

# Thermal Decomposition of Two Gaseous Perfluorocarboxylic Acids: Products and Mechanisms

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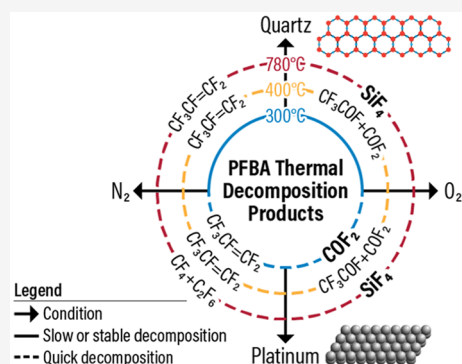
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**ABSTRACT:** The thermal decomposition products and mechanisms of per- and polyfluoroalkyl substances (PFASs) are poorly understood despite the use of thermal treatment to remediate PFAS-contaminated media. To identify the thermal decomposition products and mechanisms of perfluorocarboxylic acids (PFCAs), gaseous perfluoropropionic acid (PFPrA) and perfluorobutyric acid (PFBA) were decomposed in nitrogen and oxygen at temperatures from 200 to 780 °C. In nitrogen (i.e., pyrolysis), the primary products of PFPrA were  $\text{CF}_2=\text{CF}_2$ ,  $\text{CF}_3\text{CF}_2\text{H}$ , and  $\text{CF}_3\text{COF}$ .  $\text{CF}_3\text{CF}=\text{CF}_2$  was the dominant product of PFBA. These products are produced by HF elimination (detected as low as 200 °C).  $\text{CF}_4$  and  $\text{C}_2\text{F}_6$  were observed from both PFCAs, suggesting formation of perfluorocarbon radical intermediates. Pyrolysis products were highly thermally stable, resulting in poor defluorination. In oxygen (i.e., combustion), the primary product of both PFPrA and PFBA below 400 °C was  $\text{COF}_2$ , but the primary product was  $\text{SiF}_4$  above 600 °C due to reactions with the quartz reactor. Oxygen facilitated thermal defluorination by reacting with PFCAs and with pyrolysis products (i.e., fluoroolefins and fluorocarbon radicals). Platinum improved combustion of PFCAs to  $\text{COF}_2$  at temperatures as low as 200 °C, while quartz promoted the combustion of PFCAs into  $\text{SiF}_4$  at higher temperatures (>600 °C), highlighting the importance of surface reactions that are not typically incorporated into computational approaches.

**KEYWORDS:** PFCAs, pyrolysis, combustion, PFAS, thermal treatment

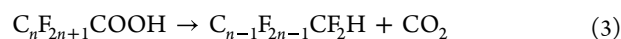
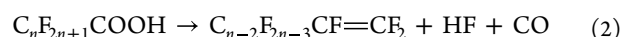
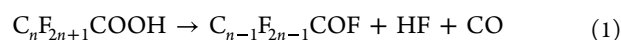


## 1. INTRODUCTION

Per- and polyfluoroalkyl substances (PFASs) are fluorinated organic chemicals used in numerous products such as cosmetics,<sup>1,2</sup> nonstick coatings, packaging, pesticides, and aqueous film-forming foams.<sup>3</sup> The widespread use of PFASs and their long half-lives in the environment have led to their detection in water, soil, and air.<sup>4–9</sup> Human exposure to PFASs is associated with multiple negative health outcomes, including cancer, decreased birth weight, and immune system and lipid metabolism disruption.<sup>3,10</sup> Similar ecological health outcomes have also been observed for animals exposed to PFASs, in part because some PFASs are highly bioaccumulative.<sup>11,12</sup> The toxicity, bioaccumulation, and persistence of PFASs have attracted the attention of the scientific and regulatory community.<sup>13</sup>

To address PFAS contamination, thermal treatments such as incineration,<sup>14–16</sup> smoldering,<sup>17</sup> pyrolysis of biosolids,<sup>18,19</sup> and regeneration of granular activated carbon (GAC)<sup>20–22</sup> are currently being studied. There is a substantial body of research focusing on computational predictions of perfluorocarboxylic acids (PFCAs) decomposition based on density functional theory (DFT). Results from these studies have suggested that the primary pyrolysis mechanism of perfluoropropionic acid (PFPrA),<sup>23</sup> perfluorobutyric acid (PFBA),<sup>24</sup> and perfluoropentanoic acid<sup>25,26</sup> is HF elimination via reaction 1, yielding

perfluoroacyl fluoride. Reactions 2 and 3 have been proposed to be minor pathways. However, when perfluorooctanoic acid (PFOA) was experimentally evaluated in a quartz or aluminum reactor, there was a divergence between the DFT-proposed and experimental products; only reactions 2 and 3 were observed.<sup>27,28</sup> Additionally, DFT models do not typically capture the surface effects of reactor materials or catalysts, highlighting the need for experimental verification of thermal decomposition mechanisms and products.



Published experimental research has focused on the temperature required to decompose PFCAs, rather than the mechanisms, and has demonstrated that decomposition begins at ~200 °C and partial mineralization (>86%) occurs above

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700 °C.<sup>21,29–32</sup> With addition of calcium hydroxide or calcium oxide above 400 °C, fluorine from perfluorooctanesulfonic acid (PFOS) (>60%) and PTFE (>80%) was mineralized and fixed as calcium fluoride.<sup>15,33</sup> When PFASs were sorbed to GAC prior to thermal treatment, the temperature required to initialize decomposition of PFASs was substantially lower than the temperature of those not sorbed to GAC. This was attributed to a catalytic effect of the relatively mobile electrons in the GAC surface,<sup>22</sup> but overall, surface effects and catalysis of PFAS thermal decomposition have been relatively poorly studied.

Volatile organofluorine (VOF) can also be formed during the thermal decomposition of PFASs<sup>21,34</sup> and can be transported long distances (>150 km),<sup>35</sup> but the formation and fate of volatile decomposition products are not well understood. Mass balances on fluorine during thermal treatment are rarely achieved,<sup>19</sup> and low mineralization rates of fluorine indicate the formation of organofluorine and/or VOF.<sup>20,34</sup> For example, the VOF yield was as high as 13% during thermal treatment of GAC loaded with PFOA<sup>34</sup> and VOF such as CF<sub>4</sub> and C<sub>2</sub>F<sub>6</sub> are extremely challenging to thermally decompose, even up to 1000 °C.<sup>36</sup> Thus, it is imperative to understand the formation of VOF during thermal treatment to identify treatment alternatives that mitigate VOF release.

Finally, both combustion (in oxygen) and pyrolysis (in an inert atmosphere) are parts of thermal treatment.<sup>18,31</sup> It is known that O<sub>2</sub> improves mineralization of PFASs, but the mechanisms are again not well understood.<sup>36,37</sup> The objectives of this study were (1) to compare the decomposition pathways proposed by computational studies to experimental results, (2) to understand the effects of temperature on thermal decomposition and/or mineralization of PFCAs and respective VOF formation, (3) to identify the role of O<sub>2</sub> in the thermal mineralization of PFCAs, and (4) to assess the effect of reactor materials, including quartz (present in nearly all previous studies of PFAS decomposition as the reactor material) and platinum (Pt), on decomposition. To achieve these goals, we investigated the fate and transformations of two model PFCAs, PFPrA and PFBA, in nitrogen (N<sub>2</sub>) and O<sub>2</sub> atmospheres at temperatures from 200 to 780 °C via Fourier transform infrared spectroscopy (FTIR). PFPrA and PFBA were selected because they were volatile enough in initial experiments to produce a FTIR signal when the headspace gas above a neat standard was introduced into the FTIR cell and because the decomposition mechanisms are likely to be similar for all PFCAs, a commonly observed group of PFASs in the environment.

## 2. MATERIALS AND METHODS

**2.1. Reagents and Materials.** PFPrA and PFBA (both in their acid forms) and 350 mg of Pt foil (0.025 mm thickness, 99.9%) were purchased from Sigma-Aldrich (St. Louis, MO). Quartz wool was purchased from COSA Xentaur (Houston, TX). Two Pt wires (0.25 mm diameter, 100 cm length) were purchased from Alfa Aesar (Tewksbury, MA).

