

## Parasite infectivity and phycotoxins

1 Infectivity of the parasite *Metschnikowia bicuspidata* is decreased by time spent as a transmission  
2 spore, but exposure to phycotoxins in the water column has no effect

3

4 Kristel F. Sánchez, Department of Ecology and Evolutionary Biology, University of Michigan, Ann Arbor,  
5 MI

6 Baili Zhong, Department of Ecology and Evolutionary Biology, University of Michigan, Ann Arbor, MI and  
7 Chemistry Department, Pomona College, Claremont, CA

8 Jorge A. Agudelo, Department of Ecology and Evolutionary Biology, University of Michigan, Ann Arbor, MI  
9 and Department of Chemistry and Biochemistry, Florida International University, Miami, FL

10 Meghan A. Duffy, Department of Ecology and Evolutionary Biology, University of Michigan, Ann Arbor,  
11 MI; ORCID: 0000-0002-8142-0802

12

13 **Keywords:** microcystin, anatoxin, host-parasite, cyanobacteria, *Daphnia*

14 **Corresponding author's email:** kfsanche@umich.edu

15

16     **Summary**

17     1. Transmission from one host to another is a crucial component of parasite fitness. For some aquatic  
18     parasites, transmission occurs via a free-living stage that spends time in the water, awaiting an encounter  
19     with a new host. These parasite transmission stages can be impacted by biotic and abiotic factors that  
20     influence the parasite's ability to successfully infect or grow in a new host.

21     2. Here we tested whether time spent in the water column and/or exposure to common cyanobacterial  
22     toxins impacted parasite transmission stages. More specifically, we tested whether the infectivity, within  
23     host growth, and virulence of the fungal parasite *Metschnikowia bicuspidata* changed as a result of time  
24     spent in the water or from exposure to cyanotoxins in the water column. We exposed parasite  
25     transmission spores to different levels of one of two ecologically important cyanotoxins, microcystin-LR  
26     and anatoxin-a, and factorially manipulated the amount of time spores were incubated in water. We  
27     removed the toxins and used those same spores to infect one genotype of the common lake zooplankton  
28     *Daphnia dentifera*.

29     3. We found that cyanotoxins did not impact parasite fitness (infection prevalence and spore yield per  
30     infected host) or virulence (host lifetime reproduction and survivorship) at the tested concentrations  
31     (10 $\mu$ g/L & 30 $\mu$ g/L). However, we found that spending longer as a transmission spore decreased a spore's  
32     chances for successful infection: spores that were only incubated for 24 hours infected approximately  
33     75% of exposed hosts, whereas spores incubated for 10 days infected less than 50% of exposed hosts.

34     4. We also found a negative relationship between the final spore yield from infected hosts and the  
35     proportion of hosts that became infected. In treatments where spores spent longer in the water column  
36     prior to encountering a host, infection prevalence was lower (indicating lower per spore infectivity), but  
37     each infected host yielded more spores at the end of infection. We hypothesize that this pattern may  
38     result from intraspecific parasite competition within the host.

39 5. Overall, these results suggest that transmission spores of this parasite are not strongly influenced by  
40 cyanotoxins in the water column, but that other aspects of spending time in the water strongly influence  
41 parasite fitness.

42 **Introduction**

43 A crucial component of parasite fitness is successfully transmitting from one host to another. For  
44 many parasites, transmission occurs via a free-living stage that spends time in the environment until it  
45 encounters a suitable host. In aquatic ecosystems, these parasite transmission spores spend time in the  
46 water column (before sinking out or moving on to their next life history stage), where they are exposed  
47 to abiotic and biotic factors that can ultimately affect the timing, size, and duration of an epidemic.

48 While we know these transmission stages are very common for aquatic parasites and that many biotic  
49 and abiotic factors impact their transmission to hosts, we know less about how prolonged exposure to  
50 environmental factors such as cyanotoxins influences their ability to infect and propagate in new hosts.

51 One factor that likely influences the fitness of parasite transmission stages is time spent in the  
52 environment. In addition to the risks of spore predation (Thielges *et al.*, 2013) and settling out of the  
53 water column (Brookes *et al.*, 2004), spores may lose viability over time, perhaps due to depleting the  
54 resources they acquired from their host (Amigó *et al.*, 1996; Malone, Gatehouse & Tregidga, 2001; Duffy  
55 & Hunsberger, 2019). Longer time in the environment also increases spore exposure to stressors such as  
56 light, temperature, secondary metabolites produced by other organisms, and anthropogenic pollutants  
57 (Heagle, 1973; Amigó *et al.*, 1996; Hallmann & Sikora, 1996; Vasemagi, Visse & Kisand, 2017; Shaw *et al.*,  
58 2020). Together, this suggests that increased time in the water column should decrease parasite fitness.

59 When they are in the water column, parasite spores are exposed to secondary metabolites. These  
60 are often released from phytoplankton cells, becoming dissolved in water (Huisman *et al.*, 2018); some  
61 are known to have antibacterial and antifungal properties (Borowitzka, 1995; Ostensvik *et al.*, 2002; Volk  
62 & Furtkert, 2006; Leflaive & Ten-Hage, 2007). While there is no clear consensus on why cyanobacteria  
63 produce these phycotoxins, one possibility is that these antimicrobial compounds are released in the  
64 water as part of an allelopathic attack by phytoplankton to deter enemies (Leflaive & Ten-Hage, 2007);

65 these same compounds may have a negative impact on the transmission spores of parasites. If so, this  
66 could impact parasite epidemics in aquatic ecosystems (Lafferty & Holt, 2003). Thus, we were interested  
67 in understanding how cyanotoxins might impact parasite transmission stages in the water.

68 Understanding the impact of cyanotoxins on parasite transmission stages is particularly important  
69 given that cyanobacterial harmful algal blooms (CHABs) are expected to increase in frequency and  
70 intensity in the coming years with climate change (Huisman *et al.*, 2018). CHABs and their toxins can  
71 increase infection rates (by making hosts more vulnerable to infection) and increase the negative effects  
72 of parasites on their hosts (Landsberg, 2002; Harvell *et al.*, 2004) including important zooplankton in  
73 freshwater ecosystems (Tellenbach *et al.*, 2016); however, positive effects on aquatic hosts of consuming  
74 chemically defended cyanobacteria have also been observed (Coopman *et al.*, 2014; Manzi *et al.*, 2019;  
75 Sánchez *et al.*, 2019). When considering the potential impact of CHABs on parasitism, to date the focus  
76 has been on one main mechanism: that toxins may alter host susceptibility or tolerance (Landsberg,  
77 2002; Harvell *et al.*, 2004; Coopman *et al.*, 2014; Penczykowski *et al.*, 2014; Andersen *et al.*, 2016;  
78 Tellenbach *et al.*, 2016; Duperron *et al.*, 2019; Lassudrie *et al.*, 2020). Here, because some cyanotoxins  
79 are known to have antimicrobial properties, we consider the possibility that cyanotoxins have direct  
80 negative effects on parasite transmission spores in the water column.

81 Using the *Daphnia-Metschnikowia* host-parasite system, we tested whether time spent in the water  
82 column and cyanotoxins have negative consequences on the parasite's ability to infect and grow within a  
83 host, both of which are key components of parasite fitness. A variety of biotic and abiotic factors are  
84 known to impact *Metschnikowia* epidemics in lakes (Cáceres *et al.*, 2006; Hall *et al.*, 2010). In this study,  
85 we were concerned with how the duration of exposure to the environment impacts the free-living stage  
86 of the parasite with an emphasis on the duration of exposure to certain cyanotoxins. We also looked for  
87 impacts on parasite virulence, specifically looking at host reproduction and lifespan, both of which are  
88 influenced by infections.

