

**Isoprene emission by plants in polluted environments**

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**Abstract**

In recent years, anthropogenic activities and climate change have significantly increased exposure of plants to environmental stresses (single or multiple) and pollutants, with negative consequences for the survival and productivity of vegetation. Plants may activate an armament of defenses against stresses. Isoprene, the most abundant biogenic volatile organic compound (BVOC) emitted by plants, is supposed to have a direct or indirect antioxidant role by quenching reactive oxygen species (ROS) or by reprogramming gene expression for antioxidant activation. On the other hand, isoprene is involved in the chemistry of troposphere, further contributing to a build up of pollutants when mixing with anthropogenic gases. In this review, we summarize present knowledge on the impact of air and soil pollutants on isoprene emission by plants, indicating possible feedback and feedforward mechanisms that may affect whole ecosystem functioning and evolution of plant species.

**Keywords:** Isoprene emission; biogenic volatile compounds; environmental stress; air chemistry, soil pollution; belowground communications, climate change

## Introduction

Plants are sessile but not passive components of the ecosystems, and they interact with the environment in several ways. Biogenic volatile organic compounds (BVOCs) are gases that are emitted by organisms in all terrestrial and marine ecosystems (Loreto et al. 2014). Plants emit worldwide more than 1 Pg C per year as BVOCs (Guenther et al. 1995, 2012), about half of which is isoprene (Guenther et al. 2006). Leaf BVOCs may be constitutively emitted (generally leaf life-long) or induced by abiotic and biotic stresses (Loreto and Schnitzler 2010). Some constitutive BVOCs may also be induced by stresses (Harrison et al. 2013). More than 1700 BVOCs have been identified, which are emitted by 90 different plant families belonging to both angiosperms and gymnosperms (Knudsen et al. 2006), and including also ferns and mosses (Hanson et al. 1999). As the detection systems get more accurate and high-throughput, the idea that all plants emit BVOCs is becoming realistic.

Synthesis of most significant BVOCs occurs through three pathways: the lipoxygenase (LOX), the shikimic acid, and the terpenoid pathways (Pichersky and Gershenzon 2002). Terpenoids or isoprenoids are the largest group of specialized plant metabolites and derive from two precursors: isopentenyl pyrophosphate (IPP) and dimethylallyl pyrophosphate (DMAPP) (McGarvey and Croteau 1995). Among terpenoids, isoprene ( $C_5H_8$ ) is the simplest and most volatile BVOC. Isoprene is formed by the chloroplastic methyl erythritol 4-phosphate (MEP) pathway through isoprene synthase (IspS), which catalyzes the removal of pyrophosphate (PPi) from DMAPP. Constitutive, light-dependent emissions of volatile isoprenoids are generally limited to large amounts of isoprene, especially from fast-growing plants (Loreto and Fineschi 2015), and of monoterpenes from some families of trees and bushes (Loreto and Schnitzler 2010). Emission of isoprene is a metabolic cost for plants, but benefits may outweigh the cost, especially under high temperature (Jardine et al. 2012) and oxidative stress (Vickers et al. 2009). Isoprene was thought to quench reactive oxidative species (ROS) such as hydrogen peroxide ( $H_2O_2$ ) (Loreto and Velikova 2001), singlet oxygen ( $^1O_2$ ), (Affek and Yakir 2002) or reactive nitrogen species (RNS) (Velikova et al. 2005), and to stabilize chloroplast membranes (Velikova et al. 2011a) facilitating photosynthetic electron transport rate (Pollastri et al. 2019). The antioxidant role of isoprene is now revised, as the capacity of isoprene emission induces a reprogramming of the entire genome and metabolome (Monson et al., 2021; Dani and Loreto,

2022) that improve the ability of plants to tolerate various stresses. The general role of isoprene as a stress protective agent remains unquestioned.

Stress tolerance is improved because of carbon allocated for constitutive BVOC biosynthesis but also because of stress-induced BVOCs (Paré and Tumlinson 1999; Mithöfer and Boland 2012), especially after herbivores or pathogens attacks (Dicke and Baldwin 2010), or after abiotic stresses such as drought, high temperatures, or oxidative pollutants (Loreto and Schnitzler 2010).