**2.2. Thermal Decomposition Experiments.** A modified Shimadzu TOC-VCSH instrument was used as a temperature-controlled furnace for the thermal treatment of PFCAs.<sup>32</sup> The temperature was controlled through the instrument software. A quartz combustion tube was packed with 0.3 g of quartz wool to prevent short circuiting in the combustion tube or Pt wire and foil to test the potential for catalytic degradation.<sup>32</sup> O<sub>2</sub> or

N<sub>2</sub> carrier gases were passed through the combustion tube at a rate of 300 mL/min. A stoppered vial containing <1 mL of pure PFPrA or PFBA was connected in parallel with the carrier gas line. The combustion tube was interfaced to a 5 m optical path FTIR gas cell. The gas downstream of the combustion module was monitored by a Shimadzu IRTracer-100 instrument equipped with a MCT (mercury–cadmium–telluride) detector. The resolution of FTIR was 2 cm<sup>−1</sup>. The number of scans to produce a spectrum was three, taking 12 s to acquire.

Each thermal decomposition experiment was conducted under four variable conditions: reactants (PFPrA and/or PFBA), temperature (200–780 °C), atmosphere (N<sub>2</sub> or O<sub>2</sub>), and packing material (quartz wool and/or Pt). Before each experiment, the stoppered vial and gas cell were purged with carrier gas at a flow rate of 300 mL/min to remove interference from air. Initially, the carrier gas bypassed the stoppered vial and flowed directly into the combustion tube (Figure 1), and

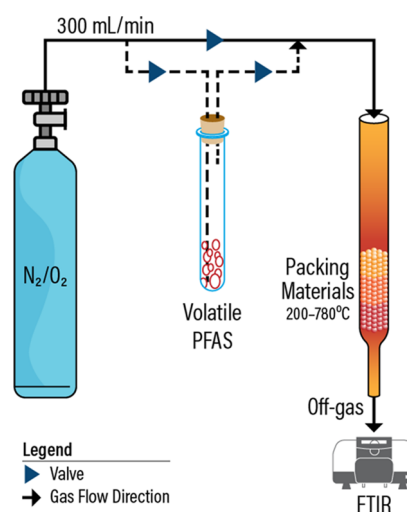
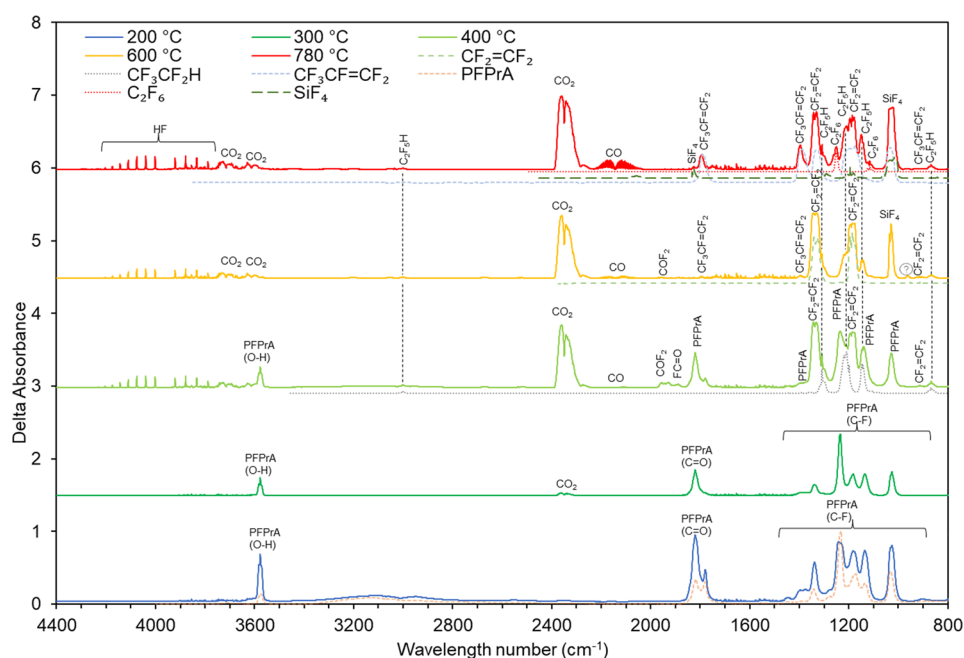


Figure 1. Schematic of the experimental approach.

the temperature of the combustion tube was set to 200 °C. When the temperature of the combustion tube and the FTIR baseline were stable, the headspace gas from the stoppered vial containing either PFPrA or PFBA was introduced by altering the carrier gas flow path. For 3 s (PFPrA) or 10 s (PFBA), to produce a sufficiently strong FTIR signal to interpret the spectra without saturating the detector (based on initial experiments), the carrier gas flowed through the headspace of the stoppered vial. PFBA required more time because of the lower vapor pressure.<sup>38</sup> The carrier gas flow path was then returned to directly flow to the combustion tube. The FTIR spectra were scanned and recorded every 1 min for the first 4 min and then every 2–3 min later until no signal remained, approximately 20 min later. The temperature was then adjusted as needed (e.g., 300, 400, 600, and 780 °C), and the FTIR baseline was reset. Spectra with the strongest peaks but having an intensity of <1 absorbance unit were utilized for product identification, and IR peak assignments are listed in Table S1.

**2.3. Product Identification Approach.** Products were identified by matching observed FTIR spectra with published spectra from the NIST Chemistry Webbook,<sup>39</sup> [spectrabase.com](https://spectrabase.com),<sup>40</sup> or journal publications.<sup>41–48</sup> Reference spectra and the spectra from this study are listed in Table S1. For the sake of clarity during comparisons, reference spectra were adjusted by



**Figure 2.** Infrared spectra acquired during pyrolysis of PFPrA in N<sub>2</sub> at temperatures from 200 to 780 °C. Dashed and dotted lines are standard spectra from the library (see Table S1).

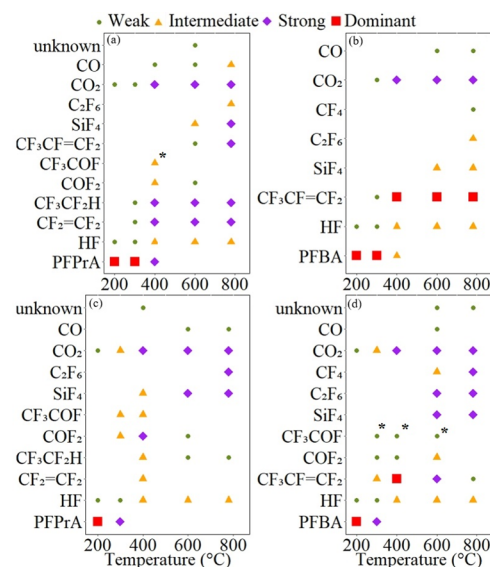
multiplying by a constant. The spectra of organofluorine compounds, including PFPrA, PFBA, 1H-perfluoroethane (CF<sub>3</sub>CF<sub>2</sub>H), CF<sub>2</sub>=CF<sub>2</sub>, CF<sub>3</sub>CF=CF<sub>2</sub>, COF<sub>2</sub>, CF<sub>3</sub>COF, and C<sub>2</sub>F<sub>6</sub>, were in some cases partially obscured due to overlapping peaks of other products. When partial overlap occurred, products were confirmed by two additional characteristic peaks that were not obscured by other products in the fingerprint region. Spectra of inorganic species, including CO<sub>2</sub>, CO, HF, and SiF<sub>4</sub>, rarely overlapped with those of other products. In one case, the SiF<sub>4</sub> peak at 1030 cm<sup>-1</sup> during the pyrolysis of PFBA above 600 °C was completely obscured by CF<sub>3</sub>CF=CF<sub>2</sub>. However, the spectra of SiF<sub>4</sub> could be clearly observed at two adjacent temperatures, 600 and 780 °C, where the spectrum of CF<sub>3</sub>CF=CF<sub>2</sub> was weaker (Figure S5). Each of the identified products was observed at least twice in separate experiments.

To describe the relative absorbance intensity of the formed products in Figures 3 and 4, we defined four terms: dominant, strong, intermediate, and weak. We define “dominant” as a peak or group of related fluorine-containing peaks from the reagent (e.g., parent PFAS) or decomposition products that were most prevalent in the spectra. In many cases, a dominant product was the only peak or set of associated peaks in the spectra (see Figure S1B at 600 °C, where C<sub>3</sub>F<sub>6</sub> is the dominant peak). “Strong” was a product with a peak absorbance of >0.3. “Intermediate” was a product with an absorbance between 0.1 and 0.3. “Weak” had an absorbance of <0.1. We have included these groupings to relate the relative intensity of the products in a succinct way, and the full spectra are provided in the Supporting Information.

