

1 **Individual-based landscape genomics for conservation: an analysis pipeline**

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3 **Running header:** Individual-based landscape genomics

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5 E. Anne Chambers^{1,2*†} | Anusha P. Bishop^{1,2†} | Ian J. Wang^{1,2}

6
7 ¹*Department of Environmental Science, Policy, and Management, University of California*
8 *Berkeley, Berkeley, CA 94720, USA*

9 ²*Museum of Vertebrate Zoology, University of California Berkeley, Berkeley, CA 94720, USA*

10
11 **Correspondence to be sent to: Department of Environmental Science, Policy, and Management,*
12 *University of California Berkeley, Berkeley, CA 94720, USA; E-mail: eachambers@berkeley.edu*

13 [†]*E. Anne Chambers and Anusha P. Bishop should be considered joint first authors.*

15 **Abstract**

16 Landscape genomics can harness environmental and genetic data to inform conservation
17 decisions by providing essential insights into how landscapes shape biodiversity. The massive
18 increase in genetic data afforded by the genomic era provides exceptional resolution for
19 answering critical conservation genetics questions. The accessibility of genomic data for non-
20 model systems has also enabled a shift away from population-based sampling to individual-based
21 sampling, which now provides accurate and robust estimates of genetic variation that can be used
22 to examine the spatial structure of genomic diversity, population connectivity, and the nature of
23 environmental adaptation. Nevertheless, the adoption of individual-based sampling in
24 conservation genetics has been slowed due, in large part, to concerns over how to apply methods
25 developed for population-based sampling to individual-based sampling schemes. Here, we
26 discuss the benefits of individual-based sampling for conservation and describe how landscape
27 genomic methods, paired with individual-based sampling, can answer fundamental conservation
28 questions. We have curated key landscape genomic methods into a user-friendly, open-source
29 workflow, which we provide as a new R package, A Landscape Genomics Analysis Toolkit in R
30 (ALGATR). The ALGATR package includes added functionality for all of the included methods and
31 extensive vignettes designed with the primary goal of making landscape genomic approaches
32 more accessible and explicitly applicable to conservation biology.

33

34 **KEYWORDS:** Conservation Genetics, Conservation Biology, Landscape Genomics, Spatial
35 Analysis, Population Structure, Genetic Diversity

36 1 | INTRODUCTION

37 The Anthropocene is an era of unprecedented global change, making the question of how
38 best to conserve biodiversity more important than ever. Although broad conservation policies are
39 typically enacted at the species level, conservation actions and management decisions regularly
40 take place at the landscape level (Fiedler et al., 2022). Protecting species diversity and the
41 processes that maintain it ultimately requires conserving the geographic regions that support the
42 ecological and evolutionary processes intrinsic to population viability (Malcom & Carter, 2021).
43 Particularly with accelerating environmental change, land-use conversion, and habitat loss and
44 degradation, biological conservation has increasingly become an inherently spatial problem
45 (Shaffer et al. 2022). Uniting landscape ecology and population genetics in a single framework,
46 landscape genetics provides a suite of spatially-explicit approaches for addressing this challenge
47 (Balkenhol et al., 2015; Keller et al., 2015; Manel et al., 2003, 2010; Segelbacher et al., 2010;
48 Shaffer et al., 2022; Storfer et al., 2007; van Strien et al., 2014). Spatial approaches in landscape
49 genetics involve the integration of geographic, environmental, and genetic data to understand
50 how spatial patterns of genetic variation are influenced by landscape features, environmental
51 factors, and dispersal barriers. Broadly, these include methods to delimit population boundaries,
52 quantify the drivers of genetic differentiation, estimate landscape resistance to gene flow,
53 characterize patterns of genetic diversity, and identify genes involved in adaptation to local
54 environmental variation (Balkenhol et al., 2015; Storfer et al., 2018; Wagner & Fortin, 2013).

55 Genomic data provide valuable information for identifying population boundaries,
56 demographics, and connectivity, all of which are important for conservation (Funk et al., 2012;
57 Hohenlohe et al., 2021; Keller et al., 2015; Manel & Holderegger, 2013; Segelbacher et al.,
58 2010). Genetic diversity itself is increasingly a target of conservation action especially because

59 of the role it can play in mitigating the impacts of ongoing climate change (Hoban et al., 2020;
60 Schmidt et al., 2023). Prior to the availability and accessibility of genomic data for non-model
61 systems, landscape genetics studies relied on population-based sampling, in which many
62 individuals from each location were genotyped for a handful of genetic markers. Now that we
63 can sequence many thousands of loci from across the genome – allowing us to capture more of
64 the complexity that exists in natural systems (Forester & Lama, 2022; Holliday et al., 2017;
65 Shafer et al., 2015; Supple & Shapiro, 2018) – landscape genetics has increasingly shifted
66 towards individual-based sampling. With genomic data, just a single individual per location can
67 provide robust inferences about intraspecific genetic variation (Hohenlohe et al., 2021; Selmoni
68 et al., 2020). This is advantageous for conservation efforts as well, as individual-based sampling
69 allows for broader geographic and environmental coverage, provides greater spatial resolution,
70 and minimizes the overall impact of sampling on each population (Shaffer et al., 2022; Wang &
71 Bradburd, 2014).

72 Despite the benefits of individual-based sampling, a broader shift to these sampling
73 schemes has been slowed, in large part, by methodological concerns. Landscape genetics is
74 replete with analytical approaches, but the very breadth of those choices, and the many decisions
75 to be made in implementing each method present a certain challenge, particularly for those
76 seeking entry into landscape genetics but even for experienced practitioners as well. Moreover,
77 because many landscape and population genetic methods were originally designed for
78 population-based sampling, their validity for individual-based sampling has sometimes come
79 into question, further complicating the question of how best to implement a landscape genomic
80 framework for conservation. Here, we present a pipeline for performing cutting-edge landscape
81 genomic analyses with individual-based sampling, discuss the key conservation-oriented

82 questions it can answer, and detail the considerations for each method it includes (Table 1). Our
83 pipeline is implemented as an R package, A Landscape Genomics Analysis Toolkit in R
84 (ALGATR), that is publicly available on GitHub (<https://github.com/TheWangLab/algatr>). ALGATR
85 includes detailed walkthroughs and documentation to make it easily accessible to anyone eager
86 to use landscape genomics to achieve actionable conservation impacts.

