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## Dynamic regulation of water potential in *Juniperus osteosperma* mediates ecosystem carbon fluxes

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Complete List of Authors:	Guo, Jessica; The University of Arizona, Arizona Experiment Station Barnes, Mallory; Indiana University, O'Neill School of Public and Environmental Affairs Smith, William; University of Arizona, School of Natural Resources and the Environment Anderegg, William; University of Utah, School of Biological Sciences; The University of Utah, Wilkes Center for Climate Science and Policy Kannenberg, Steven; Colorado State University, Department of Biology; Colorado State University, Graduate Degree Program in Ecology; West Virginia University, Biology
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Manuscripts

# 1 Dynamic regulation of water potential in *Juniperus osteosperma* mediates ecosystem carbon 2 fluxes

3 Jessica S. Guo

4 Mallory L. Barnes

5 William K. Smith

6 William R.L. Anderegg

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9      **Summary**

10     • Some plants exhibit dynamic hydraulic regulation, in which the strictness of hydraulic  
11     regulation (i.e., iso/anisohydry) changes in response to environmental conditions.  
12     However, the environmental controls over iso/anisohydry and the implications of flexible  
13     hydraulic regulation for plant productivity remain unknown.

14     • In *Juniperus osteosperma*, a drought-resistant dryland conifer, we collected a 5-month  
15     growing season timeseries of *in situ*, high temporal-resolution plant water potential ( $\Psi$ )  
16     and stand gross primary productivity (GPP). We quantified the stringency of hydraulic  
17     regulation associated with environmental covariates and evaluated how predawn water  
18     potential contributes to empirically predicting carbon uptake.

19     • *J. osteosperma* showed less stringent hydraulic regulation (more anisohydric) after  
20     monsoon precipitation pulses, when soil moisture and atmospheric demand were high,  
21     and corresponded with GPP pulses. Predawn water potential matched the timing of GPP  
22     fluxes and improved estimates of GPP more strongly than soil and/or atmospheric  
23     moisture, notably resolving GPP underestimation prior to vegetation green-up.

24     • Flexible hydraulic regulation appears to allow *J. osteosperma* to prolong soil water  
25     extraction and therefore the period of high carbon uptake following monsoon  
26     precipitation pulses. Water potential and its dynamic regulation may account for why  
27     process-based and empirical models commonly underestimate the magnitude and  
28     temporal variability of dryland GPP.

29     Keywords: carbon uptake, dryland ecosystem, hydraulic regulation, juniper woodland,  
30     iso/anisohydry, precipitation pulse dynamics, stem water potential

31

32 **Introduction**

33 Along the soil-plant-atmosphere continuum, gradients of water potential ( $\Psi$ ) drive water  
34 transport and govern the tradeoff between obtaining carbon dioxide for photosynthesis and water  
35 loss through stomata (Berry *et al.*, 2010). The concept of a ‘plant water use strategy’  
36 encompasses the numerous ways plants have evolved to confront this inescapable dilemma,  
37 including the prevalent iso/anisohydry spectrum based on the stomatal regulation of  $\Psi$  (Jones,  
38 1998; Tardieu & Simonneau, 1998). Isohydry describes a conservative stomatal strategy to  
39 minimize reductions in  $\Psi$  and preserve hydraulic conductivity, while anisohydry is a profligate  
40 stomatal strategy that prioritizes carbon gain at the expense of low  $\Psi$ . The degree of  
41 iso/anisohydry describes plant strategy in response to declining soil moisture absent other  
42 limiting factors (Novick *et al.*, 2019) and is generally operationalized as a species-level and  
43 theoretical trait. However, recent work has demonstrated that these strategies can be quite  
44 variable within a species and may arise from plant-environment interactions (Hochberg *et al.*,  
45 2018), including vapor pressure deficit (VPD), which is often decoupled from soil moisture at  
46 short timescales (Novick *et al.*, 2016). Within-species shifts in iso/anisohydry have been  
47 observed for *Larrea tridentata* (Guo *et al.*, 2020) and *Quercus douglasii* (Feng *et al.*, 2019)  
48 during different seasons, in *Quercus suber* in response to competition (Haberstroh *et al.*, 2022),  
49 *Acacia aptaneura* as a result of repeated experimental drought (Nolan *et al.*, 2017), and in  
50 multiple species between wet and dry years (Wu *et al.*, 2021).

