

## Microplastics are ubiquitous and increasing in soil of a sprawling urban area, Phoenix (Arizona)



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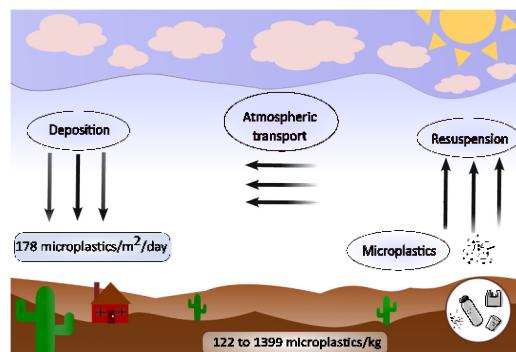
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### HIGHLIGHTS

- Microplastics occurred in all soil samples collected from a semi-arid region in AZ.
- Spatial variability of microplastics depicted substantial variability.
- A significant increase in soil microplastics was observed from 2005 to 2015.
- The average sizes of soil microplastics is smaller in 2015 compared to those in 2005.
- PE, PS, PVC, PA, PES and PP were the major polymers identified.

### GRAPHICAL ABSTRACT



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### ABSTRACT

Microplastics are environmental contaminants that have been extensively studied in marine and aquatic environments; terrestrial ecosystems, where most microplastics originate and have the potential to accumulate, typically receive less attention. This study aims to investigate the spatial and temporal soil concentrations of microplastics in a large desert metropolitan area, the Central Arizona-Phoenix Long-Term Ecological Research (CAP-LTER) area. Soil samples from the Ecological Survey of Central Arizona (ESCA) surveys (2005 and 2015) were leveraged to study spatial distributions and the temporal change of microplastic abundances. The temporal soil microplastics data were supplemented by microplastics deposition fluxes in a central location within the area (Tempe, AZ) for a period of one year (Oct 5th, 2020 to Sept 22nd, 2021). Samples were processed and microplastics were counted under an optical microscope to obtain quantitative information of their distribution in soil. Results for the spatial variation of the microplastic abundances in soil samples in Phoenix and the surrounding areas of the Sonoran Desert from 2015 depict microplastics as ubiquitous and abundant in soils (122 to 1299 microplastics/kg) with no clear trends between locations. Microplastics deposition fluxes show substantial deposition in the focal area (71 to 389 microplastics/m<sup>2</sup>/day with an average deposition flux of 178 microplastics/m<sup>2</sup>/day) but the role of resuspension and redistribution by dust storms to deposition may contribute to the unclear spatial trends. Comparison between the 2005 and 2015 surveys show a systematic increase in the abundance of microplastics and a decrease in microplastics size. Micro-Raman spectroscopy identified a variety of plastics including PE, PS, PVC, PA, PES and PP. However, a majority of microplastics

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remained chemically unfiendifiable. Polyethylene was present in 75 % of the sampling sites and was the most abundant polymer on average in all soil samples.

## 1. Introduction

Microplastics are a growing concern as pervasive environmental pollutants. They are mostly secondary in nature and result from the breakdown of larger plastics over time into microplastics, generally defined as plastic particles < 5 mm in size (Arthur et al., 2009). Primary microplastics are purposefully manufactured at this small size, tend to have uniform shapes including spherical microbeads, and originate from sources such as personal care and abrasive cleaning products (Reveillet et al., 2021). Secondary microplastics generally exhibit more diverse shapes.

Existing research shows that microplastics are found in interconnected environments; from the Arctic to Antarctica, including the depths of oceans, freshwater bodies, atmospheric deposition and soil environments (Patiño et al., 2022; Ifet al., 2023; Zhao et al., 2023). While numerous studies have investigated the occurrence and abundance of microplastics in marine and freshwater environments (Andrade, 2011; Erfksen et al., 2013; Horton et al., 2017; Bergmann et al., 2017; Klefin and Fischer, 2019; Chandranthan et al., 2023) research in terrestrial ecosystems remains less well developed despite the fact that a majority of the plastic debris reaching the oceans has originated on land (Lebreton et al., 2022).

A review study has estimated that the annual release of plastic waste to land is 4–23 times greater than that released to the oceans (Horton et al., 2017). Although soil microplastics were first reported in 2016 (Ding et al., 2022), microplastics research in terrestrial ecosystems remains underrepresented and accounts for only about 5 % of the research on microplastics (Ding et al., 2022). Street runoff, landfills and atmospheric deposition are potential input pathways of microplastics into terrestrial ecosystems, making soils a significant sink for synthetic microplastics in terrestrial ecosystems (Möller et al., 2020). Moreover, plastic mulch films and soil conditioners used in agricultural amendments are potential sources of soil microplastics (Ng et al., 2018). Soils can also act as a source of atmospheric microplastics to other environmental reservoirs through resuspension of fugitive dust, thereby contributing to the global microplastics transport cycle (Zhang et al., 2020).

Microplastics in terrestrial environments can cause changes in soil ecosystems affecting soil structure, plant growth and microbial activity (De Souza Machado et al., 2019). A meta-analysis in agricultural soils in China revealed that microplastics affect soil enzymes including glucose oxidase, urease, and cellulase activity (Liang et al., 2023). In soil environments, microplastics can degrade through microbial action as microorganisms can utilize the carbon in plastic polymer chains for their growth (Mohan et al., 2020; Huang et al., 2023). Owing to the flow of organic matter in dryland soils (Marusenko et al., 2011), soil microbes in these ecosystems may resort to metabolize anthropogenic carbon sources such as microplastics. Studies have shown microplastics to have an acute toxic effect on soil organisms such as nematodes (Klim et al., 2020). In addition to the direct toxic effects caused by microplastics, they can also act as vectors of other contaminants such as Polycyclic Aromatic Hydrocarbons (PAHs) and metals found in soils (Hildebrandt et al., 2021; Hu et al., 2022a, 2022b).

This study investigates soil concentrations and deposition fluxes of microplastics in the Phoenix, AZ, metropolitan area. The study area is vast (6400 km<sup>2</sup>) and is a rapidly growing urban metropolitan that has been intensively studied for urban ecology for over 20 years (Griffith and

Redman, 2004). Our study utilizes samples from the Ecological Survey of Central Arizona (ESCA) performed by the Central Arizona-Phoenix Long-Term Ecological Research (CAP-LTER), a large-scale field survey that is conducted in the CAP-LTER study area characterizing the urbanized, suburbanized, and agricultural areas of metropolitan Phoenix and the surrounding Sonoran Desert (Griffith and Redman, 2004). Previous studies have investigated various key ecological indicators and contaminants, including soot black carbon concentrations and isotopic compositions, PAHs and lead concentrations in soils (Marusenko et al., 2011; Zhuo et al., 2012; Hamilton and Hartnett, 2013). We leveraged the Ecological Survey of Central Arizona (ESCA) 200-point survey (2005 and 2015) to study spatial distributions and the temporal change of microplastic abundances. We supplement the temporal data in soils with the collection of atmospheric deposition fluxes of microplastics in a central location within the area (Tempe, AZ) for a period of one year (Oct 5th, 2020 to Sept 22nd, 2021). All soil and deposition samples were processed and characterized using an optical microscope to obtain quantitative information of the distribution of microplastics. Chemical characterization of microplastics was performed by micro-Raman spectroscopy to understand their chemical composition.

## 2. Materials and methods

### 2.1. Sampling of microplastics

#### 2.1.1. Sampling of microplastics in soil

Soil samples used in the study were collected by CAP-LTER from the ESCA 200-point survey (2005 and 2015). CAP-LTER program is one of twenty-eight LTER sites funded by the National Science Foundation and is one of the two LTER sites that specifically study urban ecology (Griffith and Redman, 2004). CAP-LTER studies how human activities alter the functioning of ecosystems, and subsequently urban sustainability, in Central Arizona and metropolitan Phoenix where two tributaries of the Colorado River, the Salt and Gila Rivers, merge (Griffith et al., 2013). The ESCA 200-point survey is a field survey that is conducted every 5 years at approximately 200 sample plots randomly located in the urbanized, suburbanized, and agricultural areas of metropolitan Phoenix, and the surrounding Sonoran Desert (Fig. S1) (Griffith and Redman, 2004). The sampling plots (30 m × 30 m) are randomly located using a tessellation-stratified dual-density sampling scheme. Forty-eight samples from 2015 were selected to study the spatial variation aspect, while fourteen samples from 2005 and 2015 were analyzed for the temporal variation of microplastics. Topsoil samples (2 cm depth) were used for this study as this range could be expected to have the highest abundance of microplastics accumulated over time by direct input from terrestrial sources and atmospheric deposition. Top soil samples from the same study area were sampled and analyzed in a previous study for other environmental contaminants, such as PAHs (Marusenko et al., 2011). The soils analyzed in this study are desert soils samples, mainly from urban and suburban areas of metropolitan Phoenix and remote areas in the surrounding areas of the Sonoran Desert. According to the United States Department of Agriculture – National Resource Conservation Service (USDA-NRCS, <http://soils.usda.gov/>), soils in Central Arizona are classified as arid soils (too dry for the growth of mesophytic plants) and entisols (little to no evidence of developing pedogenic horizons) (Bohn et al., 2001; Hamilton and Hartnett, 2013).

