

1 **Title Page**

2 Seasonal growth and senescence of seagrass alters sediment accumulation rates and carbon burial
3 in a coastal lagoon

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8 Running head: seasonal sediment accumulation in seagrass beds

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10 Keywords: sediment transport, carbon burial, seasonal variation, flow-vegetation interactions,
11 numerical modeling, seagrass meadows

12 **Abstract**

13 Seagrass meadows are important carbon sinks in the global coastal carbon cycle yet are
14 also among the most rapidly declining marine habitats. Their ability to sequester carbon depends
15 on flow-sediment-vegetation interactions that facilitate net deposition, as well as high rates of
16 primary production. However, the effects of seasonal and episodic variations in seagrass density
17 on net sediment and carbon accumulation have not been well quantified. Understanding these
18 dynamics provides insight into how carbon accumulation in seagrass meadows responds to
19 disturbance events and climate change. Here we apply a spatially resolved sediment transport
20 model that includes coupling of seagrass effects on flow, waves, and sediment resuspension in a
21 seagrass meadow to quantify seasonal rates of sediment and carbon accumulation in the meadow.
22 Our results show that organic carbon accumulation rates were largely determined by sediment
23 accumulation and that they both changed non-linearly as a function of seagrass shoot density.
24 While seagrass meadows effectively trapped sediment at meadow edges during spring-summer
25 growth seasons, during winter senescence low-density meadows (< 160 shoots m^{-2}) were
26 erosional with rates sensitive to density. Small variations in winter densities resulted in large
27 changes in annual sediment and carbon accumulation in the meadow; meadow-scale (hundreds
28 of square meters) summer seagrass dieback due to marine heatwaves can result in annual erosion
29 and carbon loss. Our findings highlight the strong temporal and spatial variability in sediment
30 accumulation within seagrass meadows and the implications for annual sediment carbon burial
31 rates and the resilience of seagrass carbon stocks under future climate change.

32

33 **Introduction**

34 Seagrass meadows are essential coastal habitats that offer valuable ecosystem services
35 including carbon sequestration, nutrient cycling and improved water quality, and sediment
36 stabilization (de Boer 2007; McGlathery et al. 2007). They have been recognized as important
37 carbon sinks in the global marine carbon cycle due to high rates of net primary production and
38 carbon burial in the sediment (Duarte et al. 2013), and contribute to more than 10% of the annual
39 sediment carbon burial in global oceans (Fourqurean et al. 2012). Despite their importance in
40 coastal ecosystems, seagrasses are one of the most rapidly declining marine habitats, threatened
41 by degraded water quality, temperature stress, and sea level rise (Orth et al. 2006; Waycott et al.
42 2009). The disappearance of seagrass can cause seabed erosion and the exposure of accumulated
43 sediment carbon to oxic conditions, leading to carbon emissions (Pendleton et al. 2012; Aoki et
44 al. 2021). For seagrass meadows to be considered as effective carbon sinks, the accumulated
45 carbon must be preserved in sediments for a long period (e.g., decades to centuries).

46 An accurate estimate of sediment accumulation rates within seagrass meadows is critical
47 for quantifying sediment budgets and organic carbon burial. Various methods have been used to
48 determine sediment accumulation rates in seagrass meadows worldwide, each focusing on
49 different temporal scales and spatial extents (Potouroglou et al. 2017). Repeated bathymetric
50 surveys can provide a direct measure of long-term changes in deposition at large spatial scales
51 (Walter et al. 2020), but lack the precision to capture most short-term change (e.g., annual or
52 seasonal scale). Radiometric dating methods are widely used in seagrass studies to determine
53 sedimentation rates over several decades (Duarte et al. 2013), but they have been criticized for
54 not accounting for surface mixing effects (Johannessen and Macdonald 2016). Sediment
55 traps/plates are able to measure sediment deposition over a period of several days to months, but

56 they cannot fully resolve erosion/resuspension processes on the seabed and the results are
57 susceptible to sediment loss during retrieval (Gacia and Duarte 2001; Nolte et al. 2013). Surface
58 elevation tables can provide precise short-term and long-term measurements, but are limited in
59 spatial scale and may cause local scouring of the seabed (Potouroglou et al. 2017).

60 Long-term sediment accumulation rates within seagrass meadows have been
61 characterized at a number of sites using the aforementioned approaches (Greiner et al. 2013;
62 Duarte et al. 2013; Oreska et al. 2020), but short-term dynamics of sediment accumulation in
63 seagrass meadows remain unclear. Seagrass beds show strong temporal and spatial variability in
64 erosion and deposition in response to seasonal seagrass growth and senescence (Hansen and
65 Reidenbach 2013; Zhu et al. 2021) that is difficult to interpret from long-term sedimentary
66 records. There is also growing evidence of increasing frequency of marine heatwaves that can
67 cause seasonal summer seagrass dieback and losses of accumulated carbon in seagrass sediments
68 (Arias-Ortiz et al. 2018; Aoki et al. 2021). Because these short-term processes and disturbances
69 can strongly alter annual sediment budgets and long-term dynamics of seagrass meadows, it is
70 important to understand these seasonal sediment dynamics and the associated drivers, especially
71 in the context of future climate change.

72 An alternative approach to characterizing sediment accumulation rates in seagrass
73 meadows is through modeling. Numerical models capable of resolving the synergistic effects of
74 flow-wave-vegetation-sediment interaction (Chen et al. 2007; Carr et al. 2016; Donatelli et al.
75 2018) provide a tool for understanding seasonal sediment dynamics in seagrass meadows in
76 spatially resolved settings. However, most previous modeling studies of sediment transport
77 within seagrass meadows have used idealized seagrass meadows or did not resolve seasonal
78 seagrass growth and wind patterns. To better resolve spatial variations of dynamic factors and to

79 understand the effects of seasonal growth of seagrass on sediment accumulation, we applied the
80 process-based hydrodynamic and sediment transport Delft3D model, including coupling of
81 seagrass effects on flow, waves, and sediment resuspension, to a seagrass meadow in a shallow
82 coastal bay in Virginia, USA. The model has been parameterized and extensively validated using
83 long-term data from the site (wind conditions, hydrodynamic and suspended sediment data,
84 sediment accumulation rates, and seagrass characteristics; Zhu et al. 2021).

85 In this study, the coupled model was run for 12 consecutive months with seasonally
86 varying winds and tides as well as seagrass densities. The results were analyzed to address three
87 questions. (1) How do rates of sediment accumulation and organic carbon burial within seagrass
88 meadows vary in response to seasonal variations in seagrass density? (2) How does short-term
89 disturbance in seagrass density affect annual sediment accumulation rates? (3) What are the
90 effects of seasonal variations in seagrass density on spatial sediment erosion/deposition patterns
91 in seagrass meadows?

92 **Study site**

93 The study site, South Bay, is one of the back-barrier bays within the Virginia Coast
94 Reserve Long-Term Ecological Research site (Fig. 1a). It is bordered by a barrier island to the
95 east and is connected to two tidal inlets that exchange water with the Atlantic Ocean (Fig. 1b).
96 South Bay has a mean depth of 1 m below mean sea level and an average semidiurnal tidal range
97 of 1.2 m (Fagherazzi and Wiberg 2009). Sea level in the study area is rising at a rate of 4-5 mm
98 yr^{-1} (https://tidesandcurrents.noaa.gov/sltrends/sltrends_station.shtml?id=8631044). South Bay is
99 a shallow, oligotrophic environment that provides favorable light conditions for seagrass growth
100 and is a successful seagrass restoration site where *Zostera marina* now dominates the subtidal
101 flats (McGlathery et al. 2012). Located at the southern geographical limit for *Zostera marina*

102 growth in the Western Atlantic Ocean (Aoki et al. 2020), the seagrass meadows in South Bay
103 show strong seasonal variability that significantly impacts bay dynamics (Hansen and
104 Reidenbach 2013; Rheuban et al. 2014). Seagrasses within the bay reach a maximum shoot
105 density (> 500 shoots m^{-2}) in early summer and suffer from a mid-season loss due to heat stress
106 in late summer; the density slightly increases in autumn after temperatures moderate and then
107 declines to a minimum (50-100 shoots m^{-2}) during winter senescence. When the temperature
108 increases in the next spring, seagrasses start to re-grow and the density gradually increases to 300
109 shoots m^{-2} in late spring (Fig. 1c; Hansen and Reidenbach 2013; Rheuban et al. 2014;
110 Reidenbach and Thomas 2018; Berger et al. 2020). The presence of high-density seagrass
111 significantly reduces sediment resuspension within seagrass meadows during summer, while
112 significant sediment resuspension occurs in winter when frequent and stronger northeasterly
113 winds coincide with minimum seagrass density (Hansen and Reidenbach 2013; Zhu et al. 2021).