89

90 **Methods**

91 In this study, we used the zooplankton *Daphnia dentifera*, which is common in stratified lakes in  
92 temperate North America (Tessier & Woodruff, 2002). For our experiments, we used the Midland 37  
93 (MID37) genotype, which was isolated from Midland Lake in Greene County, Indiana and has been used  
94 in several prior experiments (e.g., Auld, Hall & Duffy, 2012; Auld *et al.*, 2014). We also used the common  
95 fungal parasite *Metschnikowia bicuspidata* (“Standard” isolate, originally isolated from Baker Lake in  
96 Barry County, Michigan). *Daphnia* become infected by inadvertently consuming transmission spores they  
97 encounter in the water column when feeding. By “transmission spore”, we refer to the mature, needle-  
98 shaped ascus that contains the ascospore (Metschnikoff, 1884; Codreanu & Codreanu-Balcescu, 1981).  
99 After consumption by the host, infection can begin if the needle-shape spore crosses the gut barrier and  
100 is not fought off by a host hemocyte response (Metschnikoff, 1884; Stewart Merrill & Cáceres, 2018).  
101 Once infection has taken hold, the fungus replicates within the hemolymph of the host (Stewart Merrill  
102 & Cáceres, 2018). The parasite reduces the fecundity and lifespan of infected hosts (Auld *et al.*, 2012).  
103 *Metschnikowia* is an obligate killer, meaning it must kill its host in order to transmit to a new host (Ebert,  
104 2005); transmission spores are released into the environment after host death, after which they can be  
105 consumed by a new host, completing the parasite’s life cycle.

106 We incubated fungal transmission spores for different lengths of time in filtered lake water (Pall  
107 AE filters, 1  $\mu\text{m}$  pore size). The incubation times were 24 hours, 3 days, 5 days, 7 days, and 10 days. We  
108 also added two common cyanobacterial toxins, microcystin-LR and anatoxin-a, to the water during the  
109 incubations. We chose these cyanotoxins because they are commonly produced during blooms (Huisman  
110 *et al.*, 2018) and because prior research suggested that they reduced infection prevalence (in the case of  
111 microcystin-LR) or that production of them can increase in the presence of *Metschnikowia* (in the case of

112 anatoxin; Sánchez *et al.*, 2019). Microcystin-LR is produced by some members of the genus *Microcystis*,  
113 which has been extensively studied due to concerns over its geographical expansion and capability of  
114 producing CHABs in both marine and freshwater ecosystems (Huisman *et al.*, 2018). Anatoxin-a is  
115 produced by some members of the genus *Anabaena*. Microcystin-LR is a hepatotoxin while anatoxin-a is  
116 considered a potent neurotoxin in vertebrate models (Christensen & Khan, 2020). Both toxins are also  
117 produced by other genera of cyanobacteria such as *Plantktothrix*, *Oscillatoria*, *Aphanizomenon*,  
118 *Cylindrospermum*, and *Dolichospermum*, all capable of producing CHABs (Huisman *et al.*, 2018). We also  
119 included two types of negative controls: a solvent control of 0.01% acetic acid (see toxin preparation for  
120 explanation) and a negative control with no toxin or solvent added. The incubation times were crossed  
121 factorially with the toxin treatments, as described below. After the appropriate incubation time, we  
122 carried out infection assays in which we exposed *Daphnia* hosts to these spores and measured infection,  
123 spore production, host reproduction, and host mortality over time.

124

125 *Toxin preparation*

126 Pure microcystin-LR standard was purchased from Cayman Chemical (Ann Arbor, MI) and suspended in  
127 1mL of nano-pure water for a concentration of 1 mg/mL. Anatoxin-a standard was acquired from Abraxis  
128 (Warminster, PA). The anatoxin-a comes in a solution of 3:1 water and methanol and 0.01% acetic acid.  
129 We placed 1.5 mL of the anatoxin solution in a 2 mL Eppendorf tube and evaporated the methanol using  
130 an Eppendorf Vacufuge (Eppendorf, Hamburg, Germany) at 23 °C. Once evaporation had occurred, we  
131 added nanopure water to restore to the original volume/concentration. Toxins were frozen in between  
132 uses during the exposure period.

133

134 *Spore preparation*

135 Spores for the experiment were grown in vivo by infection of *Daphnia dentifera* ("Standard" genotype) in  
136 the laboratory. Infected animals with well-developed late-stage infections were collected from laboratory  
137 cultures and placed in 2mL Eppendorf tubes with 100-500 uL of milliQ water, then stored in the  
138 refrigerator at 4 °C. For this experiment, we used spores from animals that had been stored in the  
139 refrigerator for 2 weeks. To generate the spore slurry for experimental infections, we crushed infected  
140 animals to release spores, and then determined the density of mature ascospores using a  
141 hemocytometer (Hausser Scientific 3100, Horsham, PA, USA) and a compound microscope (Olympus  
142 BX53, Center Valley, PA, USA) at 400X magnification. Each incubation treatment was initiated with a new  
143 spore slurry (made from infected animals that were harvested live from the laboratory cultures, then  
144 stored in the refrigerator for 2 weeks) because we know that *Metschnikowia* spores lose infectivity over  
145 time, even in the refrigerator (Duffy & Hunsberger, 2019).

146

147 *Transmission spores exposure to toxins*

148 On the first day of the experiment ("day 1"), we initiated the longest incubation treatment (10 days) by  
149 placing 5000 mature transmission spores of *Metschnikowia bicuspidata* in 15 mL Falcon tubes filled with  
150 10 mL of lake water that had been filtered through a Pall AE filter (Pall Corporation, Port Washington,  
151 NY). We then added microcystin-LR, anatoxin-a, acetic acid (negative control), or no toxin/chemical; for  
152 both microcystin-LR and anatoxin-a, we had two toxin levels: 10 and 30 µg/L. The concentrations we  
153 chose for this study are below and/or well within the range observed during natural CHABs blooms (Park  
154 *et al.*, 1998; Pawlik-Skowrońska *et al.*, 2004; Ibelings *et al.*, 2005; Ha, Hidaka & Tsuno, 2009), and  
155 therefore ecologically relevant. These treatment doses are below the LC50s reported in previous  
156 *Daphnia* toxicology experiments for microcystins and anatoxin-a (DeMott, Zhang & Carmichael, 1991;  
157 Pawlik-Skowrońska, Toporowska & Mazur-Marzec, 2019). Even though we did not expose hosts to these

158 toxins, we chose these concentrations because they should not cause high levels of stress and mortality  
159 in *Daphnia*, so any impact to pathogens could have a substantial impact on parasite-host interactions.  
160 While these concentrations are likely on the high side of what spores are likely to encounter in nature, if  
161 there is no impact of the cyanotoxins at these levels, it suggests that they are unlikely to significantly  
162 impact the free-living stages of this parasite in the water column. There were 10 replicates of each  
163 treatment (including the negative control treatment of acetic acid), with the exception of the no toxin  
164 controls (0 µg/L), which had 20 replicates; this yielded a total of 70 experimental units per incubation  
165 time treatment. The tubes with spores were left uncapped for 10 days inside a large plastic tote covered  
166 with a lid at 20 °C with a 16:8 L:D photoperiod. On day 4, the same procedure from day 1 was repeated.  
167 In this treatment, spores were incubated for 7 days at 20 °C with a 16:8 L:D photoperiod, yielding the 7-  
168 day incubation treatment. On day 6, the same procedure was used to initiate the 5-day incubation  
169 treatment. Additionally, on this day, adult *Daphnia* were set-up so that we could use their offspring in  
170 the infection assays. Adult *Daphnia* of the MID37 genotype were placed in 150 mL beakers (5 adults per  
171 beaker) filled with 100 mL of filtered lake water. Beakers were placed in incubators at 20 °C with a 16:8  
172 L:D photoperiod for 24 hours. On day 7, we collected neonates (0-24 hours old) produced from mothers  
173 that had been set-up the previous day. We placed 10 neonates per beaker in 250 mL beakers with 150  
174 mL filtered lake water for a total of 400 animals. Each beaker received 2 mg C per L of *Ankistrodesmus*  
175 *falcatus* food and was placed in incubators at 20 °C with a 16:8 L:D photoperiod. After that, 2 mg C per  
176 mL of *Ankistrodesmus* was added to each beaker each day until the day the experimental animals were  
177 exposed to parasites. On days 8 and 10, we used the same procedure as described above to initiate the  
178 3- and 1-day incubation treatments, respectively.

