global scale.

Fixed effects/ moderators	Mean effect size/regression coefficient	LCI	UCI	N	Model support based on omnibus tests (Q_m , df)
MAT				1888	$Q_{m*1} = 29.446, p < .0001$
Intercept	0.2914	0.2476	0.3351		
Slope	0.0101	0.0065	0.0138		
Latitude				1909	$Q_{m*1} = 4.667, p = .0307$
Intercept	0.4199	0.3690	0.4709		
Slope	-0.0012	-0.0023	-0.0001		
AP				821	$Q_{m*1} = 7.469, p = .0063$
Intercept	0.2209	0.0908	0.3510		
Slope	0.0270	0.0076	0.0464		
TK				1658	$Q_{m*1} = 6.999, p = .0082$
Intercept	0.2257	0.1256	0.3257		
Slope	0.0934	0.0242	0.1626		
TN				1772	$Q_{m*1} = 20.416, p < .0001$
Intercept	0.4330	0.3933	0.4727		
Slope	-0.1803	-0.2585	-0.1021		
SLA				1987	$Q_{m*1} = 25.405, p < .0001$
Intercept	0.3138	0.2491	0.3784		
Slope	0.0089	0.0055	0.0124		
PLA				897	$Q_{m*1} = 6.744, p = .0094$
Intercept	0.4373	0.3878	0.4867		
Slope	0.0006	0.0002	0.0011		
LCN				1641	$Q_{m*1} = 10.608, p = .0011$
Intercept	0.5698	0.4939	0.6439		
Slope	-0.0036	-0.0058	-0.0015		
LCC				1510	$Q_{m*1} = 37.941, p < .0001$
Intercept	0.5769	0.5192	0.6346		
Slope	-0.0032	-0.0042	-0.0022		
LNC				1863	$Q_{m*1} = 34.477, p < .0001$
Intercept	0.6211	0.5556	0.6866		
Slope	-0.0997	-0.1330	-0.0664		
N.addition				1607	$Q_{m*1} = 10.087, p = .0015$
Intercept	0.3777	0.3344	0.4210		
Slope	-0.0027	-0.0044	-0.0010		
Duration				2662	$Q_{m*1} = 28.239, p < .0001$
Intercept	0.4835	0.4496	0.5174		
Slope	-0.0259	-0.0354	-0.0163		

Note: LCI and UCI represent the lower and upper bounds of the 95% confidence intervals. Significant influence of moderators is indicated in bold. Abbreviations: AP, soil available phosphorus; LCC, leaf carbon content per leaf area; LCN, leaf carbon/nitrogen ratio; LNC, leaf nitrogen content per leaf area; MAP, mean annual precipitation; MAT, mean annual temperature; N, number of effect sizes; N.addition, nitrogen addition amount (include the experimental N addition and atmospheric inorganic N deposition); Photopathway, photosynthesis pathway; PLA, leaf area per plant; SLA, specific leaf area; TK, soil total potassium; TN, soil total nitrogen.

Therefore, herbaceous plants are theoretically more responsive to changes in biomass than woody plants to N addition, which we confirmed in this study. Furthermore, the results of our study revealed

a significant positive impact of N addition on the biomass of seed plants, but not spore plants. There are several factors that may explain this outcome. Firstly, seed plants typically require higher

levels of N than spore plants due to their more intricate vascular system, which allows for more efficient nutrient uptake and transport. In contrast, spore plants lack this specialized system and, thus, may be less responsive to N addition (Xia & Wan, 2008). Secondly, seed plants possess a higher photosynthetic rate than spore plants, enabling them to use the additional N availability to produce more biomass (Sage, 2004). In contrast, spore plants have lower rates of photosynthesis and may not fully benefit from the increased N availability (Flexas & Carriqui, 2020). Overall, the disparate responses of seed and spore plants to N addition can be attributed to their differing nutrient demands, vascular systems, and photosynthetic capacities.

The results of the study indicate that deciduous trees exhibit a more pronounced growth response to N addition compared to evergreen trees. One possible reason for this is that deciduous trees are capable of promptly adjusting their photosynthetic capacity in response to fluctuations in N availability, whereas evergreen trees maintain a relatively stable photosynthetic capacity owing to their long-lived foliage (Li et al., 2020). Additionally, deciduous trees may have a greater requirement for N during the production and replacement of leaves, while evergreen trees demand less N due to their prolonged leaf life (Weng et al., 2017). Another factor that could account for the superior biomass response of deciduous trees to N addition is their greater nutrient-use efficiency (Hiremath, 2000). Research has demonstrated that deciduous trees possess higher N-use efficiency relative to evergreen trees (Liu et al., 2018), enabling them to maintain higher growth rates with reduced N inputs.