Isoprene plays several roles in atmosphere chemistry, all of which are due to its oxidation (Heald et al. 2009; Archibald et al. 2010). When anthropogenic volatile pollutants such as nitrogen oxides ( $\text{NO}_x$ ) are absent, isoprene further cleanses the atmosphere of ozone. In the presence of  $\text{NO}_x$ , however, isoprene participates in reactions leading to increased ozone formation (Fehsenfeld et al. 1992) under a well-established stoichiometry (Kanakidou et al. 2005). As the emission of isoprene to the atmosphere is so prevalent, the impact of environmental factors such as light intensity, atmospheric  $\text{CO}_2$  concentration, temperature, relative humidity, and nutrient status on isoprene emission has attracted great attention (Loreto and Schnitzler 2010; Harrison et al. 2013). Climate change impact on isoprene emission has been mainly attributed to positive long-term (enzyme-driven) and short-term (substrate-driven) feedback of rising temperature (Lehning et al. 2001; Rennenberg et al. 2006), implying that future emissions of isoprene will also increase (Arneth et al. 2008). This may be counteracted by an often large (and largely unexplained) inhibition of isoprene in rising  $\text{CO}_2$  (Rosenstiel et al. 2003; Guidolotti et al. 2011). However, the inhibitory impact of rising  $\text{CO}_2$  seems to be lost as the temperature gets higher, and the overall impact of climate change is therefore expected to lead to a heavier load of isoprene in the atmosphere (Sharkey and Monson 2017). It may be also hypothesized that, in response to increasing environmental stresses and global warming, a shift of native plants toward species and genotypes able to emit isoprene constitutively or an induced manner will occur (Lerdau 2007). Here, we focus on reviewing anthropogenic atmospheric and soil pollutants and climate change that could also affect isoprene emission by natural vegetation and thus alter further the load of isoprene in the atmosphere.

## **Focus on isoprene and air pollution**

### *a) Isoprene and the chemistry of the troposphere.*

Most of the plant BVOCs have relatively short lifetimes in the atmosphere ranging from less than a minute to few hours depending on the atmospheric conditions (Blande et al. 2014). In the case of isoprene, rapid reaction with  $\text{NO}_x$  leads not only to ozone production, but also to the appearance of secondary products of isoprene oxidation, mainly methylvinyl-ketone (MVK), methacrolein (MACR) and 2-methyltetrols like 2-methylthreitol and 2-methylerythritol that have been found in the natural aerosol of Amazonia forest (Claeys et al. 2004).

Formaldehyde is also produced by isoprene oxidation and, despite the low yield (<10%), this BVOC has been used as an important proxy of isoprene natural planetary sources by satellite inspection (Palmer et al. 2006). MVK and MACR are markers of isoprene oxidation also *in planta* and therefore it is possible for these secondary BVOCs to be directly emitted by plants and not only formed by isoprene reactions in the atmosphere (Jardine et al. 2012). Recent results, however, suggest that MVK and MACR might be produced *in planta* by pathways other than isoprene oxidation (Kai et al. 2012), and that MVK may even be further oxidized to methyl ethyl ketone (MEK), making the pattern of interactions between plant BVOCs and atmospheric chemistry even more complex (Cappellin et al. 2019).

Indeed, ozonolysis (Pinto-Zevallos et al. 2010) results in the formation of many secondary organic aerosols (SOAs) (Seinfeld and Pandis 2006; Laothawornkitkul et al. 2009) with relevant climatic impacts (Claeys et al. 2004; Paulot et al. 2009). Isoprene, monoterpenes, and other terpenoids characterized by high emission rates and high reactivity with the atmospheric oxidants that are present in polluted and urban areas ( $\text{NO}_3^-$ , ozone, hydroxyl radical ( $\text{OH}^\cdot$ )), are major contributors of SOA burden (Kanakidou et al. 2005; Goldstein and Galbally 2007). Field studies have shown that under conditions with moderate to high BVOC levels,  $\text{NO}_3^-$  predominantly reacts with BVOCs (Brown and Stutz 2012) to produce multifunctional compounds such as organic nitrates (ONs) (Nah et al. 2015; Faxon et al. 2017).