114 Overall, seagrass meadows in South Bay effectively accumulate fine particles with an
115 average sediment deposition rate of 6.3 mm yr^{-1} (based on radiometric dating at two
116 representative sites; Greiner et al. 2013; Oreska et al. 2018), and resulted in a finer sediment
117 grain size (mean = 71 μm) within the meadow than outside the meadow (mean = 124 μm ;
118 Lawson et al. 2007; McGlathery et al. 2012; Oreska et al. 2017). As a result of sediment
119 accumulation, the seagrass meadow is able to bury organic carbon at an average rate of 42 g C
120 $m^{-2} yr^{-1}$ (Oreska et al. 2020).

121 **Methods**

122 Hydrodynamic and sediment transport simulations were conducted using the process-
123 based and spatially resolved Delft3D FLOW/MOR model (Lesser et al. 2004), coupled with the
124 nearshore phase-averaged wave model SWAN (Booij et al. 1999). The domain decomposition

125 technique (Deltares 2014) was used to locally refine the model grid in South Bay to better
126 capture seagrass meadows in the bay and the bordering barrier island. The model grid consisted
127 of a small model domain covering the core study area in South Bay (Fig 1b; 305 \times 302 grid cells
128 with a spatial resolution of \sim 70 m) and a large model domain spanning the rest of the Virginia
129 Coast Reserve (148 \times 444 grid cells with a spatial resolution of 200 m). The open ocean
130 boundaries of the large model domain were forced with water levels extracted from the NOAA
131 tide gauge station at Wachapreague (Site ID:8631044) after adjusting tidal amplitude and phase
132 to generate tidal simulation results in excellent agreement with measured tides at Wachapreague
133 ($R^2 \geq 0.98$ and root mean square error ≤ 0.07 m). Wave simulations were driven by hourly wind
134 conditions from the same NOAA station and coupled with the flow model every hour. Model
135 bathymetry and bottom sediment size distributions and properties were extracted from Wiberg et
136 al. (2015). Three sediment classes were used in the model: a 32-64 μm coarse silt fraction, a < 32
137 μm medium to fine silt fraction, and a $> 64 \mu\text{m}$ sand fraction (a representative median grain size
138 of 125 μm was defined in the model for the sand fraction). To better capture the effects of
139 seagrass on sediment transport, sediment size distributions in the South Bay seagrass meadows
140 were initialized based on local surveys from Oreska et al. (2017). The active sediment layer
141 thickness, which defines the maximum erosion depth of the seabed at each model time step, was
142 set to 5 cm to avoid unrealistically high sediment availability.

143 The model was implemented in depth-averaged mode with a time step of 0.25 min.
144 Several previous studies have shown that depth-averaged Delft3D simulations are able to
145 produce reasonable results for well-mixed shallow coastal bays, including the Virginia Coast
146 Reserve, and to resolve the synergistic effects of flow-wave-vegetation-sediment interactions in
147 spatially resolved settings (Nardin and Edmonds 2014; Nardin et al. 2018; Zhu et al. 2021). In

148 order to incorporate seagrass effects on currents and waves, the Baptist vegetation model (Baptist
149 et al. 2007) and the Suzuki vegetation wave energy dissipation model (Suzuki et al. 2012) were
150 implemented in the Delft3D FLOW module and the SWAN model, respectively. These two
151 methods considered vegetation as cylindrical structures characterized by vegetation height, stem
152 diameter, shoot density, and vegetation flow drag coefficient and wave drag coefficient (Table
153 S1).

154 The coupled model has been parameterized and extensively validated using summer and
155 winter hydrodynamic and suspended sediment data during a 4-day period in January and June
156 2011 from a seagrass site and a nearby unvegetated site in South Bay (Zhu et al. 2021). Model
157 skill indices, including bias, root mean square error, and Willmott skill index (Willmott 1981),
158 were calculated for model validation parameters (water level, significant wave height, depth-
159 averaged velocity, and total suspended sediment concentration) to quantify model ability to
160 characterize hydrodynamic and suspended sediment characteristics in the bay. Values of the
161 Willmott skill index are summarized in Table S2. Excellent agreement between modeled and
162 measured water levels was obtained at both sites during each validation period. Wave height skill
163 scores (0.56-0.87) for the seagrass site were generally higher than those of the unvegetated site.
164 While generally good agreement between modeled and measured depth-averaged velocity was
165 obtained at both sites, the model slightly over-estimated peak velocity during flood tides with an
166 average bias of 0.05 m s^{-1} (Zhu et al. 2021). The model successfully captured most sediment
167 resuspension events at both sites during winter (skill scores ≥ 0.80) and predicted a strong
168 reduction of suspended sediment concentration ($> 80\%$) in the summer seagrass meadow that is
169 consistent with field observations. See Zhu et al. (2021) for further details of the model
170 parameterization and validation. Overall, the coupled model is able to provide spatially-resolved

171 simulations of flow and sediment transport patterns within/outside the seagrass meadow with
172 performance similar to that of other model studies considering seagrass effects (Chen et al. 2007;
173 Moki et al. 2020).

174 In this study, we extended the simulation period considered in Zhu et al. (2021) to a
175 complete annual cycle using winds and tidal forcing for a representative year (August 1, 2011 to
176 July 31, 2012; Fig. S1), with typical seasonal seagrass characteristics (shoot density ranging from
177 100-600 shoots m^{-2} , seagrass height ranging from 0.2-0.4 m, and stem diameter ranging from
178 0.2-0.4 cm) observed from the site (Table S1). Evaluating model sensitivity to variations in
179 seagrass characteristics, Zhu et al. (2021) found that variations in seagrass shoot density had a
180 larger impact on bay dynamics than did variations in seagrass height and stem diameter.
181 Considering the wide range of annual seagrass density variation (100-600 shoots m^{-2}) and its
182 strong impact on bay dynamics at our study site, we mainly focused on the impacts of seagrass
183 density variations in this paper.

184 In order to better represent observed spatial density gradients in the meadow, three
185 seagrass density classes were assigned in the model each month, with the highest density (N) in
186 the central meadow, an intermediate density of $0.8N$ outside the central area, and the lowest
187 density of $0.6N$ near meadow edges (Fig. 1b). A reference model case without seagrass was also
188 run for the same simulation period for comparison. The model was run for a three-month spin-up
189 period to reach a quasi-equilibrium state prior to beginning the annual simulations, and results
190 generated during the spin-up run were used as initial conditions for the annual simulations to
191 avoid disturbances caused by model initialization.

192 The coupled model is able to simulate mineral sediment transport processes but cannot
193 explicitly simulate organic carbon sequestration. In order to quantify organic carbon burial rates

194 in the meadow, a negative linear relationship between sediment organic carbon concentration
195 and sand fraction was established (Fig. S2; $R^2 = 0.75$) based on previous measurements from
196 Oreska et al. (2017) within the same meadow. Using this relationship, monthly distributions of
197 sand fraction (Fig. S3) output from model simulations were converted to maps of surface
198 sediment organic carbon concentration (Fig. S4). Then monthly distributions of organic carbon
199 burial rate [$\text{mg month}^{-1} \text{ cm}^{-2}$] in the meadow (Fig. S5) were determined by multiplying modeled
200 sediment accumulation rates [mm month^{-1}] by the surface sediment organic carbon concentration
201 [mg cm^{-3}].

202 **Results**

203 **Sediment and organic carbon accumulation rates vary with seagrass density**

204 With seasonally varying seagrass densities, our simulation results show that bed shear
205 stress and total suspended sediment concentration averaged across the seagrass meadow changed
206 non-linearly as a function of seagrass density (Fig. 2a, 2b). The most rapid changes of bed shear
207 stress and total suspended sediment concentration occurred at low seagrass densities, while there
208 was little change in bed shear stress and total suspended sediment concentration when seagrass
209 densities were > 200 shoots m^{-2} . Throughout the year, sediment resuspension in the meadow was
210 mainly controlled by seagrass density rather than wind speeds (Fig. S6).