#### 2.1.2. Sampling of microplastics in air

Total atmospheric fallout (dry and wet deposition) samples were collected by means of glass funnels in 4 L glass bottles (Fig. S2) on the observation deck of the Interdisciplinary Science & Technology Building

#4 on the Tempe campus of Arizona State University (33.4179° N, 111.9284° W). Samples were collected for 14 days each, from Oct 5th, 2020 to Sept 22nd, 2021. The suburban location is located in the city of Tempe, Arizona, approximately eight miles east of downtown Phoenix. Once collected, the samples were covered with aluminum foil to prevent contamination.

## 2.2. Sample processing

The soil samples were weighed (60 g) and sieved (8"-FH-SS-SS-US-#3-1/2, Hogentogler & Co. Inc., MD, USA) to remove material larger than 5.6 mm. The following approach was based on the methods manual from the National Oceanic and Atmospheric Administration (NOAA) for microplastics analyses in water and beach sediment samples and adapted as appropriate to process the soil samples (Masura et al., 2015; Chandranthan et al., 2023). First, sieved samples were subjected to wet peroxide oxidation to remove natural organic matter. A 10 mL aqueous 0.05 M Fe (II) solution prepared from FeSO<sub>4</sub>·7H<sub>2</sub>O (Sigma-Aldrich, MO, USA) and 10 mL of 30 % hydrogen peroxide (Sigma-Aldrich, MO, USA) were added to samples for 5 min at room temperature and subsequently heated to 60 °C on a hot plate (Fischer Scientific, NH, USA) to remove naturally occurring organic matter. Samples were subsequently made to undergo a density separation step using a 1.8 g cm<sup>-3</sup> NaI solution prepared from NaI (Sigma-Aldrich, MO, USA). The saturated NaI solution allows the separation of microplastics, including relatively denser plastics such as PVC (<1.58 g cm<sup>-3</sup>) from the sample (Nueffle et al., 2014; Zhang and Liu, 2018). After allowing solids to settle, the floating microplastics were filtered through pre-fired glass microfiber filters (Whatman GF/A, Sigma-Aldrich, MO, USA). The filters were oven-dried at 80 °C until complete dryness in a Muffle furnace (Vulcan 3-1750, CA, USA).

The sample collectors for total atmospheric fallout (funnel and bottle) were thoroughly rinsed thrice with ultrapure water (>18.4 MΩ cm, Purelab Flex, IL, USA) to ensure that all microplastics adhering to the glass surface were recovered. Samples were processed similar to an approach adopted in a recent study reporting microplastics suspended in air in Tempe, AZ (Chandranthan et al., 2023). The aqueous extracts were obtained and subjected to wet peroxide oxidation to eliminate natural organic matter. The solutions were vacuum filtered through baked glass fiber filters (Whatman GF/A, Sigma-Aldrich, MO, USA). The filters were then oven-dried at 80 °C to complete dryness in a Muffle furnace (Vulcan 3-1750, CA, USA).

Preventive measures to minimize background contamination during microplastics analyses is important to produce reliable results (Prata et al., 2021; Shanmugam et al., 2022; Hidalgo-Ruz et al., 2012). A procedural blank was run during all stages of sample processing including wet peroxide oxidation, density separation and subsequent analyses. Glassware washed with a detergent and rinsed three times with ultrapure water (>18.4 MΩ cm, Purelab Flex, IL, USA) were subsequently baked and used during all stages of sample processing and analyses. The clean equipment was covered to minimize potential airborne microplastics contamination. Working surfaces were cleaned with ultrapure water (>18.4 MΩ cm, Purelab Flex, IL, USA) and isopropyl alcohol (Fischer Scientific, MA, USA) prior to sample analysis. The usage of synthetic textiles was minimized, and a fiber was drawn from the lab coat during each day of laboratory analysis and observed under the optical microscope to check for fibers with similar appearance in samples. The recovery percentage of microplastics from filters was assessed through matrix spikes with a known number of microplastics (5 µm–5000 µm). A soil sample (60 g) sourced from the ESCA 200-point survey in 2015 was spiked with a known number of microplastics ( $n = 60$ ). The microplastics were photographed and sized using a digital microscope (Leica DM6B-Z, Germany) equipped with the Leica DFC7000 T camera and the Leica Application Suite X (LAS X) software prior to spiking, and were distinctive of color and shape. The artificially incorporated microplastics included Polyethylene (PE), Polypropylene

(PP), Polyvinyl Chloride (PVC), Polystyrene (PS) and Polyethylene terephthalate (PET) ranging in size from 5 µm - 5000 µm. The spiked soil sample was subjected to wet peroxide oxidation and subsequently density separation similar to the approach in Section 2.2. The spiked microplastics were observed on a filter under the digital microscope (Leica DM6B-Z, Germany) equipped with the Leica DFC7000 T camera and the Leica Application Suite X (LAS X) software. The recovery test showed a recovery of 95 % of spiked microplastics.

## 2.3. Optical microscopy

Filters were examined under a digital microscope (Leica DM6B-Z, Germany) equipped with the Leica DFC7000 T camera and the Leica Application Suite X (LAS X) software. The size range of interest was selected as 5 µm - 5000 µm; both the size and the morphology of each microplastic was noted to characterize each sample. The lower size limit is 5 µm due to the smallest possible size resolution of the digital microscope. As there is no standard definition for the distinction between a fiber and a fragment in the microplastics literature, we used an operational definition for the observed shapes of microplastics where a fiber was recognized to be cylindrical in shape with an aspect ratio (length/diameter) ≥ 3, while a fragment was recognized to be shard-like and flattened. Microplastics were sized along the largest dimension using the software program ImageJ (version 1.5, National Institute of Health, USA, <http://imagej.nih.gov/ij>).

The following criteria were used to identify microplastics under the optical microscope (Hidalgo-Ruz et al., 2012; Chandranthan et al., 2023). Only objects that did not display cellular or organic structures were identified as microplastics. Further, only fibers that were generally evenly thick through the entire length exhibiting 3-dimensional bending were identified as microplastics with further accounting for fiber splicing which can occur in microplastic fibers. Finally, particles that exhibit clear and homogenous color throughout the entire structure were identified as microplastics with the exception of the appearance of bleaching, biofouling and weathering which could cause patterns or stripes.

## 2.4. Statistical analysis of the temporal variation of soil microplastics (2005 and 2015)

A Mann-Whitney Test ( $p \leq 0.05$ ) was performed using OriginPro, Version 2023 (Origin Lab Corporation, Northampton, MA, USA) to compare differences between the soil microplastic abundances in 2005 and 2015.

## 2.5. Chemical characterization using Micro-Raman spectroscopy

Micro-Raman spectroscopy was used to provide chemical characterization of the polymers in identified microplastics. All microplastics counted under the optical microscope were analyzed under the micro-Raman spectrometer. Raman data was collected from 50 to 3800 cm<sup>-1</sup> using a custom-built microscope and a Mitutoyo M Plan Apo 50×/0.42 objective. An Acton Research monochromator (SpectraPro-300i) utilizing a 600 g/mm grating and coupled to a Roper Scientific CCD (model LN/CCD 1340/100-EB/1) recorded the signal. A Coherent Sapphire solid state cw laser emitting 532 nm served as the excitation source and the data was collected using cyclohexane with known peak positions. The spatial and spectral resolutions of the micro-Raman spectrometer were 1 µm and 1 cm<sup>-1</sup> respectively.