#### *b) Impact of main atmospheric determinants of climate change on isoprene.*

The two main atmospheric constituents affecting isoprene emission are carbon dioxide ( $\text{CO}_2$ ) and ozone ( $\text{O}_3$ ). Anthropogenic  $\text{CO}_2$  emission is the most important forcing variable affecting changes in climate since the beginning of the industrial era. Over time,  $\text{CO}_2$  concentrations have continued to increase in the atmosphere, reaching 417 ppm (IPCC 2021) with further increases each year. A recent meta-analysis (Feng et al. 2019) summarized decades of experimental data (e.g. Possell, Hewitt, and Beerling 2004; Rosenstiel et al. 2003) showing a largely negative

impact of rising CO<sub>2</sub> on isoprene emission, while emission of monoterpenes is substantially unaffected by CO<sub>2</sub> accumulation. The negative impact of rising CO<sub>2</sub> on isoprene has surprised scientists, as isoprene is almost totally made by photosynthetic carbon (Delwiche and Sharkey 1993), and photosynthesis is stimulated by CO<sub>2</sub> (Long et al. 2004). It has been suggested that the decrease of isoprene emission when CO<sub>2</sub> increases is related to i) photorespiration inhibition, and to the consequent reduction of pyruvate available to the MEP pathway (Rosenstiel et al. 2003); or ii) to competition for phosphoenol pyruvate (PEP) a cytosolic substrate that may support chloroplastic demand (Loreto and Fares 2007); or iii) to an inhibitory effect on IspS activity (Scholefield et al. 2004). A hypothesis that the CO<sub>2</sub> inhibition was related to a triose phosphate utilization limitation of photosynthesis was recently ruled out (Lantz et al.) Guidolotti et al. (2011) found an inverse relationship between isoprene and intercellular CO<sub>2</sub> concentration ( $C_i$ ), which holds even at currently ambient CO<sub>2</sub> concentration (> 400 ppm). This supports the notion that the CO<sub>2</sub>-dependent reduction of isoprene reflects fine biochemical adjustments. However, the inhibitory impact of rising CO<sub>2</sub> seems to be lost as the temperature gets higher (a necessary consequence of CO<sub>2</sub> accumulation as CO<sub>2</sub> is a greenhouse gas), and the overall impact of climate change is therefore expected to lead to a heavier load of isoprene in the atmosphere (Sharkey and Monson 2017). Moreover, CO<sub>2</sub> may also indirectly stimulate isoprene emission at whole canopy and ecosystem level, because of higher photosynthesis, growth rate and biomass accumulation, made possible by the increase in CO<sub>2</sub> availability (Arneth et al. 2007). However, several lines of evidence indicate that CO<sub>2</sub> does not always increase linearly with an increase in photosynthesis and plant growth/productivity. Moreover, there is a significant interspecific variability in [CO<sub>2</sub>]-responsiveness of isoprene emission that is unexplained. Such variability in emission reduction could be caused by a significant variation in the size, composition of the precursor pools responsible for isoprene emissions (Niinemets et al. 2021). Squire et al. (2014) found that climate change, which includes both rising temperature and CO<sub>2</sub>, increased isoprene emissions by natural vegetation and the effect is expected to continue as long as CO<sub>2</sub> overfertilizes plants (Squire et al. 2014). Future increase of isoprene emission by natural vegetation is expected when accounting for rising temperature only (Sanderson et al. 2003; Lathière et al. 2005; Wu et al. 2012). By modelling temperature and CO<sub>2</sub> interaction (which includes direct and indirect CO<sub>2</sub> effects) indeed it is confirmed that isoprene emissions will be stimulated over the 21<sup>st</sup> century (Arneth et al. 2007; Heald et al. 2009). A framework modeling study based on a scenario where

the effect of climate and natural vegetation changes (driven by the rising of temperature and by the expansion of broadleaf forests respectively) co-occur, suggests an increase of isoprene emission by ~42% by 2050, which drops to ~4% if CO<sub>2</sub> inhibition of isoprene emission is also included (Tai et al. 2013).

We speculate that both effects of climate change and environmental stress could lead to an increase of isoprene-emitting species in polluted environments in response to the negative effects (e.g oxidative stress) resulting from increased air and soil contaminants (Figure 1). The other gas that has received large attention for its feedback on isoprene is ozone. While generally, CO<sub>2</sub> improves plant growth (Long et al. 2004), ozone is a serious environmental stress that causes heavy damage to photosynthesis. Indeed, when ozone enters the leaf, it is degraded to other ROS, which can cause oxidative stress and damage to lipo-proteic bilayer of the photosynthetic membranes, with consequent rapid chloroplast degradation (Loreto et al. 2001).

