

1 A test of the Geographic Mosaic Theory of Coevolution: investigating widespread species of

2 Amazonian *Protium* (Burseraceae) trees, their chemical defenses, and their associated

3 herbivore faunas

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24 **Abstract**

25 Plants and their insect herbivores represent a large fraction of the species in Amazonian forests
26 and are often directly implicated in the origin and maintenance of biodiversity at local and
27 regional scales. How these interactions may change over geographic distance is unknown
28 because very few studies have investigated the herbivore fauna and defense chemicals of any
29 host plant species at multiple sites in tropical forests. One hypothesis, the Geographic Mosaic
30 Theory of Coevolution, predicts that if herbivore assemblages turn over in different parts of a
31 plant's range, then plant defense chemicals should also change, reflecting local selection
32 pressures. We tested this theory by studying 12 species of *Protium* (Burseraceae) trees that occur
33 in both Iquitos, Peru, and Manaus, Brazil, in rainforests separated by 1500 km. We surveyed all
34 insects observed directly feeding on the plants in both locations for 48 weeks in Manaus and 64
35 weeks in Iquitos. We analyzed the secondary metabolites in the leaves of all species in both
36 locations using GC/MS and HPLC. Although in both locations we found that *Protium*
37 herbivores were dominated by insects from the orders Hemiptera, Coleoptera and Lepidoptera,
38 we found almost complete turnover in the herbivore species composition in the two sites, and
39 each host plant species had a different assemblage of herbivores in each location. Comparing the
40 phylogenetic beta-diversity, we found low similarity in herbivore phylogenetic relatedness
41 between host plant species in the two locations. However, the secondary metabolites found
42 within a *Protium* species were similar across the two locations. We found no evidence that
43 individuals from a host plant species in Iquitos or Manaus expressed locally-adapted defense
44 chemicals, as individuals from geographic locations did not form clusters when looking at
45 patterns of chemical similarity. These results are not consistent with the Geographic Mosaic
46 Theory of Coevolution. The most intriguing pattern we found was a strong correlation between

47 the diversity of herbivores per host plant species in both locations. We also found that plants
48 with high chemical diversity had lower numbers of herbivore species and numbers of total
49 herbivores in both locations. We conclude that high chemical diversity is the most effective
50 strategy for *Protium* trees to reduce insect herbivore attacks. We speculate that each secondary
51 metabolite is effective at repelling only a few insect herbivores, and that different chemicals are
52 likely effective in different parts of a plants' geographic range. Future studies should investigate
53 additional locations and additional natural enemies (i.e., fungal pathogens) to test the hypothesis
54 that chemical diversity reduces attack from natural enemies and may explain the ecological and
55 evolutionary success of rainforest trees over time and space.

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57 **Key Words:** rainforest plants, secondary metabolites, insects, chemical ecology, beta-diversity,
58 phylobetadiversity

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70 **Introduction**

71 Natural enemies have been implicated in the origin of plant diversity, especially in hyperdiverse
72 tropical forests where the species richness of plants (and their enemies) reaches its apex (Price
73 2002). Insects are among the most important enemies of tropical tree species (Coley and Barone
74 1996), and of those, caterpillars, beetles, and hemipterans have consistently been reported as the
75 most common and diverse herbivores in tropical systems (Barone, 1998, Bassett 1999, Novotny
76 et al. 2005, 2006, 2007, Sam et al. 2018, Richards et al. 2015, Salazar et al. 2018).

77 In turn, plants have evolved a suite of different chemical defenses to repel attack from
78 insects, with each one of these specialized (i.e., secondary) metabolites having different efficacy
79 against different kinds of herbivores (Volf et al. 2015, Salazar et al. 2018). Across major insect
80 clades, large physiological and metabolic differences in taxa can affect the mechanisms of action
81 of specialized metabolites. For example, caterpillar guts have very high pH compared to beetles,
82 and thus the oxidative biological activity of tannins can differ significantly between these two
83 taxa (Barbehenn and Constabel 2011). The effects of specific tannins on herbivores can range
84 widely; from causing gut lesions, reducing overall nutrition, or sometimes having beneficial
85 effects (Barbehenn and Constabel 2011). Within herbivore groups, the evolution of
86 specialization can allow some species to better overcome or circumvent plant host chemical
87 defenses (Volf et al. 2015). Given the variation in specialized metabolite susceptibility across
88 herbivores, plants would be expected to evolve chemical arsenals that match a particular
89 biological context and herbivore community. In particular, plant species with wide geographical
90 ranges are likely to be challenged by an equally-wide range of biological contexts, with
91 significant changes in herbivore abundance and compositions across their range. Thus, it is
92 reasonable to expect that, if two widely separated populations of plants contained different

93 assemblages of associated herbivores, this may select for different concentrations (or different
94 qualitative mixtures) of different defensive specialized metabolites (Volf et al. 2015).