## 3. Results and discussion

### 3.1. Spatial variation of microplastics

Microplastics were found in all soil samples, thereby indicating the ubiquitous presence in soil environments in Phoenix and surrounding

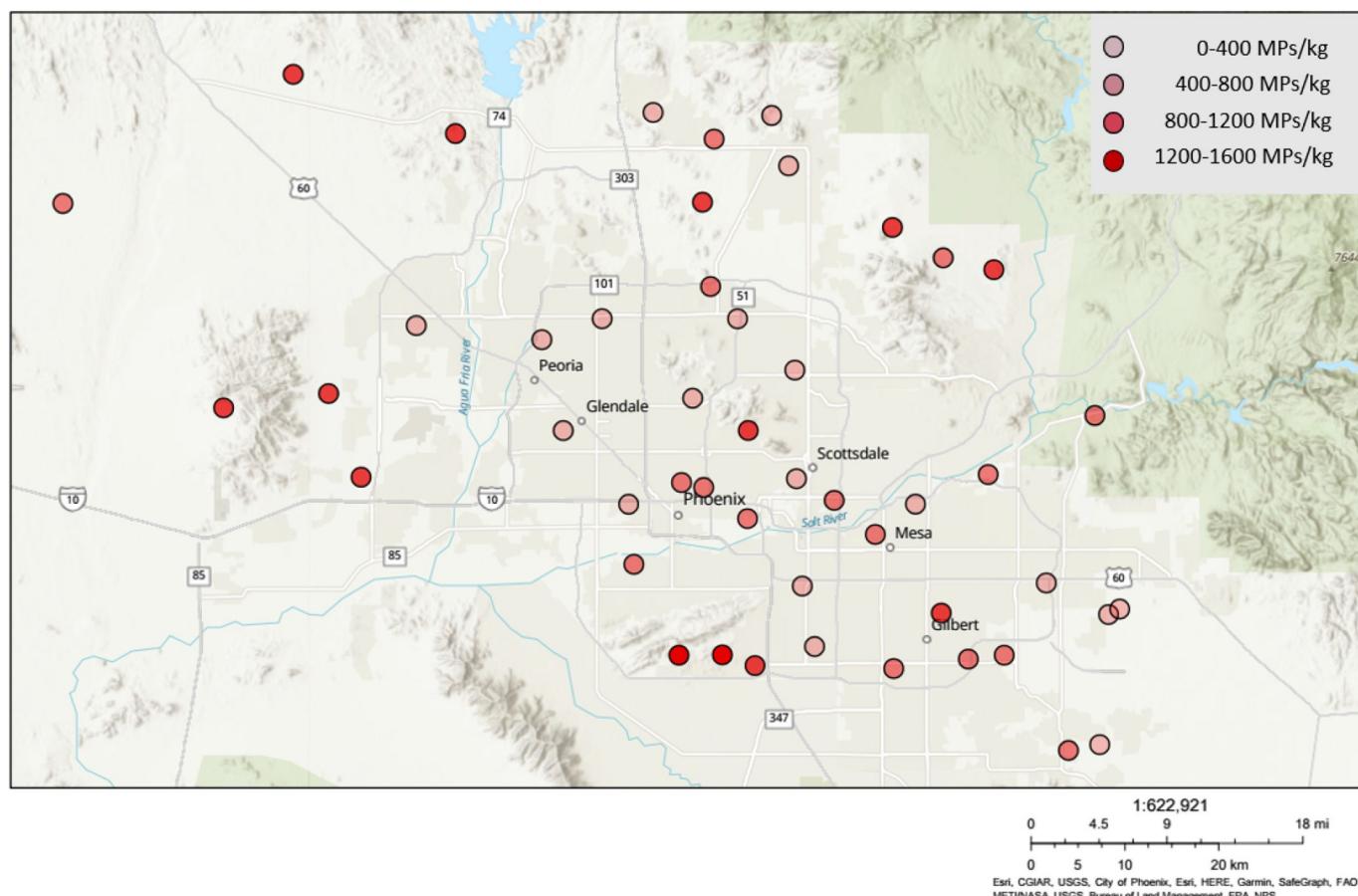
areas of the Sonoran Desert (Fig. S3). The microplastics abundance in soil samples from 2015 ranged from 122 to 1399 microplastics/kg with a heterogeneous spatial distribution depicting no clear spatial trends (Fig. 1). Three replicates sampled were analyzed for each sampling location (Table S1). The relative standard deviation of the measurements was on average 6 %. The highest microplastic abundance was found at Site 14 which is located in a residential area within the city of Phoenix, AZ. The lowest microplastic abundance was observed at sampling Site 8; a residential location within the city of Peoria, AZ.

Sampling sites K6, I14, L14, U21, AB19 had higher microplastic abundances compared to other sampling locations and were in the vicinity of roadsides or public alleys.

Comparing the observed microplastic abundances with those reported in literature is not straightforward due to the discrepancies in sampling approaches, sample processing, microplastics identification, operational definitions of microplastics and instrument detection limits. However, it can be useful to put microplastic abundances in soil into context by comparing with other existing studies (Table 1). A recent review reported that microplastics in soil globally range from 0.36 to 160,000 microplastics/kg, where a majority of the reported studies are from China (Zhang et al., 2022). Microplastic abundances reported in Lut and Kavir Deserts in Iran, Northeast Tibetan Plateau in China, Swiiss Flood plain soils and Badain Jaran Desert in China are 1–2 orders of magnitude lower than that observed in the current study (Scheurer and Biegafke, 2018; Feng et al., 2020; Abbas et al., 2021; Ifti et al., 2022). This could be attributed to the relatively higher soil size limits of interest in the aforementioned studies ( $\leq 100 \mu\text{m}$ ,  $20 \mu\text{m}$ ,  $125 \mu\text{m}$  and  $40 \mu\text{m}$ ) when compared to that of the current study ( $5 \mu\text{m}$ ) as this study observed a greater number of microplastics in the smaller size categories (Section 3.2). The observed differences in microplastic abundance could also be

**Table 1**  
Summary of soil microplastics reported in previous studies.

Location	Microplastics/kg	Size range/ $\mu\text{m}$	References
Lut and Kavir Deserts, Iran	20 (0–83)	$\leq 100$ –500	(Abbas et al., 2021)
Northeast Tibetan Plateau, China	47 (20–110)	20–5000	(Feng et al., 2020)
Swiss Flood plain soils	$\leq 593$	125–5000	(Scheurer and Biegafke, 2018)
Northern Germany farmland soils	$4 \pm 12$	1000–5000	(Harms et al., 2021)
Shanghai, China farmland soils	$78 \pm 13$	30–5000	(Ifti et al., 2018)
Shaanxi Province, China	1430–3410	Dominant <500	(Ding et al., 2020)
Southwestern China	18,760 (7100–42,960)	Dominant (50–1000)	(Zhang and Ifti, 2018)
Taklamakan desert, China	119–7292	Dominant <500	(Ifti et al., 2022)
Mu Us Desert, China	2696 (1360–4960)	0–5000	(Ding et al., 2021)
Lahore, Pakistan	4483 (1750–12,200)	50–5000	(Raffique et al., 2020)
Fars province, South Central Iran	Field 1: 380 (40–830)	$\leq 100$ –5000	(Rezaei et al., 2022)
Field 2: 510 (200–1100)			
Badain Jaran Desert, China	6 (1–12)	40–5000	(Wang et al., 2021)
Phoenix and the surrounding areas of the Sonoran Desert	122–1399	5–5000	Current study



attributed to the differences in microplastics input via dry and wet deposition, resuspension and dilution by mobile dust/sand particles, and environmental conditions that can weather microplastics to sizes that evade identification by the analytical techniques used in each study (e.g. make particles smaller). The microplastic abundance in the current study is an order of magnitude lower than that reported in Lahore, Pakistan (Rafique et al., 2020); note this study specifically links the high measured microplastic abundance to high population density in the study area.

Substantial variability was observed spatially for microplastic abundance across the study area. A previous study investigating the density of plastic trash in the Sonoran Desert, AZ states the occurrence of passively-dispersed plastic trash in the desert did not appear to dissipate with distance from potential sources of origin and were independent of road proximity (Zyflstra, 2013). Prior work suggests a majority of the microplastics in the environment originate from secondary degradation of larger plastics (Kasmurki et al., 2022). As such, it is not surprising that higher microplastic abundances were observed in relatively remote areas in our study as environmental transport and degradation can lead to secondary microplastics. An additional source such as atmospheric deposition of microplastics could be a contributing factor for microplastics in soil.

Three previous studies addressed the spatial variation of soil contaminants within the CAP-LTER area (Marusenko et al., 2011; Zhuo et al., 2012; Hamilton and Hartnett, 2013). PAHs were reported to be present in surface soils along major highways in the Phoenix, AZ metropolitan area in a previous study (Marusenko et al., 2011). A previous study reporting soot black carbon concentrations did not observe a strong correlation between soot black carbon concentrations and distance to the urban core (Hamilton and Hartnett, 2013). A study in 2012 investigated the distribution of trace elements in soil samples as part of the Central Arizona-Phoenix Long-Term Ecological Research (CAP-LTER) from the Ecological Survey of Central Arizona (ESCA) 200-point survey in 2005. A majority of trace elements depicted higher concentrations above their crustal averages. The author attributes the accumulation of elements to other additional sources such as atmospheric deposition (Zhuo et al., 2012). Consistent with these other studies, the measured microplastics in the current study also show no clear pattern relative to land use, anthropogenic activities or proximity to urban core.