211 Similarly, seasonal growth and senescence of seagrass exerted a strong influence on
212 sediment accumulation and carbon burial within seagrass meadows. Four low-density scenarios
213 (seagrass shoot density $N = 0, 25, 50$ and 100 shoots m^{-2}) were specified during winter
214 conditions to better resolve the effects of low seagrass density (Table S1). Simulation results
215 show that rates of sediment accumulation and organic carbon burial averaged across the meadow

216 varied strongly with seagrass density (Fig. 2c, 2d). When seagrass density was lower than 160
217 shoots m^{-2} , typical in winter, seagrass beds were erosional with rates that were sensitive to
218 density. When seagrass density > 200 shoots m^{-2} , rates of sediment accumulation and organic
219 carbon burial were relatively constant due to strong flow retardation (Fig. 2c, 2d, and S7). At the
220 meadow scale the average organic carbon accumulation rate was largely determined by sediment
221 accumulation (Fig. 3; $R^2 = 0.99$). On the annual time scale, simulations with typical seasonal
222 seagrass characteristics predicted a meadow-averaged sediment accumulation rate of 4.1 ± 0.5
223 (standard error) mm yr^{-1} and a carbon accumulation rate of 22 ± 1.6 (standard error) $\text{g C m}^{-2} \text{ yr}^{-1}$.

224 Based on model results showing how monthly sediment accumulation rates vary as a
225 function of seagrass density (Fig. 2c), we can design density variation scenarios by changing the
226 seagrass density in specific months and quantifying the corresponding effect (e.g., short-term
227 winter density variations and summer seagrass dieback due to marine heatwaves) on annual
228 sediment accumulation rates. Three density variation scenarios were constructed: lower-than-
229 average winter density (winter seagrass density reduced by 50 shoots m^{-2}); higher-than-average
230 winter density (winter seagrass density increased by 50 shoots m^{-2}); and a summer marine
231 heatwave scenario with summer seagrass density reduced to 50 shoots m^{-2} (Aoki et al. 2021). For
232 the lower-than-average winter density scenario, sediment accumulation rates in winter months
233 were much lower than in the normal-density simulations (red symbols in Fig. 2e) and the annual
234 sediment accumulation rate was reduced by 44% to 2.3 mm yr^{-1} . For the higher-than-average
235 winter density scenario, the seagrass meadow accumulated sediment during the entire winter
236 (gray symbols in Fig. 2e), resulting in an annual sediment accumulation rate of 6.6 mm yr^{-1} , a
237 61% increase compared with the rate under typical densities. For the summer marine heatwave
238 scenario, the meadow became erosional, with an annual sediment accumulation rate of -1.0 mm

239 yr^{-1} (Fig. 2f), thereby leading to the release of stored carbon in seagrass sediments (-4.6 g C m^{-2}
240 yr^{-1}). This modeling result of seabed erosion and sediment carbon loss is in general agreement
241 with a previous study in the same meadow documenting a net loss of 20% of sediment carbon in
242 the upper 5 cm of the bed (not including the effects of bed level changes) caused by a summer
243 marine heatwave in 2015 (Aoki et al. 2021).

244 **Spatial erosion and deposition pattern within meadows**

245 To better understand the spatial variability of seasonal sediment accumulation at the
246 meadow scale, we divided our annual simulation results into four groups according to the
247 seagrass growth cycle at our study site: summer growth and mid-season loss from June to August,
248 autumn regrowth from September to October, winter senescence from November to March, and
249 early growth from April to May. Our simulation results show that there were strong spatial
250 gradients of bed shear stress from the meadow edge toward the interior (Fig. S8), and these
251 spatial gradients had a significant impact on seasonal erosion/deposition patterns. Sediment
252 accumulation mainly occurred at meadow edges during summer when seagrass density was high
253 (Fig. 4a) and decreased rapidly with distance into the meadow interior (Fig. 4e). When seagrass
254 density decreased in autumn, the meadow still accumulated sediment with a lower deposition
255 rate, and this sediment could be transported further into the meadow interior due to weaker flow
256 attenuation by the seagrass (Fig. 4b, S8b). During the minimum densities in winter, most of the
257 meadow experienced erosion (Fig. 4c). The most severe erosion occurred near meadow edges
258 during the winter senescence period ($\sim 9 \text{ mm}$), an amount roughly equal to the mass of sediment
259 deposited at the edges in summer (Fig. 4e). When seagrass started regrowing in spring, the
260 meadow once again became depositional, with sediment accumulating at meadow edges, but at

261 lower rates compared to the summer growing season (Fig. 4d). Seasonal organic carbon
262 accumulation patterns were similar to those for sediment accumulation (Fig. S5).

263 On the annual time scale, the meadow-averaged sediment accumulation rate was low (0.4
264 $\pm 0.2 \text{ mm yr}^{-1}$) in the simulation with no seagrass, whereas the presence of a seagrass meadow
265 maintained a higher average sediment accumulation rate of $4.1 \pm 0.5 \text{ mm yr}^{-1}$ (Fig. 5a, 5b).
266 Despite the large spatial variability of sediment accumulation rates across the meadow (Fig. 5d),
267 the modeled sediment accumulation rates agreed reasonably well with rates estimated from ^{210}Pb
268 dating in previous studies at two sites of the meadow (see locations in Fig. 5b); the modeled
269 sediment accumulation rates at the central and northern sites were 5.0 and 4.3 mm yr^{-1} ,
270 respectively, while the rates estimated from ^{210}Pb dating were 6.6 and 6.0 mm yr^{-1} at these two
271 sites (Greiner et al. 2013; Oreska et al. 2018). Pronounced sediment accumulation ($> 4.0 \text{ mm yr}^{-1}$)
272 occurred at meadow edges and in the northwestern and southern portion of the meadow interior
273 (Fig. 5b), in good agreement with previous observations from Oreska et al. (2017) within the
274 same meadow. Because the amount of sediment accumulation at meadow edges in summer was
275 largely offset by winter erosion (Fig. 4e), the annual sediment accumulation rate near the edge
276 was mainly dependent on deposition during autumn regrowth and spring early growth seasons.
277 During autumn regrowth, winter senescence, and early growth seasons, sediment could be
278 transported further into the northwestern and southern portion of the meadow (Fig. 4) where
279 larger water depths resulted in lower bed shear stress and promoted sediment accumulation (Fig.
280 5b, 5c). At the meadow scale, while there was no depth preference for sediment
281 erosion/deposition in simulations without seagrass effects (Fig. 6a), deeper locations within the
282 meadow tended to receive more sediment deposition than shallow ones when seagrass effects
283 were included in model simulations (Fig. 6b).

284 **Discussion and conclusions**

285 A number of field studies (Hasegawa et al. 2008; Hansen and Reidenbach 2013) have
286 shown that high densities of seagrass in the summer growing season resulted in larger reductions
287 in bed shear stress and sediment resuspension compared with low-density meadows in winter.
288 However, the effects of seasonally varying seagrass densities on sediment accumulation have not
289 been well quantified due to the lack of monthly or seasonal measurements. Our annual
290 simulation results show that sediment accumulation rates within seagrass meadows changed non-
291 linearly between seasons as a function of seagrass density (Fig. 2c). The most rapid changes of
292 sediment accumulation rates occurred at low seagrass densities in winter (Fig. 2c), while there
293 was little change in sediment accumulation rates when seagrass densities were > 200 shoots m^{-2}
294 during other seasons due to effects of strong flow reduction at high shoot densities (Fig. 2a, S7;
295 Hansen and Reidenbach 2013).

296 Our seagrass density variation scenarios show that small variations in winter shoot
297 density can result in large changes ($> 40\%$) in annual sediment accumulation rates of the
298 meadow (Fig. 2e). Considering that most previous research has focused on flow and sediment
299 dynamics in high-density meadows (Donatelli et al. 2018), more comprehensive investigations of
300 sediment accumulation during low-density seasons are needed to better resolve the effects of
301 seagrass density variations on annual sediment accumulation. We also found that summer
302 seagrass dieback events, for example due to marine heatwaves, can change seagrass beds from
303 depositional (4.1 mm yr^{-1}) to erosional (-1.0 mm yr^{-1} ; Fig. 2f) on annual timescales. This strong
304 sensitivity of sediment accumulation rates to seagrass density variations has significant
305 implications for future scenarios of change. If seagrasses were present in much lower densities
306 due to degradation, physical disturbance, or increasing frequency of marine heatwaves, meadows

307 would inevitably become erosional and release accumulated carbon (Arias-Ortiz et al. 2018;
308 Walter et al. 2020), like the observed sediment carbon loss associated with the 2015 summer
309 marine heatwave in South Bay seagrass meadows (Aoki et al. 2021). Moreover, if seagrasses
310 were exposed to multiple years of degraded environmental conditions, including temperature
311 stress, the continuous erosion of seagrass beds associated with reduced seagrass growth would
312 increase water depth of the bay and promote sediment resuspension, thereby creating a less
313 favorable light condition for seagrass growth and potentially triggering irreversible collapse of
314 meadows (Carr et al., 2012).