95 Such coevolutionary interactions among plants and their enemies in tropical forests could
96 lead to speciation when plants experienced different combinations of enemies in different parts
97 of their range (Schemske 2009). The Geographic Mosaic Theory of Coevolution (GMTC) posits
98 that over an organism's geographic range it is likely that it will be involved in complex
99 coevolutionary relationships with a multitude of interacting partners (both mutualistic and
100 antagonistic), and such biotic interactions will vary in strength among geographically-separated
101 subpopulations, creating locally-adapted phenotypes (Thompson 2005). In terms of plant defense
102 and natural enemies, GMTC predicts that one would find natural variation in chemical defenses
103 (qualitative and/or quantitative) in plant populations exposed to significantly different qualitative
104 and quantitative herbivore pressures (Thompson 2005, Salazar and Marquis 2012), depending on
105 the beta-diversity of natural enemies across a host plant's range (Leemu et al. 2012).

106 To test this herbivore-chemistry GMTC, ideally one would evaluate defense chemistry
107 and natural enemies attacking the same species of host plants at different locations throughout
108 their geographic range (at similar elevations, habitats and latitudes to control for the effect on
109 environmental changes and other confounding factors). To date, only a few studies have
110 evaluated the beta-diversity of insect herbivores sampling the same host plant species at different
111 geographic locations, with some studies showing low beta diversity (Novotny et al. 2006), and
112 others showing the opposite pattern (Basset 2000, Novotny et al. 2005, Salazar and Marquis
113 2012, Fine et al. 2013). Even fewer studies have investigated host plant chemical defenses in
114 conjunction with insect herbivore turnover, although there are many studies showing large
115 variation in plant chemistry among populations (Mooney and Emboden 1968, Macedo and

116 Langenheim 1989, Berenbaum and Zangerl 1998, Zangerl and Berenbaum 2003, Leemu et al.
117 2012). Most of these studies have shown that plant defense chemistry within a taxon varies
118 among populations across geographic distances and habitats (but see Endara et al. 2018). Most of
119 these previous studies have limited their comparisons to one or a few insect clades (i.e.,
120 Lepidoptera) and a subset of defense chemicals (i.e., alkaloids). A comprehensive evaluation of
121 all insect herbivores and all specialized metabolites for widely-separated populations of host
122 plants is needed to better understand the link between change in herbivore composition and host
123 defense chemistry across large geographic distances. This will allow an evaluation of the
124 selection pressures of the community of herbivores as well as the local adaptation of defense
125 chemistry by the host plants.

126 *Protium* (Burseraceae) has been the subject of multiple studies regarding plant defense
127 chemistry and/or insect herbivores (Siani et al. 2012, Lokvam et al. 2015, Fine et al. 2013,
128 Salazar et al. 2018). It represents an ideal study system for investigating Amazonian tree
129 diversity because of its high species richness (Daly et al. 2022), large number of coexisting
130 species (Vleminckx et al. 2018), high relative abundance (ter Steege et al. 2013), up-to-date
131 taxonomy (Daly and Fine 2018) and phylogeny (Fine et al. 2014).

132 Recently we reported on the chemical diversity of *Protium* (Burseraceae) and its
133 associated insect herbivore fauna in the Allpahuayo-Mishana National Reserve near Iquitos, Peru
134 (Salazar et al. 2018). To test predictions of the GMTC and assess the link between plant
135 herbivore turnover and specialized metabolite diversity we investigated *Protium*-insect herbivore
136 interactions in a distant but comparable region in the Amazon basin. Here, we surveyed insect
137 herbivores on 12 *Protium* species in rain forests near Manaus, Brazil for twelve-months (on
138 average 15 *Protium* individuals per species). We selected 12 species also found in Iquitos, Peru,

139 1500 km away. We subjected each *Protium* species in Manaus to the same detailed analysis of
140 their secondary metabolite composition performed in our Iquitos study.