87

88 **2 | INDIVIDUAL-BASED LANDSCAPE GENOMICS FOR** 89 **CONSERVATION**

90 Having many independent loci is advantageous because each serves as an independent
91 instantiation of the coalescent process (Leitwein et al., 2020), so sequencing large numbers of
92 loci can provide robust estimates of genetic differentiation between localities, even when the
93 sample size at each locus is small. For this reason, the number of loci that can be captured by
94 next-generation sequencing makes individual-based sampling tenable for landscape genetics.
95 Large genomic datasets provide strong statistical power for inferring spatial patterns of genetic
96 variation, including the detection of genotype-environment associations (GEA), geographic
97 barriers and corridors, and populations or genetic clusters without *a priori* assignments (Manel et
98 al., 2003; Patterson et al., 2006). Individual-based, genomic-scale datasets also contribute
99 distinct advantages for practical conservation biology – in particular, distributing sampling
100 across more sites provides greater geographic and environmental coverage, better spatial
101 resolution, and lower impact on natural populations.

102 **2.1 | Greater geographic and environmental coverage**

103 For the same total sample size, individual-based sampling schemes are able to include
104 more localities than population-based schemes, providing greater geographic coverage and

105 environmental breadth. Broadening the spatial extent of a project captures greater landscape
106 heterogeneity and more of the genetic variation across a species' range. It also allows projects to
107 cover more areas that may be of interest for conservation and land management efforts, which is
108 valuable for evaluating currently designated protected areas, for preserving genetic diversity and
109 population connectivity, and for assessing the potential contributions of new management areas.

110 Expanding coverage over environmental space provides increased power to estimate
111 response curves that capture the relationships between a species and its environment (Shaffer et
112 al., 2022; Wang & Bradburd, 2014), including the detection of ecologically important patterns of
113 genetic variation and the environmental drivers of that variation (Manel et al., 2012; Selmoni et
114 al., 2020). This increases the likelihood of capturing alleles involved in local adaptation,
115 particularly if sampling includes environments near the edge of species' tolerance limits
116 (Lotterhos & Whitlock, 2014; Rellstab et al., 2015; Storfer et al., 2018; Stucki et al., 2017).
117 These alleles may play important roles in resilience to future climate change, making them
118 important targets for conservation action. Even analyses that identify environmental drivers of
119 neutral genetic variation can contribute important information on which environmental axes are
120 important for maintaining adaptive potential and how environmental variation influences genetic
121 structure and population connectivity.

122 **2.2 | Greater spatial resolution**

123 Individual-based sampling schemes also allow for greater sampling density, relative to
124 population-based schemes, providing finer spatial resolution for inferences of population
125 structure, gene flow, and the distribution of genetic diversity (Balkenhol & Fortin, 2015; Manel
126 et al., 2003; Shaffer et al., 2022). By reducing the gaps between sampling localities, individual-
127 based sampling designs allow for more accurate identification of where population genetic

128 elements, like genetic breaks or corridors of connectivity, occur on the landscape. By doing so,
129 we can also understand how population genetic elements intersect with landscape features, like
130 habitat transitions and potential physical barriers to gene flow. Homing in on where population
131 boundaries, corridors of connectivity, and regions supporting greater genetic diversity are found
132 can assist conservation practitioners and land managers in determining which areas are most
133 valuable for designating critical habitat and which contribute most to enabling population
134 connectivity and persistence. This is especially important because many land management
135 decisions take place over relatively small areas and prioritizing conservation efforts requires
136 understanding the precise relationships between landscape elements and biodiversity.

137 **2.3 | Lower impact on populations**

138 Many species are of conservation concern because they have small or heavily fragmented
139 populations, which puts them at greater risk from population fluctuations or disturbances. In
140 trying to understand which species or populations to prioritize for conservation management,
141 researchers must strike a balance between collecting adequate sample sizes to obtain crucial
142 information while limiting the overall impacts of sampling on population health. In the most
143 extreme cases, population-based sampling may not even be feasible for rare or highly imperiled
144 species because finding sufficient numbers of individuals may prove prohibitively difficult
145 (Supple & Shapiro, 2018). By sampling only one, or a few, individuals from each locality,
146 individual-based sampling minimizes the impact of sampling efforts on natural populations,
147 conveying clear benefits for population health and sustainability and even enabling studies of
148 some highly imperiled species that would otherwise be impossible. Sampling single individuals
149 can also reduce the effect of repeated sampling when species management plans call for repeated
150 genetic monitoring over time.

151 **2.4 | Potential downsides of individual-based sampling**

152 Because many population genetic analysis methods were developed for a traditional
153 population-based framework, their application to individual-based sampling can present certain
154 challenges. At a minimum, knowing which population genetic methods can be applied to
155 individual-based sampling requires understanding the assumptions inherent to each method, such
156 as whether they assume normal distributions of allele frequencies per locality or whether they
157 use population designation as an informative prior. Some methods can also be sensitive to the
158 metrics used as inputs, and several common metrics of genetic distance and diversity cannot be
159 calculated on an individual basis (e.g., *F*-statistics). However, there are reliable metrics that can
160 be calculated from individual-based sampling that should not bias results (e.g., Shirk et al.,
161 2017), several of which can be calculated in our pipeline (ALGATR).

162 Sampling fewer individuals per location also increases the binomial sampling noise
163 around estimates of local allele frequencies (Wang & Bradburd, 2014). This is consequential for
164 some methods that require precise allele frequency estimates. For example, genotype-
165 environment association (GEA) methods were developed for population-based sampling and
166 may not perform as well with allele frequency estimates based on only a single individual per a
167 locality (Rellstab et al., 2015). Although not exclusive to individual-based sampling, researchers
168 should also be cognizant of how the spatial distribution of their sampling may affect spatial
169 autocorrelation in their sampling design. Reflecting a common concern in spatial ecological
170 studies, when data points have a high degree of non-independence, analyses that do not include
171 an autoregressive component may produce inaccurate results (Dale & Fortin, 2014; Hurlbert,
172 1984).

173 Finally, individual-based sampling inherently limits the information acquired about the
174 variation present within a population. This represents a reasonable tradeoff for landscape
175 genetics, which is primarily concerned with between-population variation, and even for
176 conservation efforts that are concerned with landscape-level processes, but it does present a
177 limitation for some objectives, such as those concerned with demographic processes. For
178 example, individual-based sampling can only provide coarse data on population sizes and would
179 likely not be an optimal choice for genetic monitoring of within-population changes through
180 time. Despite these limitations, the set of methods that have assumptions compatible with
181 individual-based sampling leaves researchers with many strong analytical options.

182

183 **3 | A LANDSCAPE GENOMICS PIPELINE FOR CONSERVATION**

184 Researchers must choose from a wealth of methodological options when designing a
185 landscape genomic study. The best path will depend on a complex set of factors, ranging from
186 the goals and questions of the study to the assumptions and requirements for each potential
187 method to be employed to computational costs and limitations that balance available resources
188 and methods with dataset size. Important methodological considerations in landscape genomics
189 begin with the broad consideration of which type of analysis to include, followed by deciding
190 which method to perform, and then making decisions on parameter settings and options for that
191 specific method. In addition to methodological considerations, input data must often be
192 processed (e.g., calculating genetic or environmental distances). Understanding and prioritizing
193 tradeoffs can be challenging even for experienced landscape geneticists and especially for
194 conservation practitioners or other biologists who may be approaching landscape genetics for the

195 first time. We aim to address these issues by providing a user-friendly, accessible workflow
196 which is implemented in the ALGATR pipeline.