315 Although previous studies have shown that seagrasses can effectively trap sediment and
316 promote sediment deposition (Gacia and Duarte 2001), there have been few direct meadow-scale
317 observations of spatial erosion/deposition patterns within seagrass meadows. Very little is known
318 about the spatial sediment deposition pattern in response to seasonal seagrass density variations.
319 Our simulation results show that edge effects play an important role in seasonal patterns of
320 sediment accumulation and carbon burial at a meadow scale (Fig. 4, S5). This is supported by
321 spatial autoregressive analyses in the same meadow that found that edge proximity was more
322 important than shoot density and meadow age in determining sediment organic carbon content
323 (Oreska et al. 2017). Sediment accumulation mainly occurred at meadow edges in spring and
324 summer growth seasons when seagrass density was relatively high (Fig. 4a, 4d). The modeled
325 sediment accumulation at meadow edges in the present study was in good agreement with other
326 model results considering seagrass edge effects at high shoot densities (Chen et al. 2007; Carr et
327 al. 2016). During the winter senescence period, severe erosion (~9 mm) was observed near
328 meadow edges, an amount roughly equal to 40%-50% of the mass of sediment deposited at the
329 edges in other seasons (Fig. 4).

330 During autumn regrowth, winter senescence, and early growth seasons, lower densities of
331 seagrass allow sediment to be advected further into the meadow (Fig. 4b, 4c, and 4d), providing
332 the primary mechanism for sediment deposition in the interior of the meadow. Our simulation
333 results show that water depth exerted a strong influence on the spatial pattern of sediment
334 accumulation in the meadow interior (Fig. 6b). Together with edge effects, depth variations can
335 produce strong spatial variability in sediment accumulation and carbon burial across the meadow
336 (Fig. 5; Samper-Villarreal et al. 2016; Oreska et al. 2017). This has implications for site selection
337 and timing of sediment sampling to characterize sediment accumulation and carbon burial in
338 seagrass meadows. Ideally, sampling sites should include the meadow edge and interior, as well
339 as deeper and shallower sites. Sampling during the autumn regrowth period may provide the best
340 representation of the annual spatial pattern of deposition (Fig. 4b, 5b).

341 Organic carbon burial rates within the meadow were obtained by multiplying sediment
342 accumulation rate by surface sediment organic carbon concentration at each model grid location.
343 At the meadow scale, although there were spatial gradients of sediment organic carbon
344 concentration from meadow edges to the interior (varying from 2 to 6 mg cm⁻³; Fig. S4), the
345 spatial pattern of organic carbon burial was largely determined by sediment accumulation (Fig. 3,
346 Fig. 4 vs. Fig. S5) because of its strong spatial variability across the meadow (Fig. 5d). This is in
347 agreement with a recent study indicating the strong impacts of sediment accumulation rates on
348 organic carbon burial at a basin scale (Johannessen and Macdonald 2016). In some other systems
349 where spatial distributions of sediment accumulation rate are relatively uniform, like those in
350 lacustrine environments (Lin et al., 2022), the spatial pattern of organic carbon burial will more
351 likely be controlled by spatial distributions of sediment organic carbon concentration.

352 On an annual time scale, our simulations with seasonal seagrass characteristics predicted
353 a meadow-wide average sediment accumulation rate of $4.1 \pm 0.5 \text{ mm yr}^{-1}$. This value is similar to
354 the rapid rates of sea-level rise (4-5 mm yr⁻¹) at our study site, but is less than the average long-
355 term sediment accumulation rate (6.3 mm yr⁻¹) determined by ²¹⁰Pb dating at two representative
356 sites within the meadow (Greiner et al. 2013; Oreska et al. 2018). In future studies, more
357 spatially distributed sediment accumulation rate measurements will be needed to constrain model
358 simulation results and improve our understanding of spatial erosion/deposition patterns within
359 seagrass meadows.

360 The modeled carbon accumulation rate averaged across the meadow ($22 \pm 1.6 \text{ g C m}^{-2} \text{ yr}^{-1}$)
361 was lower than the rate obtained from the meadow-scale carbon stock estimate in the same
362 meadow based on 16 spatially distributed sampling sites ($42 \text{ g C m}^{-2} \text{ yr}^{-1}$; Oreska et al. 2020).
363 The underestimation of sediment accumulation and carbon burial rates is likely due to the
364 absence of primary production in model simulations. A more realistic approach would be to
365 incorporate vegetation growth dynamics and organic matter trapping (e.g., Best et al. 2018;
366 Brückner et al. 2019) into the coupled model. In addition, our model grid size (~70 m) was too
367 coarse to resolve seagrass patchiness, which has been shown to play a significant role in
368 distributions of bed shear stress and sediment accumulation (Ricart et al. 2015; Carr et al. 2016).
369 Another limitation of this study is that we did not simulate temperature in the coupled model and
370 therefore could not resolve the effects of biodegradation and mineralization on sediment carbon
371 burial, which have been found to be important factors affecting carbon storage in other seagrass
372 ecosystems (Sohma et al. 2018).

373 Our model was able to produce reasonable spatially-resolved simulations of sediment
374 accumulation in seagrass meadows under seasonally varying winds and tides and seagrass

375 densities. This is one of the first modeling attempts to examine in detail the extent and drivers of
376 temporal and spatial variability in sediment accumulation and carbon burial within seagrass
377 meadows. Our results have significant implications for seagrass ecosystems under future climate
378 change. In a warming climate, seagrasses in temperate regions in the northern hemisphere will
379 shift northward and the southernmost populations, like the meadows at our site, will likely face
380 reduced growth and meadow loss due to temperature stress (Wilson and Lotze 2019). Marine
381 heatwaves have already impacted seagrass meadows in this region (Berger et al. 2020; Aoki et al.
382 2021), and these are predicted to increase in frequency in future warming oceans (Oliver et al.
383 2019). Our modeling results of sediment accumulation in seagrass meadows and field studies
384 from the same region (Aoki et al. 2021) indicate that under these circumstances, seabed erosion
385 and carbon emissions will increase. Conservation actions are therefore urgently needed to
386 mitigate the effects of climate change on seagrass ecosystems.

387 **References**

- 388 Aoki, L. R., K. J. McGlathery, P. L. Wiberg, and A. Al-Haj. 2020. Depth Affects Seagrass
389 Restoration Success and Resilience to Marine Heat Wave Disturbance. *Estuaries and Coasts*
390 **43**: 316–328. doi:10.1007/s12237-019-00685-0
- 391 Aoki, L. R., K. J. McGlathery, P. L. Wiberg, M. P. J. Oreska, A. C. Berger, P. Berg, and R. J.
392 Orth. 2021. Seagrass Recovery Following Marine Heat Wave Influences Sediment Carbon
393 Stocks. *Front. Mar. Sci.* **7**: 1170. doi:10.3389/fmars.2020.576784
- 394 Arias-Ortiz, A., O. Serrano, P. Masqué, and others. 2018. A marine heatwave drives massive
395 losses from the world's largest seagrass carbon stocks. *Nat. Clim. Chang.* **8**: 1–7.
396 doi:10.1038/s41558-018-0096-y
- 397 Baptist, M. J., V. Babovic, J. R. Uthurburu, M. Keijzer, R. E. Uittenbogaard, A. Mynett, and A.
398 Verwey. 2007. On inducing equations for vegetation resistance. *J. Hydraul. Res.* **45**: 435–
399 450. doi:10.1080/00221686.2007.9521778
- 400 Berger, A. C., P. Berg, K. J. McGlathery, and M. L. Delgard. 2020. Long-term trends and
401 resilience of seagrass metabolism: A decadal aquatic eddy covariance study. *Limnol.*
402 *Oceanogr.* **65**: 1423–1438. doi:10.1002/lno.11397
- 403 Best, S. N., M. Van der Wegen, J. Dijkstra, P. W. J. M. Willemsen, B. W. Borsje, and D. J. A.
404 Roelvink. 2018. Do salt marshes survive sea level rise? *Modelling wave action*,