A recent study investigating the spatial distribution of soil microplastics in Hainan Island, China reports a heterogeneous distribution with high variability between sampling sites (Khan et al., 2023). Also, a previous study investigating the microplastics distribution in a Central Asian Desert (an enclosed area with almost no human activity) attributes the source of microplastics to atmospheric transport and airborne microplastics deposition (Wang et al., 2021). The distribution of soil microplastics in arid and semi-arid environments in Fars province, south-central Iran, was reported to be heterogeneous between fields and sampling locations with no clear trend (Rezaei et al., 2022). This study suggests that the region experiences significant dust storms, which is a similar occurrence in the region of interest in our study (Eagar et al., 2017). Eagar and coworkers estimated that haboob dust storms alone might account for 75 % of the dust deposition locally (Eagar et al., 2017). Wind-blown dust is an important factor for the transport of anthropogenic contaminants such as microplastics in arid and semi-arid regions (Ashrafi et al., 2014). A recent study in the metropolis of Shiraz, Iran estimated that a majority of microplastics (>90 %) in deposited dust were derived from outside of the city during an intense dust storm event in May 2018 (Abbasfi et al., 2022). This suggests that dust storms could affect the local site-specific soil microplastic abundances through both deposition and resuspension, and hence environmental processes would redistribute microplastics throughout the region.

### 3.1.1. Deposition fluxes of microplastics

The deposition fluxes (collected as 2-week averages) of microplastics in Tempe, AZ for a one-year period ranged from 71 to 389 microplastics/

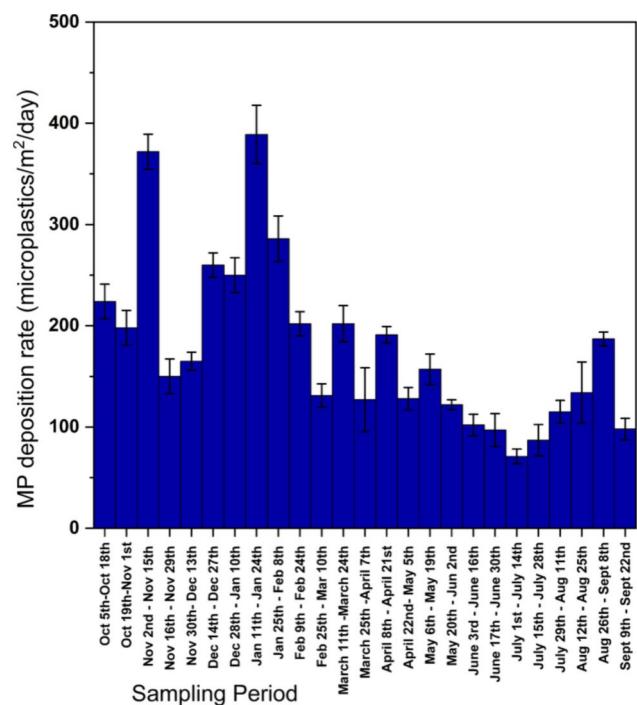


Fig. 2. The deposition flux of microplastics (MP) in Tempe, AZ. The error bars represent the standard deviation of the mean obtained from three replicate measurements.

$\text{m}^2/\text{day}$  with an average deposition flux of 178 microplastics/ $\text{m}^2/\text{day}$  (Fig. 2). Microplastics were identified in all deposition samples collected through the one-year period. Higher deposition fluxes were observed during (Nov 2nd- Nov 15th, 2020) and (Jan 11th-Jan 24th, 2021) compared to other biweekly sampling periods. Rainfall was not observed and recorded from the weather station located at the sampling site (Earth Networks Inc., an AEM company, <https://www.earthnetworks.com/>) during the aforementioned two bi-weekly periods. Since Tempe, AZ received very limited rainfall during the sampling periods, it was not possible to perform a correlation analysis between microplastics deposition fluxes and precipitation events. The lowest deposition flux was observed during (July 1st - July 14th, 2021). Wind-rose diagrams did not show any correlation between microplastics deposition fluxes and wind speed/direction. A recent study investigating suspended airborne microplastics in Tempe, AZ during the same period reports a similar observation with no clear seasonal and meteorological influence on microplastics concentrations and is consistent with the observations in the current study (Chandranthan et al., 2023). Deposition fluxes measured in Tempe show substantially higher, but variable, rates and may not directly govern the distribution of soil microplastic abundances. Wind-blown dust storms common in arid regions (Eagar et al., 2017; Abbasfi et al., 2022) result in re-entrainment and further degradation of microplastics captured by deposition studies. These types of processes likely impact incorporation of microplastics into local soil deposits and may explain discrepancies between prior deposition studies and the current measurements.

The deposition flux in the current study is approximately three times higher when compared to that reported in a recent study at a Southern China metropolis (Yuan et al., 2023). This may be attributed to the differences in the flower size limits for quantification of microplastics between the two studies (Table 2). The mean deposition fluxes reported at Hamburg in Germany, Lanzhou City in China and Central London, UK are approximately 1.5, 2, and 4 times higher than that in our study (Klefin and Fischer, 2019; Wright et al., 2020; Liu et al., 2022). The authors attribute the relatively higher deposition fluxes observed in Lanzhou City, China to local sources including a local waste recycling

**Table 2**

Summary of microplastic deposition fluxes reported in previous studies.

Location	Microplastics/ m <sup>2</sup> /day	Size range/μm	References
Lanzhou City, China	354 (57–689)	50–5000	(Liu et al., 2022)
Hamburg, Germany	275 (136–512)	< 63 – < 300	(Klefin and Fischer, 2019)
Central London, UK	712 (575–1008)	20–5000	(Wright et al., 2020)
Pyrenees mountains, France	365	< 25–5000	(Affen et al., 2019)
São Paulo megacity, Brazil	123 ± 48	Dominant (100–200)	(Amato-Lourenço et al., 2022)
South Central Ontario, Canada	7 (4–9)	20–5000	(Welsch et al., 2022)
Ho Chi Minh City, Vietnam	71–917	300–5000	(Truong et al., 2021)
Hafizhu Districts, Southern China	66 ± 8 (21–109)	13–334	(Yuan et al., 2023)
Tropical sites in Malaysia	114–689	≤5–5000	(Hee et al., 2023)
Jakarta, Indonesia	15 (3–40)	358–925	(Purwiyanto et al., 2022)
Tempe, USA	178 (71–389)	5–5000	Current Study

site. The high deposition fluxes in Central London, UK is attributed to the higher density population in the study area of interest.

### 3.2. Temporal variation of microplastics

Results for the temporal variability of microplastics in soil indicate a systematic increase in the abundance of microplastics from 2005 to 2015 at all the studied sampling sites (Fig. 3). The microplastic abundances are 1.3 to 5.2 times higher in 2015 compared to 2005. Results of a Mann-Whitney Test ( $p \leq 0.05$ ) performed indicated a statistically significant difference between the microplastic abundances in 2005 and 2015 ( $p = 0.006$ ). A study investigating the temporal variation of microplastics in sediment samples from the Chuckle Sea over 5 years reports an increase in microplastics over time (Fang et al., 2022). While the temporal variations of microplastics in aquatic systems have been relatively well reported (Kobayashi et al., 2021; Munari et al., 2021; Taibot et al., 2022), studies on the temporal variation in terrestrial areas

environments are limited.

Microplastics enter the environment mostly through secondary degradation processes of large plastics (Wong et al., 2020). Therefore, increasing amounts of plastics in the environment can give rise to increasing numbers of microplastics in the environment over time. A similar observation is reported in a study that investigated the microplastic abundances in sediment samples from the Chuckle Sea over 5 years (Fang et al., 2022). The authors attributed this increase in smaller microplastics over time to the breakdown of large plastics during long-range transport. The substantial atmospheric deposition of microplastics occurring in the region could also be a contributing factor to the observed increase of microplastics after the lapse of 10 years in the current study (Section 3.1.1).

The size distributions of microplastics in soil samples from 2005 and 2015 were analyzed to compare sizes of microplastics over time. Normalized size distributions of microplastics present in all soil samples collected in 2005 and 2015 were analyzed (Fig. 4) to understand which size classes of the microplastics were most prevalent in the two years. The size distributions represent the averages of all normalized size distributions of samples (Fig. 4).