197 The ALGATR pipeline includes a curated set of methods that can be applied to individual-
198 based sampling, providing robust results under most realistic scenarios, and covering the core
199 areas of landscape genomics that are of particular interest for conservation. To make landscape
200 genomics methodology more approachable, we have included transparent documentation in
201 ALGATR for each method, including vignettes that provide step-by-step guidance for processing
202 input data, running analyses, producing relevant summary statistics, interpreting results, and
203 generating figures. Although ALGATR makes use of (for the most part) existing packages, we
204 have added new functionality to each of these methods and adapted, when necessary, methods to
205 accommodate individual-based sampling (Table S1).

206 Two of the main barriers to performing any analysis are ensuring input files are in the
207 correct format and determining the proper parameter settings for respective landscape genomic
208 methods; ALGATR provides functions for formatting input files and testing parameters for each
209 method included in the pipeline. Performing any of the methods in ALGATR requires only a file
210 with variant calls (i.e., a VCF file) and sampling coordinates. ALGATR also includes a set of
211 utility functions that provide users with options for customization if they so desire, including
212 functions for downloading bioclimatic data, pruning SNPs based on linkage disequilibrium,
213 imputing missing genotypes, and calculating a variety of genetic and geographic distance metrics
214 (Table S1). We provide guidance on the application of these utility functions so that users can
215 visualize and evaluate the effects of different ways of processing their data.

216 Below, we outline four key questions for conservation that landscape genomic
217 approaches can answer and describe methods for doing so that are included in the ALGATR

218 pipeline (Table 1). For each method, we describe its key components, practical considerations,
219 and relevance to conservation questions.

220 **3.1 | How do we delineate population units for management?**

221 Species management plans frequently call for the designation of evolutionarily significant
222 units, populations, or genetic groups that contain some particular value to the species as a whole
223 (Allendorf et al., 2022; Turbek et al., 2023). For example, some conservation plans target the
224 protection of populations that harbor unique genetic variants, subspecies that exhibit geographic
225 variation (e.g., Teixeira & Huber, 2021), or groups that maximize adaptive potential (Funk et al.,
226 2019). A critical first step in delimiting conservation units is inferring genetic clusters based on
227 population structure or genetic similarity.

228 Identifying genetic clusters is straightforward when individuals are structured into
229 discrete, isolated populations, but many species instead exhibit continuous distributions of
230 individuals (Bradburd & Ralph, 2019; Hohenlohe et al., 2021; Manel et al., 2003). Methods to
231 infer genetic clusters, including approaches like ADMIXTURE (Alexander et al., 2009),
232 STRUCTURE (Pritchard et al., 2000) sNMF (Frichot et al., 2014), TESS (Caye et al., 2016), and
233 conStruct (Bradburd et al., 2018), use clustering techniques (e.g., *K*-means clustering) to assign
234 individuals to genetic clusters based on how neutral genetic variation is distributed. The
235 approaches that are most apt for landscape genetics are those that are spatially explicit, often
236 using sample location as prior information. The ALGATR pipeline implements one such method,
237 TESS (Caye et al., 2016), which infers the optimal numbers of clusters and then calculates an
238 assignment probability or ancestry coefficient of each individual to each cluster. We have added
239 functions to ALGATR to generate interpolated maps of ancestry coefficients using the *autoKrige*
240 function from the AUTOMAP package (Hiemstra et al., 2009), which differs from the interpolation

241 done by default in TESS in that it produces raster maps that can then be used in downstream
242 analyses (Figure 1).

243 **3.2 | How is genetic variation distributed?**

244 Genetic diversity provides the raw material on which selection can act, allowing
245 populations or species to respond to new threats or perturbations, ranging from novel pests and
246 pathogens to ongoing climate change (Hoffmann et al., 2017). Loss of genetic diversity can lead
247 to inbreeding depression, diminished adaptive potential, and increased extinction risk, making it
248 a key target for conservation action (Kardos et al., 2021; O’Grady et al., 2004). Understanding
249 how genetic diversity is distributed across the landscape is, therefore, critical for examining the
250 drivers of genetic variation, identifying areas that harbor greater genetic diversity, and
251 quantifying regional vulnerability to environmental change (Segelbacher et al., 2010; Sommer et
252 al., 2013).

253 One challenge for individual-based landscape genomics has been how to calculate
254 metrics of genetic diversity, which are traditionally calculated at the population level, when
255 samples are distributed across the landscape without discretely-bounded clusters. ALGATR uses a
256 recently developed method for mapping local estimates of genetic diversity based on continuous
257 individual samples implemented in the R package WINGEN (Bishop et al., 2023). WINGEN
258 generates continuous maps of heterozygosity, allelic richness, and nucleotide diversity using
259 moving windows and spatial interpolation (Figure 1). WINGEN also includes options to account
260 for any unevenness in sampling design using rarefaction and allows users to customize the size
261 of the moving window and the resolution of the output maps, which are provided as rasters that
262 can be imported for further analysis in any commonly-used GIS software. These rasters thus

263 provide detailed views of genomic diversity across the landscape that can inform land
264 management strategies aimed at protecting genetic diversity.

265 **3.3 | What are the drivers of population connectivity?**

266 One of the primary goals of systematic conservation planning is to ensure that protected
267 areas promote population viability and persistence, and key to this is identifying how and where
268 critical population processes occur on the landscape (Nielsen et al., 2022; Segelbacher et al.,
269 2010). Understanding the drivers of gene flow and genetic differentiation can help conservation
270 practitioners understand the impact of habitat fragmentation, land-use conversion, and
271 environmental change on population connectivity (Hall & Beissinger, 2014; van Strien et al.,
272 2014). The population connectivity between points on a landscape can diminish when gene flow
273 is reduced over greater geographic distances, resulting in a pattern of isolation-by-distance (IBD;
274 Wright, 1943), or when gene flow is restricted between regions with increasingly different
275 environmental conditions (due to selective or non-selective mechanisms), leading to a pattern of
276 isolation-by-environment (IBE; Wang & Bradburd, 2014). The extent to which geographic and
277 environmental isolation shape genetic divergence provides important information for
278 maintaining functional connectivity, which is a key component of population viability, across the
279 landscape (Segelbacher et al., 2010).