51 The implications of temporally-variable hydraulic strategies on ecosystem carbon (C) fluxes  
52 have not been fully elucidated. This knowledge gap may limit accurate modeling of carbon and  
53 water cycle dynamics, which in turn restricts our ability to predict and mitigate climate change  
54 impacts (Kennedy *et al.*, 2019a; Novick *et al.*, 2022). Particularly in dryland ecosystems,  
55 persistent water limitation and episodic precipitation promote tight coupling between carbon and  
56 water cycles (Biederman *et al.*, 2016), resulting in added temporal complexity that can be  
57 difficult to model (Noy-Meir, 1973; Loik *et al.*, 2004; Ogle & Reynolds, 2004; Feldman *et al.*,  
58 2018). Dryland ecosystems are largely responsible for the interannual variability of the global  
59 carbon sink (Poulter *et al.*, 2014; Ahlström *et al.*, 2015), yet dynamic global vegetation models  
60 have been found to significantly underestimate the interannual variability of C uptake in dryland  
61 regions (Biederman *et al.*, 2017; MacBean *et al.*, 2021a). Understanding the temporal dynamics

62 and environmental sensitivity of plant hydraulic strategies may be critical to improving  
63 predictive forecasts of the global carbon cycle (Eller *et al.*, 2020; Sabot *et al.*, 2020, 2022).

64 Despite its importance, plant hydraulic stress is often notably absent from large-scale estimates  
65 of ecosystem productivity (Smith *et al.*, 2019). Such models commonly combine remotely-  
66 sensed indices of vegetation greenness and light use efficiency [LUE; Running *et al.* (2004);  
67 Zeng *et al.* (2022)], defined as the slope of the relationship between biomass and cumulative  
68 intercepted photosynthetically active radiation (Monteith *et al.*, 1977). Greenness indices can  
69 represent the structural capacity for photosynthesis on a seasonal basis (Wang *et al.*, 2022), but  
70 do not capture the sub-daily constraints imposed by soil and atmospheric drought, such that  
71 productivity seasonality is much weaker in remotely-sensed than tower-based fluxes (Garbulsky  
72 *et al.*, 2008; Biedermaier *et al.*, 2017; Smith *et al.*, 2019; Pierrat *et al.*, 2021). Instead, water stress  
73 effects are typically incorporated into estimates of LUE using moisture scalars derived from  
74 estimates of VPD [e.g., MODIS LUE; Zhao & Running (2010)], remotely sensed vegetation or  
75 evaporative indices [e.g., eddy covariance; (EC)-LUE model Yuan *et al.* (2007)], or combined  
76 VPD and soil moisture [e.g., CFLUX; King *et al.* (2011)]. However, the range of  
77 ecophysiological responses to moisture stress are too complex for a single environmentally-  
78 derived indicator or function to adequately represent (Zhang *et al.*, 2015). Plant water potential, a  
79 direct metric of plant water stress that integrates soil and atmospheric drivers, may thus provide a  
80 key physiological constraint on ecosystem productivity, which could improve our ability to  
81 represent drought impacts and quantify interannual variability of C uptake.