The highest count for microplastics on average was observed between 5 μm – 50 μm in soil samples from 2015. Contrary to this observation from 2015, microplastics were predominantly found between 50 μm – 100 μm on average in soil samples from 2005. This finding is consistent with microplastics entering the environment through secondary formation with various weathering processes, including physical abrasion, resulting in more, smaller microplastics over time. This can result in their degradation to even smaller sized microplastics as time progresses during long-range transport (Fang et al., 2022). A study exposing micrometer-sized polyethylene and polypropylene pellets to UV radiation and mechanical forces at temperatures typical of a dry (beach) environment report the production of smaller microplastics, of which the abundances increased with decreasing size (Song et al., 2017). Therefore, it is not surprising that significant degradation and weathering of microplastics can occur to produce smaller sized microplastics over longer time scales under relatively dry, high-temperature conditions in the Sonoran Desert. Our results on size distribution patterns are an indication of breakdown processes of microplastics that could occur over the span of ten years in this location.

Fibrous microplastics accounted for the majority (≥87 % in 2005 and ≥ 98 % in 2015) of the microplastics in soil samples. The overall increase in relatively smaller sized fibers in 2015 soil samples is likely an indication of the degradation of larger fibrous microplastics to produce smaller fibers over time. Fragments were observed as the only other morphology present and no spherical pellets and beads were observed in the samples.

### 3.3. Chemical characterization of microplastics in soil

Micro-Raman spectroscopy was used to identify the chemical composition of polymers contained in the microplastics. Raman characterization for microplastics in 2005 and 2015 soil samples revealed an array of polymers including PE, PS, PVC, Polyamide (PA), Polyester (PES), PP depicting a variety of polymers. A majority of microplastics remain chemically unidentified (Fig. 5 and Fig. S4). This observation is consistent with the chemical characterization results of airborne microplastics in Tempe, AZ reported in a recent study (Chandranthan et al., 2023). The aforementioned study conducted lab experiments and postulated that weathering of microplastics over time could potentially alter surface properties thereby rendering them undetectable using μ-Raman spectroscopy. Polyethylene was present in 75 % of the sampling sites and was the most abundantly identified polymer on average in all soil samples. A previous study investigating the density of plastic trash in the Sonoran Desert, AZ during 2005–2006, states polyethylene bags were found in substantial numbers during field crew surveys (Zylstra, 2013). This is reflective of the high overall global production of

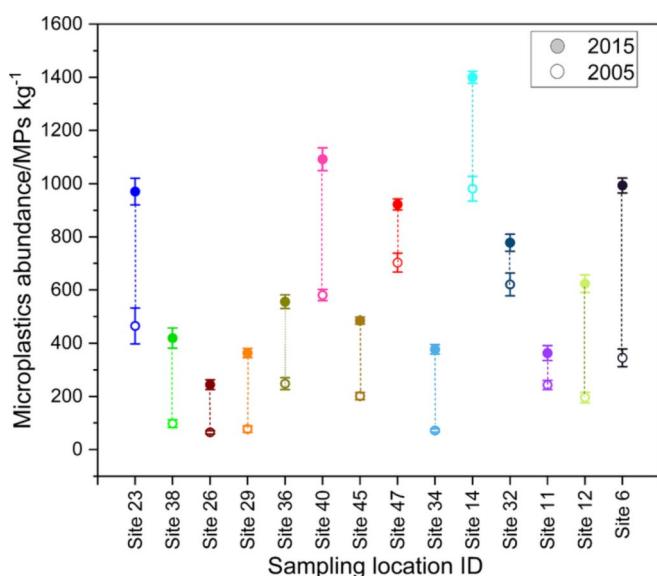


Fig. 3. Temporal changes in microplastics in soil samples from 2005 and 2015. Each error bar represents the standard deviation of the mean obtained from three replicate measurements. Sampling location IDs were randomly assigned.

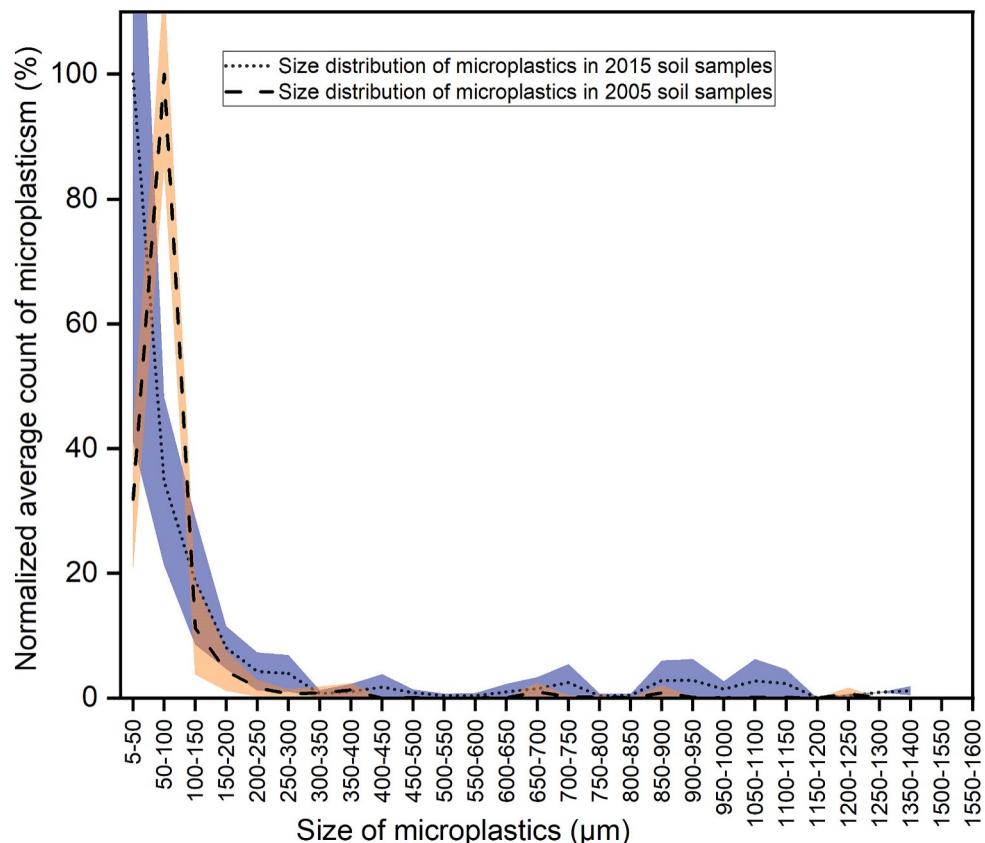


Fig. 4. Normalized percent size distributions of microplastics in soil samples from 2005 and 2015. Shaded error bars represent the standard deviation of the mean.

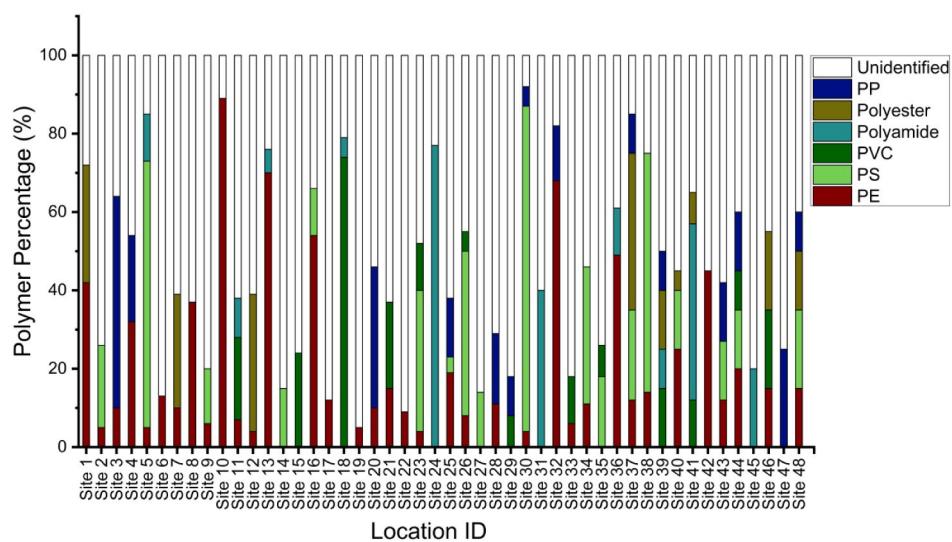


Fig. 5. Raman Characterization of microplastics of soil samples from 2015.

polyethylene which is the most produced thermoplastic worldwide (Nerland et al., 2014.; Zhong et al., 2018) and mismanagement of plastic bags in landfills and waste streams is widely reported. Previous studies reporting the occurrence and chemical characterization of microplastics in terrestrial environments also showed polyethylene as one of the most prevalent polymers found in samples (Scheurer and Biegalek, 2018; Ding et al., 2021; Khan et al., 2023). The types of identified polymers have not changed from 2005 to 2015 for soil microplastics. Additionally, there is no substantial increase in the abundance of each identified polymer from 2005 to 2015. This could possibly be due to the large fractions of

microplastics that are chemically unidentifiable in soil samples. Weathering of microplastics by complex degradation processes in the environment can cause significant alteration rendering them chemically unidentifiable (Chandranthan et al., 2023). A large majority of microplastics were chemically unidentifiable in all sites, ranging from 8% to as high as 95% of total microplastics with an average of 54% in 2015 soil samples. The average of chemically unidentified microplastics was even higher in 2005 (67%).