141 We asked: 1) How much change in herbivore species composition can be observed in the tree
142 genus *Protium* between Iquitos and Manaus?

143 2) How different is the *Protium* secondary metabolite composition in species occurring in
144 both sites?

145 3) How are chemical diversity and herbivore diversity related in the two locations?

146 Given the large distance between the two sites, we predicted to find significant
147 differences in the insect herbivore fauna associated with *Protium* species. Because insect
148 herbivores should contribute to divergent natural selection for effective chemical defenses in
149 *Protium* species, we also predicted that the magnitude of the herbivore species turnover would be
150 mirrored by similar changes in the host species defense chemistry across the two sites.

151

152 **Methods**

153 **Background**

154 In a recently published work, the authors studied 860 individuals from 31 sympatric
155 species of *Protium* within a 10 ha area of intact rainforest in Iquitos, Peru, and surveyed the
156 insect herbivores feeding on the plants for 62 weeks (Salazar et al. 2018). We analyzed the
157 secondary metabolite composition of these plants, finding almost 600 specialized metabolites.
158 We found 231 different insect herbivore taxa out of 4214 individual confirmed feeding records,
159 mostly from the orders Hemiptera, Lepidoptera and Coleoptera. *Protium* species with high
160 chemical diversity had significantly fewer numbers and fewer species of insect herbivores

161 feeding on them (Salazar et al. 2018). Here we leveraged and expanded this dataset to a second
162 site and used the combined dataset to investigate a new set of research questions.

163 **Trees**

164 *Protium* (Burseraceae) includes more than 140 species of mostly Neotropical trees,
165 ranging from subcanopy treelets to canopy trees (Daly and Fine 2018). The highest diversity and
166 abundance of *Protium* is found in the Amazon basin where single sites can harbor up to 35
167 sympatric species (Rankin-de-Morona et al. 1992, Daly et al. 2022), and where *Protium* is often
168 among the top five genera in terms of abundance throughout the basin (ter Steege et al. 2013). A
169 phylogenetic analysis has proposed a well-resolved phylogeny of *Protium* (Fine et al. 2014).
170 This work includes all species used for the current study, although some of the species names
171 have changed. Thus, some of the species we reported in Salazar et al. (2018) and Fine et al.
172 (2014) have been updated, and we include a table in the Appendix that lists all species in our
173 study with how their names may have changed since the previous publications (table S1).

174

175 ***Protium* herbivore surveys in Peru**

176 Complete details on this dataset are published in Salazar et al. (2018). In short, we
177 sampled 31 species of *Protium* in the Allpahuayo-Mishana Reserve using forest transects. To
178 balance our sampling effort across all *Protium* species, 18–25 individuals of each species were
179 sampled (a total of 860 *Protium* plants). Individuals were censused for herbivores at least once a
180 week for 64 consecutive weeks (May 2013 to September 2014). Each plant census included two
181 steps: observation and collection. The observation step was implemented to record and collect
182 only the herbivores that were observed feeding directly on the host plant. Collected herbivores
183 were preserved in 80% ethanol and later assigned to a morphospecies using an “*in-house*”

184 reference collection (deposited in the Essig museum of Entomology, UC Berkeley; Berkeley,
185 CA). In total, the herbivore surveys quantified 4,214 feeding events from 231 herbivore
186 morphospecies. Of all morphospecies in the dataset, 90 were singletons (found only once) and
187 the 50 most abundant species accounted for 87% of total records. This sampling scheme was
188 designed to accurately assess the quantitative and qualitative herbivore pressure experienced by
189 the focal host-plant species, not to determine the ‘complete’ diet breadth of each herbivore
190 observed.