### 3. RESULTS AND DISCUSSION

**3.1. Pyrolysis of PFCAs.** **3.1.1. Pyrolysis in the Presence of Quartz.** PFPrA and PFBA were pyrolyzed between 200 and 780 °C in a quartz reactor partially filled with quartz wool to impede short circuiting. An example FTIR spectrum during PFPrA pyrolysis is shown in Figure 2, and a summary of the

products of both PFPrA (CF<sub>3</sub>CF<sub>2</sub>COOH) and PFBA (CF<sub>3</sub>CF<sub>2</sub>CF<sub>2</sub>COOH) is shown in Figure 3. The products fell



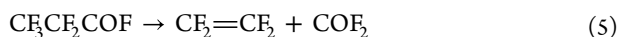
**Figure 3.** Pyrolysis products in N<sub>2</sub> at various temperatures: (a) PFPrA (b) PFBA, (c) PFPrA with Pt, and (d) PFBA with Pt. FTIR spectra are provided in Figures S1 and S2. Weak, intermediate, strong, and dominant are defined in Materials and Methods. An asterisk indicates that only the COF group was confirmed.

into three categories: inorganic fluorine (i.e., HF and SiF<sub>4</sub>), identified and unidentified organofluorine, and mineralized carbon (i.e., CO and CO<sub>2</sub>). At 200 °C, organofluorine products were not observed for either PFPrA or PFBA. HF was, however, observed at 200 °C (Figure S6), suggesting that the decomposition of PFPrA and PFBA by HF elimination is initiated at relatively low temperatures. Beginning at 300 °C, olefins were present (Figure S7), and at ≥400 °C, most or all



of reactants were decomposed to their respective olefins (i.e., PFPrA  $\rightarrow$  CF<sub>2</sub>=CF<sub>2</sub>; PFBA  $\rightarrow$  CF<sub>3</sub>CF=CF<sub>2</sub>) and CF<sub>3</sub>CF<sub>2</sub>H (only from PFPrA). Above 600 °C, SiF<sub>4</sub> and C<sub>2</sub>F<sub>6</sub> were also formed from both PFPrA and PFBA.

Prior computational studies have suggested that **reaction 1** dominates pyrolysis of PFCAs,<sup>23,24,26</sup> but during pyrolysis of PFBA, the corresponding CF<sub>3</sub>CF<sub>2</sub>COF was not detected at any temperature. For PFPrA at 400 °C, we observed C=O stretching at 1898 cm<sup>-1</sup>, which can be attributed to the acyl fluoride group (-COF) in CF<sub>3</sub>COF<sup>45</sup> as predicted by **reaction 1** (Figure 2). However, other computational and experimental research has suggested that the observed acyl fluoride functional group may also be formed during the pyrolysis of PFASs via reactions with HF (**reaction 4**), forming a corresponding CF<sub>3</sub>CF<sub>2</sub>COF from PFPrA rather than CF<sub>3</sub>COF (**reaction 1**).<sup>25,49</sup> The -COF peak was relatively small, and C-F stretching was somewhat obscured by other products. However, considering that some HF formation was concurrently present in the spectra and that some COF<sub>2</sub> was also concurrently observed (likely formed via **reaction 5**, which requires CF<sub>3</sub>CF<sub>2</sub>COF),<sup>50,51</sup> the -COF peak is likely associated with CF<sub>3</sub>CF<sub>2</sub>COF. Overall, we find that **reaction 1** is unlikely to play a major role in the pyrolysis of PFCAs below ~600 °C in a quartz reactor. This is further supported by the lack of observed C<sub>6</sub>F<sub>13</sub>COF being formed via **reaction 1** during PFOA pyrolysis in two other publications.<sup>27,28</sup>



**Reactions 2** and **3** have been proposed to be minor pathways by a computational study,<sup>23</sup> but here we note that these pathways appear to be principal below ~780 °C. The primary products of PFPrA pyrolysis were HF, CO<sub>2</sub>, CF<sub>2</sub>=CF<sub>2</sub>, and CF<sub>3</sub>CF<sub>2</sub>H above 200 °C. These products are explained well by HF elimination via **reactions 2** and **3**. However, for PFBA, CF<sub>3</sub>CF<sub>2</sub>CF<sub>2</sub>H was not observed at any temperature, and therefore, **reaction 3** does not appear to play a major role. Instead, CF<sub>3</sub>CF=CF<sub>2</sub> was the dominant product at all temperatures, suggesting the dominant decomposition pathway is **reaction 2**. In another publication, the pyrolysis of PFBA sodium and potassium salts (e.g., CF<sub>3</sub>CF<sub>2</sub>CF<sub>2</sub>COONa and CF<sub>3</sub>CF<sub>2</sub>CF<sub>2</sub>COOK, respectively) yielded >97% CF<sub>3</sub>CF=CF<sub>2</sub>,<sup>52</sup> agreeing well with our findings using protonated PFBA. PFBA has only one additional fluorinated carbon (CF<sub>2</sub>) compared to PFPrA, but the decomposition products are substantially different, highlighting that for these two small molecules, small changes to the molecular structure can have large impacts on outcomes. Although these changes are substantial, we expect that for longer chain PFCAs the primary decomposition pathways will be among the three pathways identified here (**reactions 1–3**).

Above ~600 °C, there appeared to be a substantial change in mechanism. For example, CF<sub>3</sub>CF=CF<sub>2</sub> was formed at  $\geq 600$  °C during the pyrolysis of PFPrA, likely from the thermal transformation of CF<sub>2</sub>=CF<sub>2</sub>, which was also observed at  $\geq 400$  °C.<sup>53,54</sup> This transformation became more apparent at higher temperatures (Figure 2). It is well documented that CF<sub>2</sub>=CF<sub>2</sub> is highly reactive, and some of the reactions can produce other VOF, including perfluorocyclobutane (cyclo-C<sub>4</sub>F<sub>8</sub>), perfluoroisobutene [(CF<sub>3</sub>)<sub>2</sub>C=CF<sub>2</sub>], perfluoroethane (C<sub>2</sub>F<sub>6</sub>), and tetrafluoromethane (CF<sub>4</sub>), although these products were not apparent in this study.<sup>53,54</sup>

Above ~600 °C, CO and SiF<sub>4</sub> were also significant products. The reaction between HF and quartz forming SiF<sub>4</sub> is well-known, but HF is unlikely to be the primary source of SiF<sub>4</sub> because the system is hydrogen limited. It is more likely that CO and SiF<sub>4</sub> were formed from reactions between perfluorocarbon radicals and quartz, as reported by others.<sup>55,56</sup> The formation of C<sub>2</sub>F<sub>6</sub> and CF<sub>4</sub> at 780 °C from both PFPrA and PFBA (Figure S1) also suggests the formation of perfluorocarbon radicals from homolysis of C–C and C–F bonds, which then terminate in these more stable molecules.

It is worth noting that the source of oxygen in CO and CO<sub>2</sub> not only is from the carboxyl group by **reactions 1–3** but also may be from quartz (SiO<sub>2</sub>) by reaction with fluorocarbon radicals.<sup>55–57</sup> CO<sub>2</sub> can also be formed during the pyrolysis of fluorocarbons (e.g., CHF<sub>3</sub> and C<sub>4</sub>F<sub>10</sub>) without supplemental oxygen in an aluminum reactor, and the oxygen was thought to be from aluminum.<sup>58,59</sup>

In summary, temperature is critical and delineates PFCA decomposition into two regions during pyrolysis. (1) From ~200 to ~600 °C, PFCA decomposition begins at the least stable functional group, the carboxylic acid, via HF elimination forming more stable fluorinated olefins and 1H-perfluoroalkanes. (2) Above ~600 °C, perfluorocarbon radicals are produced, resulting in the formation of CF<sub>4</sub> and C<sub>2</sub>F<sub>6</sub>. Perfluorocarbon radicals also reacted rapidly with the quartz reactor, resulting in the formation of SiF<sub>4</sub>. The reaction between radicals and the quartz reactor has implications for thermal treatment, where co-occurring waste materials may substantially alter the products from those that would be expected from computational analysis alone.