The observed differences for abundances of each polymer in soil samples may possibly be due to the differences in the inherent

resistances to weathering mechanisms in the environment (Abbasfi et al., 2021). Mechanical and oxidative weathering can degrade less strong polymeric materials with relatively lower tensile strengths. Accordingly, PP that has a relatively lower tensile strength of 40 MPa was observed in only 29% (in 2015) and 21% (in 2005) of the sampling sites. Polyamide with a relatively higher tensile strength (70 MPa) was dominant in Site 24, mechanically amounting to as high as 77% of the total microplastics in the soil sample.

A recent study deployed sediment traps in an arid region in Sarakhs, Iran to investigate the enrichment of microplastics in sediments at different heights from the ground (Abbasfi et al., 2023). The chemical characterization results of the aforementioned study are consistent with the findings of the current study where PE (a relatively low-density polymer;  $< 1 \text{ g cm}^{-3}$ ) was the most abundant polymer present in samples with fibers as the most dominant shape of microplastics.

#### 4. Conclusions

Microplastics were ubiquitous in the soils of Phoenix and the surrounding areas of the Sonoran Desert. In 2015 soil samples, microplastics concentrations varied an order of magnitude with a relatively heterogeneous distribution with no clear spatial trends. The results for microplastics deposition fluxes show substantial microplastics deposition occurring in Tempe, AZ and this route may influence the unclear spatial trend for the abundance of soil microplastics as local dust storms could even continually redistribute microplastics from the surface.

The Ecological Survey of Central Arizona (ESCA) performed by the Central Arizona–Phoenix Long-Term Ecological Research (CAP-LTER) followed for a temporal comparison between 2005 and 2015. At the same sampling locations over a 10-year span, all samples showed a statistically significant increase in microplastics indicative of the increasing amounts of microplastic accumulation in the environment. The deposition fluxes can account for part of this and could be enhanced by the degradation of larger plastics into smaller microplastics over time. The size class depicting the average highest count for microplastics was noticeably smaller in soil samples from 2015 while microplastics measured by our protocol were concentrated in relatively larger sizes on average in soil samples from 2005. This implies that secondary degradation processes of microplastics can be a larger contributive factor towards the temporal increase in microplastics. Raman characterization for microplastics in 2005 and 2015 soil samples revealed an array of polymers including PE, PS, PVC, PA, PES and PP. A large majority of the microplastics remain chemically unidentified. Weathering of microplastics over time could potentially change them thereby rendering them unidentified using  $\mu$ -Raman spectroscopy. Polyethylene was dominantly present in a majority of the sampling sites and was the most abundantly identified polymer on average in all soil samples indicative of the large production of polyethylene on a global scale.

#### CRediT authorship contribution statement

Kanchana Chandranthan: Conceptualization, Methodology, Data curation, Formal analysis, Funding acquisition, Writing original draft, Writing – review & editing.

Matthew Fraser: Supervision, Writing – review & editing.

Pierre Herkes: Writing – review and editing, Validation, Supervision, Project administration.

#### Declaration of competing interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

#### Data availability

Data will be made available on request.

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#### Appendix A. Supplementary data

Supplementary data to this article can be found online at <https://doi.org/10.1016/j.scitotenv.2023.167617>.

#### References

Abbasfi, S., Turner, A., Hoseini, M., Amiri, H., 2021. Microplastics in the Lut and Kavir Deserts, Iran. *Environ. Sci. Technol.* 55, 5993–6000. <https://doi.org/10.1021/acs.est.1c00615>.

Abbasfi, S., Rezaei, M., Ahmad, F., Turner, A., 2022. Atmospheric transport of microplastics during a dust storm. *Chemosphere* 292. <https://doi.org/10.1016/j.chemosphere.2021.133456>.

Abbasfi, S., Rezaei, M., Mina, M., Sameni, A., Ofleszczuk, P., Turner, A., Ritzema, C., 2023. Enrichment and horizontal atmospheric transport of microplastics from soil. *Chemosphere* 322. <https://doi.org/10.1016/j.chemosphere.2023.138150>.

Afflen, S., Afflen, D., Phoenix, V.R., Le Roux, G., Durández Jiménez, P., Sfimonneau, A., Bfinet, S., Gaflop, D., 2019. Atmospheric transport and deposition of microplastics in a remote mountain catchment. *Nat. Geosci.* 12, 339–344. <https://doi.org/10.1038/s41561-019-0335-5>.

Amato-Lourenço, L.F., dos Santos Gaflão, L., Wfiebeck, H., Carvalho-Oliveira, R., Mauad, T., 2022. Atmospheric microplastic fallout in outdoor and indoor environments in São Paulo megacity. *Sci. Total Environ.* 821. <https://doi.org/10.1016/j.scitotenv.2022.153450>.

Andrade, A.L., 2011. Microplastics in the marine environment. *Mar. Pollut. Bull.* 62, 1596–1605. <https://doi.org/10.1016/j.marpolbul.2011.05.030>.

Arthur, C., Baker, J., Bamford, H. (Eds.), 2009. Proceedings of the International Research Workshop on the Occurrence, Effects and Fate of Microplastic Marine Debris. Sept 9–11, 2008. NOAA Technical Memorandum NOS-OR&R-30.

Ashrafi, K., Shaffiour-Motlagh, M., Aslmand, A., Ghaferi, S., 2014. Dust storm simulation over Iran using HYSPLIT. *Iran. J. Environ. Health Sci. Eng.* 12, 9. <https://doi.org/10.1186/2052-336X-12-9>.

Bergmann, M., Wfizberger, V., Krumpen, T., Lorenz, C., Primpke, S., Tekman, M.B., Gerdtz, G., 2017. High quantities of microplastic in Arctic Deep-Sea sediments from the HAUSGARTEN observatory. *Environ. Sci. Technol.* 51, 11000–11010. <https://doi.org/10.1021/acs.est.7b03331>.

Bohn, H.L., McNeafl, B.L., O'Connor, G.A., 2001. *Soil Chemistry*. John Wiley & Sons.

Cachier, H., Buatmenard, P., Fontugne, M., Rancher, J., 1985. Source Terms and Source Strengths of the Carbonaceous Aerosol in the Tropics. *Journal of Atmospheric Chemistry*, 3, pp. 469–489.

Chandranthan, K., Fraser, M.P., Herkes, P., 2023. Airborne microplastics in a suburban location in the desert southwest: occurrence and identification challenges. *Atmos. Environ.* 298. <https://doi.org/10.1016/j.atmosenv.2023.119617>.

De Souza Machado, A.A., Lau, C.W., Kflos, W., Bergmann, J., Bacheflier, J.B., Pafit, E., Becker, R., Gförl, A.S., Ritzema, M.C., 2019. Microplastics can change soil properties and affect plant performance. *Environ. Sci. Technol.* 53, 6044–6052. <https://doi.org/10.1021/acs.est.9b01339>.

Ding, L., Zhang, S., Wang, X., Yang, X., Zhang, C., Qf, Y., Guo, X., 2020. The occurrence and distribution characteristics of microplastics in the agricultural soils of Shaanxi Province, in north-western China. *Sci. Total Environ.* 720. <https://doi.org/10.1016/j.scitotenv.2020.137525>.

Ding, L., Wang, Xfingf, Ouyang, Z., Chen, Y., Wang, Xuexia, Lf, D., Lf, S., Yang, X., Jia, H., Guo, X., 2021. The occurrence of microplastic in Mu Us Sand Land soils in northwest China: different soil types, vegetation cover and restoration years. *J. Hazard. Mater.* 403. <https://doi.org/10.1016/j.jhazmat.2020.123982>.

Ding, L., Huang, D., Ouyang, Z., Guo, X., 2022. The effects of microplastics on soil ecosystem: A review. *Curr. Opin. Environ. Sci. Health.* <https://doi.org/10.1016/j.coesh.2022.100344>.

Eagar, J.D., Herkes, P., Hartnett, H.E., 2017. The characterization of haboobs and the deposition of dust in Tempe, AZ from 2005 to 2014. *Aeolian Res.* 24, 81–91. <https://doi.org/10.1016/j.aeolia.2016.11.004>.