191

192 ***Protium* herbivore surveys in Brazil**

193 In Manaus, we surveyed juvenile *Protium* trees at two locations, the Reserva Ducke and
194 Reserva da Campina. The Reserva da Campina contains white-sand forests similar to the white-
195 sand forests found in the Allpahuayo-Mishana Reserve in Iquitos; the Reserva Ducke has terra
196 firme forests on sandy soil and clay soils similar to Allpahuayo-Mishana terra firme forests. We
197 selected 12 *Protium* species for weekly monitoring also found in Iquitos. Data and specimen
198 collection mirrored the protocol followed in Iquitos, Peru (Salazar et al. (2018), see above). In
199 short, we surveyed 8-17 individuals per *Protium* species once a week (from November 2014 to
200 October 2015) and collected insects that were observed feeding directly on the plants. Reserva da
201 Campina is 36 km north of Reserva Ducke and was included in the weekly monitoring to
202 increase the number of plant species surveyed. We sampled the two areas on the same day and in
203 sequence but with alternating visit order. One week we visited Reserva Ducke first; the
204 following week, it was Reserva da Campina to avoid sampling the plants at the same time of the
205 day. We monitored 68 plants in Reserva da Campina and 112 in Reserva Ducke. Two species
206 were only found at Campinas: *P. heptaphyllum* and *P. subserratum* morphotype 3. Only *P.*

207 *calanense* and *P. hebetatum* were sampled at both places. We observed 966 feeding records from
208 163 insect morphospecies.

209

210 **Insect molecular work (Manaus and Iquitos)**

211 Insect DNA barcoding was conducted at the “morphospecies” level. For each putative
212 insect species (morphospecies) we sequenced at least 5 individuals (more samples were used for
213 morphologically variable taxa and recalcitrant morphospecies). A small sample from each insect
214 (a leg or a small fraction the abdomen) was washed with distilled water and dried to remove all
215 traces of ethanol. DNA extraction and PCR amplification of the COI gene were accomplished
216 using standard methods (see Salazar et al. (2018)). Alignment was done using the MUSCLE
217 algorithm (Edgar 2004). To determine the best substitution model for our data we used the R
218 package Phangorn via AIC (Schliep 2011). Phylogenetic reconstruction was carried out using 10
219 independent runs under a GTR+G+I model for 3 million generations (four chains, 25% burn-in,
220 sampled every 100 generations) using MrBayes 3.2.6 (Ronquist and Huelsenbeck 2003).

221

222 **Plant secondary metabolite analysis (Iquitos and Manaus):**

223 The Burseraceae (Frankincense and Myrrh family), are well known for the production of
224 terpene resins, and *Protium* is no exception. Before this study, almost 100 different terpene
225 compounds had been characterized for *Protium* by gas-chromatography mass-spectroscopy
226 (Rudiger et al. 2007). These compounds have been shown to have medicinal properties (Rudiger
227 et al. 2007, Siani et al. 2012) and are widely believed to have defensive function against
228 pathogens and insect herbivores (Langenheim 2003). *Protium* species also produce a wide array

229 of non-volatile secondary compounds, including polymeric flavans, flavones, oxidized terpenes,
230 and quinic acids, among others (Siani et al. 2012, Lokvam and Fine 2012, Lokvam et al. 2015).
231 Chemical analysis was conducted at the plant species level. Between six and nine different
232 individuals per species were analyzed. Young and mature leaf samples were collected in silica
233 gel from the same marked individuals within the herbivore observation transects and transported
234 to the University of California, Berkeley for analysis. To assess the maximum number of
235 secondary compounds, we performed separate analyses for high- and low-molecular-weight
236 metabolites (GCMS and HPLC-MS respectively) using a non-targeted metabolomic approach.
237 Find more details of the methods and chemical composition of the study species in Salazar et al.
238 (2018).

239 ***Low-molecular-weight metabolite chemistry (GC/MS)***

240 Of the twelve species surveyed for herbivores, only 11 could be used for the final GC/MS
241 analysis. *Protium rhoifolium* was not included due to having fewer than 3 samples. GC/MS
242 methods follow Salazar et al. (2018). In short, 100 mg of dry leaf material was pulverized and
243 extracted using 150 microliters of a 1:4 solution of ethanol: dichloromethane with 0.075 mg/l of
244 Piperine as an internal standard. We injected 2.5 microliters of plant extract. The inlet was kept
245 at a constant temperature of 275°C (split injection 1:40). The oven was programmed as follows:
246 85° C, hold for 2 min; ramp 1: 10° C /min; 155° C, hold for 1 min; ramp 2: 6° C /min; 260° C,
247 hold for 1 min; ramp 3: 2° C /min; 300° C hold for 14 min (total run time 60 min). MS
248 conditions: EI @ 70eV, scanning range 40-550 amu. Raw GC/MS data were processed in R with
249 the package XCMS (script available upon request) using the approach by Salazar et al. (2016a,
250 b). Chromatographic features from species were putatively classified using AMDIS (Automated
251 Mass Spectral Deconvolution and Identification System) and the NIST/EPA/NIH and MassBank

252 Databases (Horai et al. 2010). Metabolites that did not have a match from the available mass
253 spectra databases or in the available literature were classified as unknown.