The findings from the study indicate that broadleaved trees experience a more substantial increase in biomass following N addition compared to coniferous trees. This phenomenon can be attributed to various factors, including N utilization efficiency and leaf traits. Broadleaved trees typically exhibit higher N utilization efficiency than coniferous trees, which refers to their ability to efficiently

acquire and use N for growth and development (Wyka et al., 2012). Furthermore, broadleaved trees possess a more extensive and shallow root system that enables them to take up N from a larger volume of soil (Schenk & Jackson, 2002). In addition, broadleaved trees have a higher leaf N concentration that enhances their photosynthetic capacity and growth rate (Sardans et al., 2011). Leaf traits such as SLA and leaf lifespan can also affect plant response to N addition. Broadleaved trees generally have higher SLA and shorter leaf lifespan compared to coniferous trees (Adler et al., 2014). The higher SLA allows for rapid leaf growth, which results in greater carbon fixation and biomass accumulation following N addition (Augusto & Boča, 2022). The shorter leaf lifespan reduces the carbon costs associated with maintaining old leaves, enabling faster replacement with new, more efficient leaves. Overall, our findings highlight the potential of broadleaved trees as a target for N management strategies aiming at increasing carbon sequestration in terrestrial ecosystems.

4.2 | Leaf functional traits affect plant biomass allocation to nitrogen addition

Plants with high SLA and PLA are capable of high photosynthetic capacity and are considered nitrogen-use efficient as they have a higher capacity to capture nutrients in response to higher N availability in soils (Laliberté et al., 2012). A higher LNC is indicative of more effective N absorption, retention, or both (Gornish & Prather, 2014). We found that the positive effect of N addition on biomass decreased as plant N capture or storage capacity (i.e., LNC) improved, suggesting that when plants have high N-use efficiency or reabsorption efficiency, N resources are not the main source of stress-limiting plant growth. By contrast, the higher the SLA and PLA of terrestrial plants, the higher the positive effect of N addition on biomass (Figure 6a; Table 1), leading to a significantly increased leaf

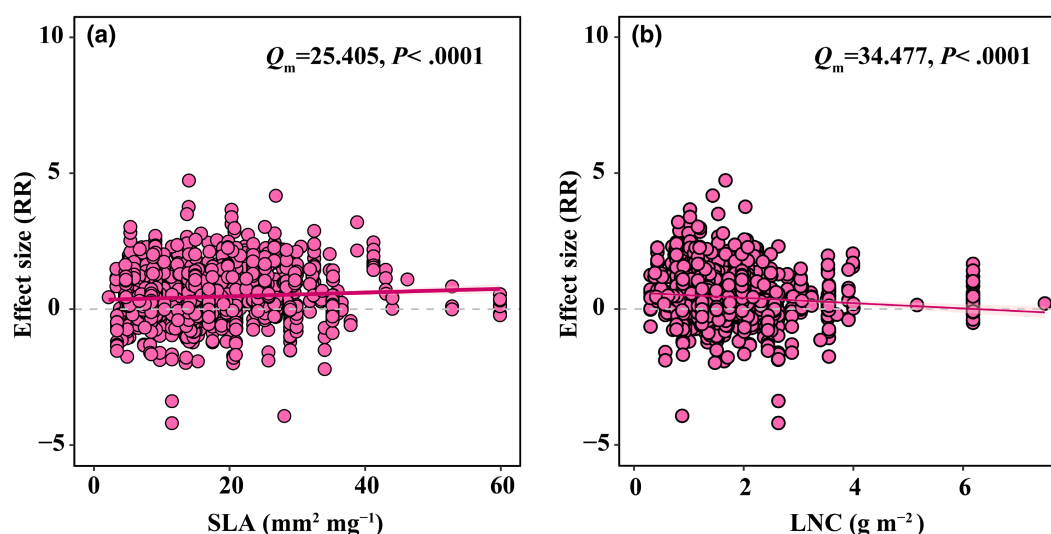


FIGURE 6 Specific leaf area (SLA; a) and leaf N content per leaf area (LNC; b) affecting plant biomass response to N addition. Circles represent the individual experiments under experimental N addition in the meta-analysis. RR, log response ratio. RR=0, dashed grey line; predicted mean effect size (with 95% CI in grey dashed lines), medium violet red lines. [Colour figure can be viewed at [wileyonlinelibrary.com](https://onlinelibrary.wiley.com/doi/10.1111/gcb.16731)]

mass fraction and a decreased root-shoot ratio. Therefore, terrestrial plants are projected to increase SLA and PLA to be more competitive for light in environments where the N supply is sufficient.