- 405 morphodynamics and vegetation dynamics. *Environ. Model. Softw.* **109**: 152–166.
406 doi:10.1016/j.envsoft.2018.08.004
- 407 Booij, N., R. C. Ris, and L. H. Holthuijsen. 1999. A third-generation wave model for coastal
408 regions 1. Model description and validation. *J. Geophys. Res. Ocean.* **104**: 7649–7666.
409 doi:10.1029/98JC02622
- 410 Brückner, M. Z. M., C. Schwarz, W. M. Dijk, M. Oorschot, H. Douma, and M. G. Kleinhans.
411 2019. Salt Marsh Establishment and Eco-Engineering Effects in Dynamic Estuaries
412 Determined by Species Growth and Mortality. *J. Geophys. Res. Earth Surf.* **124**: 2962–2986.
413 doi:10.1029/2019JF005092
- 414 Carr, J. A., P. D'Odorico, K. J. McGlathery, and P. L. Wiberg. 2012. Stability and resilience of
415 seagrass meadows to seasonal and interannual dynamics and environmental stress. *J.*
416 *Geophys. Res. Biogeosci.* **117**: 1007. doi:10.1029/2011JG001744
- 417 Carr, J. A., P. D'Odorico, K. J. McGlathery, and P. L. Wiberg. 2016. Spatially explicit feedbacks
418 between seagrass meadow structure, sediment and light: Habitat suitability for seagrass
419 growth. *Adv. Water Resour.* **93**: 315–325. doi:10.1016/j.advwatres.2015.09.001
- 420 Chen, S. N., L. P. Sanford, E. W. Koch, F. Shi, and E. W. North. 2007. A nearshore model to
421 investigate the effects of seagrass bed geometry on wave attenuation and suspended
422 sediment transport. *Estuaries and Coasts* **30**: 296–310. doi:10.1007/BF02700172
- 423 Deltares. 2014. Delft3D-FLOW User Manual (Version: 3.15.34158): Simulation of multi-
424 dimensional hydrodynamic flows and transport phenomena, including sediments.
- 425 de Boer, W. F. 2007. Seagrass-sediment interactions, positive feedbacks and critical thresholds
426 for occurrence: A review. *Hydrobiologia* **591**: 5–24. doi:10.1007/s10750-007-0780-9
- 427 Donatelli, C., N. K. Ganju, S. Fagherazzi, and N. Leonardi. 2018. Seagrass Impact on Sediment
428 Exchange Between Tidal Flats and Salt Marsh, and The Sediment Budget of Shallow Bays.
429 *Geophys. Res. Lett.* **45**: 4933–4943. doi:10.1029/2018GL078056
- 430 Duarte, C. M., H. Kennedy, N. Marbà, and I. Hendriks. 2013. Assessing the capacity of seagrass
431 meadows for carbon burial: Current limitations and future strategies. *Ocean Coast. Manag.*
432 **83**: 32–38. doi:10.1016/j.ocecoaman.2011.09.001
- 433 Fagherazzi, S., and P. L. Wiberg. 2009. Importance of wind conditions, fetch, and water levels
434 on wave-generated shear stresses in shallow intertidal basins. *J. Geophys. Res.* **114**: F03022.
435 doi:10.1029/2008JF001139
- 436 Fourqurean, J. W., C. M. Duarte, H. Kennedy, and others. 2012. Seagrass ecosystems as a
437 globally significant carbon stock. *Nat. Geosci.* **5**: 505–509. doi:10.1038/ngeo1477
- 438 Gacia, E., and C. M. Duarte. 2001. Sediment retention by a Mediterranean *Posidonia oceanica*
439 meadow: The balance between deposition and resuspension. *Estuar. Coast. Shelf Sci.* **52**:
440 505–514. doi:10.1006/ecss.2000.0753
- 441 Greiner, J. T., K. J. McGlathery, J. Gunnell, and B. A. McKee. 2013. Seagrass Restoration
442 Enhances “Blue Carbon” Sequestration in Coastal Waters. *PLoS One* **8**: e72469.
443 doi:10.1371/journal.pone.0072469

- 444 Hansen, J., and M. A. Reidenbach. 2013. Seasonal Growth and Senescence of a *Zostera marina*
445 Seagrass Meadow Alters Wave-Dominated Flow and Sediment Suspension Within a
446 Coastal Bay. *Estuaries and Coasts* **36**: 1099–1114. doi:10.1007/s12237-013-9620-5
- 447 Hasegawa, N., M. Hori, and H. Mukai. 2008. Seasonal changes in eelgrass functions: Current
448 velocity reduction, prevention of sediment resuspension, and control of sediment-water
449 column nutrient flux in relation to eelgrass dynamics. *Hydrobiologia* **596**: 387–399.
450 doi:10.1007/s10750-007-9111-4
- 451 Johannessen, S. C., and R. W. Macdonald. 2016. Geoengineering with seagrasses: Is credit due
452 where credit is given? *Environ. Res. Lett.* **11**: 113001. doi:10.1088/1748-
453 9326/11/11/113001
- 454 Lawson, S. E., P. L. Wiberg, K. J. McGlathery, and D. C. Fugate. 2007. Wind-driven sediment
455 suspension controls light availability in a shallow coastal lagoon. *Estuaries and Coasts* **30**:
456 102–112. doi:10.1007/BF02782971
- 457 Lesser, G. R., J. A. Roelvink, J. A. T. M. T. M. van Kester, and G. S. Stelling. 2004.
458 Development and validation of a three-dimensional morphological model. *Coast. Eng.* **51**:
459 883–915. doi:10.1016/j.coastaleng.2004.07.014
- 460 Lin, Q., E. Liu, E. Zhang, R. Bindler, B. Nath, K. Zhang, and J. Shen. 2022. Spatial variation of
461 organic carbon sequestration in large lakes and implications for carbon stock quantification.
462 *CATENA* **208**: 105768. doi:10.1016/J.CATENA.2021.105768
- 463 McGlathery, K. J., L. K. Reynolds, L. W. Cole, R. J. Orth, S. R. Marion, and A. Schwarzschild.
464 2012. Recovery trajectories during state change from bare sediment to eelgrass dominance.
465 *Mar. Ecol. Prog. Ser.* **448**: 209–221. doi:10.3354/meps09574
- 466 McGlathery, K., K. Sundbäck, and I. Anderson. 2007. Eutrophication in shallow coastal bays
467 and lagoons: the role of plants in the coastal filter. *Mar. Ecol. Prog. Ser.* **348**: 1–18.
468 doi:10.3354/meps07132
- 469 Moki, H., K. Taguchi, Y. Nakagawa, S. Montani, and T. Kuwae. 2020. Spatial and seasonal
470 impacts of submerged aquatic vegetation (SAV) drag force on hydrodynamics in shallow
471 waters. *J. Mar. Syst.* **209**: 103373. doi:10.1016/J.JMARSYS.2020.103373
- 472 Nardin, W., and D. A. Edmonds. 2014. Optimum vegetation height and density for inorganic
473 sedimentation in deltaic marshes. *Nat. Geosci.* **7**: 722–726. doi:10.1038/NGEO2233
- 474 Nardin, W., L. Larsen, S. Fagherazzi, and P. Wiberg. 2018. Tradeoffs among hydrodynamics,
475 sediment fluxes and vegetation community in the Virginia Coast Reserve, USA. *Estuar.
476 Coast. Shelf Sci.* **210**: 98–108. doi:10.1016/j.ecss.2018.06.009
- 477 Nolte, S., E. C. Koppenaal, P. Esselink, K. S. Dijkema, M. Schuerch, A. V. De Groot, J. P.
478 Bakker, and S. Temmerman. 2013. Measuring sedimentation in tidal marshes: A review on
479 methods and their applicability in biogeomorphological studies. *J. Coast. Conserv.* **17**: 301–
480 325. doi:10.1007/s11852-013-0238-3
- 481 Oliver, E. C. J., M. T. Burrows, M. G. Donat, and others. 2019. Projected Marine Heatwaves in
482 the 21st Century and the Potential for Ecological Impact. *Front. Mar. Sci.* **6**: 734.
483 doi:10.3389/fmars.2019.00734