High doses of ozone could cause an initial stimulation of isoprene emission (Velikova et al. 2005) due to higher expression of IspS and its activity (Fares et al. 2006). It seems that this upregulation is more evident in leaves developing under enriched O<sub>3</sub> atmosphere and that build up a better resistance to pollutants (Fares et al. 2006). However, plants adapted to high O<sub>3</sub> are not able to further stimulate isoprene emission when exposed to a following stimulus (Calfapietra et al. 2008). If O<sub>3</sub> does not damage photosynthesis, isoprene emission is maintained and as already mentioned, often stimulated. Under these conditions, isoprene may scavenge ROS made by O<sub>3</sub> (Velikova et al. 2004) or may make the photosynthetic membranes more resistant to ROS-dependent oxidation (Pollastri et al. 2019; Velikova et al. 2011a). As isoprene is also stimulated by high temperatures, when high O<sub>3</sub> episodes are more frequent, the antioxidant role of isoprene becomes more relevant when associated with heat waves (Jardine et al. 2012). A first line of defense against O<sub>3</sub> is stomatal closure. Loreto and Fares (2007) showed that leaf damage is associated to O<sub>3</sub> concentration inside leaves rather than to the atmospheric O<sub>3</sub>. However, under prolonged or chronic exposure to O<sub>3</sub> that overcomes the epidermal barrier, photosynthesis is severely impaired and consequently also isoprene emission is restrained. Indeed, O<sub>3</sub> often irreversibly damages plant tissues leading to reduced crop yields and forest growth (Mills et al. 2010). The uptake of O<sub>3</sub> inside mesophyll causes oxidation of cell wall components, damages photosynthetic apparatus with detrimental effects on growth rate and biomass production, and accelerates leaf senescence (Ashmore 2005; Fares et al. 2006; Wittig et al. 2009). Meta-analysis

data show that isoprene and photosynthesis are reduced to similar extent (10%) by high O<sub>3</sub> exposure (Feng et al. 2019). However, isoprene emission is significantly increased by exposure of leaves to high UV-b (Harley et al. 2006; Tiiva et al. 2007) and UV-a (Pallozzi et al. 2013) radiation, which is a requisite for O<sub>3</sub> formation in the atmosphere. Thus, the overall impact of air pollution on isoprene emission needs additional field testing where all factors dynamically interact together.

Figure 1 summarizes the interaction between plant isoprene and atmospheric pollutants in cities and industrial areas, which may have two effects: on one hand this interaction may increase the O<sub>3</sub> load and high O<sub>3</sub> episodes may exacerbate environmental stresses; on the other hand, this same interaction may favor evolution of a vegetation that is resistant to O<sub>3</sub> pollution and associated oxidative stresses. As isoprene is involved in antioxidant protection (Loreto and Schnitzler 2010) this may lead to higher isoprene emission by both native and alien (invasive) species (Lerdau 2007). The two effects may feedback on each other, and the loop may cause unpredictable consequences. Llusia et al. (2010) suggested that protection against multiple environmental stress conferred by high capacity to emit terpenoids accounted for the success of invasive plant species in Hawaii. Similarly, establishment and proliferation of *Artemisia vulgaris* in a new habitat seems to be related to its capacity to emit BVOCs (Barney et al. 2005). On the other hand, it is also conceivable that human-driven land use change, by replacing natural vegetation with agricultural crops, has also selected against high isoprene emitters (Loreto and Fineschi 2015). Changes in natural vegetation (reduction of isoprene emitting species) could affect air quality (Tai et al. 2013; Hantson et al. 2017). Clearly, understanding the future effects of climate change on isoprene emission is a very complex task, because of the wide range of multiple and concomitant environmental factors that could have synergistic or antagonistic effects.