254 ***High-molecular-weight metabolite chemistry (HPLC-MS).***

255 Leaf samples from our study species were pulverized and 100 mg samples were extracted using
256 1.5 mL of hexane (three times), 4:1 (v/v) ethanol/aq 0.5% acetic acid (4 times) and 7:3 (v/v)
257 acetone/aq 0.5% acetic acid (3 times). Aqueous ethanol and acetone extracts were combined and
258 dried under nitrogen. The mass of each extract was recorded before analysis by HPLC. Samples
259 were dissolved in 1:1 methanol and methyl sulfoxide (0.1% (v/v) formic acid) at the rate of 100
260 $\mu\text{g}/\mu\text{L}$. The sample solvent included gallic acid, 4-hydroxy benzoic acid, sinapinic acid, flavone,
261 chrysin and α -tocopherol as internal standards. Injection volumes were 3. HPLC analyses were
262 carried out using an Agilent 1100 system (ESIMS) using an ion trap mass detector (LCQ Fleet,
263 Thermo Scientific, Waltham, MA, USA). Separation was done using an Atlantis T3 2×150 mm
264 3μ ODS HPLC column (@ 40° C; Waters Corp., Milford, MA, USA). Gradient elution was
265 performed as follows: (Solvent **A**) 1:1 methanol and acetonitrile and (Solvent **B**) water, both
266 acidified with 0.1% (v/v) formic acid. At time (t, min) = 0, 5% **A** in **B**; at t = 10, 15% **A** in **B**; at t
267 = 60, 30% **A** in **B**; at t = 75, 70% **A** in **B**; at t = 100, 95% **A** in **B**; at t = 135, 100% **A**. This was
268 followed by an isocratic step of 100% **A** to t = 150. ESIMS data were collected in the negative
269 ion mode from t = 0-80, positive ion mode t = 80-114 and negative ion mode t = 114-150. HPLC
270 raw data was processed using MZmine (Pluskal et al. 2010).

271

272 **Data analysis**

273 ***Protium herbivore and chemical turnover***

274 To visualize the differences of *Protium* herbivore composition between our two sites we
275 used R (packages Picante, Ape, and Bipartite) to map the number of observed and confirmed
276 feeding records of each morphospecies on the herbivore phylogeny. Similarly, we used R to
277 generate a Sankey Network to visualize the per-host species differences in herbivore community
278 composition between Manaus and Iquitos. We assessed the herbivore community turnover by
279 calculating the length-based phylogenetic beta diversity using the Phylosor algorithm (Bryant et
280 al. 2008, Leprieur et al. 2012). To determine if our empirical values of phylogenetic beta
281 diversity were within the random expectation, for each *Protium* species/site combination we
282 created a null distribution by estimating the phylogenetic beta diversity on a set of 1000
283 randomly-generated herbivore communities (phylogeny label swapping). To quantify the
284 turnover in *Protium* species chemical composition we calculated the Morisita chemical similarity
285 index (Salazar et al. 2016a, b). In short, we used the MZmine workflow to create matrices of
286 shared chromatographic features across all samples of the same species. These matrices are made
287 by comparing the m/z data across the retention time component. These matrices were compared
288 both using presence absence and abundance-weighted data. To visualize the changes in chemical
289 composition of the focus *Protium* species between our two distance sites we performed chemical
290 hierarchical clustering of multiple individuals of each species/site combination.

291 ***Relationship between herbivore community and host species chemistry***

292 We performed a simple linear regression between plant species chemical richness
293 (number of chemical compounds detected across our two analytical approaches) and the plant
294 host herbivore species load (total number of herbivore species observed feeding on a *Protium*
295 species). Finally, to assess the similarity between herbivore loads for each species across our two

296 sites we regressed the *Protium* species herbivore richness found in Manaus with the one found in
297 Iquitos.