179

180 *Infection assays*

181 On day 11, we carried out infection assays, in which we exposed *Daphnia* to transmission spores that  
182 had been incubated for different time lengths and with varying levels of exposure to toxins/chemicals. All  
183 Falcon tubes, containing spores with different exposure times and toxin levels, were collected and placed  
184 in a centrifuge (Sorvall ST 16, Thermo Scientific, Waltham, MA, USA) to spin down the spores. The tubes  
185 were spun at 3000 rpm for 10 min. We decanted the tubes using a 10mL pipette without disrupting the  
186 spore pellet at the bottom. In a pilot experiment, we confirmed that the original concentration of spores  
187 was recovered from vials after a 24-hour incubation followed by centrifugation. After the water was  
188 removed (which also removed the toxin), we resuspended the spores by adding 10 mL of filtered lake  
189 water and disturbing the pellet by vigorously pipetting the water in the tube. Then, we placed one 6-7-  
190 day old *Daphnia* in each Falcon tube and allowed the tubes to incubate again at 20 °C with 16:8 L:D  
191 photoperiod. Because we had originally placed 5000 mature transmission spores in each tube, this  
192 yielded an exposure dose of 500 spores/mL. Hosts were fed 1 mg C per mL of *Ankistrodesmus* on this  
193 day; using this lower level of food on the day of exposure is common in infection assays because it  
194 promotes infection. After 24 hours (that is, on day 12), each *Daphnia* individual was removed from the  
195 tubes with spores and placed in a 50 mL beaker filled with 30 mL filtered lake water that did not contain  
196 spores (one animal per beaker). Animals were fed 2 mg/L C of *Ankistrodesmus falcatus ad libidum* for  
197 the rest of the experiment (20 days post-infection, 32 days from day 1 of entire experiment).

198 During those 20 days, we tracked mortality in each of the beakers five days per week (Mondays  
199 through Fridays). Water changes were done twice a week; during these, we counted offspring in each  
200 beaker; offspring were removed from beakers and discarded. Any animals that died during the trial were  
201 placed in 2 mL Eppendorf tubes with 100 µL of nanopure water and stored at 4 °C for later spore counts.  
202 At the end of the experiment (20 days post-infection, 32 days from day 1 of entire experiment), any  
203 remaining live animals (130 total) were placed in a 2mL Eppendorf tube with 100 µL of nanopure water.

204 Animals were then ground to release spores, and spores in the ascus stage were counted using a  
205 hemocytometer under a compound microscope at 400X magnification.

206

207 *Data analysis*

208 For all our models, we analyzed data from the microcystin-LR and anatoxin-a treatments separately. This  
209 means that for each analysis described below, there was one performed for the microcystin-LR relevant  
210 data and another for the anatoxin-a data. The same no-toxin control data (0 µg/L; 20 replicates per  
211 incubation time) were used for the two sets of data (microcystin-LR and anatoxin-a). When analyzing the  
212 anatoxin-a data, the acetic acid treatments were included in the analyses (as an additional negative  
213 control) and were treated in our analyses as a low concentration (0.01%); statistical analysis did not find  
214 differences between the no toxin controls and the acetic acid controls (data not shown). Prior to  
215 analyses, the data was checked for normality using the Shapiro-Wilk test. Data that did not meet  
216 normality were analyzed using appropriate family error distribution link functions and checking for  
217 overdispersion. All analyses were carried out in R Studio Version 1.2.1335 using the stats v3.6.1 package.

218 We assessed environmental effects on the parasite's ability to infect by comparing differences in  
219 the number of animals that developed terminal infections (that is, infections that produced asci (Stewart  
220 Merrill & Cáceres, 2018)). In this analysis, our response variable had two outcomes (terminal infection or  
221 not). We performed a generalized linear model (GLM) with incubation time and toxin concentration as  
222 explanatory variables, using a binomial family error distribution.

223 We also evaluated if toxin concentration and incubation time (that is, the time spent as a  
224 transmission spore in the water) affected the number of mature transmission spores produced per  
225 infected host individual. For this analysis, we ran a GLM; here the number of spores was our response

226 variable and, similarly to the analysis of infections, toxin concentration and incubation time were used as  
227 the explanatory variables. In this model, we used a Gaussian error distribution.

228 To evaluate the effects of incubation time and toxin exposure on the parasite's virulence, we  
229 measured host offspring production, host survival over the 20-day experiment, and proportion of hosts  
230 who died within 20 days of parasite exposure. For these analyses, we ran GLMs with toxin concentration  
231 and time of exposure as explanatory variables; for the two former response variables—the number of  
232 offspring each host individual produced and the number of days each host survived (up to 20 days post  
233 infection)—we used a Poisson family error distribution. For the proportion of dead hosts at the end of  
234 the experiment, we used a binomial family distribution.

235 We ran two linear models analyzing the relationship between the proportion of infected  
236 individuals and the mean number of spores produced at the end of infection. In the first, we averaged  
237 across the different toxin exposure treatments, yielding one value for each incubation time ( $n = 5$ ). In the  
238 second, we averaged within the toxin treatments, yielding five values per toxin treatment (one per  
239 incubation time treatment, total  $n = 20$ ).

240

241 We also ran a second experiment to evaluate whether plastic absorbed considerable amounts of  
242 microcystin in the water and therefore negated the toxic effects on the parasites transmission spores.  
243 Studies have reported adsorption of microcystins by plastics (Hyenstrand *et al.*, 2001; Moura *et al.*,  
244 2022), which raised the possibility that the results of our first experiment may have been due to toxin  
245 concentrations that were lower than we intended. Therefore, in the second experiment, we evaluated  
246 whether the plastic vessels we used in the first block adsorbed microcystin-LR from the water and  
247 whether using glass vs. plastic vessels for the incubations impacted infections. Additional methods and  
248 results from that experiment can be found in the supplementary materials. Briefly, we did not find any

249 effect of vessel (that is, glass vs. plastic tubes) on toxin concentration, infection prevalence, or spore  
250 production.

251

252 **Results**

253 *Time spent in the water decreased infectivity, but toxin exposure did not*

254 Transmission spores that spent longer in the water were less infectious, but toxin exposure did not  
255 significantly influence infectivity (Figure 1, Table 1 “Infection prevalence”). The number of hosts that  
256 became infected decreased with increased incubation time: spores that were only incubated for 24  
257 hours infected around 75% of exposed hosts, whereas spores incubated for 10 days infected less than  
258 50% of exposed hosts. The reductions in infectivity with increasing incubation time were consistent  
259 across different toxin concentrations (as indicated by a non-significant concentration x incubation  
260 interaction term in the GLM: microcystins,  $Z = 1.160$ ,  $p = 0.246$ ; anatoxin,  $Z = 0.816$ ,  $p = 0.415$ ).

261

262 *Spore yield was not affected by toxin exposure*

263 Neither microcystin nor anatoxin dose significantly influenced the number of spores produced per  
264 infected host (Figure 2; Table 1). However, in the anatoxin treatments (but not the microcystin  
265 treatments), infected hosts exposed to spores that were incubated for longer periods of time yielded  
266 more spores at death (or, for those that did not die within 20 days of exposure, at 20 days post-exposure;  
267 Figure 2; Table 1).

268

269 *Virulence was not affected by toxin concentration or incubation time*

270 Neither toxin concentration nor incubation time impacted host offspring production, lifespan, or  
271 survivorship measured up to 20 days post-infection. Host lifetime offspring production was consistent  
272 across the microcystin-LR treatments (Figure 3, left panel) and across the anatoxin-a treatments (Figure  
273 3, right panel), and across incubation times (Figure 3, Table 1). Similarly, neither toxin concentration nor  
274 incubation time significantly influenced the number of days each host survived (up to 20 days post-  
275 infection) or the proportion of dead individuals at day 20 (Table 1). Thus, there is no evidence that  
276 exposure of transmission spores to toxins in the water, or the duration of time spent in the water  
277 column, influenced the virulence of this fungal parasite.