The future rising of environmental stresses (known to trigger oxidative stresses) related to anthropogenic processes could lead to a positive feedback for isoprene (both constitutive and induced, Figure 1) and other BVOCs biosynthesis and emission, acting as a plant-defense-system in response to climate change and warming (Peñuelas and Llusia 2003). For example, field measurements showed that white oak tree canopies have higher isoprene emission rates when exposed to more sunlight, reduced water availability, and high temperature (Sharkey et al. 1996). Interestingly, these plants did not show any anomalies in their growth and an increased

thermotolerance (Singsaas et al. 1997) and it is suggested that the quenching of ROS by isoprene could be an effective way to reduce the negative effects of oxidative stress compounds (Velikova et al. 2005).

#### **Focus on isoprene and soil pollution**

Soils may also contribute to the exchange of BVOCs, as sinks or sources, depending on the very diverse soil composition in terms of microorganisms, flora and fauna, thus expanding functional considerations on trophic interactions from the aboveground to the belowground plant compartment (Penuelas et al. 2014). While the complex relationships between isoprene and atmospheric pollution have been largely investigated, much less is known about the impact of soil pollution on isoprene. One could speculate that isoprene antioxidant action also improves resistance to soil pollution. “Soil pollution” refers to the presence of a chemical or substance out of place and/or present at a higher-than-normal concentration that has adverse effects on any non-targeted organism (FAO and ITPS 2015). Soil pollution acts as an abiotic stress on plants. It triggers ROS production, leading to adaptive plant responses including the improvement of the primary antioxidant redox system and the increase of the biosynthesis of secondary metabolites.

Except natural areas with specific geological conditions, the major soil pollutants are related to anthropogenic (industrial) activities that release different kinds of pollutants, from complex hydrocarbons released by oil industries to very simple chemical elements such as heavy metals released as byproducts of several processes or nutrient elements such as nitrates from excess fertilization or phosphates from commercial cleaning industries.

Heavy metals soil pollution is a problem of major importance for plant productivity and survival (Salt et al. 1998; Foy et al. 2003; Fargasová and Molnárová 2010). There are several cases in which the effect of heavy metal pollution on BVOC emissions has been investigated. Velikova et al. (2011b) suggested that heavy metal (Ni) pollution increases both constitutive (isoprene) and induced (monoterpenes and sesquiterpenes) isoprenoid emissions. Indeed, other reports indicate that high doses of Cu could induce BVOCs (Obara et al. 2003; Mithöfer et al. 2004) some of which characterize the interplay plants-herbivores (Winter et al. 2012). Soil cadmium stress seems to increase total leaf VOC emission (Lin et al. 2022). A time course with cadmium stress induced an upregulation of isoprene synthase (Li et al. 2017). Moreover, isoprene affects heavy metals detoxification transcriptome (Zuo et al. 2019).

The impact of nutrients on isoprene emission has received even wider coverage. Nutrient excess is often a consequence of pollution and over-fertilization, and finally eutrophication (Shortall 2013). Nitrates seem to generally elicit production and emission of isoprene, possibly making more N available for isoprene synthase biosynthesis (Litvak et al. 1996; Fernández-Martínez et al. 2017). On the other hand, excess of phosphorus in soils has a clear inhibitory effect on isoprene emission, assessed in different experiments (Fares et al. 2008; Cocozza et al. 2019, 2020) but never explained physiologically. Intuitively, high phosphorous should be beneficial for the synthesis of a molecule that requires large inputs of phosphorylated substrates, like ATP and NADPH (Sharkey and Yeh 2001). High phosphorous also stimulates photosynthesis which supplies carbon for isoprene synthesis. However, uncoupling of isoprene synthesis and photosynthesis is often observed, for example under elevated CO<sub>2</sub> (cfr. see above), or under water and salt stress (Loreto and Schnitzler 2010). It was proposed that competition with mitochondrial respiration for pyruvate or phosphoenolpyruvate is responsible for the inhibition of isoprene emission under high phosphorous nutrition (Fares et al. 2008), similar to what may occur under elevated CO<sub>2</sub> (Rosenstiel et al. 2003), although we think that all pyruvate for isoprene synthesis comes from the Calvin-Benson cycle Sharkey et al. 2020.