82 Pinyon-juniper woodlands are broadly distributed in the southwestern United States and provide  
83 a well-studied test system for how hydraulic strategies like iso/anisohydry can modulate  
84 productivity and mortality (McDowell *et al.*, 2008). Pinyon mortality following the 2002-2003  
85 drought was likely associated with differences in plant hydraulic regulation (Breshears *et al.*,  
86 2009; Plaut *et al.*, 2012); juniper survival was largely attributed to a less hydraulically vulnerable  
87 xylem and thus a greater ability to withstand low water potentials (McDowell *et al.*, 2008).  
88 Although generally considered anisohydric, *Juniperus monosperma* exhibited strong stomatal  
89 control and negligible xylem embolism under drought manipulation (Garcia-Forner *et al.*,  
90 2016b), thereby challenging the hypothesis that anisohydric species are more prone to hydraulic  
91 failure. As the southwestern US megadrought persists (Williams *et al.*, 2022) and induces

92 mortality even among resilient *Juniperus* spp. (Kannenberg *et al.*, 2021), it is imperative to  
93 examine how flexible hydraulic strategies interact with plant productivity and survival.

94 In this study, we utilize a five-month time-series of plant  $\Psi$  and gross primary productivity  
95 (GPP) in a juniper woodland to evaluate the temporal dynamics of hydraulic strategy and  
96 incorporate plant water stress into a common GPP framework. Previous work by Guo *et al.*  
97 (2020) examined dynamic hydraulic strategy in *Larrea tridentata* but lacked a co-located  
98 timeseries of ecosystem carbon fluxes. By contemporaneously measuring plant  $\Psi$  and GPP  
99 continuously at daily resolution, we can directly investigate the implications of  $\Psi$  regulation and  
100 hydraulic status for productivity in an iconic southwestern species. We ask:

101 1) Does plant hydraulic regulation vary over time in *J. osteosperma*?  
102 2) How are temporal patterns in hydraulic regulation related to GPP over a growing season?  
103 3) Can GPP prediction be enhanced by plant water potential?

104 **Materials and Methods**

105 This study was conducted at an early-successional pinyon-juniper woodland (37.5241 N,  
106 109.7471 W, 1866 m a.s.l.) in southeastern Utah. Local climate conditions include cold winters  
107 and hot, dry summers, with high interannual variability in summer precipitation due to its  
108 location at the northern boundary of the North American Monsoon. The locally flat topography  
109 is dominated by Utah juniper (*Juniperus osteosperma*, 92% tree basal area) and two-needle  
110 pinyon (*Pinus edulis*, 8% tree basal area), with sparse understory comprising big sagebrush  
111 (*Artemisia tridentata*), prickly pear cactus (*Opuntia* spp.), and bunchgrasses. Mean growing  
112 season leaf area index was 0.4, and the site was chained in the 1960s, resulting in a relatively  
113 even-aged and sized population of *J. osteosperma*. See Kannenberg *et al.* (2023) for further site  
114 description and processing of eddy covariance variables.

115 *Plant water potential*

116 Stem water potential of seven mature *J. osteosperma* within the tower footprint (< 20 m) was  
117 monitored with both automated and manual measurements between May 24 and November 5,  
118 2021. Half-hourly water potential was monitored with stem psychrometers (ICT International  
119 PSY1) calibrated prior to installation. Two instruments per tree were installed by removing the

120 bark and phloem to expose a flat xylem surface. Psychrometer sensor heads were attached with  
 121 self-adhesive silicone tape to maintain a tight seal and wrapped in reflective insulation to  
 122 minimize temperature gradients. Because plant wounding responses can fill the sensor chamber,  
 123 each psychrometer was uninstalled, cleaned with chloroform, and reinstalled on a new branch  
 124 every 4-5 weeks. The day after reinstallation, the xylem water potential was measured manually  
 125 with a Scholander-type pressure chamber (PMS 610) by excising a needle cluster with diameter  
 126 between 2 and 4 mm and measuring within 2 minutes of collection; psychrometer water  
 127 potentials generally matched pressure chamber values (Kannenberg *et al.*, 2023).