280 Generally, landscape genomics methods that investigate IBD and IBE rely on matrix
281 regression, in which matrices of geographic and environmental distances serve as explanatory
282 variables, and the response variable is a genetic distance matrix. For fitted models, the ratios
283 between standardized regression coefficients (beta coefficients) for each explanatory variable can
284 approximate the relative contributions of each environmental variable (or aggregated
285 environmental distance) and geographic distance to explain variation in genetic distances.

286 Variable selection can be performed using backwards elimination in which explanatory variables
287 are incrementally removed and the model refitted until only variables with statistically
288 significant effects remain (Ferrier et al., 2002; Wang, 2013).

289 ALGATR implements two methods for estimating IBD and IBE: generalized dissimilarity
290 modeling (GDM; Ferrier, 2002; Ferrier et al., 2002, 2007) and multiple matrix regression with
291 randomization (MMRR; Wang, 2013) (Figure 1). GDM models turnover in the compositional
292 dissimilarity between pairs of sites by transforming explanatory variables using a set of I-spline
293 basis functions (Ferrier, 2002; Ferrier et al., 2002, 2007; Fitzpatrick & Keller, 2015). GDM's
294 main advantage is that it can account for nonlinear relationships between variables (Mokany et
295 al., 2022). The shape of the resulting relationships can then be used to identify threshold values
296 that may represent important areas of genetic turnover because the shape of the I-spline functions
297 indicate how genetic dissimilarity changes across an environmental gradient for each explanatory
298 variable (Storfer et al., 2018). The environmental data layers used in the analysis can also be
299 transformed based on these fitted relationships and combined into a map of genetic
300 compositional dissimilarity across the study region (see Figure 1).

301 Similarly, MMRR performs linear matrix regression on genetic and environmental
302 distances and allows for multiple independent variables (environmental and geographic
303 distances) to be examined simultaneously. This approach provides less flexibility than GDM
304 (Table S2) but may be less prone to overfitting, a tradeoff some researchers may prefer. MMRR
305 performs significance testing through random permutations of the rows and columns of the
306 dependent matrix, which is necessary because of the non-independence of values in pairwise
307 distance matrices. MMRR provides individual regression coefficients and *p*-values for each
308 explanatory variable and for the fitted model.

309 One remaining set of analyses in this category are those that seek to understand the extent
310 to which landscape features act to restrict gene flow, a phenomenon that results in a pattern
311 known as isolation-by-resistance (IBR; McRae, 2006). The workflow for investigating IBR
312 includes first generating a resistance surface based on prior information or a hypothesis about
313 how landscape elements contribute to differential resistance to movement for the study organism.
314 Resistance distances can then be calculated using least-cost path analysis (Wang et al., 2009;
315 Wang & Shaffer, 2017) or circuit theory analysis (Dickson et al., 2019; McRae & Beier, 2007).
316 Because the parameterization of the resistance surface is highly complex, requiring decisions that
317 often have a significant impact on the estimate of resistance distances (Koen et al., 2012; Spear
318 et al., 2010; Spear et al., 2015), and because the optimal approach to resistance surface
319 parameterization depends strongly on the study design and objectives (Peterman et al., 2019;
320 Spear et al., 2010; Zeller et al., 2012), ALGATR does not include an automated approach for
321 generating resistance surfaces. However, if a user has generated a resistance surface with their
322 preferred method, which could include parameterization based on habitat suitability models
323 (Wang et al., 2008), genetic algorithms (Peterman, 2018), or gradient forest analysis of allele
324 frequencies (Vanhove & Launey, 2023), then resistance distances can be calculated using
325 ALGATR's *geo_dist()* function, specifying "resistance" with the type argument (type =
326 "resistance"). This function uses the *commuteDistance()* function in the GDISTANCE package (van
327 Etten & Hijmans, 2010). The resulting resistance distances can then be used in downstream
328 analyses, including using them in MMRR or GDM to quantify IBR.

329 **3.4 | How can we identify and protect adaptive genetic variation?**

330 The accessibility of genomic data for natural systems brings not only greater resolution
331 but also opportunities to examine spatial patterns of adaptive genetic variation (Manel et al.,

2010; Parisod & Holderegger, 2012; Schoville et al., 2012). Identifying ecologically-important, functional genetic variation – genes that convey fitness benefits under different conditions – and the environmental forces underlying that variation provides a new dimension of valuable information for conservation efforts. Characterizing genes involved in adaptation to local environmental conditions can help guide local reintroduction efforts, assess the feasibility of genomic rescue for genetically depauperate populations, evaluate strategies for assisted gene flow, and quantify climate resilience and vulnerability (Browne et al., 2019; Frankham et al., 2014; Seaborn et al., 2021; Thurman et al., 2020).

A critical step for these objectives is to identify genes associated with specific environmental variables. Genotype-environment association (GEA) analyses quantify statistical associations between allele frequencies and environmental variables to test the hypothesis that allelic variation at a locus reflects adaptation to the local environment (Capblancq et al., 2018; Lotterhos, 2023). Because spurious genotype-environment associations could result from a variety of factors other than local adaptation, including population structure and demography, the signal of environmental selection must be parsed from the background level of neutral divergence resulting from population structure (Ahrens et al., 2018; De Mita et al., 2013; Lotterhos & Whitlock, 2014, 2015; Rellstab et al., 2015; Storfer et al., 2018). Approaches for doing so include using latent factors to represent population structure (Caye et al., 2019; Rellstab et al., 2015), conditioning on neutral variation using principal components (Duforet-Frebourg et al., 2016), using Moran eigenvector maps to decompose spatial relationships (Forester et al., 2018), and identifying sets of loci that are putatively neutral and incorporating them as covariates in the model (Dauphin et al., 2022; Meirmans, 2015; Storfer et al., 2018).

354 Two GEA methods are included in ALGATR, latent factor mixed modeling (LFMM; Caye
355 et al., 2019) and redundancy analysis (RDA; Figure 1). LFMM is a univariate method that uses
356 latent factors to account for unobserved variables, including population structure (Caye et al.,
357 2019). This approach is advantageous because it provides controls on factors, other than
358 selection, that may incidentally covary with allele frequencies. The challenge is determining the
359 appropriate number of latent factors for any given dataset (called K values). ALGATR provides
360 four options for doing so: a Tracy-Widom test (Frichot et al., 2013), a “quick elbow” test (an
361 approach similar to examining a scree plot), the TESS clustering algorithm (in which latent
362 factors correspond to some measure of population structure), or K-means clustering (Jombart et
363 al., 2010).