**3.1.2. Catalyzed Pyrolysis with Pt.** To understand the potential of surface effects to change the decomposition mechanisms, the quartz reactor was filled with large-surface area Pt foil and wire. The products of PFPrA and PFBA decomposition are shown in panels c and d of Figure 3, respectively. At 200 °C, the product spectra were similar to those produced without Pt; most reactants were not decomposed. The greatest difference between quartz and Pt packing was the disappearance of olefins at increased temperatures for both PFPrA and PFBA. For example, for PFPrA, CF<sub>2</sub>=CF<sub>2</sub> was not observed above 400 °C with Pt but was present even at 780 °C when no Pt was present. For PFBA, CF<sub>3</sub>CF=CF<sub>2</sub> was again present below 400 °C, but its level decreased with an increase in temperature, despite being the dominant product at all temperatures without Pt. For both PFPrA and PFBA, oxygen-containing products (e.g., COF<sub>2</sub> and CF<sub>3</sub>COF) were formed to a greater extent in the presence of Pt, likely due to oxidation of the respective olefins by surface-bound reactive oxygen species<sup>60–63</sup> present from residual O<sub>2</sub> and H<sub>2</sub>O sorbed when the Pt was not in use and exposed to air. CF<sub>3</sub>CF<sub>2</sub>H was also formed to a lesser extent from PFPrA in the presence of Pt compared to without Pt, although the difference in mechanism is not clear.

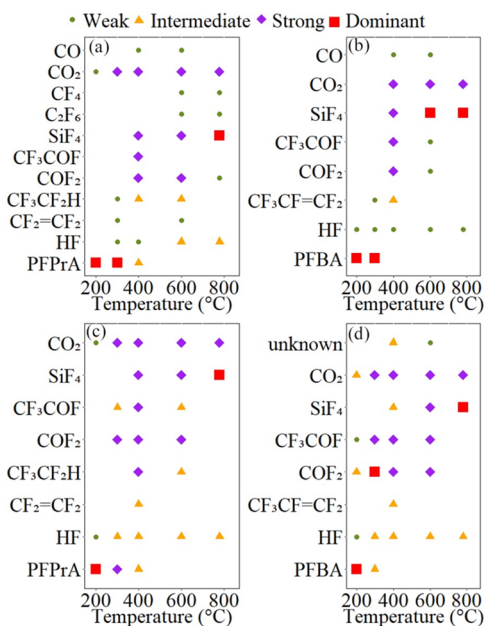
At 780 °C, C<sub>2</sub>F<sub>6</sub> and CF<sub>4</sub> were formed to a greater extent in the presence of Pt, especially for PFBA (Figure S2). Formation of these two products over the olefin (CF<sub>3</sub>CF=CF<sub>2</sub>, when only quartz was present) indicates that perfluorocarbon radicals formed to a greater extent on the Pt surface than when no Pt was present. This is because the formation of C<sub>2</sub>F<sub>6</sub> and CF<sub>4</sub> is primarily from reactions between perfluorocarbon radicals (i.e., radical termination).<sup>59,64</sup>

The signal associated with the -COF group observed during the pyrolysis of PFPrA without Pt was again observed with Pt,

and now the  $\text{CF}_3\text{COF}$  could clearly be identified from observed C–F stretching at 300 and 400 °C (Figure S2a). This suggests that although reaction 1 played a minor role without Pt, it was more important when Pt is present and indicates that surface effects can alter the decomposition pathways. This was particularly evident at 300 °C, where  $\text{CF}_3\text{COF}$  was a principal product, and little or no  $\text{CF}_3\text{CF}_2\text{H}$  or  $\text{CF}_2=\text{CF}_2$  was present (Figure S2b), despite being formed when Pt was not present in the reactor (Figure 3c). Given that reaction 1 appears to be occurring with Pt, we also expected to find the product CO. However, no CO was observed. This is likely because CO was catalytically oxidized to  $\text{CO}_2$  by oxygen or hydroxyl present on the Pt surface.<sup>65</sup> How Pt promotes reaction 1 is not clear, but it is apparent that the addition of Pt to the reaction mixture substantially changed the decomposition mechanisms of PFPrA.

For PFBA, the characteristic -COF peak at 1898  $\text{cm}^{-1}$  was observed but it is not clear if it was from  $\text{CF}_3\text{CF}_2\text{COF}$ , which would be expected on the basis of the findings from PFPrA, due to multiple overlapping signals in the C–F region.  $\text{COF}_2$  was also present simultaneously (Figure S2b), and therefore, this signal is more likely attributed to  $\text{CF}_3\text{COF}$  from the oxidation of  $\text{CF}_3\text{CF}=\text{CF}_2$  on the surface of Pt by reaction 6.

**3.2. Combustion of PFCAs.** 3.2.1. *Combustion with Quartz Wool.* Combustion (i.e., in the presence of  $\text{O}_2$ ) products of PFPrA and PFBA with a quartz reactor and quartz packing are shown in panels a and b of Figure 4, respectively.



**Figure 4.** Combustion products in  $\text{O}_2$ : (a) PFPrA, (b) PFBA, (c) PFPrA with Pt, and (d) PFBA with Pt. The FTIR spectra are provided in Figures S3 and S4. Weak, intermediate, strong, and dominant are defined in Materials and Methods.

At 200 and 300 °C,  $\text{O}_2$  did not substantially participate in the decomposition of PFPrA and PFBA; the products were similar to or the same as those produced by pyrolysis. One notable difference at 200 °C was that HF was formed less than by pyrolysis based on signal intensity, or not formed at all (Figure S6). It is not clear why HF was formed to a lesser extent.

At 400 °C, the extent of formation of  $\text{CF}_3\text{CF}_2\text{H}$  and  $\text{CF}_2=\text{CF}_2$  from PFPrA was substantially reduced compared to that of

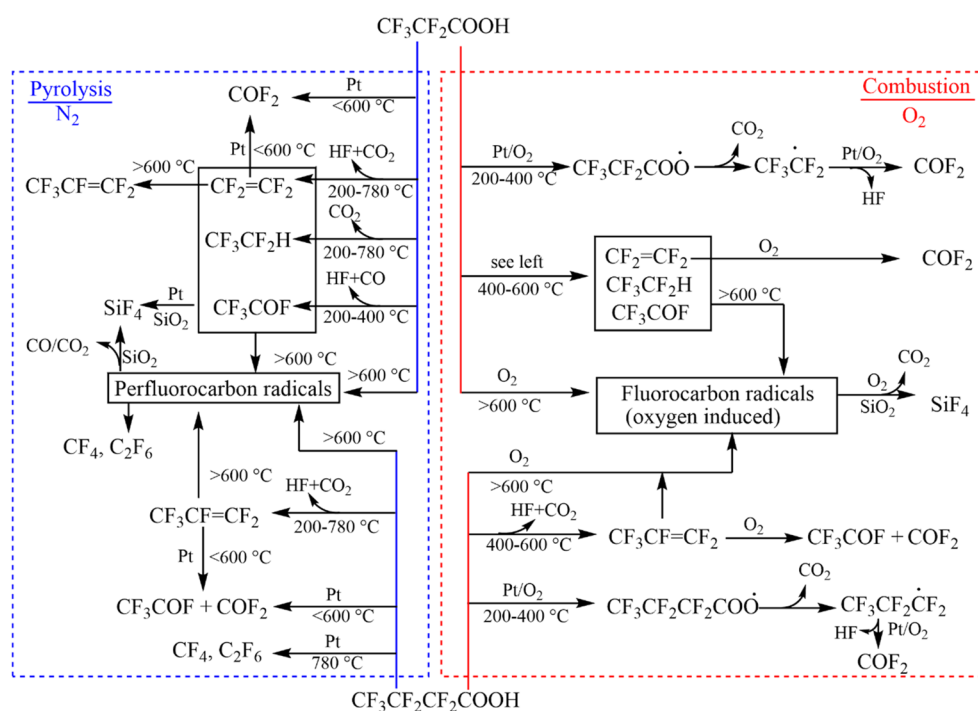
pyrolysis (Figures S1a and S3a), but the production of  $\text{COF}_2$  increased, based on the absorbance ratio of  $\text{CF}_3\text{CF}_2\text{H}$  or  $\text{CF}_2=\text{CF}_2$  to  $\text{COF}_2$ . This may be because  $\text{O}_2$  reacts with the olefin  $\text{CF}_2=\text{CF}_2$ , resulting in the formation of  $\text{COF}_2$  via reaction 7, where  $\text{CF}_2=\text{CF}_2$  was produced from PFPrA by reaction 2. When PFBA was combusted at 400 °C,  $\text{CF}_3\text{COF}$  and  $\text{COF}_2$  were formed to a greater extent than during pyrolysis, and the extent of formation of the product  $\text{CF}_3\text{CF}=\text{CF}_2$  was reduced (Figures S1b and S3b). This is explained well by reaction 6, and thus, it is likely that the observed  $\text{COF}_2$  produced by PFPrA is produced through a similar reaction with the intermediate olefin (reaction 6), rather than direct reactions between PFPrA and  $\text{O}_2$ .