298

299 **RESULTS**

300 **Turnover in the *Protium* herbivore assemblages between Iquitos and Manaus**

301 We found very high species turnover between sites. Only four herbivore species were
302 found at both sites (3 Lepidoptera and one Hemiptera, see Figure 1). However, three of these
303 species were only found once at one of the two sites. Therefore, we found only one common
304 species at both sites, a tortricid leaf roller (morphospecies “Cabeza Negra -LJ04-“, Tortricidae).
305 At both sites the most diverse and abundant group of herbivores were the Hemipterans (mainly
306 leaf hoppers – Cicadellidae-), followed by Coleopterans (mainly leaf beetles -Chrysomelidae)
307 and caterpillars (Lepidoptera – various families-; Figure 1).

308 Given the high species turnover we assessed branch-length based phylogenetic beta
309 diversity (Phylosor) (Bryant et al. 2008, Leprieur et al. 2012). Here, the phylogenetic beta
310 diversity for an herbivore assemblage in the same host species in the two locations was lower
311 than the average random expectation, although these differences were not statistically significant
312 for most *Protium* species (Fig S1). Differences in herbivore composition for one site are not
313 based on subsets of herbivore composition of the other site; they derive from completely
314 different herbivore species that are found at one site or the other (Fig S2). Most insect herbivores
315 were associated with multiple species of *Protium* (Figure 2).

316

317 **Change in *Protium* defense chemical composition between Iquitos and Manaus**

318 Contrary to our expectations, we did not find large differences in secondary metabolite
319 composition between the Manaus and Iquitos populations for the twelve focal species. Although
320 we did find some intraspecific variation, the variation in secondary metabolites between
321 individuals of different populations (Iquitos and Manaus) was not larger than the variation within
322 the same population. A non-targeted metabolomic analysis of both GC/MS and HPLC-MS data
323 shows that individuals of one population appear nested within the clusters of the other population
324 (Figure 3). The presence/absence chemical similarity composition (Morisita chemical similarity
325 index) for the most dominant chromatographic features (specialized metabolites that comprise at
326 least 0.1% of the total chromatogram integrated area) is 1 for most sample pairs (range=0.97-1).
327 This result highlights that, in the vast majority of cases, the specialized metabolites expressed by
328 a *Protium* species in Iquitos are the same metabolites expressed in Manaus (in Fig S3, we
329 provide as an example of chromatograms of *Protium ferrugineum* from both sites) Similarly,
330 when we accounted for the changes in relative abundance of specialized secondary compounds
331 between our two sites, similarity between samples ranged from 0.74 to 0.98, with no obvious
332 difference between and across populations. (Table S2).

333

334 **The link between *Protium* herbivore community and host species chemistry**

335 The most notable finding of this study is that, despite the large changes in herbivore community
336 composition, the relationship between a *Protium* species' chemical diversity and its herbivore
337 species richness was very consistent in both Iquitos and Manaus. In both locations we found a
338 negative relationship between host species chemical richness and the number of herbivore
339 species found feeding on a particular *Protium* taxon (Figure 4). Moreover, our most fascinating
340 finding was, despite the extremely large herbivore species turnover, the herbivore species

341 richness of a *Protium* species in one site predicts the species richness of the same species in a
342 different site (Figure 4).

343

344 **DISCUSSION**

345 We found almost complete turnover in insect herbivore species composition between
346 hosts of the same species of *Protium* in Iquitos and Manaus. In both locations, insects from the
347 orders Hemiptera, Coleoptera and Lepidoptera were the most common herbivores feeding on the
348 plants, in that rank order (Figure 1). However, there was minimal phylogenetic similarity in
349 herbivore assemblages within a species across the two sites. These differences in herbivore
350 faunas in host plants in the two locations therefore do not represent a simple replacement by
351 herbivore sister taxa or close relatives. For example, insects associated with *Protium hebetatum*
352 in Iquitos were phylogenetically dissimilar to insects associated with *Protium hebetatum* in
353 Manaus; much more dissimilar than a random expectation (Figure S1). This pattern was repeated
354 in all of the twelve species that we surveyed in the two locations. This result is contrary to the
355 expectation of the co-evolutionary arms race hypothesis (Ehrlich and Raven 1964) where closely
356 related specialist herbivores would be associated with the same (or closely-related) host plant
357 species and instead is more in agreement with the host-tracking hypothesis of Endara et al.
358 (2017). Also similar in both sites was the pattern of insect herbivores feeding on multiple species
359 of *Protium*. Strict monophagy was extremely rare in both locations (Figure 2).