- 484 Oreska, M. P. J., K. J. McGlathery, L. R. Aoki, A. C. Berger, P. Berg, and L. Mullins. 2020. The
485 greenhouse gas offset potential from seagrass restoration. *Sci. Rep.* **10**: 1–15.
486 doi:10.1038/s41598-020-64094-1
- 487 Oreska, M. P. J., K. J. McGlathery, and J. H. Porter. 2017. Seagrass blue carbon spatial patterns
488 at the meadow-scale. *PLoS One* **12**: e0176630. doi:10.1371/journal.pone.0176630
- 489 Oreska, M. P. J., G. M. Wilkinson, K. J. McGlathery, M. Bost, and B. A. McKee. 2018. Non-
490 seagrass carbon contributions to seagrass sediment blue carbon. *Limnol. Oceanogr.* **63**: S3–
491 S18. doi:10.1002/lno.10718
- 492 Orth, R. J., T. J. B. Carruthers, W. C. Dennison, and others. 2006. A Global Crisis for Seagrass
493 Ecosystems. *Bioscience* **56**: 987–996. doi:10.1641/0006-
494 3568(2006)56[987:AGCFSE]2.0.CO;2
- 495 Pendleton, L., D. C. Donato, B. C. Murray, and others. 2012. Estimating Global “Blue Carbon”
496 Emissions from Conversion and Degradation of Vegetated Coastal Ecosystems. *PLoS One*
497 **7**: e43542. doi:10.1371/journal.pone.0043542
- 498 Potouroglou, M., J. C. Bull, K. W. Krauss, and others. 2017. Measuring the role of seagrasses in
499 regulating sediment surface elevation. *Sci. Rep.* **7**: 1–11. doi:10.1038/s41598-017-12354-y
- 500 Reidenbach, M. A., and E. L. Thomas. 2018. Influence of the Seagrass, *Zostera marina*, on Wave
501 Attenuation and Bed Shear Stress Within a Shallow Coastal Bay. *Front. Mar. Sci.* **5**: 397.
502 doi:10.3389/fmars.2018.00397
- 503 Rheuban, J. E., P. Berg, and K. J. McGlathery. 2014. Ecosystem metabolism along a
504 colonization gradient of eelgrass (*Zostera marina*) measured by eddy correlation. *Limnol.*
505 *Oceanogr.* **59**: 1376–1387. doi:10.4319/lo.2014.59.4.1376
- 506 Ricart, A. M., P. H. York, M. A. Rasheed, M. Pérez, J. Romero, C. V. Bryant, and P. I.
507 Macreadie. 2015. Variability of sedimentary organic carbon in patchy seagrass landscapes.
508 *Mar. Pollut. Bull.* **100**: 476–482. doi:10.1016/j.marpolbul.2015.09.032
- 509 Samper-Villarreal, J., C. E. Lovelock, M. I. Saunders, C. Roelfsema, and P. J. Mumby. 2016.
510 Organic carbon in seagrass sediments is influenced by seagrass canopy complexity,
511 turbidity, wave height, and water depth. *Limnol. Oceanogr.* **61**: 938–952.
512 doi:10.1002/LNO.10262
- 513 Sohma, A., H. Shibuki, F. Nakajima, A. Kubo, and T. Kuwae. 2018. Modeling a coastal
514 ecosystem to estimate climate change mitigation and a model demonstration in Tokyo Bay.
515 *Ecol. Modell.* **384**: 261–289. doi:10.1016/J.ECOLMODEL.2018.04.019
- 516 Suzuki, T., M. Zijlema, B. Burger, M. C. Meijer, and S. Narayan. 2012. Wave dissipation by
517 vegetation with layer schematization in SWAN. *Coast. Eng.* **59**: 64–71.
518 doi:10.1016/j.coastaleng.2011.07.006
- 519 Walter, R. K., J. K. O’Leary, S. Vitousek, M. Taherkhani, C. Geraghty, and A. Kitajima. 2020.
520 Large-scale erosion driven by intertidal eelgrass loss in an estuarine environment. *Estuar. Coast. Shelf Sci.* **243**: 106910. doi:10.1016/j.ecss.2020.106910
- 522 Waycott, M., C. M. Duarte, T. J. B. Carruthers, and others. 2009. Accelerating loss of seagrasses

523 across the globe threatens coastal ecosystems. *Proc. Natl. Acad. Sci. U. S. A.* **106**: 12377–
524 12381. doi:10.1073/pnas.0905620106

525 Wiberg, P. L., J. A. Carr, I. Safak, and A. Anutaliya. 2015. Quantifying the distribution and
526 influence of non-uniform bed properties in shallow coastal bays. *Limnol. Oceanogr.*
527 *Methods* **13**: 746–762. doi:10.1002/lom3.10063

528 Willmott, C. J. 1981. On the validation of models. *Phys. Geogr.* **2**: 184–194.
529 doi:10.1080/02723646.1981.10642213

530 Wilson, K. L., and H. K. Lotze. 2019. Climate change projections reveal range shifts of eelgrass
531 *Zostera marina* in the Northwest Atlantic. *Mar. Ecol. Prog. Ser.* **620**: 47–62.
532 doi:10.3354/meps12973

533 Zhu, Q., P. L. Wiberg, and M. A. Reidenbach. 2021. Quantifying Seasonal Seagrass Effects on
534 Flow and Sediment Dynamics in a Back-Barrier Bay. *J. Geophys. Res. Ocean.* **126**:
535 e2020JC016547. doi:10.1029/2020JC016547

536

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541 providing sediment organic carbon concentration and sand fraction data at the study site.

542

543 **Conflict of Interest**

544 None declared.

545

546 **Figure Legends**

547 **Fig. 1** (a) Aerial image of the study area, (b) Model grid and bathymetry in South Bay, and (c)
548 Typical monthly seagrass shoot density (N) at the central meadow throughout the year. Red
549 dashed lines in (b) represent boundaries of three seagrass density classes (N , $0.8N$, and $0.6N$)
550 used in the model (Table S1). The seagrass shoot density data shown in (c) were compiled from
551 previous seasonal seagrass observations in South Bay (Hansen and Reidenbach 2013; Rheuban et
552 al. 2014; Reidenbach and Thomas 2018; Berger et al. 2020).

553 **Fig. 2** (a) Bed shear stress, (b) Total suspended sediment concentration (SSC), (c) Sediment
554 accumulation rates, and (d) Organic carbon accumulation rates as a function of seagrass density,
555 (e) Bar plots of seagrass density and changes of sediment accumulation rate in seagrass meadows
556 under normal-density conditions (blue), lower-than-average winter density scenario (red), and
557 higher-than-average winter density scenario (gray), and (f) Bar plots of seagrass density and
558 changes of sediment accumulation rate in seagrass meadows under normal-density conditions
559 and summer marine heatwave scenario (yellow). Red symbols in (c) and (d) are model results
560 output from low-density scenarios using a seagrass shoot density of 0, 25, and 50 shoots m^{-2} ,
561 respectively. Details about model settings and seagrass characteristics used in model simulations
562 are provided in Table S1.

563 **Fig. 3** Relationship between modeled surface sediment organic carbon (C_{org}) accumulation rate
564 and modeled sediment accumulation rate averaged across the meadow.

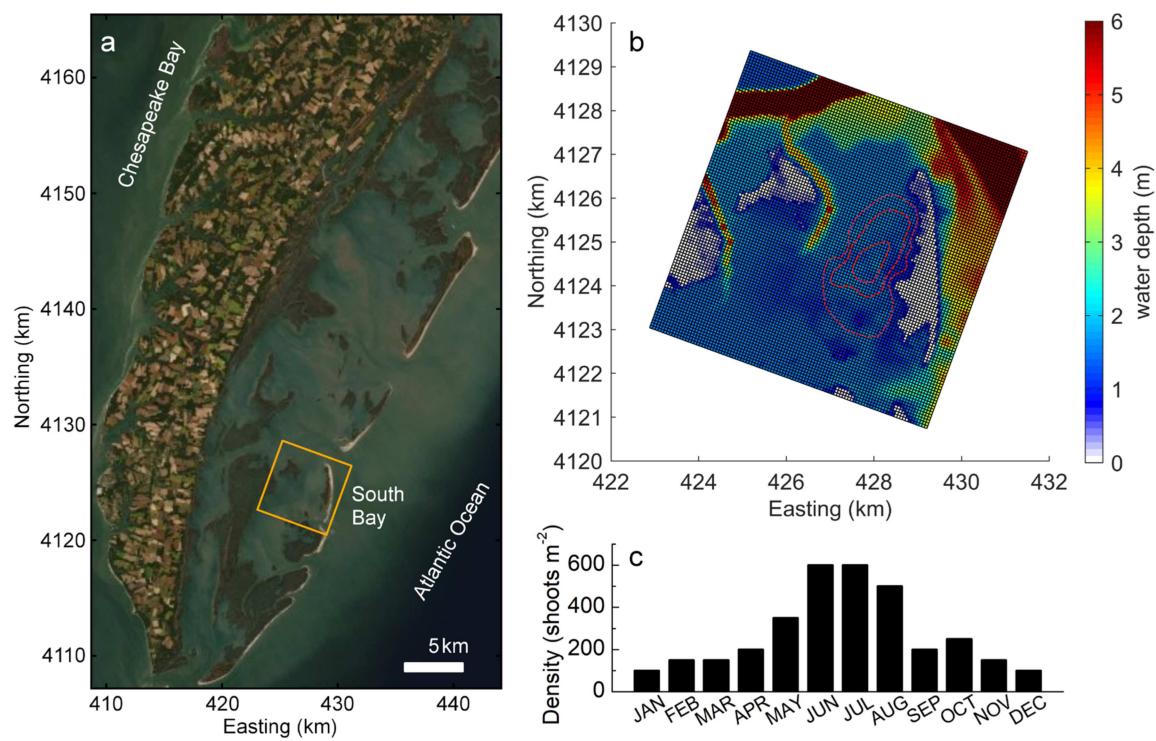
565 **Fig. 4** Average seasonal sediment accumulation rates output from model simulations with typical
566 seasonal seagrass characteristics: (a) summer growth and mid-season loss, (b) autumn regrowth,
567 (c) winter senescence, and (d) early growth. (e) Box plots of cumulative sediment accumulation
568 during summer growth and mid-season loss (SG) and winter senescence (WS) as a function of

569 distance to the meadow edge. The black line in (a)-(d) shows the meadow outline. The data
570 shown in (e) were extracted from 28 interior transects perpendicular to the northern, western, and
571 southern edges of the meadow.