278

279 *Greater parasites infectiousness is associated with lower spore yield*

280 Treatments that had higher infection prevalence yielded fewer spores per infected host (Figure 4). When  
281 averaged across the different toxin treatments within incubation times, longer incubation times had  
282 fewer infections that yielded more spores per infected host ( $R^2 = 0.80$ ,  $p = 0.02$ ). If we average within  
283 toxin treatments by incubation times, we observe a similar pattern but with a lower amount of variance  
284 explained ( $R^2 = 0.40$ ,  $p = 0.001$ ).

285

## 286 **Discussion**

287 In this study, we found that the amount of time transmission spores spend in the water  
288 impacted their fitness. If spores encountered a new host within a few days of being released from their  
289 dead host, the number of hosts that became infected was high, suggesting high per-spore infectivity.  
290 However, the high prevalence did not translate into high spore yield; instead, hosts from treatments with  
291 high infectivity produced low numbers of mature transmission spores. On the other hand, spores that

292 spent longer in the water column after being released from a dead host infected relatively few hosts, but  
293 each infected host yielded more mature transmission spores as compared to a host infected with spores  
294 that spent less time in the water column.

295 A notable finding from our study is the absence of toxic effects of two common cyanotoxins on  
296 parasite transmission spores while in the water column. We predicted that long exposures to high  
297 concentrations of cyanotoxin would decrease parasite fitness and virulence. However, our analyses did  
298 not detect any effect of toxins. Previous studies have found that consuming toxin-producing  
299 cyanobacteria can make hosts more susceptible to parasites (Tellenbach *et al.*, 2016), can decrease the  
300 number of spores produced within the host (Manzi *et al.*, 2019), and in certain cases protects hosts from  
301 infection (Sánchez *et al.*, 2019). However, the conditions of parasite toxin exposure in the previous  
302 experiments were different from our current study. In the previous studies, parasites were exposed to  
303 these cyanotoxins when hosts consumed toxic cyanobacteria as part of their diet, so interactions  
304 between the toxins and parasites would occur in the host gut. The key difference is that, in our study, the  
305 toxin exposure only occurred in the water column, and hosts were not exposed to these compounds.  
306 This suggests that the location of exposure of spores to cyanotoxins is important. One possible reason  
307 why we did not observe effects of toxins on parasite fitness might be due to the structure of the  
308 transmission spore. *Metschnikowia* transmission spores are composed of an ascus, which is an outer  
309 structure surrounding the ascospore; the ascospore is the actual transmission spore that pierces the host  
310 gut and therefore causes the infection. One possibility is that the ascus, which protects the spore from  
311 chemical and physical damage (Lachance *et al.*, 1976), also shields against secondary metabolites in the  
312 water, which would explain the lack of effects from the cyanotoxins in this study. Taken together, the  
313 results of our study combined with other studies that have found effects of toxin and non-toxin  
314 producing cyanobacteria on parasite fitness (Coopman *et al.*, 2014; Tellenbach *et al.*, 2016; Manzi *et al.*,  
315 2019; Sánchez *et al.*, 2019) suggest that the interaction between parasites and secondary metabolites

316 may be particularly important within the gut of the host. If true, this would suggest that, for parasites  
317 with a protective structure that envelops transmission spores, the interaction of toxins with parasites  
318 might primarily be important in parasite life stages that lack the protective coating.

319 There was an apparent tradeoff between per-spore infectivity and the number of spores yielded  
320 per infected host. One mechanism that could explain this result is within-host spore competition (Ebert,  
321 Zschokke-Rohringer & Hans, 2000). Spores that spent little time in the water column had a high per-  
322 spore infectivity, which likely led to multiple spores infecting an individual host; individual *Daphnia* can  
323 be infected by multiple *Metschnikowia* spores (Stewart Merrill & Cáceres, 2018). Infection by multiple  
324 spores would result in resource competition among spores inside the host, decreasing parasite spore  
325 production. In contrast, spores that spent longer in the water column were less viable; this would mean  
326 fewer spores were able to infect an individual host, decreasing resource competition and increasing  
327 within host reproduction. Prior work has shown that not all spores that make it to the gut pierce the  
328 gut—a critical first step of infection that must occur prior to growing in the host hemolymph (Stewart  
329 Merrill & Cáceres, 2018). Our results and this hypothesized mechanism are consistent with an earlier  
330 study using the same parasite but a different species of *Daphnia*; that study found parasite reproductive  
331 success decreased with increased spore dosage at the time of infection (Ebert *et al.*, 2000). It is also  
332 possible that spores that survived longer were, on average, ‘stronger’ than the spores that died during  
333 the incubation, meaning a non-random sample of spores generated the infections in the longer  
334 incubation periods.

335 It remains to be determined why spores lost infectivity so rapidly in the water column. One  
336 possibility is that light played a role. Our parasite incubations for the toxin exposures were done in the  
337 light (using a 16:8 light:dark photoperiod), and other studies have found that light, including UV and  
338 photosynthetically active radiation (PAR), decreases infectivity of *Metschnikowia* spores (Overholt *et al.*,  
339 2012; Shaw *et al.*, 2020). Therefore, spores may have retained high infectivity for longer if our

340 experiment had been done in the dark. However, in a previous experiment, even spores that were  
341 maintained in the refrigerator in the dark rapidly lost infectivity (Duffy & Hunsberger, 2019); thus, it is  
342 possible that multiple factors explain the rapid loss in infectivity of *Metschnikowia* spores in the water  
343 column.

344 One limitation of our study is that we only used a single host genotype and a single parasite  
345 genotype. There is substantial diversity in the susceptibility of *Daphnia dentifera* to *Metschnikowia* (e.g.,  
346 Duffy & Sivars-Becker, 2007; Duffy *et al.*, 2012; Auld, Searle & Duffy, 2017). While there is much less  
347 diversity in *Metschnikowia*, there is some (Shaw, 2019). Future work exploring whether *Metschnikowia*  
348 genotypes vary in their sensitivity to phycotoxins and/or time spent in the water column would be  
349 valuable, as would studies assessing whether the patterns we found in this study are consistent when  
350 other host genotypes (and species) are exposed. Interestingly, a new study that incubated  
351 *Metschnikowia* spores with cyanobacterial extracts or a control solution, then exposed them to two  
352 genotypes of *Daphnia galeata* x *longispina* (Manzi *et al.*, 2022) suggests that these results might hold  
353 broadly. Consistent with our study, they did not find reduced infectivity of spores that had been  
354 incubated with the cyanobacterial extract; in fact, for one of the two host genotypes, infectivity of spores  
355 that had been incubated with the cyanobacterial extract was actually higher.

356 Here, we found that parasite spores rapidly lose infectivity in the water, but that two common  
357 cyanotoxins had no detectable effect on spores. We propose that *Metschnikowia* spores may be  
358 protected from toxins in the environment by their ascus structure. We also found that, while spores that  
359 had spent more time in the water were less infectious, they yielded more spores per infected host,  
360 results that might be driven by resource competition. This raises the intriguing possibility that there  
361 might be an intermediate spore age (or dose) that is most likely to fuel large epidemics. Given that  
362 spores spend substantial time in the sediment between outbreaks (Decaestecker *et al.*, 2004), the joint  
363 impacts of these mechanisms are likely to be important drivers of *Metschnikowia* epidemics in lakes.

364

365 **Acknowledgments**

366 We would like to thank M. Hunter for providing lab equipment and space and for his valuable input, P.  
367 Yajnik for his advice on statistical analysis, and G. Kling and four anonymous reviewers for helpful  
368 feedback on earlier drafts of this manuscript. K. Sánchez was supported by an NSF Graduate Research  
369 Fellowship. B. Zhong and J. Agudelo were supported by the Doris Duke Conservation Scholars Program at  
370 the University of Michigan. This work was support by NSF grant DEB-1655856 to MAD, Block Grant funds  
371 from the University of Michigan Department of Ecology & Evolutionary Biology, and by the Moore  
372 Foundation (GBMF9202; DOI: <https://doi.org/10.37807/GBMF9202>). The authors report no conflicts of  
373 interest.