We expected similar mechanisms at higher temperatures (i.e.,  $\text{O}_2$  oxidizes products formed through pyrolytic mechanisms, such as olefins into  $\text{CF}_3\text{COF}$  and  $\text{COF}_2$ ) because during pyrolysis the mechanisms, involving the production of olefins, were similar at all temperatures (Figure 3). However, the dominant fluorinated product was  $\text{SiF}_4$  for both PFPrA and PFBA above 600 °C (Figure 4). During pyrolysis,  $\text{SiF}_4$  was a product of reactions between perfluorocarbon radicals and quartz, and most organofluorine products persisted up to 780 °C. However, with  $\text{O}_2$  present at the same temperature, no organofluorine was observed in the gas phase from PFBA, and the majority of organofluorine was also mineralized for PFPrA. The complete mineralization of PFCAs to  $\text{SiF}_4$  at these high temperatures makes it difficult to explain the mechanisms due to the lack of observed intermediates, which likely have extremely short half-lives and react with quartz or terminate through reactions with other radicals. PFPrA and PFBA likely form some intermediates produced through reactions with  $\text{O}_2$  (possibly oxygen-containing fluorocarbon radicals), and these intermediates then preferentially react with quartz to form  $\text{SiF}_4$ . These reactions would also compete with perfluorocarbon radical recombination, explaining the lack of observable  $\text{C}_2\text{F}_6$  and  $\text{CF}_4$ .

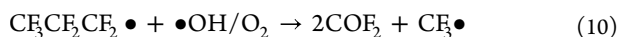
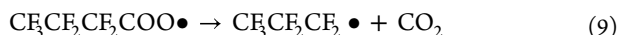
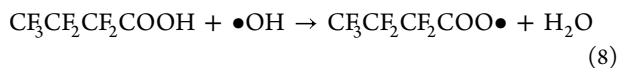
Overall, there was some overlap between the pyrolysis and combustion mechanisms. Below 300 °C,  $\text{O}_2$  played a negligible role in thermal decomposition; between 400 and 600 °C,  $\text{O}_2$  reacted with pyrolysis products (i.e., olefins), and above 600 °C,  $\text{SiF}_4$  was the dominant product. We believe these two PFCAs yielded unidentified fluorocarbon radical intermediates through reactions with  $\text{O}_2$ , which then preferentially reacted with the quartz reactor to form  $\text{SiF}_4$  as primary products. This, in turn, suppresses the formation of VOF (i.e.,  $\text{CF}_4$  and  $\text{C}_2\text{F}_6$ ), which is difficult to destroy, and has implications for thermal treatment processes operating with excess oxygen, which may help to control VOF emissions at high temperatures.

**3.2.2. Catalyzed Combustion with Pt.** When Pt was introduced with  $\text{O}_2$ , the products of PFPrA were similar at 200 °C (Figure 4c) to those when only quartz was present. One important difference was that some PFBA was oxidized to  $\text{COF}_2$  at 200 °C, and this did not appear to occur at this temperature without Pt (Figure 4b,d). At an only slightly higher temperature (300 °C),  $\text{COF}_2$  was the primary product for both PFPrA and PFBA. Notably, when Pt was present,  $\text{COF}_2$  did appear to originate from not only the oxidation of olefins but also the direct oxidation of PFCAs at or below 300 °C. We base this on the following. (1) Olefins were not



**Figure 5.** Proposed decomposition pathways of PFPrA (top) and PFBA (bottom) in the presence of quartz ( $\text{SiO}_2$ , reactor material) or Pt and quartz. Note that Pt can provide  $\bullet\text{OH}/\bullet\text{O}$  through reactions with surface-sorbed  $\text{O}_2$  and  $\text{H}_2\text{O}$ . The combustion intermediates “fluorocarbon radicals” were not directly observed but are proposed on the basis of the products.

observed, and therefore, there is little or no olefin present to be oxidized to  $\text{COF}_2$ . (2) If the expected olefin  $\text{CF}_3\text{CF}=\text{CF}_2$  was to be oxidized by  $\text{O}_2$ , it would result in  $\text{COF}_2$  and  $\text{CF}_3\text{COF}$  absorbance spectra of similar intensity, similar to those at 400 and 600  $^\circ\text{C}$  shown in Figure S4b and reaction 7, but the  $\text{COF}_2$  signal intensity was substantially greater than that of  $\text{CF}_3\text{COF}$  below 300  $^\circ\text{C}$ . Using DFT, Altarawneh et al. reported that the abstraction of hydrogen from the carboxylic acid group by  $\bullet\text{OH}$  (reactions 8 and 9) yields  $\text{CF}_3\text{CF}_2\text{CF}_2\bullet$ , which has the lowest activation energy of the abstractable sites.<sup>66</sup>  $\text{CF}_3\text{CF}_2\text{CF}_2\bullet$  can then react with  $\text{O}_2$  and  $\bullet\text{OH}$  to form  $\text{COF}_2$  (reactions 10 and 11),<sup>67</sup> facilitated by  $\bullet\text{OH}$  present on the surface of Pt as a result of splitting of residual water and/or  $\text{O}_2$ .<sup>60,62,63</sup> These catalyzed thermal decomposition pathways were also derived on the basis of the results from other processes (i.e., photolysis, electrochemistry, and advanced oxidation) and have been compiled in a recently published review.<sup>67</sup> Below  $\sim 300$   $^\circ\text{C}$ , it appears that this pathway is dominant when Pt is present.



Between 400 and 600  $^\circ\text{C}$ , reactions 8–11, yielding  $\text{COF}_2$ , were less dominant. Instead, for both PFPrA and PFBA, unimolecular decomposition yielding olefins and 1H-perfluoroalkanes dominated, similar to pyrolysis. The intermediate olefins appeared to be terminally oxidized by  $\text{O}_2$  to  $\text{COF}_2$  and/or  $\text{CF}_3\text{COF}$ . Above 600  $^\circ\text{C}$ , mineralized products tended to dominate, including  $\text{CO}_2$ , HF, and  $\text{SiF}_4$ .  $\text{CF}_3\text{COF}$  and  $\text{COF}_2$ ,

which were the primary oxidized products at lower temperatures when Pt was present, were surprisingly not present at 780  $^\circ\text{C}$ . It is possible that at higher temperatures, PFPrA and PFBA were decomposed and formed  $\text{SiF}_4$  in the top of a quartz tube via the perfluorocarbon radical pathway (section 3.2.1), before reaching the Pt placed in the lower portion of the tube, explaining why the products are the same as those from combustion with quartz.

**3.3. Summary of Decomposition Pathways.** A summary of the primary observed decomposition pathways of PFPrA and PFBA is illustrated in Figure 5 on the basis of discussions presented above. It should be noted that the temperatures labeled in Figure 5 do not indicate a cutoff of decomposition mechanisms. Those decomposition pathways are likely to happen at all temperatures with varying kinetics, and the representation presented here embodies only the decomposition pathways that tend to outcompete other pathways in the labeled temperature ranges. The primary pyrolysis products, olefins and  $\text{CF}_3\text{CF}_2\text{H}$ , are highly stable up to 780  $^\circ\text{C}$  but readily react with  $\text{O}_2$ . In an inert atmosphere, Pt facilitated the decomposition of PFPrA and PFBA, and their pyrolysis products, although these mechanisms appeared to rely on surface-sorbed oxygen and thus are not truly pyrolysis. With  $\text{O}_2$ , Pt catalyzed the oxidation of PFBA at temperatures as low as 200  $^\circ\text{C}$ . Above 600  $^\circ\text{C}$ , quartz and  $\text{O}_2$  together promoted the mineralization of PFCAs into  $\text{SiF}_4$ .

## 4. IMPLICATIONS

Thermal treatment presents an opportunity to end the PFAS cycle by mineralization to fluoride and carbon dioxide, but the mechanisms and thus potential to produce gaseous organofluorine products are not well understood. We show here that temperatures of  $<780$   $^\circ\text{C}$  or the absence of oxygen results in relatively recalcitrant gaseous PFAS (e.g., olefins,  $\text{CF}_4$ , and



C<sub>2</sub>F<sub>6</sub>) from two PFCAs. Without further treatment of the off-gas, PFASs may be transferred from contaminated media to the atmosphere. Relevant systems include thermal desorption,<sup>51,68,69</sup> PFAS-laden GAC regeneration,<sup>20,21,30,34</sup> pyrolysis of biosolids,<sup>18,19</sup> and systems with slow heat or flame propagation. Post-treatment of produced gases during such treatment is likely to require additional high-temperature treatment (>780 °C) in the presence of excess O<sub>2</sub>. In addition, combustion is usually conducted in an air (21% O<sub>2</sub>) atmosphere, which may influence the efficiency of mineralization of PFAS. Further thermodynamic and kinetic studies of gaseous PFAS degradation and the potential to degrade or capture the products of incomplete combustion are warranted.