364 RDA is a constrained ordination method that models linear, multivariate relationships
365 (Rellstab et al., 2015). By simultaneously testing multiple loci against multiple environmental
366 variables, RDA is able to detect multilocus selection (Forester et al., 2018; Rellstab et al., 2015)
367 while also accounting for covariation in allele frequencies between loci. Optionally, RDA can
368 perform variable selection using forward selection until a specified threshold is met (typically
369 using a permutation test and adjusted R^2 values). The RDA method can also utilize covariates,
370 such as population structure or geographic distance, alongside environmental predictors, an
371 approach known as a partial RDA (pRDA). Finally, variance partitioning can be performed using
372 RDA to quantify the independent contributions of each explanatory variable as well as variation
373 that is explained by a combination of explanatory variables (i.e., confounded variance;
374 Capblancq & Forester, 2021). ALGATR includes RDA and pRDA with and without variance
375 partitioning, and it implements two approaches for determining which loci should be considered

376 significant outliers, one that uses Z-scores (Forester et al., 2018) and another that transforms
377 RDA loadings into *p*-values (Capblancq, Morin, et al., 2020; Capblancq & Forester, 2021).

378 **3.5 | Caveats**

379 Like any analysis, landscape genomic methods have various limitations and assumptions,
380 and even preparing data for these methods carries some potential pitfalls. Below, we briefly
381 discuss some potentially important concerns that may arise during landscape genomic analysis
382 and describe how options and solutions for addressing these issues can be implemented (see also
383 Table S2). The ALGATR documentation aims to provide all of the information necessary to run
384 each of the analyses and guidance on the decisions that must be made for the options offered by
385 each method. It also provides references to several extensive and recent reviews that explore the
386 methodology, assumptions, and theory underlying many of these methods (e.g., Capblancq,
387 Fitzpatrick, et al., 2020; Capblancq & Forester, 2021; Fenderson et al., 2020; Forester et al.,
388 2018; Lotterhos, 2023; Rellstab et al., 2015). The package itself also contains tools that help to
389 evaluate different options for each method implemented in the pipeline.

390 GDM and MMRR take in genetic, environmental, and geographic distance matrices as
391 input – various distance metrics are available – and the choice of metric may influence
392 downstream results. Different genetic distance metrics, in particular, may result in different
393 outcomes. ALGATR provides options to calculate several metrics of genetic, geographic, and
394 environmental distances, and users should consider and test the potential impacts of different
395 metrics on their results (Shirk et al., 2017; Wang, 2020).

396 Several landscape genomic methods, including RDA, do not allow for any missing
397 genotypes, meaning that studies must either remove loci containing missing data or perform
398 imputation to fill in missing genotypes. Because removing sites containing missing data often

399 results in greatly reduced dataset size, imputation is commonly performed to maximize data
400 retention. Different types of imputation exist, including population structure-based imputation
401 (Caye et al., 2016), maximum likelihood-based imputation (D'Angelo et al., 2010), or imputation
402 based on the mean, median, or most common genotype at each site (e.g., Capblancq & Forester,
403 2021), each of which introduces different assumptions (e.g., Money et al., 2015; Shi et al., 2018;
404 Yi & Latch, 2022). To deal with missing values, ALGATR performs two types of imputation. The
405 first is a per site median-based imputation (similar to that of Capblancq & Forester, 2021),
406 although we do not recommend this simplified approach for general use because artificially
407 inflated p -values can result if missing values are non-random (i.e., if there is allelic bias in
408 missing data). We have also implemented a population structure-based imputation method that
409 uses non-negative matrix factorization (sNMF; Frichot et al., 2014)) to assign missing values.
410 This utilizes functions in the LEA package (Frichot & François, 2015) and provides a more
411 sophisticated (albeit computationally slower) imputation method compared to the median-based
412 approach. As with many similar population structure methods, it also requires a user-selected
413 value for K clusters.

414 Researchers should keep in mind that determining the number of K values that best
415 describe their data should be done with care. This is relevant to selecting the number of latent
416 factors (for LFMM) and K clusters (for TESS). For TESS, ALGATR can perform manual and
417 automatic K -selection. Automatic K -selection is provided mainly for simulation studies or meta-
418 analyses where K -values have to be selected for a large quantity of datasets such that manual
419 selection is not feasible. For LFMM, ALGATR implements four methods for automatically
420 selecting the number of latent factors (also represented using the term K): by performing a

421 Tracy-Widom test, a “quick elbow” test, using cross-validation scores from TESS, or K-means
422 clustering using adegenet’s *find.clusters* function.

423 Linkage disequilibrium (LD) results in collinearity among SNPs, a pattern that can
424 misinform landscape genomic analyses. For example, estimates of population structure can be
425 overinflated as more SNPs appear to independently support the same pattern (Rellstab et al.,
426 2015). This is, fortunately, fairly straightforward to address by pruning sites that are putatively in
427 LD. A common approach involves calculating correlations between SNPs in sliding windows of
428 custom sizes across the genome (Ahrens et al., 2018) which can be done before or after
429 performing the GEA analysis (Capblancq & Forester, 2021). ALGATR performs LD-pruning using
430 the SNPRELATE package (Zheng et al., 2012), which also avoids the need for phased input data,
431 and provides options to specify window size, window overlap, and the LD threshold.

432 **4 | CONCLUSIONS**

433 The implementation of genetically-informed conservation actions is critical to understand
434 how best to conserve biodiversity. Landscape genomic approaches can provide important insight
435 into how genetic variation is spatially distributed, allowing conservation practitioners to
436 prioritize areas for conservation efforts. Genomic-scale data have enabled a shift towards
437 individual- rather than population-based sampling because they provide more genetic resolution
438 than was previously feasible. This shift towards individual-based sampling reduces the impact on
439 populations that may already be in rapid decline, while also allowing researchers to capture
440 greater geographic and environmental coverage for their datasets and to achieve higher
441 resolution results.

442 Individual-based sampling provides a number of distinct benefits for biodiversity
443 conservation, and landscape genomic approaches are poised to seize upon these advantages to

444 provide actionable information for conservation and management efforts. Our R package,
445 ALGATR, provides an easily accessible and user-friendly pipeline that uses individual-based
446 genomic datasets to provide fine-scale genomic and spatial resolution to answer fundamental
447 conservation questions. Landscape genomics provides an ever-increasing resolution into the
448 processes and patterns that shape the genetic variation of Earth’s biodiversity, thereby increasing
449 our understanding of how best to protect it.

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461 **AUTHOR CONTRIBUTIONS**

462 All authors conceived of the study; APB and EAC wrote the ALGATR package, with
463 contributions from IJW; EAC wrote and APB co-wrote the manuscript; all authors reviewed and
464 edited the manuscript.