128 The half-hourly stem water potential time series were subjected to quality control by visual  
 129 assessment and aggregated to daily values. After removing data during the maintenance period  
 130 (+ 1 day) and outliers that were > 0.5 MPa from adjacent points, data that met the following  
 131 criteria were also discarded: 1) a step change in the magnitude of water potential not attributable  
 132 to a precipitation event; 2) loss of diurnal pattern in water potential. On average, data from 10  
 133 out of 14 psychrometers were available during a given period. Half-hourly stem water potential  
 134 was summarized to predawn ( $\Psi_{PD}$ , 2 hours prior to sunrise) and midday ( $\Psi_{MD}$ , 2 hours following  
 135 solar noon) for each logger. In addition, site-level means of predawn and midday water potential  
 136 were calculated and missing values (10 and 8, respectively) were imputed using Kalman  
 137 Smoothing via the R package ‘imputeTS’ (Moritz & Bartz-Beielstein, 2017).

138 *Vegetation indices and fAPAR*

139 We adopted a Monteith light use efficiency framework (Monteith, 1972) to estimate plant  
 140 productivity. This framework conceptualizes GPP as the product of absorbed photosynthetically  
 141 active radiation (APAR) and the efficiency with which light is converted to fixed carbon (LUE).  
 142 APAR is represented as a product of photosynthetically active radiation (PAR) and the fraction  
 143 of PAR absorbed by plant canopies (fAPAR). The foundational equation from Monteith’s  
 144 framework can be expressed as:

145 
$$GPP = LUE \cdot PAR \cdot fAPAR \quad (1)$$

146 Many current models for estimating GPP are grounded in this framework or its variations. Here,  
 147 we used the near-infrared reflectance of vegetation index (NIRv Badgley *et al.*, 2017, 2019)

148 as a proxy for fAPAR. This choice was informed by the strong correlation between NIRv and  
 149 modeled fAPAR across various soil reflectances and its robustness at low vegetation cover  
 150 (Badgley *et al.*, 2017; Wang *et al.*, 2022).

151 We calculated NIRv from Moderate Resolution Imaging Spectroradiometer (MODIS) nadir  
 152 bidirectional reflectance distribution function adjusted daily reflectance product (MCD43A, 1 d,  
 153 500 m, collection 6.1) using the point extraction tool AppEARS. MODIS bands 1 (620-670 nm)  
 154 and 2 (841-876 nm) were combined with background soil reflectance of 0.08 to represent NIRv  
 155 following Badgley *et al.* (2017). The pixel containing the coordinates of US-CdM was filtered to  
 156 include only the highest quality observations (MODIS quality flag = 0). Resulting values were  
 157 smoothed using a Savitzky-Golay filter of derivative order 0, filter order 3, and window length 5.

158 *Model description - hydraulic regulation*

159 To specify the hydraulic regulation model, we used the Martinez-Vilalta *et al.* (2014) equation to  
 160 relate  $\Psi_{MD}$  to  $\Psi_{PD}$ :

161 
$$\Psi_{MD} = \sigma \cdot \Psi_{PD} + \lambda \quad (2)$$

162 where  $\sigma$  represents the stringency of hydraulic regulation and  $\lambda$  describes the pressure drop when  
 163 soil moisture is not limiting. Plant hydraulic regulation can be described as isohydry if  $\sigma < 1$ ,  
 164 anisohydry if  $\sigma \approx 1$ , and extreme anisohydry if  $\sigma > 1$  (Martínez-Vilalta *et al.*, 2014).