While different nutrients may have opposite effects on isoprene emission, a reduction of the intensity of emitted BVOCs in plants cultivated under high level of fertilization seems to be a convergent result (Fernández-Martínez et al. 2017), which may also explain why storage of BVOCs into reservoirs is a lost trait in recently evolved angiosperm crops. Evolution against emission of BVOCs may have an important trade-off in terms of improved plant productivity in absence of stress, but losing the capacity to synthesize and emit BVOCs may not pay off when plants must defend themselves from abiotic stresses or must communicate with other organisms. We eventually hypothesize that native and pioneer plants of polluted areas emit more isoprene (constitutive and induced) and speculate that the capacity to produce large amounts of isoprene may confer an adaptive advantage in a rapidly changing climate characterized by more frequent extreme events and pollution episodes.

Water and salt are important components of soils. Isoprene synthesis and emission continues even under drought or high salinity, despite concurrent photosynthesis inhibition (Brilli et al. 2007). The massive literature covering the impact of these stresses on isoprene was

often reviewed (e.g. Loreto and Schnitzler 2010) and the topic is beyond the scope of this work aiming at reviewing only impacts of soil pollutants.

Soil structure and composition influence the development and morphology of the root system. The root is the anchorage system of plants and is critical for the uptake of nutrients required for plant growth and physiology, including the isoprene pathway. If for the aboveground part of plants (leaves) there is much scientific evidence on the effects and activities of in-situ emission of isoprene, less well studied is whether belowground heterotrophic tissues (roots) can also emit isoprene, and if roots are also influenced when plants acquire or enhance their capacity to emit isoprene. Although it is mainly emitted from leaves, there is evidence that the root systems of poplar (Ghirardo et al. 2011) and transgenic *Arabidopsis* (Loivamäki et al. 2007; Miloradovic van Doorn et al. 2020) emit a small amount of isoprene. It is shown that the constitutive promoter of isoprene synthase (PcISPS) is present and active in specific regions of roots (Cinege et al. 2009; Miloradovic van Doorn et al. 2020).

There are reports of isoprene affecting root development. Miloradovic van Doorn et al. (2020) recently proposed a ROS-related role of isoprene in roots, showing an altered lateral root development and differences in ROS accumulation in roots. ROS are known to be involved in many pathways, especially under a challenging environment, being signals able to activate defenses responses also coordinating the developmental processes with environmental conditions (Locato et al. 2018). An interplay between ROS and hormones, in particular auxin, ethylene, and abscisic acid has also been reported (Xia et al. 2015). Isoprene also seems to interfere with many hormones, especially those sharing the same MEP pathway (cytokinins and abscisic acid) but possibly also with auxins (Dani and Loreto 2022). The effect of these interactions on roots is unknown, but a possible scheme of signaling function of internal isoprene in roots in relation with ROS was proposed (Miloradovic van Doorn et al. 2020). ROS signaling affects hormonal networks and signaling processes that regulate response to environmental drivers (Mittler 2017). It was proposed that isoprene could adjust ROS by quenching (direct) or changing gene expression (indirect) and so regulate all ROS-related pathways, including those involved phyto-hormones. Even if the molecular mechanism of this interplay is still relatively unknown, this could influence the growth of the root system in particular the lateral roots (Miloradovic van Doorn et al. 2020). Isoprene affects also root proteome and many of the proteins affected are involved in redox and stress responses (Miloradovic van Doorn et al. 2020).

Soil pollutants directly affect root development and the determinants of root system architecture (Lombardi et al. 2021) and this might also contribute to change BVOC emission by aboveground and belowground plant organs (Figure 1,2), in turn altering plant capacity to cope with pollution and environmental constraints (e.g. drought stress). Finally, root volatiles often are key elements of plant-plant communication and for interactions of plants with soil microbiome (Figure 2), with positive consequences on priming defensive responses, facilitating root nutrient uptake or counteracting the negative effects of pollutants. For example, BVOCs emitted from roots may facilitate interactions with arbuscular mycorrhizal fungi, expanding their beneficial functions, from improving resistance to soil stresses to enhancing nutrient availability (Asensio et al. 2012). However, whether these same functions may be attributed to isoprene is unclear. We hypothesize that this beneficial interaction (BVOCs-soil microbiome) could lead to better plant tolerance to soil stress and have positive feedback on plant biomass (Figure 2). In leaves, isoprene does not seem to be a messenger able to induce priming of receiving plants (Giordano et al. 2021), and does not influence insect feeding (Brilli et al. 2009). This may very well be the case in roots as well, where the emission of isoprene by plants is also elusive, and possibly tiny.