465 **CONFLICT OF INTEREST**

466 The authors declare that they have no competing interests.

467 **DATA AVAILABILITY STATEMENT**

468 The ALGATR package can be installed from GitHub:

469 <https://github.com/TheWangLab/algatr>. A containerized version of the package can also be run

470 using Docker; instructions for its installation can be found in the package's README on

471 GitHub.

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791 **TABLES**

792 **TABLE 1** Summary of key conservation-oriented questions that our landscape genomic analysis pipeline (ALGATR) can answer,
 793 including input requirements for the ALGATR package and recent empirical examples of each method.
 794

Question	Analysis category	Method	Input requirements	Empirical examples
How do we delineate population units for management?	Population structure	TESS (Caye et al., 2016)	Genotype dosage matrix (preferably LD-pruned), sampling coordinates, <i>optional: raster layer for mapping</i>	Ogbonna et al., 2021
How is genetic variation distributed?	Genetic diversity	wingen (Bishop et al., 2023)	VCF, sampling coordinates, <i>optional: raster layer for mapping</i>	
What are the drivers of population connectivity?	Isolation by distance and isolation by environment	MMRR (Wang, 2013)	Genetic distance matrix, environmental distance matrices or environmental layers plus sampling coordinates	Ortego et al., 2015; Sexton et al., 2016; Zhang et al., 2016
		GDM (Ferrier et al., 2007; Fitzpatrick & Keller, 2015; Freedman et al., 2010)		Bay et al., 2018; Medina et al., 2021; Shryock et al., 2015; Wogan et al., 2020
How can we identify and protect adaptive genetic variation?	Genotype-environment association methods	RDA (Capblancq & Forester, 2021)	Genotype dosage matrix, environmental layers, sampling coordinates	Forester et al., 2022
		LFMM (Caye et al., 2019)		Carvalho et al., 2021; Cortellari et al., 2021

795

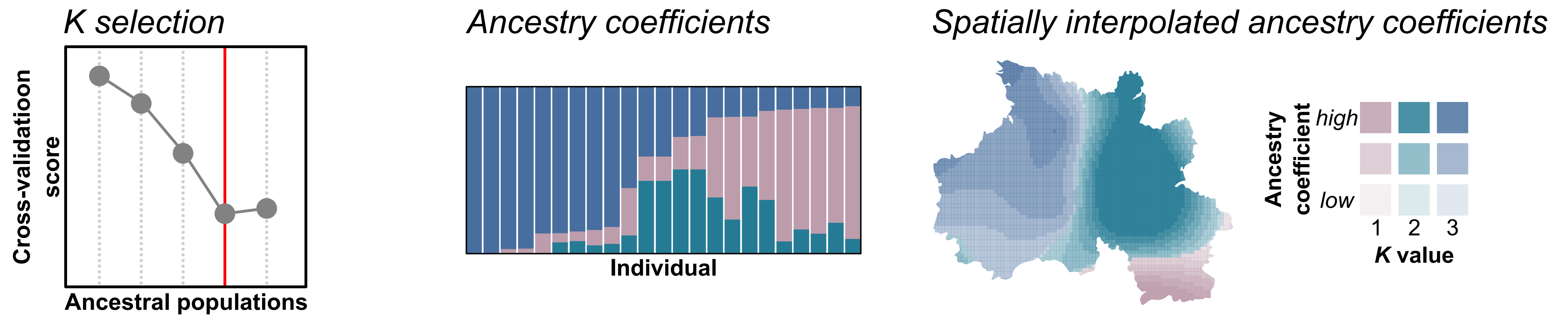
796

FIGURE LEGENDS

797 FIGURE 1 Summary of outputs produced by the ALGATR pipeline with suggested methods to
798 answer conservation questions. Methods included in ALGATR's workflow are TESS (Caye et al.,
799 2016), wingen (Bishop et al., 2023), multiple matrix regression with randomization (MMRR;
800 Wang, 2013), generalized dissimilarity modeling (GDM; Ferrier et al., 2007; Fitzpatrick &
801 Keller, 2015; Freedman et al., 2010), redundancy analysis (RDA; Capblancq & Forester, 2021),
802 and latent factor mixed modeling (LFMM; Caye et al., 2019).

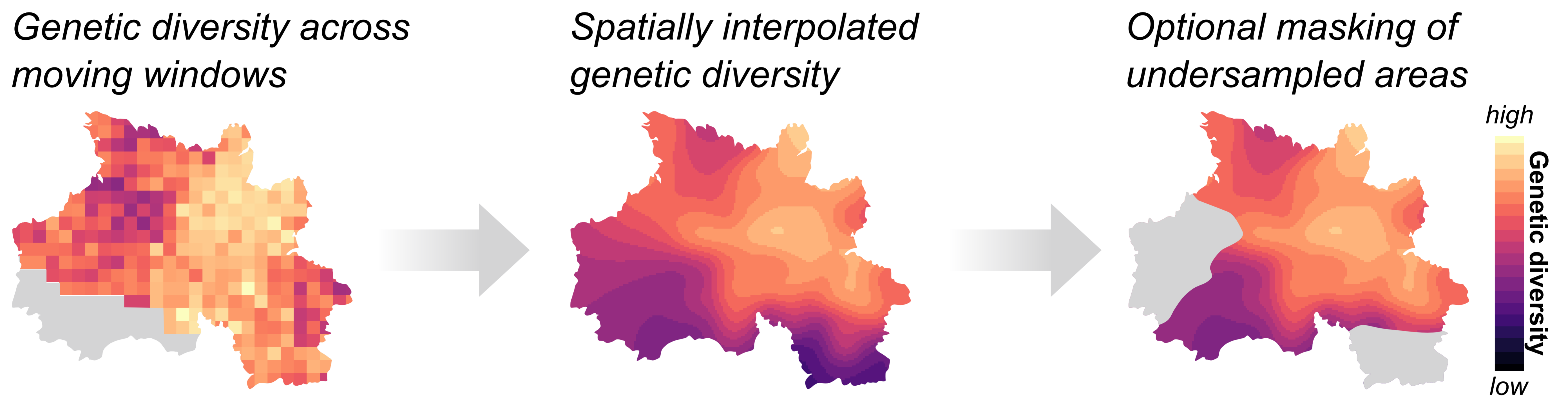
How do we delineate population units for management?