Formation of CF<sub>4</sub> and C<sub>2</sub>F<sub>6</sub> during pyrolysis at high temperatures may be particularly problematic as they are difficult to decompose once formed. One strategy for avoiding their formation may be to conduct pyrolysis below 600 °C, forming olefins or polyfluorinated products that are decomposed more readily in a thermal process with excess O<sub>2</sub>. Addition of a Pt catalyst may be beneficial at temperatures below ~600 °C but does not provide a substantial benefit above 600 °C when oxygen and quartz are present and PFCAs appear to be entirely mineralized.

One mineralized product, SiF<sub>4</sub>, appears to be a product of reactions between organofluorine radicals and the quartz reactor, and thus, other coincident materials present in the reactor (e.g., soil silica) should be considered in a computational evaluation of PFAS thermal decomposition. Terminal products such as HF, COF<sub>2</sub>, and SiF<sub>4</sub> are likely to be removed well by alkaline scrubbers present in air pollution control devices because they are relatively soluble and hydrolyze under alkaline conditions.

In addition to the implications for thermal treatment, the finding that at lower temperatures with O<sub>2</sub> present, PFCAs can be selectively oxidized to COF<sub>2</sub> on the surface of Pt is valuable in terms of total organofluorine measurements. Organofluorine measurements are currently hampered by interference from inorganic fluorine present in environmental samples. COF<sub>2</sub> is a product that is generated only from combustion of PFASs, cannot be formed from inorganic fluorine (e.g., CaF<sub>2</sub> and NaF), and therefore may be valuable as an organofluorine surrogate, avoiding interference from inorganic fluorine. Selective oxidation of the perfluorocarbon chain (CF<sub>2</sub>)<sub>n</sub> to COF<sub>2</sub> likely applies to other PFAS species as long as perfluorocarbon radicals or fluorinated olefins are formed during combustion so that COF<sub>2</sub> is subsequently formed through reactions 6–11. In ongoing unpublished work by our team, we have already observed this product from PFOS combustion, and we believe that a COF<sub>2</sub> surrogate measurement of PFASs presents an opportunity for a robust new analytical method.

## ■ ASSOCIATED CONTENT

### SI Supporting Information

The Supporting Information is available free of charge at <https://pubs.acs.org/doi/10.1021/acs.est.2c08210>.

Assignment of IR peaks (4400–600 cm<sup>-1</sup>) recorded during thermal decomposition of PFPrA and PFBA (Table S1), infrared spectra acquired during pyrolysis of PFCAs in N<sub>2</sub> at varying temperatures [(a) PFPrA and (b) PFBA (Figure S1)], infrared spectra of catalyzed pyrolysis products of PFCAs in N<sub>2</sub> at varying temper-

atures with Pt [(a) PFPrA and (b) PFBA (Figure S2)], infrared spectra acquired during combustion of PFCAs in O<sub>2</sub> at varying temperatures [(a) PFPrA and (b) PFBA (Figure S3)], infrared spectra of catalyzed combustion products of PFCAs in O<sub>2</sub> at varying temperatures with Pt [(a) PFPrA and (b) PFBA (Figure S4)], infrared spectra of products of PFBA at 600 and 780 °C in N<sub>2</sub> (Figure S5), formation of HF during PFPrA and PFBA decomposition at 200 °C (Figure S6), and infrared spectra of products of PFCAs at 300 °C in N<sub>2</sub> and O<sub>2</sub> [(a) PFPrA and (b) PFBA (Figure S7)] (PDF)

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### Notes

The authors declare no competing financial interest.

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## ■ REFERENCES

- (1) Whitehead, H. D.; Venier, M.; Wu, Y.; Eastman, E.; Urbanik, S.; Diamond, M. L.; Shalin, A.; Schwartz-Narbonne, H.; Bruton, T. A.; Blum, A.; Wang, Z.; Green, M.; Tighe, M.; Wilkinson, J. T.; McGuinness, S.; Peaslee, G. F. Fluorinated Compounds in North American Cosmetics. *Environmental Science & Technology Letters* **2021**, 8 (7), 538–544.
- (2) Schultes, L.; Vestergren, R.; Volkova, K.; Westberg, E.; Jacobson, T.; Benskin, J. P. Per- and polyfluoroalkyl substances and fluorine mass balance in cosmetic products from the Swedish market: implications for environmental emissions and human exposure. *Environmental Science: Processes & Impacts* **2018**, 20 (12), 1680–1690.