374 **Author contribution statement**

375 Conceptualization: KFS, MAD. Developing methods: KFS, MAD. Conducting the research: KFS, BZ, JAA.  
376 Data analysis: KFS, BZ, JAA, MAD. Data interpretation: KFS, BZ, JAA, MAD. Preparation figures & tables:  
377 KFS, BZ, JAA. Writing: KFS, BZ, JAA, MAD.

378 **Data availability statement**

379 Data is accessible through DRYAD digital repository <https://doi.org/10.5061/dryad.612jm6420>

380

381 **References**

382 Amigó J.M., Gracia M.P., Rius M., Salvadó H., Maillo P.A. & Vivarés C.P. (1996). Longevity and effects of  
383 temperature on the viability and polar-tube extrusion of spores of *Glugea stephani*, a  
384 microsporidian parasite of commercial flatfish. *Parasitology Research* **82**, 211–214.  
385 <https://doi.org/10.1007/s004360050097>

386 Andersen N.G., Lorenzen E., Boutrup T.S., Hansen P.J. & Lorenzen N. (2016). Sublethal concentrations of  
387 ichthyotoxic alga *Prymnesium parvum* affect rainbow trout susceptibility to viral haemorrhagic  
388 septicaemia virus. *Diseases of Aquatic Organisms* **117**, 187–195. <https://doi.org/10.3354/dao02946>

389 Auld S.K.J.R., Hall S.R. & Duffy M.A. (2012). Epidemiology of a Daphnia-multiparasite system and its  
390 implications for the Red Queen. *PLoS ONE* **7**, 1–6. <https://doi.org/10.1371/journal.pone.0039564>

391 Auld S.K.J.R., Hall S.R., Ochs J.H., Sebastian M. & Duffy M.A. (2014). Predators and Patterns of Within-  
392 Host Growth Can Mediate Both Among-Host Competition and Evolution of Transmission Potential  
393 of Parasites. *The American Naturalist* **184**, S77-90. <https://doi.org/10.1086/676927>

394 Auld S.K.J.R., Searle C.L. & Duffy M.A. (2017). Parasite transmission in a natural multihost-multiparasite  
395 community. *Philosophical Transactions of the Royal Society B: Biological Sciences* **372**, 1–10.  
396 <https://doi.org/10.1098/rstb.2016.0097>

397 Borowitzka M.A. (1995). Microalgae as sources of pharmaceuticals and other biologically active  
398 compounds. *Journal of Applied Phycology* **7**, 3–15. <https://doi.org/10.1007/BF00003544>

399 Brookes J.D., Antenucci J., Hipsey M., Burch M.D., Ashbolt N.J. & Ferguson C. (2004). Fate and transport  
400 of pathogens in lakes and reservoirs. *Environment International* **30**, 741–759.  
401 <https://doi.org/10.1016/j.envint.2003.11.006>

402 Cáceres C.E., Hall S.R., Duffy M.A., Tessier A.J., Helmle C., Macintyre S., *et al.* (2006). Physical Structure of

403 Lakes Constrains Epidemics in Daphnia Populations. *Ecology* **87**, 1438–1444

404 Christensen V.G. & Khan E. (2020). Freshwater neurotoxins and concerns for human, animal, and  
405 ecosystem health: A review of anatoxin-a and saxitoxin. *Science of the Total Environment* **736**,  
406 139515. <https://doi.org/10.1016/j.scitotenv.2020.139515>

407 Codreanu R. & Codreanu-Balcescu D. (1981). On Two *Metschnikowia* Yeast Species Producing Infections  
408 in *Daphnia magna* and *Artemia salina* (Crustacea, Phyllopoda) from Romania. *Journal of*  
409 *Invertebrate Pathology* **37**, 22–27

410 Coopman M., Muylaert K., Lange B., Reyserhove L. & Decaestecker E. (2014). Context dependency of  
411 infectious disease: the cyanobacterium *Microcystis aeruginosa* decreases white bacterial disease in  
412 *Daphnia magna*. *Freshwater Biology* **59**, 714–723. <https://doi.org/10.1111/fwb.12420>

413 Decaestecker E., Lefever C., De Meester L. & Ebert D. (2004). Haunted by the past: Evidence for dormant  
414 stage banks of microparasites and epibionts of *Daphnia*. *Limnology and Oceanography* **49**, 1355–  
415 1364. [https://doi.org/10.4319/lo.2004.49.4\\_part\\_2.1355](https://doi.org/10.4319/lo.2004.49.4_part_2.1355)

416 DeMott W.R., Zhang Q. & Carmichael W.W. (1991). Effects of toxic cyanobacteria and purified toxins on  
417 the survival and feeding of a copepod and three species of *Daphnia*. *Limnology and Oceanography*  
418 **36**, 1346–1357

419 Duffy M.A. & Hunsberger K.K. (2019). Infectivity is influenced by parasite spore age and exposure to  
420 freezing: do shallow waters provide *Daphnia* a refuge from some parasites? *Journal of Plankton*  
421 *Research* **41**, 12–16. <https://doi.org/10.1093/plankt/fby046>

422 Duffy M.A., Ochs J.H., Penczykowski R.M., Civitello D.J., Klausmeier C.A. & Hall S.R. (2012). Size and  
423 Parasite-Driven Evolution. *Science* **335**, 1636–1638

424 Duffy M.A. & Sivars-Becker L. (2007). Rapid evolution and ecological host-parasite dynamics. *Ecology*

425        *Letters* **10**, 44–53. <https://doi.org/10.1111/j.1461-0248.2006.00995.x>

426        Duperron S., Halary S., Habiballah M., Gallet A., Huet H., Duval C., *et al.* (2019). Response of Fish Gut  
427        Microbiota to Toxin-Containing Cyanobacterial Extracts: A Microcosm Study on the Medaka  
428        (*Oryzias latipes*). *Environmental Science and Technology Letters* **6**, 341–347.  
429        <https://doi.org/10.1021/acs.estlett.9b00297>

430        Ebert D. (2005). *Ecology, Epidemiology, and Evolution of Parasitism in Daphnia*. National Center for  
431        Biotechnology Information, Bethesda, MD.

432        Ebert D., Zschokke-Rohringer C.D. & Hans J.C. (2000). Dose Effects and Density-Dependent Regulation of  
433        Two Microparasites of *Daphnia magna*. *Ecology* **122**, 200–209

434        Ha J.H., Hidaka T. & Tsuno H. (2009). Quantification of toxic *microcystis* and evaluation of its dominance  
435        ratio in blooms using real-time PCR. *Environmental Science and Technology* **43**, 812–818.  
436        <https://doi.org/10.1021/es801265f>

437        Hall S.R., Smyth R., Becker C.R., Duffy M.A., Knight C.J., MacIntyre S., *et al.* (2010). Why Are Daphnia in  
438        Some Lakes Sicker? Disease Ecology, Habitat Structure, and the Plankton. *BioScience* **60**, 363–375.  
439        <https://doi.org/10.1525/bio.2010.60.5.6>

440        Hallmann J. & Sikora R.A. (1996). Toxicity of fungal endophyte secondary metabolites to plant parasitic  
441        nematodes and soil-borne plant pathogenic fungi. *European Journal of Plant Pathology* **102**, 155–  
442        162

443        Harvell D., Aronson R., Baron N., Connell J., Dobson A., Ellner S., *et al.* (2004). The Rising Tide of Ocean  
444        Diseases: Unsolved Problems and Research Priorities. *Frontiers in Ecology and the Environment* **2**,  
445        375–382

446        Heagle A.S. (1973). Interactions Between Air Pollutants and Plant Parasites. *Annual Review of*

447        *Phytopathology* **11**, 365–388. <https://doi.org/10.1146/annurev.py.11.090173.002053>

448        Huisman J., Codd G.A., Paerl H.W., Ibelings B.W., Verspagen J.M.H. & Visser P.M. (2018). Cyanobacterial  
449        blooms. *Nature Reviews Microbiology* **16**, 471–483. <https://doi.org/10.1038/s41579-018-0040-1>