TESS



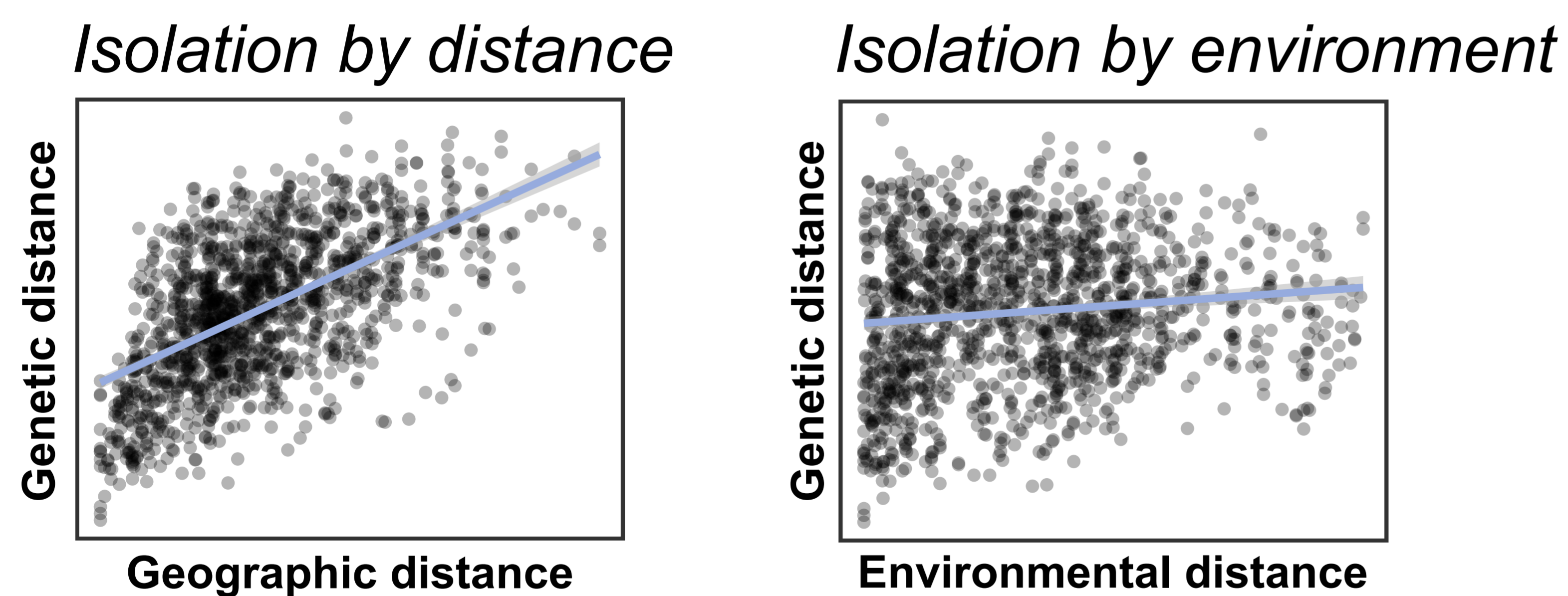
How is genetic variation distributed?

wingen

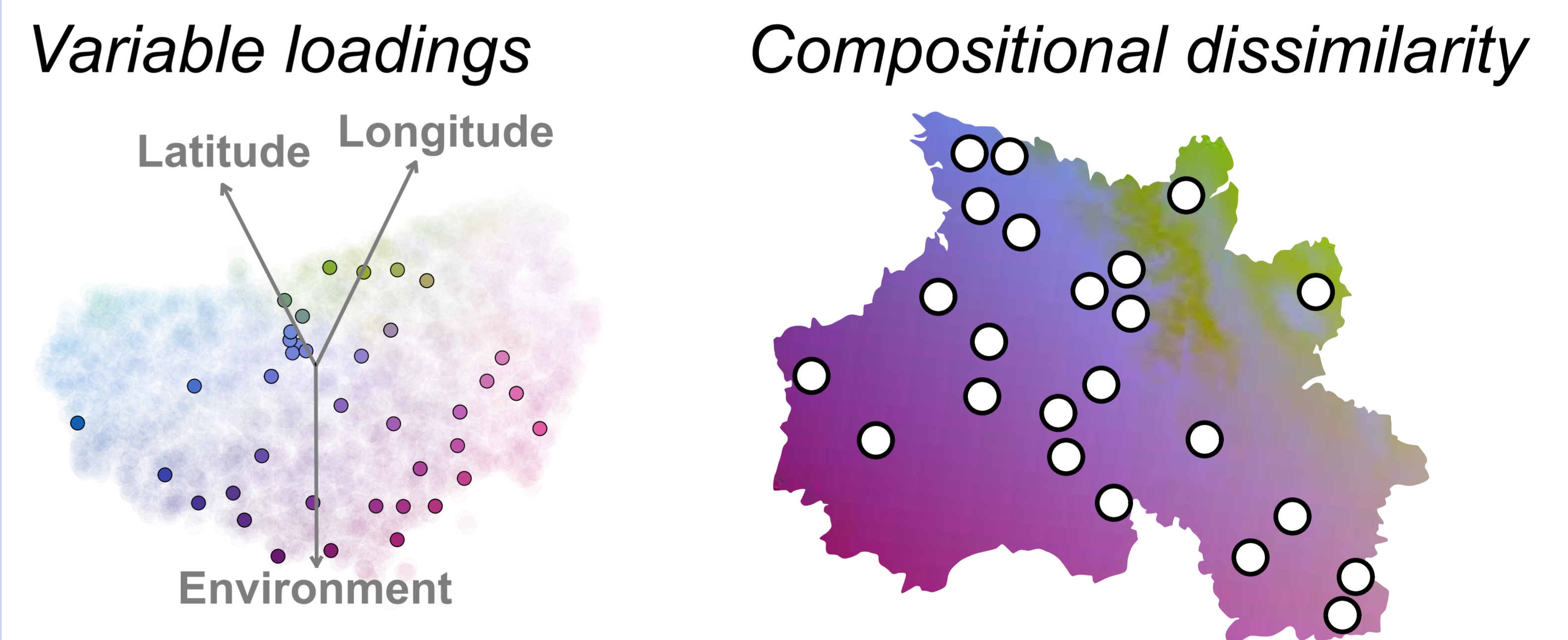


What are the drivers of population connectivity?