Erliksen, M., Mason, S., Wfifson, S., Box, Z., Zeffler, A., Edwards, W., Farley, H., Amato, S., 2013. Microplastic pollution in the surface waters of the Laurentian Great Lakes. *Mar. Pollut. Bull.* 77, 177–182. <https://doi.org/10.1016/j.marpolbul.2013.10.007>.

Fang, C., Zhang, Y., Zheng, R., Hong, F., Zhang, M., Zhang, R., Mou, J., Mu, J., Lfin, L., Bo, J., 2022. Spatio-temporal variation of microplastic pollution in the sediment from the Chukchi Sea over five years. *Sci. Total Environ.* 806, 150530 <https://doi.org/10.1016/j.scitotenv.2021.150530>.

Feng, S., Lu, H., Tfian, P., Xue, Y., Lu, J., Tang, M., Feng, W., 2020. Analysis of microplastics in a remote region of the Tibetan Plateau: implications for natural environmental response to human activities. *Sci. Total Environ.* 739, 140087 <https://doi.org/10.1016/j.scitotenv.2020.140087>.

Grimm, N.B., Redman, C.L., 2004. Approaches to the study of urban ecosystems: the case of Central Arizona—Phoenix. *Urban Ecosyst.* 7, 199–213. <https://doi.org/10.1023/b:ueco.0000044036.59953.a1>.

Grimm, Redman, C.L., Boone, C.G., Chiflders, D.L., Harlan, S.L., Turner, B.L., 2013. Viewing the Urban Socio-ecological System Through a Sustainability Lens: Lessons and Prospects from the Central Arizona-Phoenix LTER Programme. In: Long Term Socio-Ecological Research. Springer Netherlands, 2, pp. 217–246. [https://doi.org/10.1007/978-94-007-1177-8\\_10](https://doi.org/10.1007/978-94-007-1177-8_10).

Hamilton, G.A., Hartnett, H.E., 2013. Soot black carbon concentration and isotopic composition in soils from an arid urban ecosystem. *Org. Geochem.* 59, 87–94. <https://doi.org/10.1016/j.orggeochem.2013.04.003>.

Harms, I.K., Diefkötter, T., Troegel, S., Lenz, M., 2021. Amount, distribution and composition of large microplastics in typical agricultural soils in Northern Germany. *Sci. Total Environ.* 758 <https://doi.org/10.1016/j.scitotenv.2020.143615>.

Hee, Y.Y., Hanif, N.M., Weston, K., Latif, M.T., Suratman, S., Rusli, M.U., Mayes, A.G., 2023. Atmospheric microplastic transport and deposition to urban and pristine tropical locations in Southeast Asia. *Sci. Total Environ.* 902 <https://doi.org/10.1016/j.scitotenv.2023.166153>.

Hidalgo-Ruz, V., Gutow, L., Thompson, R.C., Thiel, M., 2012. Microplastics in the marine environment: a review of the methods used for identification and quantification. *Environ. Sci. Technol.* 46, 3060–3075. <https://doi.org/10.1021/es2031505>.

Hilfdebrandt, L., Nack, F.L., Zimmermann, T., Pröfrock, D., 2021. Microplastics as a Trojan horse for trace metals. *J. Hazard. Mater.* 2, 100035 <https://doi.org/10.1016/j.hazmat.2021.100035>.

Horton, A.A., Walton, A., Spurgeon, D.J., Lahive, E., Svendsen, C., 2017. Microplastics in freshwater and terrestrial environments: evaluating the current understanding to identify the knowledge gaps and future research priorities. *Sci. Total Environ.* 586, 127–141. <https://doi.org/10.1016/j.scitotenv.2017.01.190>.

Hu, J., He, D., Zhang, X., Ifi, X., Chen, Y., Wefi, G., Zhang, Y., Ok, Y.S., Luo, Y., 2022a. National-scale distribution of micro(mesop)plastics in farmland soils across China: implications for environmental impacts. *J. Hazard. Mater.* 424 <https://doi.org/10.1016/j.jhazmat.2021.127283>.

Hu, X., Yu, Q., Gatheru, Wafiqi, M., Lfing, W., Qfin, C., Wang, J., Gao, Y., 2022b. Microplastics-sorbed phenanthrene and its derivatives are highly bioaccessible and may induce human cancer risks. *Environ. Int.* 168 <https://doi.org/10.1016/j.envint.2022.107459>.

Huang, H., Mohamed, B.A., Ifi, L.Y., 2023. Accumulation and fate of microplastics in soils after application of biosolids on land: a review. *Environ. Chem. Lett.* <https://doi.org/10.1007/s10311-023-01577-3>.

Kasmurri, N., Tamizzi, N.A.A., Mojtiri, A., 2022. Occurrence, impact, toxicity, and degradation methods of microplastics in environment—a review. *Environ. Sci. Pollut. Res.* <https://doi.org/10.1007/s11356-021-18268-7>.

Khan, M.A., Huang, Q., Khan, S., Wang, Q., Huang, J., Fahad, S., Sajjad, M., Ifin, Y., Mašek, O., Ifi, X., Wang, J., Song, X., 2023. Abundance, spatial distribution, and characteristics of microplastics in agricultural soils and their relationship with contributing factors. *J. Environ. Manag.* 328, 117006 <https://doi.org/10.1016/j.jenvman.2022.117006>.

Kfin, S.W., Walldman, W.R., Kfin, T.Y., Riffing, M.C., 2020. Effects of different microplastics on nematodes in the soil environment: tracking the extractable additives using an ecotoxicological approach. *Environ. Sci. Technol.* 54, 13868–13878. <https://doi.org/10.1016/acs.est.0c04641>.

Klefin, M., Fischer, E.K., 2019. Microplastic abundance in atmospheric deposition within the Metropolitan area of Hamburg, Germany. *Sci. Total Environ.* 685, 96–103. <https://doi.org/10.1016/j.scitotenv.2019.05.405>.

Kobayashi, T., Yagi, M., Kawaguchi, T., Hata, T., Shifmizu, K., 2021. Spatiotemporal variations of surface water microplastics near Kyushu, Japan: a qualitative-quantitative analysis. *Mar. Pollut. Bull.* 169, 112563 <https://doi.org/10.1016/j.marpolbul.2021.112563>.

Lebreton, L., Royer, S.J., Peaytav, A., Strietman, W.J., Smedding-Zuurendonk, I., Egger, M., 2022. Industrialised fishing nations largely contribute to floating plastic pollution in the North Pacific subtropical gyre. *Sci. Rep.* 12 <https://doi.org/10.1038/s41598-022-16529-0>.

Ifi, W., Wang, S., Wufuer, R., Duo, J., Pan, X., 2022. Microplastic contamination in urban, farmland and desert environments along a highway in Southern Xinjiang, China. *Int. J. Environ. Res. Public Health* 19. <https://doi.org/10.3390/jipher19158890>.

Ifi, Y., Lu, Q., Xing, Y., Ifin, Kafi, Lfing, W., Yang, J., Yang, Q., Wu, T., Zhang, J., Pefi, Z., Gao, Z., Ifi, X., Yang, F., Ma, H., Ifin, Kehan, Zhao, D., 2023. Review of research on migration, distribution, biological effects, and analytical methods of microfibers in the environment. *Sci. Total Environ.* 855, 158922 <https://doi.org/10.1016/j.scitotenv.2022.158922>.

Liang, Y., Cao, X., Mo, A., Jifang, J., Zhang, Y., Gao, W., He, D., 2023. Micro(nano)plastics in plant-derived food: source, contamination pathways and human exposure risks. *TrAC Trends Anal. Chem.* 165, 117138 <https://doi.org/10.1016/j.trac.2023.117138>.

Liu, M., Lu, S., Song, Y., Lefi, L., Hu, J., Lv, W., Zhou, W., Cao, C., Shfi, H., Yang, X., He, D., 2018. Microplastic and mesoplastic pollution in farmland soils in suburbs of Shanghai, China. *Environ. Pollut.* 242, 855–862. <https://doi.org/10.1016/j.envpol.2018.07.051>.

Ifin, Z., Bafi, Y., Ma, T., Ifin, X., Wefi, H., Meng, H., Fu, Y., Ma, Z., Zhang, L., Zhao, J., 2022. Distribution and possible sources of atmospheric microplastic deposition in a valley basin city (Lanzhou, China). *Ecotoxicol. Environ. Saf.* 233 <https://doi.org/10.1016/j.ecoenv.2022.113353>.

Marusenko, Y., Herckes, P., Haffl, S.J., 2011. Distribution of polycyclic aromatic hydrocarbons in soils of an arid urban ecosystem. *Water Air Soil Pollut.* 219, 473–487. <https://doi.org/10.1007/s11270-010-0721-5>.