- (3) Sznajder-Katarzyńska, K.; Surma, M.; Cieślík, I. A Review of Perfluoroalkyl Acids (PFAAs) in terms of sources, applications, human exposure, dietary intake, toxicity, legal regulation, and methods of determination. *J. Chem.* **2019**, *2019*, 2717528.
- (4) Abusallout, I.; Holton, C.; Wang, J.; Hanigan, D. Henry's Law constants of 15 per- and polyfluoroalkyl substances determined by static headspace analysis. *Journal of Hazardous Materials Letters* **2022**, *3*, 100070.
- (5) Buck, R. C.; Franklin, J.; Berger, U.; Conder, J. M.; Cousins, I. T.; de Voogt, P.; Jensen, A. A.; Kannan, K.; Mabury, S. A.; van Leeuwen, S. P. Perfluoroalkyl and polyfluoroalkyl substances in the environment: terminology, classification, and origins. *Integr Environ. Assess Manag* **2011**, *7* (4), 513–41.
- (6) Washington, J. W.; Rankin, K.; Libelo, E. L.; Lynch, D. G.; Cyterski, M. Determining global background soil PFAS loads and the fluorotelomer-based polymer degradation rates that can account for these loads. *Sci. Total Environ.* **2019**, *651* (2), 2444–2449.
- (7) Barzen-Hanson, K. A.; Roberts, S. C.; Choyke, S.; Oetjen, K.; McAlees, A.; Riddell, N.; McCrindle, R.; Ferguson, P. L.; Higgins, C. P.; Field, J. A. Discovery of 40 classes of per- and polyfluoroalkyl substances in historical aqueous film-forming foams (AFFFs) and AFFF-impacted groundwater. *Environ. Sci. Technol.* **2017**, *51* (4), 2047–2057.
- (8) Brusseau, M. L.; Anderson, R. H.; Guo, B. PFAS concentrations in soils: Background levels versus contaminated sites. *Sci. Total Environ.* **2020**, *740*, 140017.
- (9) Lin, H.; Taniyasu, S.; Yamazaki, E.; Wei, S.; Wang, X.; Gai, N.; Kim, J. H.; Eun, H.; Lam, P. K. S.; Yamashita, N. Per- and Polyfluoroalkyl Substances in the Air Particles of Asia: Levels, Seasonality, and Size-Dependent Distribution. *Environ. Sci. Technol.* **2020**, *54* (22), 14182–14191.
- (10) Cordner, A.; De La Rosa, V. Y.; Schaidler, L. A.; Rudel, R. A.; Richter, L.; Brown, P. Guideline levels for PFOA and PFOS in drinking water: the role of scientific uncertainty, risk assessment decisions, and social factors. *J. Expo Sci. Environ. Epidemiol* **2019**, *29* (2), 157–171.
- (11) Huang, K.; Li, Y.; Bu, D.; Fu, J.; Wang, M.; Zhou, W.; Gu, L.; Fu, Y.; Cong, Z.; Hu, B.; Fu, J.; Zhang, A.; Jiang, G. Trophic Magnification of Short-Chain Per- and Polyfluoroalkyl Substances in a Terrestrial Food Chain from the Tibetan Plateau. *Environmental Science & Technology Letters* **2022**, *9* (2), 147–152.
- (12) Evich, M. G.; Davis, M. J. B.; McCord, J. P.; Acrey, B.; Awkerman, J. A.; Knappe, D. R. U.; Lindstrom, A. B.; Speth, T. F.; Tebes-Stevens, C.; Strynar, M. J.; Wang, Z.; Weber, E. J.; Henderson, W. M.; Washington, J. W. Per- and polyfluoroalkyl substances in the environment. *Science* **2022**, *375* (6580), No. eabg9065.
- (13) Wang, Z.; DeWitt, J. C.; Higgins, C. P.; Cousins, I. T. A Never-Ending Story of Per- and Polyfluoroalkyl Substances (PFASs)? *Environ. Sci. Technol.* **2017**, *51* (5), 2508–2518.
- (14) Winchell, L. J.; Ross, J. J.; Wells, M. J. M.; Fonoll, X.; Norton Jr, J. W.; Bell, K. Y. Per- and polyfluoroalkyl substances thermal destruction at water resource recovery facilities: A state of the science review. *Water Environment Research* **2021**, *93* (6), 826–843.
- (15) Wang, F.; Lu, X.; Li, X. Y.; Shih, K. Effectiveness and mechanisms of defluorination of perfluorinated alkyl substances by calcium compounds during waste thermal treatment. *Environ. Sci. Technol.* **2015**, *49* (9), 5672–80.
- (16) Aleksandrov, K.; Gehrmann, H. J.; Hauser, M.; Matzing, H.; Pigeon, D.; Stapf, D.; Wexler, M. Waste incineration of Polytetrafluoroethylene (PTFE) to evaluate potential formation of per- and Poly-Fluorinated Alkyl Substances (PFAS) in flue gas. *Chemosphere* **2019**, *226*, 898–906.
- (17) Duchesne, A. L.; Brown, J. K.; Patch, D. J.; Major, D.; Weber, K. P.; Gerhard, J. I. Remediation of PFAS-Contaminated Soil and Granular Activated Carbon by Smoldering Combustion. *Environ. Sci. Technol.* **2020**, *54* (19), 12631–12640.
- (18) Thoma, E. D.; Wright, R. S.; George, I.; Krause, M.; Presezzi, D.; Villa, V.; Preston, W.; Deshmukh, P.; Kauppi, P.; Zemek, P. G. Pyrolysis processing of PFAS-impacted biosolids, a pilot study. *J. Air Waste Manag Assoc* **2022**, *72* (4), 309–318.
- (19) Kundu, S.; Patel, S.; Halder, P.; Patel, T.; Hedayati Marzbali, M.; Pramanik, B. K.; Paz-Ferreiro, J.; de Figueiredo, C. C.; Bergmann, D.; Surapaneni, A.; Megharaj, M.; Shah, K. Removal of PFASs from biosolids using a semi-pilot scale pyrolysis reactor and the application of biosolids derived biochar for the removal of PFASs from contaminated water. *Environmental Science: Water Research & Technology* **2021**, *7* (3), 638–649.
- (20) Watanabe, N.; Takata, M.; Takemine, S.; Yamamoto, K. Thermal mineralization behavior of PFOA, PFHxA, and PFOS during reactivation of granular activated carbon (GAC) in nitrogen atmosphere. *Environmental Science and Pollution Research* **2018**, *25* (8), 7200–7205.
- (21) Xiao, F.; Sasi, P. C.; Yao, B.; Kubátová, A.; Golovko, S. A.; Golovko, M. Y.; Soli, D. Thermal stability and decomposition of perfluoroalkyl substances on spent granular activated carbon. *Environmental Science & Technology Letters* **2020**, *7* (5), 343–350.
- (22) Sonmez Baghizade, B.; Zhang, Y.; Reuther, J. F.; Saleh, N. B.; Venkatesan, A. K.; Apul, O. G. Thermal Regeneration of Spent Granular Activated Carbon Presents an Opportunity to Break the Forever PFAS Cycle. *Environ. Sci. Technol.* **2021**, *55* (9), 5608–5619.
- (23) Ge, Y.; Liu, Z.-z.; Liu, H.-x.; Feng, J.-K.; Liu, D.-s.; Ge, X.-w. Theoretical study on the degradation reaction mechanism of elimination hydrogen fluoride from perfluoropropionic acid. *Computational and Theoretical Chemistry* **2014**, *1029*, 33–40.
- (24) Altarawneh, M. A theoretical study on the pyrolysis of perfluorobutanoic acid as a model compound for perfluoroalkyl acids. *Tetrahedron Lett.* **2012**, *53* (32), 4070–4073.
- (25) Altarawneh, M.; Almatarneh, M. H.; Dlugogorski, B. Z. Thermal decomposition of perfluorinated carboxylic acids: Kinetic model and theoretical requirements for PFAS incineration. *Chemosphere* **2022**, *286* (2), 131685.
- (26) Wu, J.; Liu, Z.-z.; Liu, H.-x.; Liu, D.-s.; Ge, X.-w. Theoretical Study on Decomposition Mechanism of Perfluoropentanoic Acid. *Neimenggu Daxue Xuebao, Ziran Kexueban* **2013**, No. 3, 17.
- (27) Krusic, P. J.; Marchione, A. A.; Roe, D. C. Gas-phase NMR studies of the thermolysis of perfluorooctanoic acid. *J. Fluorine Chem.* **2005**, *126* (11), 1510–1516.
- (28) Stockenhuber, S.; Weber, N.; Dixon, L.; Lucas, J.; Grimison, C.; Bennett, M.; Stockenhuber, M.; Mackie, J.; Kennedy, E. Thermal Degradation of Perfluorooctanoic Acid (PFOA). 16th International Conference on Environmental Science and Technology Rhodes, Greece, 2019.
- (29) Xiao, F.; Sasi, P. C.; Alinezhad, A.; Golovko, S. A.; Golovko, M. Y.; Spoto, A. Thermal Decomposition of Anionic, Zwitterionic, and Cationic Polyfluoroalkyl Substances in Aqueous Film-Forming Foams. *Environ. Sci. Technol.* **2021**, *55* (14), 9885–9894.
- (30) Sasi, P. C.; Alinezhad, A.; Yao, B.; Kubátová, A.; Golovko, S. A.; Golovko, M. Y.; Xiao, F. Effect of Granular Activated Carbon and Other Porous Materials on Thermal Decomposition of Per- and Polyfluoroalkyl Substances: Mechanisms and Implications for Water Purification. *Water Res.* **2021**, *200*, 117271.
- (31) Wang, J.; Lin, Z.; He, X.; Song, M.; Westerhoff, P.; Doudrick, K.; Hanigan, D. Critical Review of Thermal Decomposition of Per- and Polyfluoroalkyl Substances: Mechanisms and Implications for Thermal Treatment Processes. *Environ. Sci. Technol.* **2022**, *56* (9), 5355–5370.
- (32) Wang, J.; Abusallout, I.; Song, M.; Marfil-Vega, R.; Hanigan, D. Quantification of per- and polyfluoroalkyl substances with a modified total organic carbon analyzer and ion chromatography. *AWWA Water Sci.* **2021**, *3* (4), No. e1235.
- (33) Wang, F.; Lu, X.; Shih, K.; Liu, C. Influence of calcium hydroxide on the fate of perfluorooctanesulfonate under thermal conditions. *Journal of Hazardous Materials* **2011**, *192* (3), 1067–71.
- (34) Watanabe, N.; Takemine, S.; Yamamoto, K.; Haga, Y.; Takata, M. Residual organic fluorinated compounds from thermal treatment of PFOA, PFHxA and PFOS adsorbed onto granular activated carbon