Soil constraints increase ROS formation, and it was shown that ROS, and ROS scavenging enzymes, play crucial roles in early-stage root-mycorrhiza interaction (Baptista et al. 2007; Nanda et al. 2010; Ditengou et al. 2015). Several studies shown that the intensity of ROS burst is important for root-microbes (mutualistic or pathogen) interaction and contact (Baptista et al. 2007; Nanda et al. 2010) and plant redox balance could be fundamental to differentiate between the various microbes. ROS adjustment by isoprene (direct or indirect) could be crucial for regulating root redox balance and root isoprene emission could facilitate the interaction and communication with soil microbiome (Figure 2). It is known that volatiles are a signal for aboveground plant communication (Ninkovic et al. 2021) but it was reviewed that root volatiles might play a role in belowground communication (plant-plant; plant-soil microbiome) (Abbas et al. 2022).

The role of isoprene in belowground communication is unclear and unknown while for root volatiles (BVOCs) with their diversity and their specificity are have been extensively analyzed in the belowground communication (Schenkel et al. 2015; Abbas et al. 2022). So, we suggest that in polluted environments, root-induced BVOCs might lead to a further increase in

this communication between plants via belowground, and thus they might ameliorate the negative effect of soil constraints (Figure 2).

## **Conclusion and future directions**

Over the past few years, isoprene emission has been well-studied for its effects on atmospheric pollution and on plant defense, but the reciprocal impact of isoprene, soil and atmospheric pollutants is more elusive and complicated, especially when considering that direct measurements of isoprene emission are missing (soil) and that long-term responses at whole plant and community level have not been extensively investigated (both in soil and air).

Future studies should evaluate the long-term effect of pollutants, including evolutionary impacts on the composition of natural and semi-natural forests around cities and industrial areas where anthropogenic pollution may be persistent over time. This allows a better evaluation on how policies of re-forestation and afforestation of these areas may impact on air quality, considering climate change pressures, which may lead to regional expansion of broadleaf forests, the main emitters of isoprene (Wu et al. 2012), and in boreal areas. Finally, soil pollution impacts on isoprene is largely uninvestigated, and the impact of soil microorganisms is also basically unknown, although preliminary experiments indicate that beneficial microorganisms such as mycorrhiza (Pollastri et al. 2018) and plant growth-promoting rhizobacteria (Brunetti et al. 2021) may stimulate isoprene emission.