450        Hyenstrand P., Metcalf J., Beattie K. & Codd G.A. (2001). Effects of adsorption to plastics and solvent  
451        conditions in the analysis of the cyanobacterial toxin microcystin-LR by high performance liquid  
452        chromatography. *Water Research* **35**, 3508–3511. [https://doi.org/10.1016/S0043-1354\(01\)00068-9](https://doi.org/10.1016/S0043-1354(01)00068-9)

454        Ibelings B.W., Bruning K., de Jonge J., Wolfstein K., Pires L.M.D., Postma J., *et al.* (2005). Distribution of  
455        Microcystins in a Lake Foodweb: No Evidence for Biomagnification. *Microbial Ecology* **49**, 487–500.  
456        <https://doi.org/10.1007/s00248-004-0014-x>

457        Lachance M.-A., Miranda M., Miller M.W. & Phaff H.J. (1976). Dehiscence and active spore release in  
458        pathogenic strains of the yeast *Metschnikowia bicuspidata* var. *australis*: possible predatory  
459        implication. *Canadian Journal of Microbiology* **22**, 1756–1761

460        Lafferty K.D. & Holt R.D. (2003). How should environmental stress affect the population dynamics of  
461        disease? *Ecology Letters* **6**, 654–664

462        Landsberg J.H. (2002). The Effects of Harmful Algal Blooms on Aquatic Organisms. *Reviews in Fisheries*  
463        *Science* **10**, 1064–1262. <https://doi.org/10.1080/20026491051695>

464        Lassudrie M., Hégaret H., Wikfors G.H. & da Silva P.M. (2020). Effects of marine harmful algal blooms on  
465        bivalve cellular immunity and infectious diseases: A review. *Developmental and Comparative*  
466        *Immunology* **108**, 103660. <https://doi.org/10.1016/j.dci.2020.103660>

467        Leflaive J. & Ten-Hage L. (2007). Algal and cyanobacterial secondary metabolites in freshwaters: A  
468        comparison of allelopathic compounds and toxins. *Freshwater Biology* **52**, 199–214.

469 <https://doi.org/10.1111/j.1365-2427.2006.01689.x>

470 Malone L.A., Gatehouse H.S. & Tregidga E.L. (2001). Effects of time, temperature, and honey on Nosema  
471 apis (Microsporidia: Nosematidae), a parasite of the honeybee, *Apis mellifera* (Hymenoptera:  
472 Apidae). *Journal of Invertebrate Pathology* **77**, 258–268. <https://doi.org/10.1006/jipa.2001.5028>

473 Manzi F., Agha R., Lu Y., Ben-Ami F. & Wolinska J. (2019). Temperature and host diet jointly influence the  
474 outcome of infection in a Daphnia-fungal parasite system. *Freshwater Biology*, 1–11.  
475 <https://doi.org/10.1111/fwb.13464>

476 Manzi F., Agha R., Mühlenhaupt M. & Wolinska J. (2022). Prior exposure of a fungal parasite to  
477 cyanobacterial extracts does not impair infection of its Daphnia host. *Hydrobiologia* **849**, 2731–  
478 2744. <https://doi.org/https://doi.org/10.1007/s10750-022-04889-7>

479 Metschnikoff V.E. (1884). Ueber eine Sprosspilzkrankheit der Daphnien Beitrag zur Lehre über den Kampf  
480 der Phagocyten gegen Krankheitserreger. *Archiv für Pathologische Anatomie und Physiologie und*  
481 *für Klinische Medicin* **96**, 177–195. <https://doi.org/10.1007/BF02361555>

482 Moura D.S., Pestana C.J., Moffat C.F., Hui J., Irvine J.T., Edwards C., *et al.* (2022). Adsorption of  
483 cyanotoxins on polypropylene and polyethylene terephthalate: Microplastics as vector of eight  
484 microcystin analogues. *Environmental Pollution* **303**, 119135.  
485 <https://doi.org/https://doi.org/10.1016/j.envpol.2022.119135>

486 Ostensvik O., Skulberg O.M., Underdal B. & Hormazabal V. (2002). Antibacterial properties of extracts  
487 from selected planktonic freshwater cyanobacteria - a comparative study of bacterial bioassays.  
488 *Journal of Applied Microbiology* **84**, 1117–1124

489 Overholt E.P., Hall S.R., Williamson C.E., Meikle C.K., Duffy M.A. & Cáceres C.E. (2012). Solar radiation  
490 decreases parasitism in Daphnia. *Ecology Letters* **15**, 47–54. <https://doi.org/10.1111/j.1461->

491 0248.2011.01707.x

492 Park H.D., Kim B., Kim E. & Okino T. (1998). Hepatotoxic microcystins and neurotoxic anatoxin-a in  
493 cyanobacterial blooms from Korean lakes. *Environmental Toxicology and Water Quality* **13**, 225–  
494 234. [https://doi.org/10.1002/\(SICI\)1098-2256\(1998\)13:3<225::AID-TOX4>3.0.CO;2-9](https://doi.org/10.1002/(SICI)1098-2256(1998)13:3<225::AID-TOX4>3.0.CO;2-9)

495 Pawlik-Skowrońska B., Skowroński T., Pirszel J. & Adamczyk A. (2004). Relationship between  
496 cyanobacterial bloom composition and anatoxin-A and microcystin occurrence in the eutrophic  
497 dam reservoir (se Poland). *Polish Journal of Ecology* **52**, 479–490

498 Pawlik-Skowrońska B., Toporowska M. & Mazur-Marzec H. (2019). Effects of secondary metabolites  
499 produced by different cyanobacterial populations on the freshwater zooplankters *Brachionus*  
500 *calyciflorus* and *Daphnia pulex*. *Environmental Science and Pollution Research* **26**, 11793–11804.  
501 <https://doi.org/10.1007/s11356-019-04543-1>

502 Penczykowski R.M., Lemanski B.C.P., Sieg R.D., Hall S.R., Housley Ochs J., Kubanek J., *et al.* (2014). Poor  
503 resource quality lowers transmission potential by changing foraging behaviour. *Functional Ecology*  
504 **28**, 1245–1255. <https://doi.org/10.1111/1365-2435.12238>

505 Sánchez K.F., Huntley N., Duffy M.A. & Hunter M.D. (2019). Toxins or medicines? Phytoplankton diets  
506 mediate host and parasite fitness in a freshwater system. *Proceedings of the Royal Society B:  
507 Biological Sciences* **286**, 20182231. <https://doi.org/10.1098/rspb.2018.2231>

508 Shaw C.L. (2019). *Drivers of Epidemic Timing and Size in a Natural Aquatic System*. University of  
509 Michigan.

510 Shaw C.L., Overholt E., Williamson C., Cáceres C.E., Hall S.R. & Duffy M.A. (2020). Shedding light on  
511 environmentally transmitted parasites: Lighter conditions within lakes restrict epidemic size.  
512 *Ecology* **101**, e03168. <https://doi.org/10.1002/ECY.3168>

513 Stewart Merrill T.E. & Cáceres C.E. (2018). Within-host complexity of a plankton-parasite interaction.

514 *Ecology* **99**, 2864–2867. <https://doi.org/10.1002/ecy.2483>

515 Tellenbach C., Tardent N., Pomati F., Keller B., Hairston N.G., Wolinska J., *et al.* (2016). Cyanobacteria

516 facilitate parasite epidemics in Daphnia. *Ecology* **97**, 3422–3432. <https://doi.org/10.1002/ecy.1576>

517 Tessier A.J. & Woodruff P. (2002). Cryptic Trophic Cascade along a Gradient of Lake Size. *Ecology* **83**,

518 1263–1270

519 Thielges D.W., Amundsen P.-A., Hechinger R.F., Johnson P.T.J., Lafferty K.D., Mouritsen K.N., *et al.* (2013).

520 Parasites as prey in aquatic food webs: implications for predator infection and parasite

521 transmission. *Oikos* **122**, 1473–1482. <https://doi.org/10.1111/j.1600-0706.2013.00243.x>

522 Vasemagi A., Visse M. & Kisand V. (2017). Effect of environmental factors and al emerging parasitic

523 disease on gut microbiome of wild salmonid fish. *mSphere* **2**, e0018-17.