165 To allow hydraulic regulation and GPP to vary over the growing season, we specified a  
 166 hierarchical Bayesian model similar to Guo *et al.* (2020), which estimated  $\sigma$  and  $\lambda$  as linear  
 167 functions of environmental drivers. Here, we used maximum daily VPD (D) and volumetric soil  
 168 water content at 10 cm ( $W_{10}$ ), which had the highest correlation with plant  $\Psi$  and GPP  
 169 (Kannenberg *et al.*, 2023). Furthermore, we implemented the stochastic antecedent model (Ogle  
 170 *et al.*, 2015) to quantify the influence of past environmental conditions. The data model for  
 171 hydraulic regulation describes the likelihood of each observed  $\Psi_{MD}$ , which was normally  
 172 distributed for each observation  $i$  ( $i = 1, 2, \dots, 1425$ ):

173 
$$\Psi_{MDi} \sim \text{Normal}(\hat{\Psi}_{MDi}, \sigma_{\Psi}^2) \quad (3)$$

174 where  $\hat{\Psi}_{MDi}$  is the predicted or mean midday water potential and  $\sigma_{\hat{\Psi}}^2$  represents the observation  
 175 variance.  $\hat{\Psi}_{MDi}$  was modeled according to Eqn. 3, where all terms were allowed to vary over  
 176 time, either as direct observations ( $\Psi_{MD}, \Psi_{PD}$ ) or as modeled parameters ( $\sigma, \lambda$ ). The time-varying  
 177 estimates of hydraulic regulation,  $\sigma$  and  $\lambda$  were indexed by  $i$  and modeled as linear combinations  
 178 of two antecedent covariates and their interaction:

$$179 \quad \begin{aligned} \sigma_i &= \beta_0 + \beta_1 \cdot D_i^{ant} + \beta_2 \cdot W_{10i}^{ant} + \beta_3 \cdot D_i^{ant} \cdot W_{10i}^{ant} + E_{\sigma,t(i)} \\ \lambda_i &= \alpha_0 + \alpha_1 \cdot D_i^{ant} + \alpha_2 \cdot W_{10i}^{ant} + \alpha_3 \cdot D_i^{ant} \cdot W_{10i}^{ant} + E_{\lambda,t(i)} \end{aligned} \quad (4)$$

180 The  $\beta$  and  $\alpha$  parameters were estimated for all trees.  $E_{\sigma}$  and  $E_{\lambda}$  represent the random effects of  
 181 each tree, where  $t(i)$  indicates tree  $t$  associated with each observation  $i$ .  $D_{max}$  and  $W_{10}$  were  
 182 scaled using the 2021 mean and standard deviation so that regression coefficients could be  
 183 compared and  $\beta_0$  and  $\alpha_0$  could be interpreted as  $\sigma$  and  $\lambda$ , respectively, under mean environmental  
 184 conditions. Antecedent VPD ( $D^{ant}$ ) and soil water content ( $W_{10}^{ant}$ ) were constructed using daily  
 185 time series of each scaled environmental variable (Ogle *et al.*, 2015):

$$186 \quad \begin{aligned} D_i^{ant} &= \sum_{p=0}^{T_{lag}} \omega_{D_p} \cdot D_{t(i)-p} \\ W_{10i}^{ant} &= \sum_{p=0}^{T_{lag}} \omega_{W_p} \cdot W_{t(i)-p} \end{aligned} \quad (5)$$

187 where  $p$  indicates the time step,  $T_{lag}$  represents the total number of past time-steps considered,  
 188  $\omega_{D_p}$  and  $\omega_{W_p}$  indicates the weight or relative importance of the  $p$ th time step into the past, and  
 189  $D_{t(i)-p}$  and  $W_{t(i)-p}$  are the observed value of each variable at  $p$  time steps ago. Antecedent  
 190 covariates are weighted averages of past covariate values, where the weights are stochastically  
 191 determined by the data. Here,  $D^{ant}$  was constructed using daily values from the current day to 4  
 192 days ago ( $p = 1, T_{lag} = 5$ ), while  $W_{10}^{ant}$  was constructed using three-day averages of  $W_{10}$  from the  
 193 current day to 20 days ago ( $p = 3, T_{lag} = 7$ ).