- (GAC). *Journal of Material Cycles and Waste Management* **2016**, *18* (4), 625–630.
- (35) D'Ambro, E. L.; Pye, H. O. T.; Bash, J. O.; Bowyer, J.; Allen, C.; Efsthathiou, C.; Gilliam, R. C.; Reynolds, L.; Talgo, K.; Murphy, B. N. Characterizing the Air Emissions, Transport, and Deposition of Per- and Polyfluoroalkyl Substances from a Fluoropolymer Manufacturing Facility. *Environ. Sci. Technol.* **2021**, *55* (2), 862–870.
- (36) Krug, J. D.; Lemieux, P. M.; Lee, C. W.; Ryan, J. V.; Kariher, P. H.; Shields, E. P.; Wickersham, L. C.; Denison, M. K.; Davis, K. A.; Swensen, D. A.; Burnette, R. P.; Wendt, J. O. L.; Linak, W. P. Combustion of C1 and C2 PFAS: Kinetic modeling and experiments. *J. Air Waste Manag. Assoc.* **2022**, *72* (3), 256–270.
- (37) Aro, R.; Eriksson, U.; Karrman, A.; Reber, I.; Yeung, L. W. Y. Combustion ion chromatography for extractable organofluorine analysis. *iScience* **2021**, *24* (9), 102968.
- (38) Bamdad, H.; Papari, S.; Moreside, E.; Berruti, F. High-Temperature Pyrolysis for Elimination of Per- and Polyfluoroalkyl Substances (PFAS) from Biosolids. *Processes* **2022**, *10* (11), 2187.
- (39) Linstrom, P. J.; Mallard, W. G., Eds. *NIST Chemistry WebBook, NIST Standard Reference Database Number 69*; National Institute of Standards and Technology: Gaithersburg, MD, 20899 DOI: 10.18434/T4D303 (accessed 2022-03-09).
- (40) SpectraBase. <https://spectrabase.com/> John Wiley & Sons, Inc. (accessed 2022-03-09).
- (41) Roehl, C.; Boglu, D.; Brühl, C.; Moortgat, G. Infrared band intensities and global warming potentials of CF<sub>4</sub>, C<sub>2</sub>F<sub>6</sub>, C<sub>3</sub>F<sub>8</sub>, C<sub>4</sub>F<sub>10</sub>, C<sub>5</sub>F<sub>12</sub>, and C<sub>6</sub>F<sub>14</sub>. *Geophysical research letters* **1995**, *22* (7), 815–818.
- (42) Stepanov, S.; Gabriel, O.; Meichsner, J. In Correlation between CF<sub>2</sub>- and C<sub>2</sub>F<sub>4</sub> concentrations in pulsed capacitively coupled CF<sub>4</sub>/H<sub>2</sub> rf plasma. *XXVIII International Conference on Phenomena in Ionised Gases (ICPIG 2007)*; July 15–20, 2007, Prague, Czech Republic; Institute of Physics, 2007; pp 1602–1605.
- (43) Stevens, J. E.; Macomber, L. D.; Davis, L. W. IR spectra and vibrational modes of the hydrofluoroethers CF<sub>3</sub>OCH<sub>3</sub>, CF<sub>3</sub>OCHF<sub>2</sub>H, and CF<sub>3</sub>OCF<sub>2</sub>CF<sub>2</sub>H and corresponding alkanes CF<sub>3</sub>CH<sub>3</sub>, CF<sub>3</sub>CF<sub>2</sub>H, and CF<sub>3</sub>CF<sub>2</sub>CF<sub>2</sub>H. *Open Physical Chemistry Journal* **2010**, *4* (1), 17–27.
- (44) Center, A. T.; Ground, A. P. Measurement of Carbonyl Fluoride, Hydrogen Fluoride, and Other Combustion Byproducts During Fire Suppression Testing by Fourier Transform Infrared Spectroscopy. Halon Options Technical Working Conference; 1998.
- (45) Mashino, M.; Ninomiya, Y.; Kawasaki, M.; Wallington, T. J.; Hurley, M. D. Atmospheric Chemistry of CF<sub>3</sub>CF = CF<sub>2</sub>: Kinetics and Mechanism of Its Reactions with OH Radicals, Cl Atoms, and Ozone. *J. Phys. Chem. A* **2000**, *104* (31), 7255–7260.
- (46) Liang, J.; Safriet, A.; Briley, S.; Roselius, M. Chemisorption study of trifluoroacetyl fluoride on FeF<sub>3</sub> by FT-IR spectroscopy. *J. Fluorine Chem.* **1996**, *79* (1), 53–57.
- (47) Mosiadz, M.; Juda, K. L.; Hopkins, S. C.; Soloducho, J.; Glowacki, B. A. An in-depth in situ IR study of the thermal decomposition of yttrium trifluoroacetate hydrate. *J. Therm. Anal. Calorim.* **2012**, *107* (2), 681–691.
- (48) Malanca, F. E.; Bierbrauer, K. L.; Chiappero, M. S.; Argüello, G. A. Photochemistry of fluorinated compounds. Kinetics of the reaction between CF<sub>3</sub>CF<sub>2</sub> and FCO radicals. *J. Photochem. Photobiol., A* **2002**, *149*, 9.
- (49) Jollie, D.; Harrison, P. An in situ IR study of the thermal decomposition of trifluoroacetic acid. *J. Chem. Soc., Perkin Trans.* **1997**, *2* (8), 1571–1576.
- (50) Moore, L. O. Pyrolysis of heptafluorobutyric anhydride. *Journal of Organic Chemistry* **1970**, *35* (9), 3201–3202.
- (51) Weber, N. H.; Stockenhuber, S. P.; Delva, C. S.; Abu Fara, A.; Grimison, C. C.; Lucas, J. A.; Mackie, J. C.; Stockenhuber, M.; Kennedy, E. M. Kinetics of Decomposition of PFOS Relevant to Thermal Desorption Remediation of Soils. *Ind. Eng. Chem. Res.* **2021**, *60* (25), 9080–9087.
- (52) LaZerte, J. D.; Hals, L. J.; Reid, T. S.; Smith, G. H. Pyrolyses of the Salts of the Perfluoro Carboxylic Acids. *J. Am. Chem. Soc.* **1953**, *75* (18), 4525–4528.
- (53) Atkinson, B.; Trenwith, A. 424. The thermal decomposition of tetrafluoroethylene. *Journal of the Chemical Society (Resumed)* **1953**, 2082–2087.
- (54) Atkinson, B.; Atkinson, V. 401. The thermal decomposition of tetrafluoroethylene. *Journal of the Chemical Society (Resumed)* **1957**, 2086–2094.
- (55) Butler, J. N. The Thermal Decomposition of Octafluorocyclobutane. *J. Am. Chem. Soc.* **1962**, *84* (8), 1393–1398.
- (56) Corbett, P.; Whittle, E. 604. The thermal decomposition of perfluoroacetic anhydride. *Journal of the Chemical Society (Resumed)* **1963**, 3247–3251.
- (57) Steunenbergh, R. K.; Cady, G. H. Pyrolysis of fluorocarbons I. *J. Am. Chem. Soc.* **1952**, *74* (16), 4165–4168.
- (58) Yu, H.; Kennedy, E. M.; Mackie, J. C.; Dlugogorski, B. Z. An Experimental and Kinetic Modeling Study of the Reaction of CHF<sub>3</sub> with Methane. *Environ. Sci. Technol.* **2006**, *40* (18), 5778–5785.
- (59) Yu, H.; Kennedy, E. M.; Ong, W.-H.; Mackie, J. C.; Han, W.; Dlugogorski, B. Z. Experimental and Kinetic Studies of Gas-phase Pyrolysis of n-C<sub>4</sub>F<sub>10</sub>. *Ind. Eng. Chem. Res.* **2008**, *47* (8), 2579–2584.
- (60) Donadio, D.; Ghiringhelli, L. M.; Delle Site, L. Autocatalytic and cooperatively stabilized dissociation of water on a stepped platinum surface. *J. Am. Chem. Soc.* **2012**, *134* (46), 19217–22.
- (61) van Spronsen, M. A.; Frenken, J. W. M.; Groot, I. M. N. Observing the oxidation of platinum. *Nat. Commun.* **2017**, *8* (1), 429.
- (62) Gland, J. L.; Sexton, B. A.; Fisher, G. B. Oxygen interactions with the Pt(111) surface. *Surf. Sci.* **1980**, *95* (2), 587–602.
- (63) Wilf, M.; Dawson, P. T. The adsorption and desorption of oxygen on the Pt(110) surface; A thermal desorption and LEED/AES study. *Surf. Sci.* **1977**, *65* (2), 399–418.
- (64) Peterson, S. D.; Francisco, J. S. Theoretical study of the thermal decomposition pathways of 2-H heptafluoropropane. *J. Phys. Chem. A* **2002**, *106* (13), 3106–3113.
- (65) Palmer, R. L.; Smith Jr, J. N. Molecular beam study of CO oxidation on a (111) platinum surface. *J. Chem. Phys.* **1974**, *60* (4), 1453–1463.
- (66) Altarawneh, M.; Almatarneh, M. H.; Dlugogorski, B. Z. Thermal decomposition of perfluorinated carboxylic acids: Kinetic model and theoretical requirements for PFAS incineration. *Chemosphere* **2022**, *286*, 131685.
- (67) Wang, J.; Lin, Z.; He, X.; Song, M.; Westerhoff, P.; Doudrick, K.; Hanigan, D. Critical Review of Thermal Decomposition of Per- and Polyfluoroalkyl Substances: Mechanisms and Implications for Thermal Treatment Processes. *Environ. Sci. Technol.* **2022**, *56* (9), 5355–5370.
- (68) Crownover, E.; Oberle, D.; Kluger, M.; Heron, G. Perfluoroalkyl and polyfluoroalkyl substances thermal desorption evaluation. *Remediation Journal* **2019**, *29* (4), 77–81.
- (69) Sorengard, M.; Lindh, A. S.; Ahrens, L. Thermal desorption as a high removal remediation technique for soils contaminated with per- and polyfluoroalkyl substances (PFASs). *PLoS One* **2020**, *15* (6), No. e0234476.