524 <https://doi.org/10.1128/mSphere.00418-17>

525 Volk R.B. & Furtak F.H. (2006). Antialgal, antibacterial and antifungal activity of two metabolites

526 produced and excreted by cyanobacteria during growth. *Microbiological Research* **161**, 180–186.

527 <https://doi.org/10.1016/j.micres.2005.08.005>

528

529

530

531 **Tables**

532 Table 1. Statistical results of generalized linear models on the effects of toxin concentration and  
 533 incubation time on the parasites fitness and virulence. Note that host reproduction and host survivorship  
 534 are used here as proxies of virulence. “Conc\*Incub” indicates the toxin concentration \* incubation time  
 535 interaction term.

<i>Variable</i>	<i>Estimate</i>	<i>Std. Error</i>	<i>Z-value</i>	<i>P-value</i>
<i>Infection prevalence</i>				
<i>Microcystin-LR</i>				
Toxin concentration	-0.012	0.025	-0.497	0.619
Incubation time	-0.382	0.083	-4.630	<0.0001*
Conc*Incub	0.005	0.005	1.160	0.246
<i>Anatoxin-a</i>				
Toxin concentration	-0.003	0.026	-0.107	0.915
Incubation time	-0.527	0.087	-6.071	<0.0001*
Conc*Incub	0.004	0.005	0.816	0.415
<i>Spore reproduction</i>				
<i>Microcystin-LR</i>				
Toxin concentration	466.76	386.33	1.208	0.231
Incubation time	2230.73	1528.03	1.460	0.149
Conc*Incub	-87.23	85.90	-1.016	0.313
<i>Anatoxin-a</i>				
Toxin concentration	326.10	355.22	0.918	0.361
Incubation time	4337.31	1508.30	2.876	<0.01*
Conc*Incub	-21.27	83.29	-0.255	0.799
<i>Host lifetime reproduction</i>				
<i>Microcystin-LR</i>				
Toxin concentration	-0.0007	0.005	-0.140	0.889
Incubation time	0.012	0.019	0.643	0.520
Conc*Incub	0.0008	0.001	0.790	0.430

*Anatoxin-a*

Toxin concentration	-0.003	0.005	-0.753	0.452
Incubation time	0.007	0.019	0.353	0.724
Conc*Incub	0.0005	0.001	0.506	0.613

*Host survivorship**Microcystin-LR*

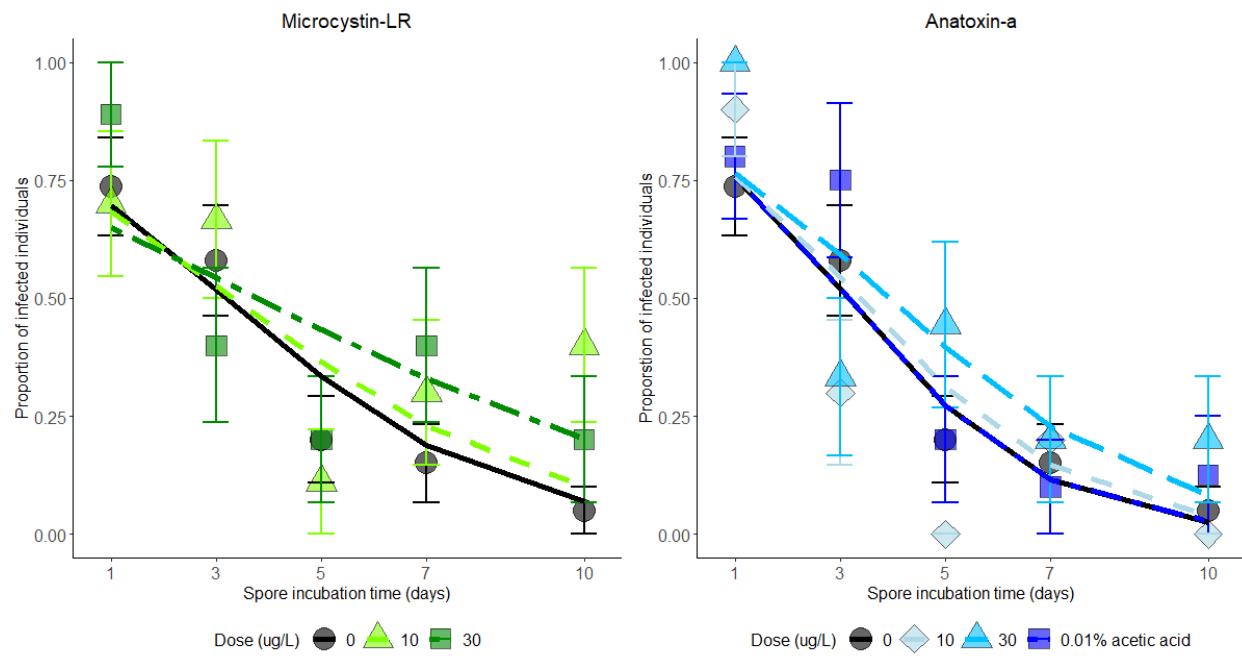
Toxin concentration	0.0008	0.004	0.227	0.820
Incubation time	0.013	0.015	0.854	0.393
Conc*Incub	-0.0002	0.0008	-0.327	0.744

*Anatoxin-a*

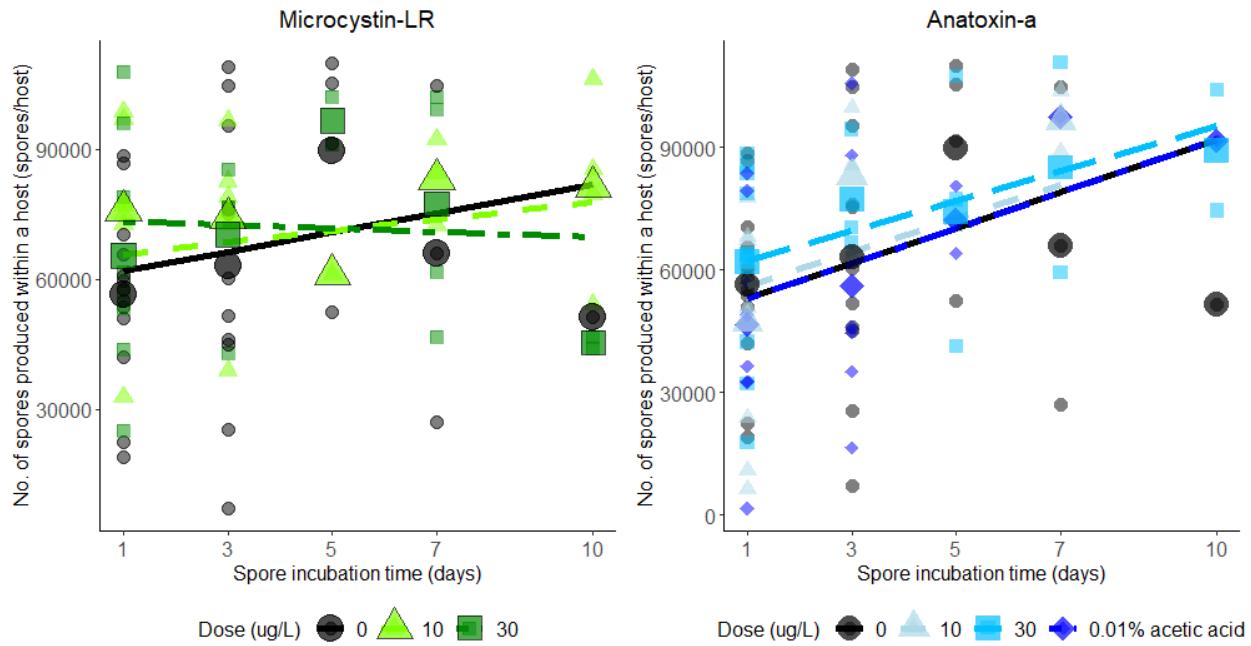
Toxin concentration	7.074e-04	3.485e-03	0.203	0.839
Incubation time	1.451e-02	1.447e-02	1.003	0.316
Conc*Incub	-9.278e-05	7.997e-04	-0.116	0.908