194 To complete this model, a zero-centered hierarchical normal prior was specified for tree random  
 195 effects:

196 
$$\begin{aligned} E_{\sigma,t} &\sim \text{Normal}(0, \sigma_{\sigma}^2) \\ E_{\lambda,t} &\sim \text{Normal}(0, \sigma_{\lambda}^2) \end{aligned} \quad (6)$$

197 where reparameterization by sweeping was employed to ensure identifiability between the  
 198 intercepts ( $\beta_0, \alpha_0$ ) and the random effects (Vines *et al.*, 1996).

199 All remaining parameters were given standard priors following Gelman *et al.* (2014). The  
 200 regression coefficients were assigned relatively non-informative normal priors centered at zero  
 201 with large variance. Antecedent importance weights, vectors of length  $T_{\text{lag}}$  (Eqn. 5), were given  
 202 non-informative Dirichlet priors that assume *a priori* that each past time step has equal  
 203 importance, and that constrain weights for each covariate to sum to 1 across all time steps,  $p$ . The  
 204 standard deviation of tree random effects ( $\sigma_{\sigma}$  and  $\sigma_{\lambda}$ ) were given relatively non-informative Unif  
 205 orm(0,1) priors, while the measurement error precision ( $1/\sigma_{\Psi}^2$ ) was assigned a conjugate,  
 206 relatively non-informative Gamma(0.1,0.1) prior.

207 *Model description - GPP*

208 To assess the drivers of daily ecosystem productivity, we developed a two-part model based on  
 209 the Monteith (1972) framework. In this model, daily GPP was modeled sequentially, first as a  
 210 function of NIRv and incoming PAR. The residuals of this model were considered indicative of  
 211 variation in LUE.

212 Typically, LUE is conceptualized as the product of its theoretical maximum ( $LUE_0$ ) and a  
 213 function of environmental stressors that reduce optimal light-use efficiency. Given that  $LUE_0$  is a  
 214 theoretical construct assumed to remain constant within our study (e.g., within a season), the  
 215 GPP model residuals can be interpreted as 1) impacts of environmental stressors on LUE and 2)  
 216 random noise or uncertainty inherent in the data. Thus, while we evaluated GPP model residuals  
 217 as functions of water stress indicators, including VPD, soil moisture, and predawn water  
 218 potential, we also acknowledge that they include data uncertainty and random noise.

219 The likelihood of observed GPP was normally distributed for each observation  $j$  ( $j = 1, 2, \dots,$   
 220 166):

221 
$$GPP_j \sim \text{Normal}(\widehat{GPP}_j, \sigma_{GPP}^2) \quad (7)$$

222 where  $\widehat{GPP}_j$  is the predicted or mean daily GPP and the variance  $\sigma_{GPP}^2$  represents the uncertainty  
 223 in observed GPP.  $\widehat{GPP}_j$  was modeled as a linear function of NIRv, PAR, and their interaction,  
 224 representing the photosynthetic-capacity component of the Monteith (1972) formulation.

225 
$$\widehat{GPP}_j = \gamma_0 + \gamma_1 \cdot NIRv_j + \gamma_2 \cdot PAR_j + \gamma_3 \cdot NIRv_j \cdot PAR_j \quad (8)$$

226 All remaining parameters were given standard priors as previously described.

227 To interpret the remaining GPP as LUE, we calculated the residuals of the above model as the  
 228 posterior mean of  $GPP_j - \widehat{GPP}_j$ ; residuals were scaled for improved model mixing. The likelihood  
 229 of the residual model described scaled resid as normally distributed for observations k ( $k = 1, 2,$   
 230  $\dots, 166$ ):

231 
$$resid_k \sim \text{Normal}(\widehat{resid}_k, \sigma_{resid}^2) \quad (9)$$