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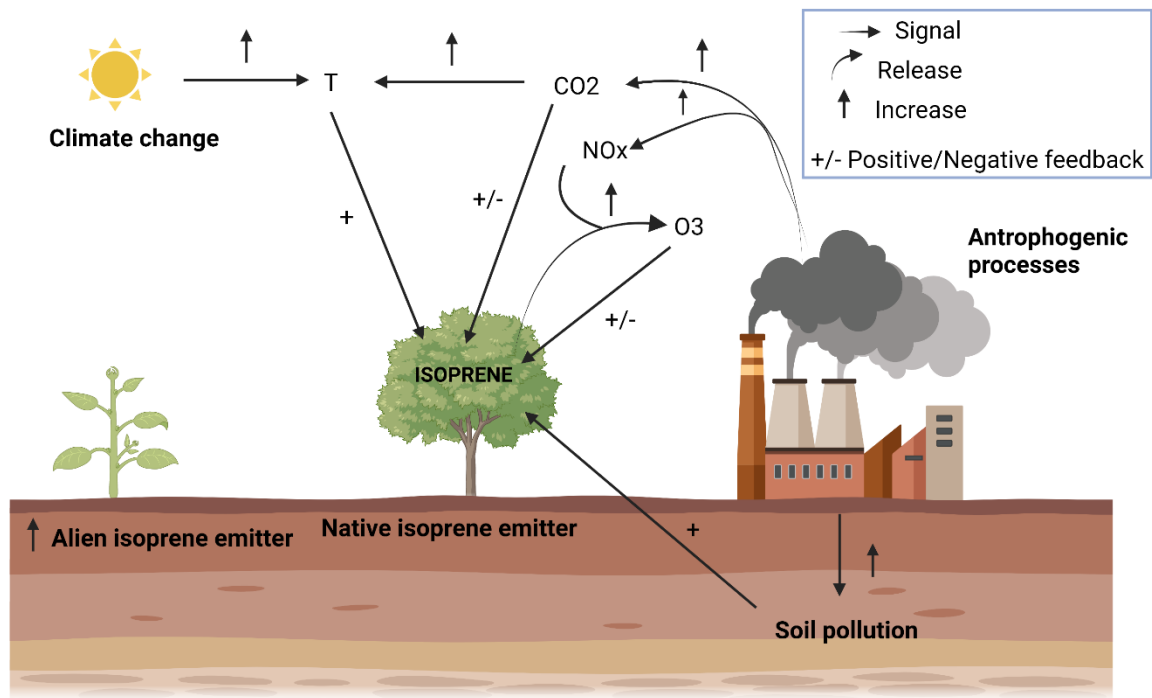
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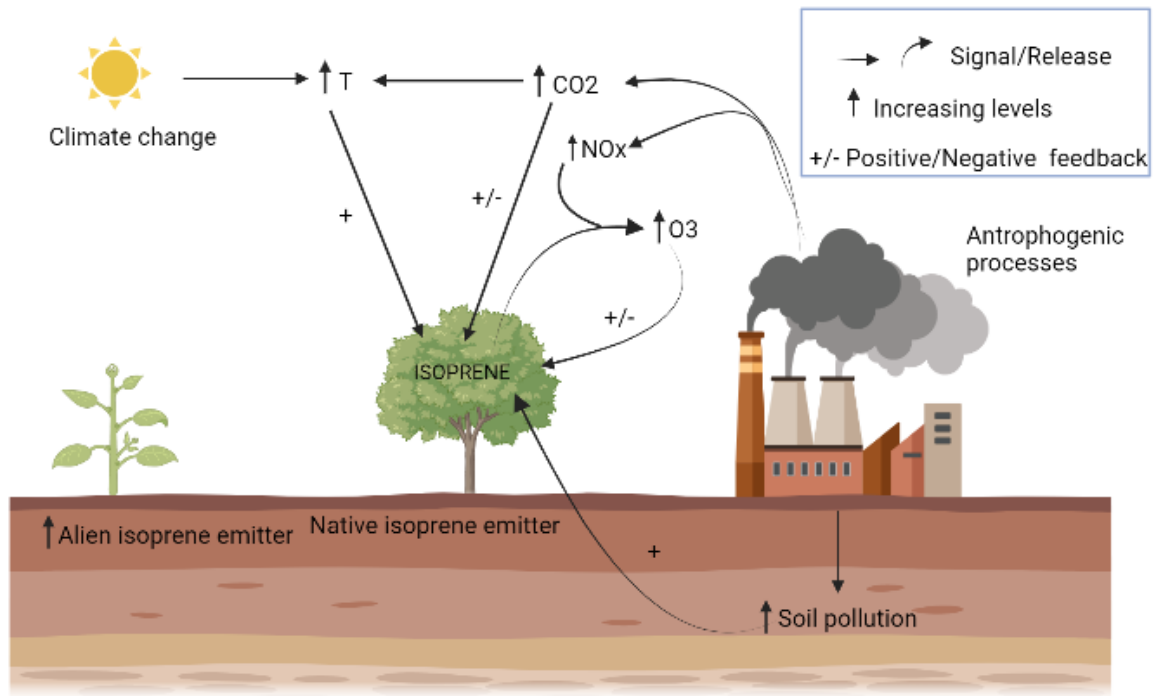
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**Fig. 1** Dynamic of plant isoprene emission and settlement of alien isoprene-emitting species induced by anthropogenic processes determining environmental stresses or by climate change. Positive or negative feedback (+/-) refers to the direct effects of pollutants on the isoprene emission capacity. ↑ represent the direct effects of anthropogenic processes and climate change





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*Figure 1 Dynamic of plant isoprene emission and settlement of alien isoprene emitting species induced by anthropogenic processes determining environmental stresses or by climate change. + or – refer to the direct effects of contaminants on the emission of isoprene while  $\uparrow$  represent the effects of anthropogenic processes and climate change on air and soil determinants and settlement of alien species.*