572 **Fig. 5** Annual sediment erosion and deposition patterns (a) without seagrass effects and (b) with
573 seagrass effects, (c) Water depth of the meadow, and (d) Histogram of modeled annual sediment
574 accumulation rates within the meadow. The asterisk and triangle in (b) show locations of
575 sediment cores for ^{210}Pb dating collected by Oreska et al. (2018) and Greiner et al. (2013),
576 respectively. The red line in (d) represents the mean modeled sediment accumulation rate, while
577 the black dashed line shows the average sediment accumulation rate by ^{210}Pb dating.

578 **Fig. 6** Box plots of annual sediment erosion/deposition within seagrass meadows as a function of
579 depth without seagrass effects (a) and with seagrass effects (b).

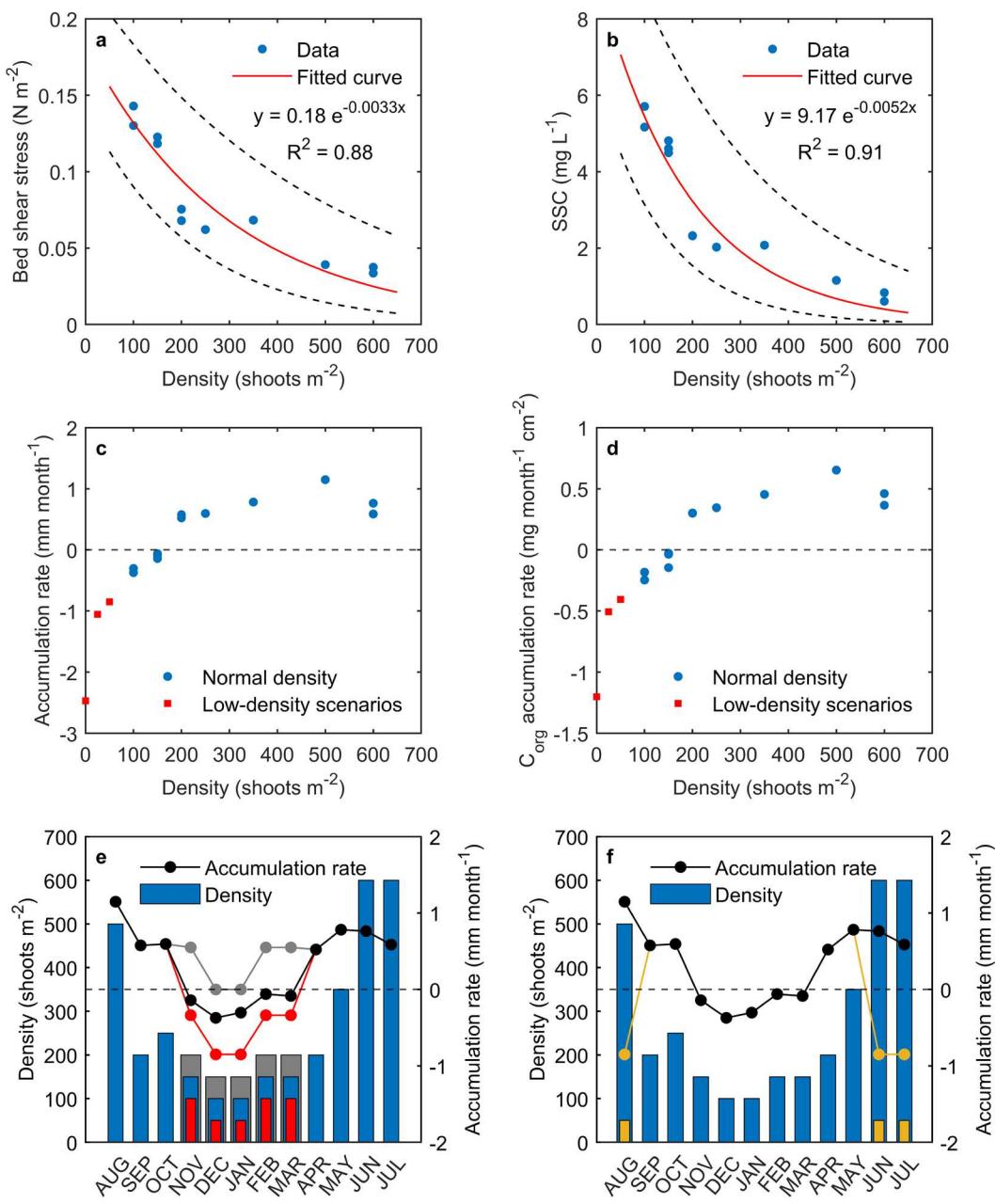
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582 Fig. 1

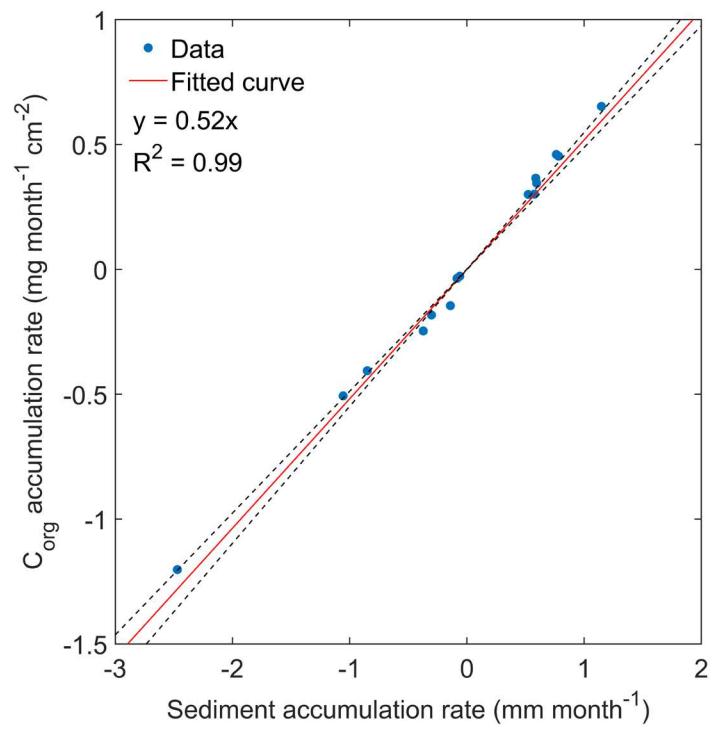
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585 Fig. 2