536

537

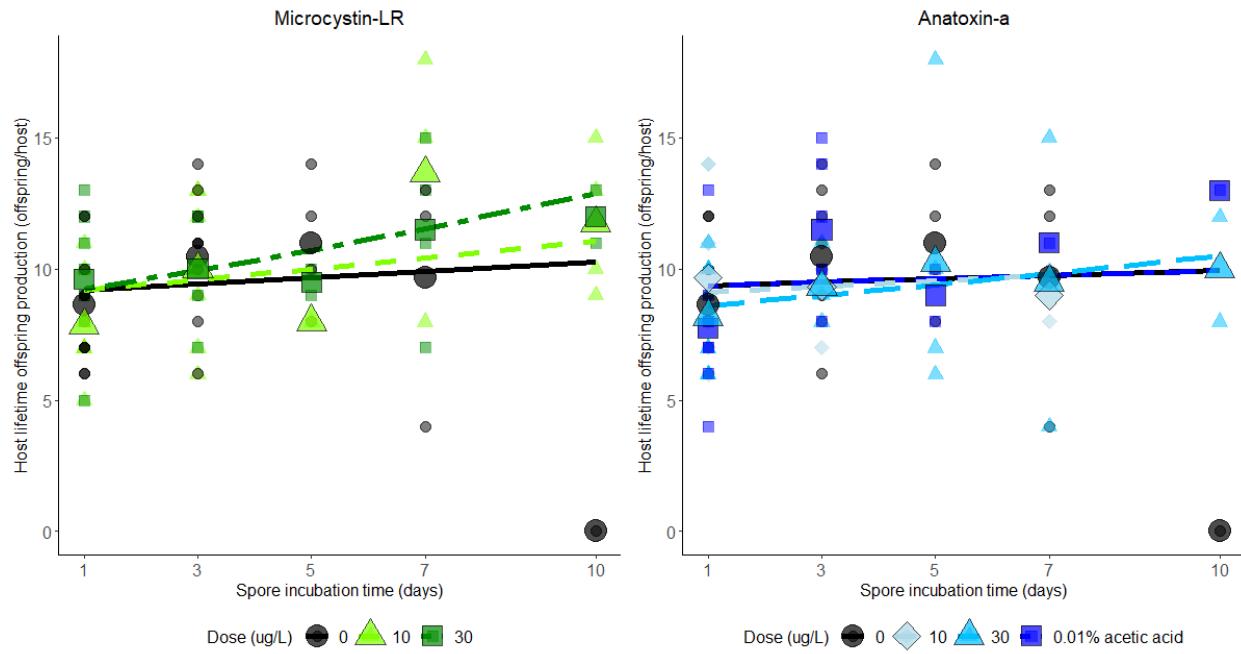


540 **Figure 1. Infection prevalence of hosts exposed to transmission spores treated with different**  
 541 **concentrations of pure cyanotoxins and incubated for different amounts of time.** This figure shows the  
 542 proportion of individuals that became infected in each treatment group and standard errors. For both  
 543 toxins and in all concentration treatments, the number of individual hosts that became infected  
 544 decreased with the amount of time the transmission spores spent in the water before being used for  
 545 host exposure. Note that the 0 µg/L concentration treatment is the same set of data for both panels.



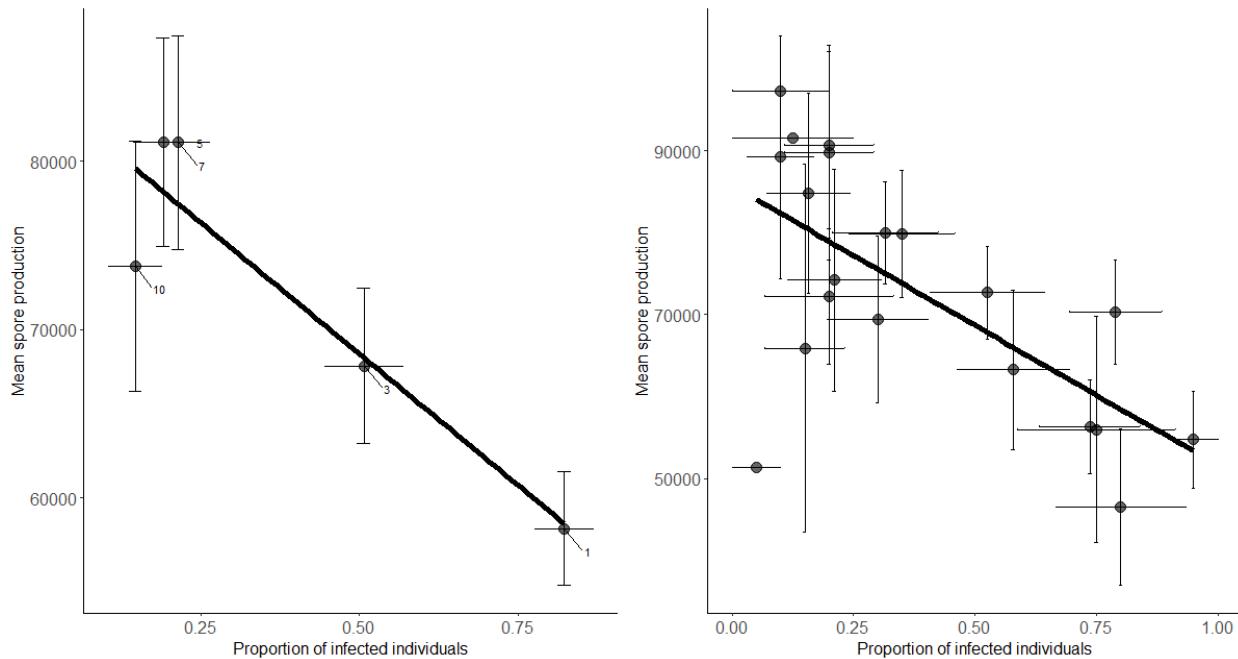
546

547 **Figure 2. Parasite reproduction within infected hosts exposed to transmission spores treated with**  
 548 **different concentrations of pure cyanotoxins that were incubated for different lengths of time. (Left**  
 549 *panel)* Spores that were treated with microcystin-LR did not appear to be affected by the concentration  
 550 of toxin nor the amount of time these spores spend in the water. (*Right panel*) The number of mature  
 551 transmission spores produced in all concentration treatments increased with longer periods of  
 552 incubation time for transmission spores. Note that the 0 µg/L concentration treatment is the same set of  
 553 data for both panels. Small symbols represent individual data points while large symbols represent  
 554 means for a given treatment group.



555

556 **Figure 3. Lifetime reproduction of hosts that were infected with transmission spores treated with**  
 557 **different concentrations of pure cyanotoxins and incubated for different lengths of time.** Virulence of  
 558 the parasite did not change based on the transmission spore treatments. Note that the 0  $\mu\text{g/L}$   
 559 concentration treatment is the same set of data for both panels. Small symbols represent individual data  
 560 points while large symbols represent means for a given treatment group.



561

562 **Figure 4. Linear regressions of the proportion of individuals that became infected versus the mean**  
 563 **number of mature transmission spores produced in each treatment group. (Left panel)** The data points  
 564 in this regression are averaged by incubation time of transmission spores in all toxin treatments. The  
 565 number in the label for each point represents the number of days those spores were incubated,  $R^2 =$   
 566  $0.80$ ,  $p = 0.02$ . **(Right panel)** This correlation was done by averaging the data by toxin treatment and  
 567 incubation period,  $R^2 = 0.40$ ,  $p = 0.001$ . For both correlations we observed a decreasing number of spores  
 568 produced within a host as the number of individuals that become infected in a treatment increased.  
 569 Error bars in both panels represent standard error for infection prevalence and spore reproduction for  
 570 the horizontal and vertical bars respectively.