Biomass allocation has been well reported to be a plastic trait of plants adapted to various habitats (Yan et al., 2016). The leaf economics spectrum (Wright et al., 2004) refers to the continuous transition in leaf traits from thin, N-rich, short-lived leaves with high photosynthetic rates, known as the exploitative strategy (Grime, 1977), to thicker, more fibrous, N-poor, long-lived leaves with lower photosynthetic rates, and is known as the conservative growth strategy (Lavorel & Grigulis, 2012). Previous studies have shown that SLA and LNC are effective predictors of where a plant is situated along the leaf economic spectrum. High SLA and LNC imply an exploitative strategy for plants, while low SLA and LNC imply a conservative strategy (Wright et al., 2004). We found that when plants adopt conservative strategies (low SLA and LNC), N addition has a consistently positive effect on plant biomass. However, when plants adopt an exploitative strategy, the positive effect of N addition on plant biomass decreases as LNC increases. It then becomes negative when LNC is greater than 6.23 g m^{-2} (Figure 6b). As a result, whereas flexibility in biomass allocation might improve plant adaptation to N addition, the underlying leaf functional traits may dictate how much plant species modify their allocation patterns in response to N addition.

4.3 | Uncertainties and implications for future studies

This global meta-analysis enhances our understanding of the effects of N addition-driven environmental change on biomass accumulation and allocation of terrestrial plants and provides a useful reference for future policymaking and implementation of ecological and environmental protection. However, there are some caveats to be aware of when interpreting the results. First, this meta-analysis focused on the impact of N addition on plant biomass and its components in the terrestrial ecosystem and does not include interactions between other global change factors (e.g., increasing temperature and changing precipitation). Second, a meta-analysis of the effects of N addition on terrestrial plant biomass and its composition was carried out based on the most comprehensive dataset; however, the uneven distribution of field-observed datasets (i.e., the lack of observed data in eastern Africa, Russia, and northern Asia) may cause a bias in the worldwide distribution impacts of increasing N availability. Therefore, more field observations in relevant regions are urgently required to enhance our understanding of the consequences of global N deposition on terrestrial ecosystems.

5 | CONCLUSIONS

Our global synthesis quantified the effects of reactive N addition on terrestrial plant biomass and its allocation in plants, as well as identifying the key drivers of terrestrial plant biomass response to reactive

N addition on a global scale. Our findings revealed that N addition had a positive effect on plant biomass, but the positive effect gradually decreased as N load and fertilization duration increased. On the global scale, more plants tend to increase vegetative growth and reduce reproductive allocation under increasing N decomposition. Furthermore, the effects of N addition on plant biomass increased significantly with increasing MAT, AP, TK, SLA, and PLA. However, they decreased significantly with increasing TN, amount of N addition, LCN, LCC, LNC, and duration of the addition. By accelerating growth and reducing reproductive allocations, terrestrial plants may alter their reproductive trade-offs to enhance their competitiveness in response to enhanced competition in N-added plots. Therefore, when simulating global changes related to N deposition and terrestrial carbon sequestration, Earth System Models should be improved to predict the response of plant biomass and its allocations to future N deposition in combination with plant leaf functional traits and environmental variables.

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CONFLICT OF INTEREST STATEMENT

The authors declare no competing interests.

DATA AVAILABILITY STATEMENT

The data that support the findings of this study are openly available in figshare at <https://doi.org/10.6084/m9.figshare.22584481>.

ORCID

Huili Feng  <https://orcid.org/0000-0001-9838-9787>

Jiahuan Guo  <https://orcid.org/0000-0002-3440-6696>

Changhui Peng  <https://orcid.org/0000-0003-2037-8097>

Daniel Kneeshaw  <https://orcid.org/0000-0003-2585-8436>

Gabrielle Roberge  <https://orcid.org/0009-0000-7543-8312>

Chang Pan  <https://orcid.org/0000-0002-0236-994X>

Xuehong Ma  <https://orcid.org/0000-0001-9658-1089>

Dan Zhou  <https://orcid.org/0009-0008-6582-2095>

Weifeng Wang  <https://orcid.org/0000-0002-9752-6185>

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