360 The degree of host specialization and spatial turnover of herbivorous insect communities
361 in tropical forests remains a subject of intense interest, ever since Erwin (1982) published his
362 estimate of total herbivore species richness based on fogging several canopies of a single species
363 of tropical tree. Novotny et al. (2007) found that in contrast to Erwin's expectation, there was

364 very little species turnover in four guilds of herbivorous insects even across more than 500 km of
365 Papua New Guinea lowland forest. However, when sampling across an elevational gradient,
366 Novotny et al. (2005) found strong differences in herbivore assemblages in three widespread
367 species of *Ficus*. Each tree species was associated with a different dominant lepidopteran
368 herbivore species in the lowlands compared to montane habitats. Similar results of high beta-
369 diversity across elevational habitats have been reported in lepidopteran herbivores specialized on
370 *Piper* (Rodríguez-Castañeda et al. 2010).

371 Fine et al. (2013) studied two morphotypes of *Protium subserratum* (the same two as in
372 this paper) in four geographic sites. Two of these sites were white-sand forests and two were
373 terra firme forests. They found strong patterns of high beta-diversity of insect herbivores
374 between habitats and also between the same habitat on opposite sides of the Ucayali River, a
375 large biogeographic barrier. Even though the sites were only 150 km apart, terra-firme vs terra-
376 firme comparisons were 62% dissimilar and white-sand vs white-sand comparisons were 70%
377 dissimilar. This represented a greater dissimilarity than the white-sand vs terra firme
378 comparisons on the same side of the river (37% and 51% dissimilarity, respectively). These
379 findings and the findings in the current study are very different from the low beta-diversity of
380 insect assemblages reported in Novotny et al. (2006) in Papua New Guinea lowland forest.
381 However, although the lowland forest sites in the Papua New Guinea study were 500 km apart
382 they did not cross any major biogeographic barriers like wide rivers or mountains.

383

384 **Chemical defense turnover is much lower than herbivore assemblage turnover.** We found a
385 striking similarity of secondary metabolite composition within *Protium* species in qualitative and
386 quantitative investment in two sites separated by 1500 km. The twelve *Protium* species have

387 much more similar chemical defenses across the two sites than their associated insect
388 assemblages. In addition, there were no population-level signatures in chemical composition
389 (Figure 3). However, we did find some individual-level differences in both unique specialized
390 metabolites and concentrations of shared metabolites, and at this point we do not know how
391 important these differences are or whether they deter herbivores differently. We can confirm that
392 the “herbivore active metabolites” (HAMs) that we identified with the least absolute shrinkage
393 and selection operator (LASSO) approach in Iquitos (Salazar et al. 2018) also do appear in
394 Manaus. Still, we cannot draw any conclusions about whether these metabolites are also
395 associated with repelling herbivores in Manaus. The LASSO method requires a larger sample
396 size than the limited subset of species we surveyed in Manaus as compared to Iquitos (Hastie et
397 al. 2009).

398 Similar consistent chemical defense allocation within species across large distances has
399 been reported in other tropical tree genera. Endara (et al. 2018) characterized metabolomes from
400 eight species of *Inga* (Fabaceae) that had widespread geographic ranges and found that a
401 majority of them had very similar secondary metabolite profiles even comparing conspecifics as
402 far away as Panama, southern Peru, French Guiana, Manaus, and Amazonian Ecuador. *Bursera*
403 showed very little differences in terpenes in populations of the same species across 18 different
404 locations in Mexican tropical dry forest (Becerra et al 2007). Trees, with a much slower
405 generation time than insect herbivores, may not be able to adapt to local herbivore communities,
406 which, with their faster generation times, are likely to change rapidly across time and space.
407 Moreover, Neotropical trees have a great dispersal ability and many appear able to maintain gene
408 flow across huge distances (Dexter et al. 2017).

409

410 **Evolution of chemical diversity.**

411 Comparing two forests 1500km apart, we found strong patterns of turnover in insect
412 herbivores feeding on the same species of plant hosts but correspondingly small differences in
413 secondary metabolites. This does not fit predictions of the Geographic Mosaic Theory of
414 Coevolution perhaps in large part because the most important herbivores in this system appear to
415 be generalists, not specialists. Previous studies reporting evidence of GMTC have studied
416 specialist herbivores and pollinators and reported variation in chemical defenses known to
417 mediate specialist interactions (Thompson 1997, Berenbaum and Zangerl 2006). The scenario
418 underlying the GMTC assumes that either local selective pressures or
419 geographical/environmental barriers will be strong enough to prevent gene flow over generations
420 across populations of host plants. If this is not the case, gene flow could help prevent speciation
421 events in host plants while spreading new defensive alleles/genes across the range. The outcome
422 of such a mechanism would be species with large geographical ranges, high regional (gamma)
423 chemical diversity (but low beta chemical diversity), and a high beta diversity for their
424 associated herbivore and pathogen communities, which is consistent with the findings of this
425 study.