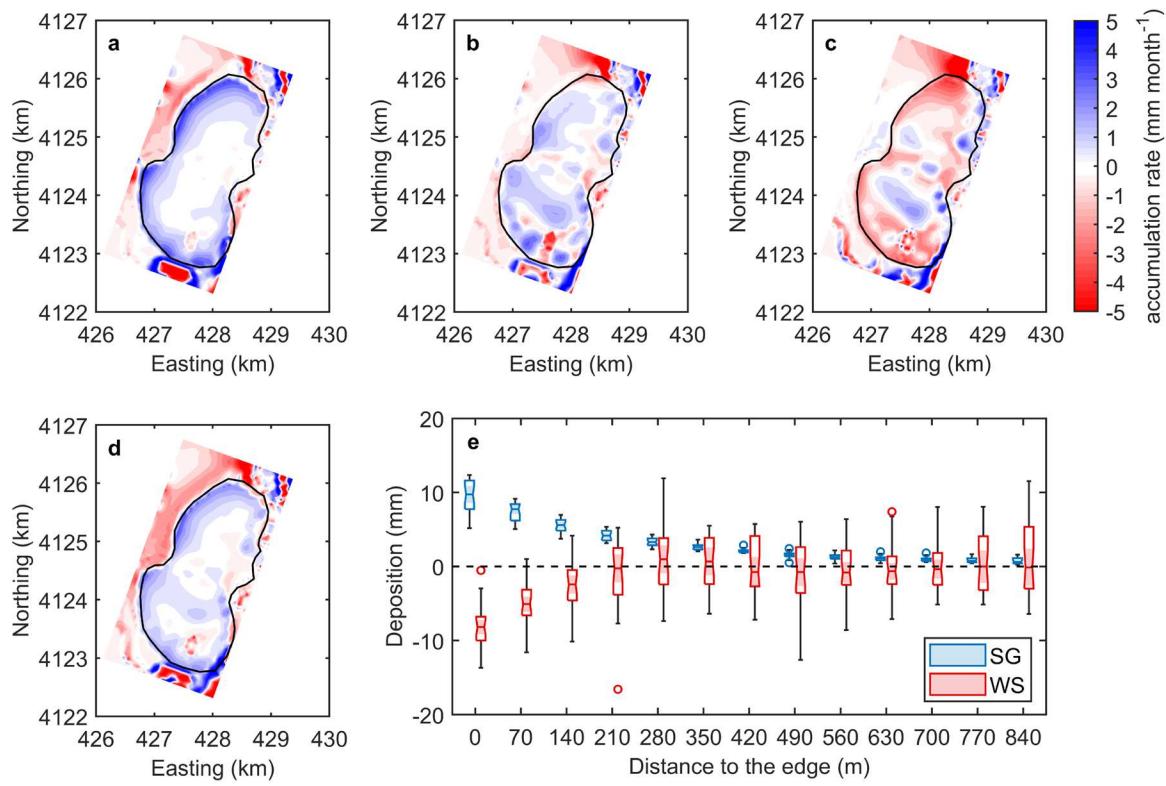
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588 Fig. 3

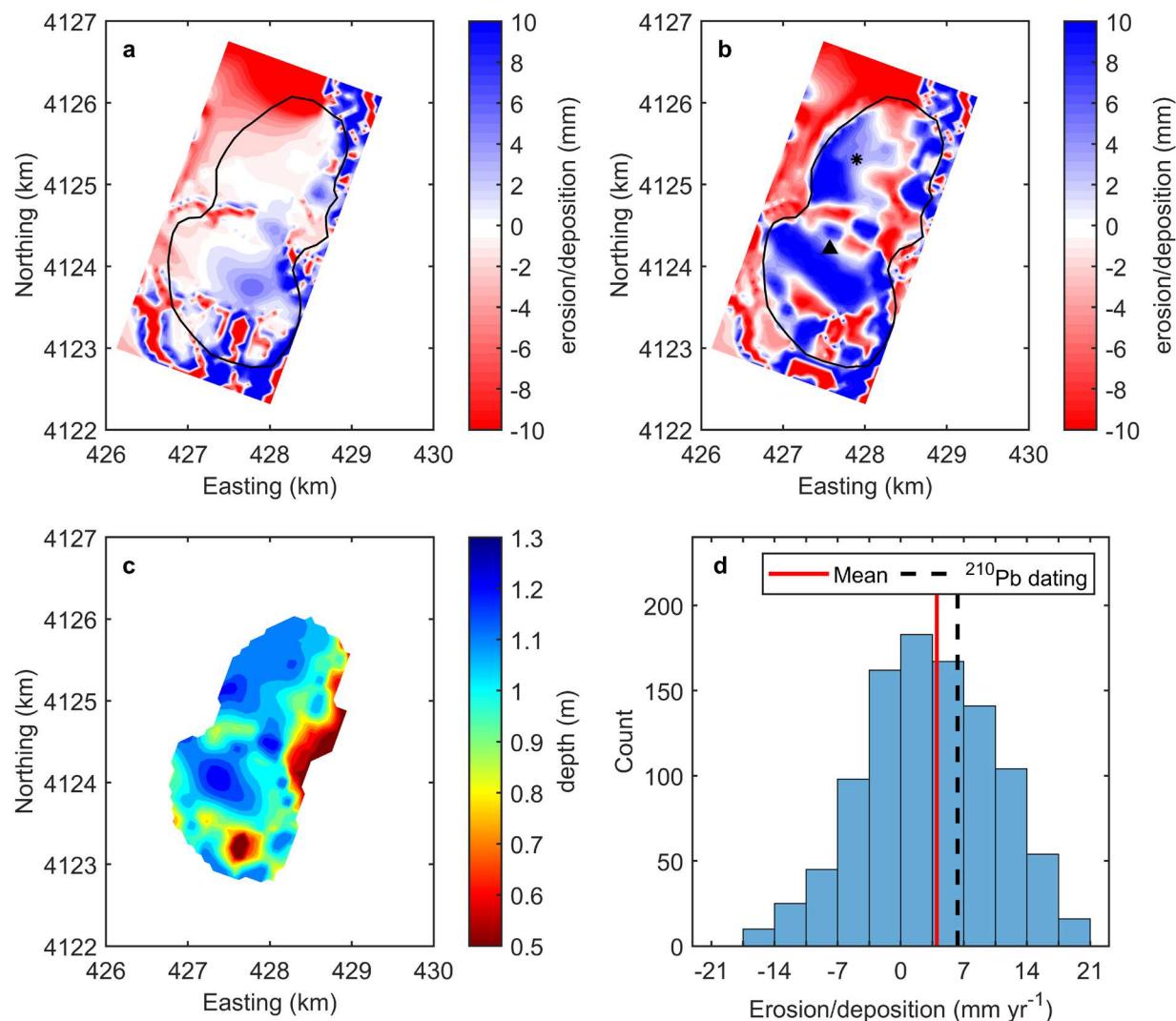
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591 Fig. 4

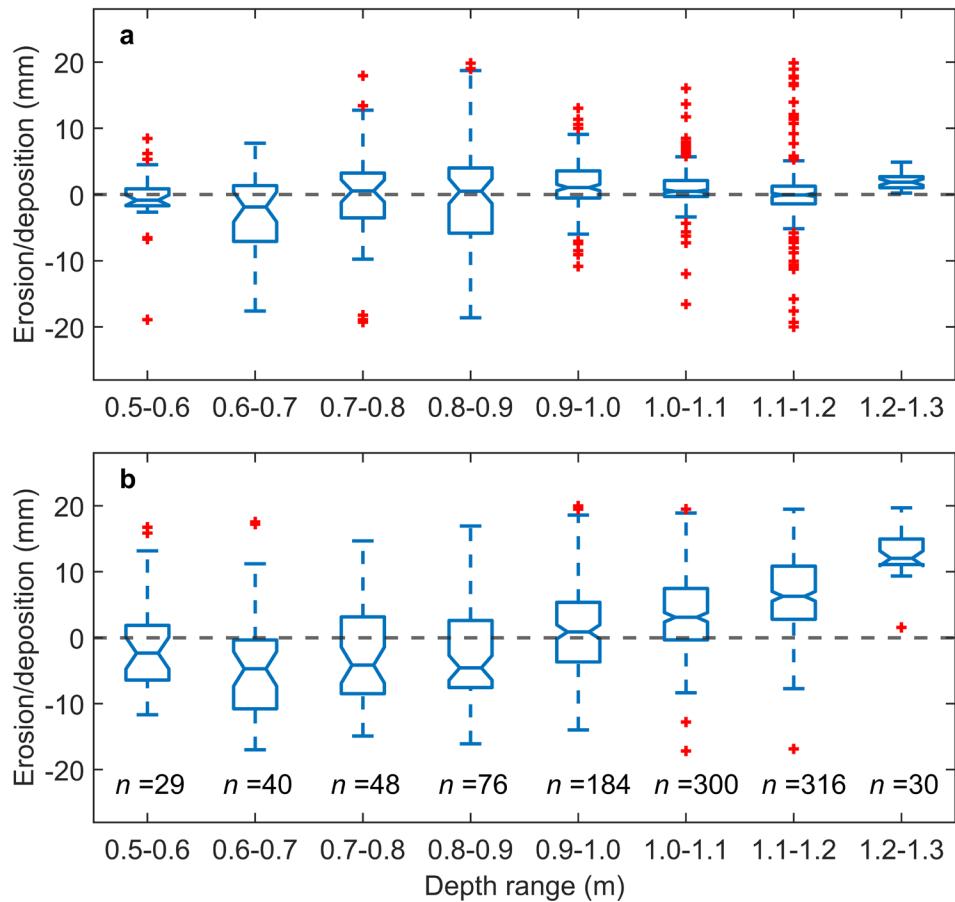
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594 Fig. 5

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597 Fig. 6