Masura, J., Baker, J., Foster, G., Arthur, C., 2015. Laboratory methods for the analysis of microplastics in the marine environment: recommendations for quantifying synthetic particles in waters and sediments. *Springer* 31pp. (NOAA Technical Memorandum NOS-OR&R-48).

Mohan, N., Montazer, Z., Sharma, P.K., Levitt, D.B., 2020. Microbial and enzymatic degradation of synthetic plastics. *Front. Microbiol.* 11, 580709 <https://doi.org/10.3389/fmicb.2020.580709>.

Moflsh, J.N., Loder, M.G.J., Laforsch, C., 2020. Finding microplastics in soils: a review of analytical methods. *Environ. Sci. Technol.* 54, 2078–2090. <https://doi.org/10.1021/acs.est.9b04618>.

Munari, C., Scoponi, M., Sfriso, A.A., Sfriso, A., Aifelio, J., Casoni, E., Mifistri, M., 2021. Temporal variation of floatable plastic particles in the largest Italian river, the Po. *Mar. Pollut. Bull.* 171, 112805 <https://doi.org/10.1016/j.marpolbul.2021.112805>.

Neerland, I., Hafslund, C., Aifan, I., Thomas, K., 2014. Microplastics in Marine Environments: Occurrence, Distribution and Effects. *Tech. Rep.* 6754–2014. Norwegian Institute for Water Research, Oslo.

Ng, E.L., Huerta Lwanga, E., Eiffridge, S.M., Johnston, P., Hu, H.W., Gefissen, V., Chen, D., 2018. An overview of microplastic and nanoplastic pollution in agroecosystems. *Sci. Total Environ.* 627, 1377–1388. <https://doi.org/10.1016/j.scitotenv.2018.01.341>.

Nueifl, M.T., Dekfif, J.H., Remy, D., Fries, E., 2014. A new analytical approach for monitoring microplastics in marine sediments. *Environ. Pollut.* 184, 161–169. <https://doi.org/10.1016/j.envpol.2013.07.027>.

Patifl, P.B., Mafit, S., Sarkar, A., 2022. Potential human health risk assessment of microplastic exposure: current scenario and future perspectives. *Environ. Monit. Assess.* 194 <https://doi.org/10.1007/s10661-022-10539-1>.

Prata, J.C., Refis, V., da Costa, J.P., Mouneyrac, C., Duarte, A.C., Rocha-Santos, T., 2021. Contamination issues as a challenge in quality control and quality assurance in microplastics analytics. *J. Hazard. Mater.* 403 <https://doi.org/10.1016/j.jhazmat.2020.123660>.

Purwiyanto, A.I.S., Prartono, T., Rfiani, E., Naufitla, Y., Cordova, M.R., Koropitan, A.F., 2022. The deposition of atmospheric microplastics in Jakarta-Indonesia: the coastal urban area. *Mar. Pollut. Bull.* 174 <https://doi.org/10.1016/j.marpolbul.2021.113195>.

Raffique, A., Irfan, M., Mumtaz, M., Qadri, A., 2020. Spatial distribution of microplastics in soil with context to human activities: a case study from the urban center. *Environ. Monit. Assess.* 192 <https://doi.org/10.1007/s10661-020-08641-3>.

Revelifl, L.E., Kuma, P., Le, R., E.C., Somervilfle, W.R.C., Gao, S., 2021. Direct radiative effects of airborne microplastics. *Nature* 598, 462–467. <https://doi.org/10.1038/s41586-021-03864-x>.

Rezaei, M., Abbasi, S., Pourmahmood, H., Ofleszczuk, P., Rfisema, C., Turner, A., 2022. Microplastics in agricultural soils from a semi-arid region and their transport by wind erosion. *Environ. Res.* 212 <https://doi.org/10.1016/j.envres.2022.113213>.

Scheurer, M., Bifagl, M., 2018. Microplastics in Swfis floodplains soils. *Environ. Sci. Technol.* 52, 3591–3598. <https://doi.org/10.1021/acs.est.7b06003>.

Shanmugam, S.D., Praveena, S.M., Sarkar, B., 2022. Quality assessment of research studies on microplastics in soils: a methodological perspective. *Chemosphere* 296. <https://doi.org/10.1016/j.chemosphere.2022.134026>.

Song, Y.K., Hong, S.H., Jang, M., Han, G.M., Jung, S.W., Shifm, W.J., 2017. Combined effects of UV exposure duration and mechanical abrasion on microplastic fragmentation by polymer type. *Environ. Sci. Technol.* 51, 4368–4376. <https://doi.org/10.1016/acs.est.6b06155>.

Talbot, R., Granek, E., Chang, H., Wood, R., Brander, S., 2022. Spatial and temporal variations of microplastic concentrations in Portland's freshwater ecosystems. *Sci. Total Environ.* 833, 155143 <https://doi.org/10.1016/j.scitotenv.2022.155143>.

Tuong, T.N.S., Strady, E., Kieu-Le, T.C., Tran, Q.V., Le, T.M.T., Thuong, Q.T., 2021. Microplastic in atmospheric fallouts of a developing Southeast Asian megacity under tropical climate. *Chemosphere* 272. <https://doi.org/10.1016/j.chemosphere.2021.129874>.

Wang, F., Lefi, Z., Peng, G., Luo, L., Ifin, K., Huang, X., Xu, Y., Shen, Q., Ifi, D., 2021. Microplastic abundance and distribution in a Central Asian desert. *Sci. Total Environ.* 800, 149529 <https://doi.org/10.1016/j.scitotenv.2021.149529>.

Weflsh, B., Aherne, J., Paterson, A.M., Yao, H., McConnefl, C., 2022. Atmospheric deposition of anthropogenic particles and microplastics in south-central Ontario, Canada. *Sci. Total Environ.* 835, 155426 <https://doi.org/10.1016/j.scitotenv.2022.155426>.

Wong, J.K.H., Lee, K.K., Tang, K.H.D., Yap, P.S., 2020. Microplastics in the freshwater and terrestrial environments: prevalence, fates, impacts and sustainable solutions. *Sci. Total Environ.* 719, 137512 <https://doi.org/10.1016/j.scitotenv.2020.137512>.

Wrigh, S.L., Ufke, J., Font, A., Chan, K.L.A., Kefly, F.J., 2020. Atmospheric microplastic deposition in an urban environment and an evaluation of transport. *Environ. Int.* 136, 105411 <https://doi.org/10.1016/j.envint.2019.105411>.

Yuan, Z., Pefi, C., Ifi, H., Ifin, L., Ifin, S., Hou, R., Lfiao, R., Xu, X., 2023. Atmospheric microplastics at a southern China metropolis: occurrence, deposition flux, exposure risk and washout effect of rainfall. *Sci. Total Environ.* 869, 161839 <https://doi.org/10.1016/j.scitotenv.2023.161839>.

Zhang, G.S., Ifin, Y.F., 2018. The distribution of microplastics in soil aggregate fractions in southwestern China. *Sci. Total Environ.* 642, 12–20. <https://doi.org/10.1016/j.scitotenv.2018.06.004>.

Zhang, Q., Xu, E.G., Ifi, J., Chen, Q., Ma, L., Zeng, E.Y., Shfi, H., 2020. A review of microplastics in table salt, drinking water, and air: direct human exposure. *Environ. Sci. Technol.* 54, 3740–3751. <https://doi.org/10.1021/acs.est.9b04535>.

Zhang, Y., Zhang, X., Ifi, X., He, D., 2022. Interaction of microplastics and soil antimicrobials in agricultural ecosystems. In: Current Opinion in Environmental Science and Health, 26, p. 100327. <https://doi.org/10.1016/j.coesh.2022.100327>.

Zhao, X., Zhou, Y., Liang, C., Song, J., Yu, S., Liao, G., Zou, P., Tang, K.H.D., Wu, C., 2023. Airborne microplastics: occurrence, sources, fate, risks and mitigation. *Sci. Total Environ.* 858, 159943 <https://doi.org/10.1016/j.scitotenv.2022.159943>.

Zhong, X., Zhao, X., Qian, Y., Zou, Y., 2018. Polyethylene plastic production process. *Insight Mater. Sci.* 1, 1–11. <https://doi.org/10.18282/fims.v1f1.104>.

Zhuo, X., Boone, C.G., Shock, E.L., 2012. Soil lead distribution and environmental justice in the Phoenix metropolitan region. *Environ. Justice* 5, 206–213. <https://doi.org/10.1089/env.2011.0041>.

Zyflstra, E.R., 2013. Accumulation of wind-dispersed trash in desert environments. *J. Arid Environ.* 89, 13–15. <https://doi.org/10.1016/j.jaridenv.2012.10.004>.