426 The most intriguing result in our study is the strong, significant correlation showing that a
427 *Protium* species that was attacked by the fewest herbivore species in Iquitos was also attacked by
428 the fewest herbivore species in Manaus, despite the fact that almost 100% of the insect
429 herbivores were different in the two locations (Figure 4). Our hypothesis to explain this
430 phenomenon is that having high chemical diversity is the best strategy available for *Protium*
431 plants to prevent attack by a large number of different natural enemies. Having a high chemical
432 diversity correlates with having the largest number of effective chemicals (Salazar et al. 2018),

433 and these effective chemicals are likely to be effective only in a subset of a plant's geographic
434 distribution, meaning that plants with high chemical diversity are more likely to be successful in
435 more biotic contexts across space and time.

436 Other studies on diverse Neotropical woody plant clades have concluded that secondary
437 metabolites are key to understanding local plant diversity and abundance (Salazar et al. 2012,
438 Endara et al. 2015, Kursar et al. 2009, Coley and Kursar 2014, Sedio et al. 2017, Coley et al
439 2018). Richards et al. (2015) found chemical diversity to be key in understanding local diversity
440 patterns of *Piper*, although the relationship between chemical abundance and herbivore diversity
441 they found is opposite of the patterns we report in *Protium*. A key point to underline is that the
442 *Protium* system is dominated by generalists and therefore, evolutionary pressure on particular
443 defense chemicals is not likely as strong as it would be with a plant clade mainly attacked by
444 specialist herbivores. We encourage investigators of other tropical plant clades to expand their
445 sampling of natural enemies to include non-Lepidopteran herbivores in order to be able to make
446 comparisons with the patterns we have uncovered in *Protium*.

447 We still need more evidence to test the hypothesis that high chemical diversity is an
448 effective strategy to reduce attack from natural enemies, as we have only been able to associate a
449 small number of the secondary metabolites in *Protium* with deterring particular herbivores
450 (Salazar et al. 2018). Yet, it is well known that there are other important natural enemies which
451 are deterred by secondary metabolites (Langenheim 2003). For example, future studies could test
452 how *Protium* secondary metabolites are associated with fungal pathogens across their geographic
453 distribution to get a more comprehensive picture of the forces influencing the evolution of
454 chemical diversity. Finally, it will be important to test the hypothesis that high chemical diversity
455 drives plant relative abundance and geographic distribution by studying these species' enemy-

456 secondary metabolite interactions at more locations throughout the Amazon basin and relating
457 this to local and regional abundance.

458

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696

697 **Figure Captions**

698

699 **Figure 1:** Phylogenetic structure of the herbivore community assemblies for the Iquitos and
700 Manaus field sites. Bar and clade color represent the taxonomic order of the herbivores (Green=
701 Coleoptera, Orange=Lepidoptera, Purple=Hemiptera. Bars represent the relative abundance (%)
702 of total abundance) of each herbivore species. Red horizontal lines indicate the four species
703 found at both sites.

704

705 **Figure 2:** Plant-herbivore community networks from Iquitos and Manaus. The left column
706 shows herbivores from Iquitos, the central column shows *Protium* species sampled in both sites,
707 and the right column shows herbivores from Manaus. Data shows only herbivores above the 25th
708 abundance percentile, to exclude singletons and less abundant species. Green = Hemiptera,
709 Orange = Coleoptera, Purple = Lepidoptera.

710

711 **Figure 3. Chemical similarity of *Protium* species found in both Manaus and Iquitos.** A
712 hierarchical clustering (Ward's method) based on the chemical composition of each individual
713 sample in the two populations. Samples are colored by site.

714

715 **Figure 4:** Panel A: association between *Protium* species herbivore species richness in Iquitos
716 and Manaus (12 species shown). Panels B (Manaus) and C (Iquitos): The relationship between a
717 *Protium* species metabolite richness and the species' herbivore richness (11 species shown).
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