MMRR

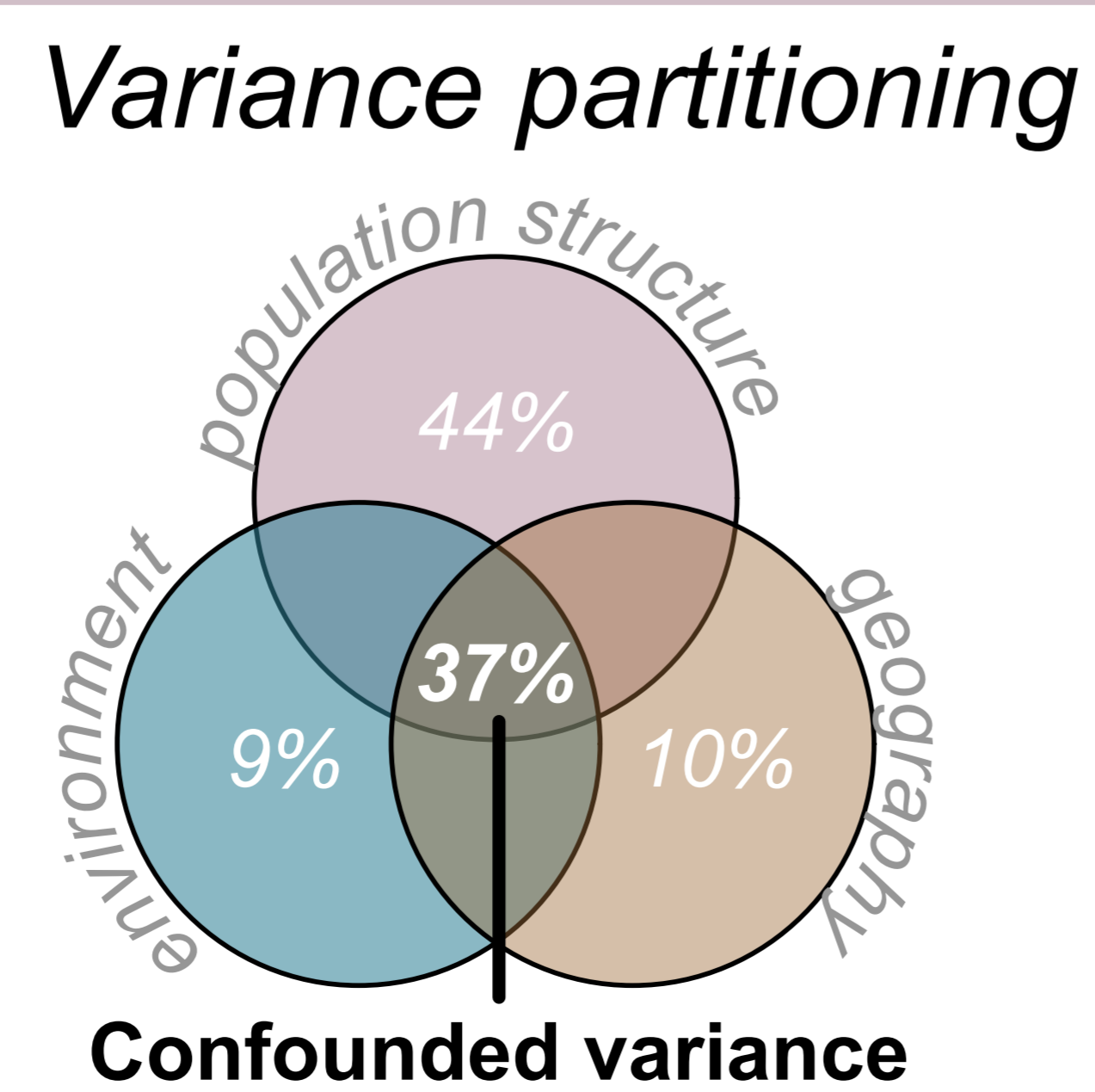
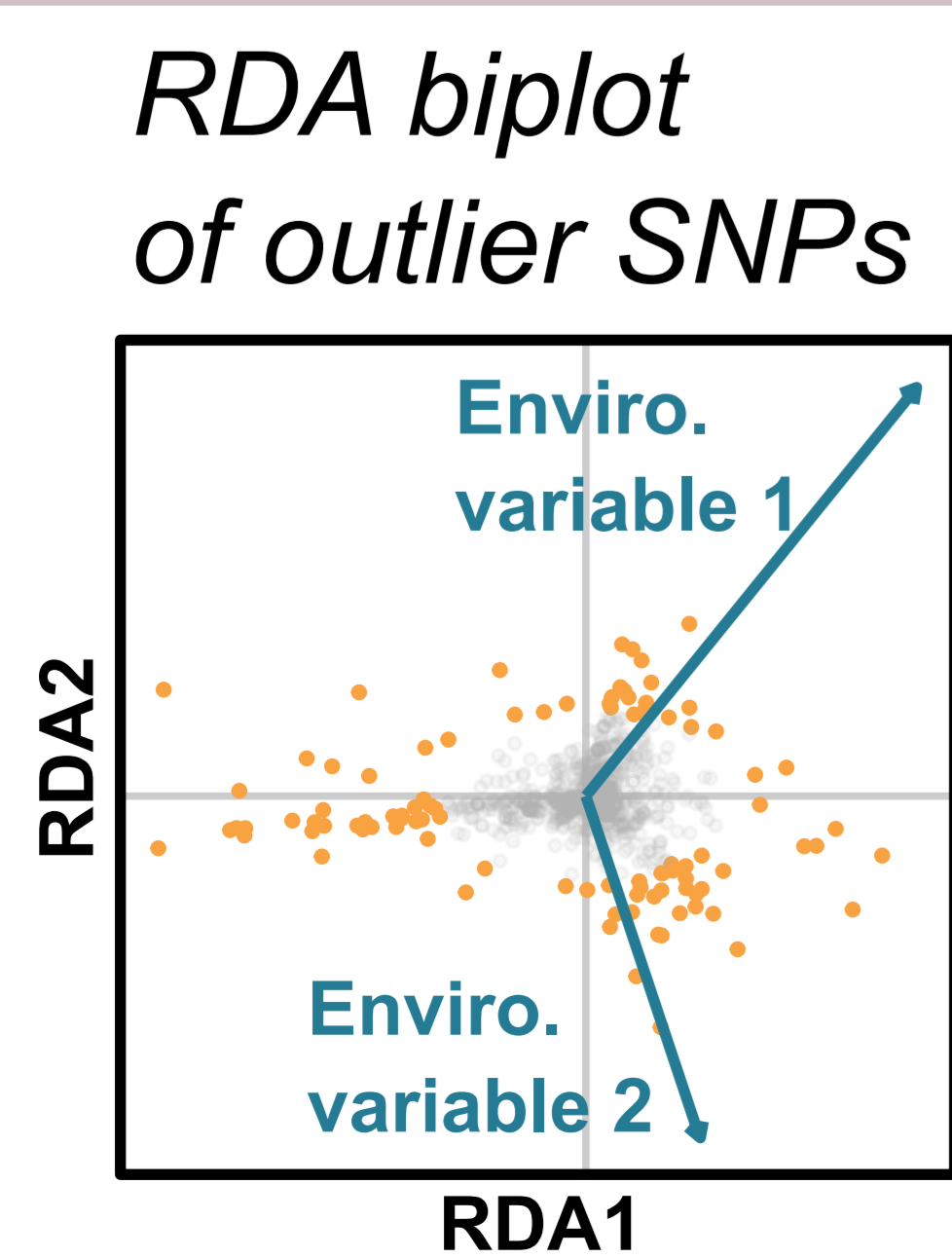


GDM



How can we identify and protect adaptive genetic variation?

RDA



LFMM