232 where  $\widehat{resid}_k$  is the predicted residual between observed and modeled (Eqn. 7, Eqn. 8) GPP, and  
 233 the variance  $\sigma_{resid}^2$  represents the uncertainty in observed resid.  $\widehat{resid}_k$  is interpreted as a dynamic  
 234 LUE constraint on GPP after vegetation greenness and light interception is accounted for. We  
 235 devised three LUE formulations that account for the combined impact of concurrent VPD and  
 236 antecedent soil moisture (Eqn. 10), antecedent soil moisture alone (Eqn. 11), and concurrent  
 237 predawn water potential (Eqn. 12):

238 
$$\widehat{resid}_k = \delta_0 + \delta_1 \cdot D_{maxk} + \delta_2 \cdot W_{10k}^{ant} + \delta_3 \cdot D_{maxk} \cdot W_{10k}^{ant} \quad (10)$$

239 
$$\widehat{resid}_k = \eta_0 + \eta_1 \cdot W_{10k}^{ant} \quad (11)$$

240 
$$\widehat{resid}_k = \theta_0 + \theta_1 \cdot \overline{\Psi_{PDk}} \quad (12)$$

241 where  $\overline{\Psi_{PD}}$  were the gapfilled means of predawn water potential at the site level. Antecedent  
 242 weights for  $W_{10}^{ant}$  were constructed identically to Eqn. 5, with unique weights determined by the  
 243 GPP residuals.

244 *Model implementation and interpretation*

245 The above models were implemented in JAGS 4.3.0 (Plummer, 2003) using R 4.1.1 and 'rjags'  
246 4.13 (R Core Team, 2021; Plummer, 2022). For each model, three parallel Markov chain Monte  
247 Carlo sequences were initiated with dispersed starting values; initial iterations were run until  
248 model convergence, as indicated by the Gelman and Rubin (Gelman & Rubin, 1992) statistic.  
249 Models were then run for 150,000 iterations and thinned by 50 to reduce within-chain  
250 autocorrelation and storage requirements, yielding a total of 9000 relatively independent  
251 posterior samples for each quantity of interest, including the regression coefficients and  
252 antecedent weights. Posterior distributions were summarized by their means and 95% highest  
253 posterior density credible intervals (henceforth, 95% CIs). Covariate effects were significant if  
254 the 95% CI did not overlap zero. To quantify seasonal variation in  $\sigma$ , the posterior samples of the  
255  $\beta$  regression coefficients and antecedent weights were combined with the time-series of scaled  
256  $D_{max}$  and  $W_{10}$  to produce posterior means and 95% CIs. Model comparison criteria for the three  
257 forms of GPP residual models included posterior predictive loss (Gelfand & Ghosh, 1998) and  
258 the coefficient of determination ( $R^2$ ) between observed and predicted values.

259 **Results**260 *Seasonal dynamics of  $\Psi$ , GPP, and  $\sigma$* 

261 Predawn ( $\Psi_{PD}$ ) and midday ( $\Psi_{MD}$ ) stem water potentials responded dynamically to moisture  
262 inputs, particularly during the monsoon period (Fig. 1). Monsoon onset, determined as the day on  
263 which on the 10% of the total July, August, and September precipitation was accumulated  
264 (Grantz *et al.*, 2007), occurred on 2021-07-23. Prior to monsoon onset, the generally high VPD  
265 and low soil moisture yielded relatively consistent mean  $\Psi_{PD}$  between -2 and -4 MPa. During the  
266 monsoon period, VPD and soil moisture were less extreme than during the more arid mid-  
267 summer period, though highly variable as a result of three major pulse-drydown events (Fig. 1).  
268 Beginning with the first major pulse event (39 mm on 2021-07-27), mean  $\Psi_{PD}$  remained above -  
269 2 MPa continuously for 46 days; however, minimum mean  $\Psi_{PD}$  was similar in premonsoon and  
270 monsoon periods due to rapid decline in  $\Psi_{PD}$  following the third major pulse-drydown. Finally,  
271 mean  $\Psi_{PD}$  stayed above -2 MPa during the fall season, likely due to cooling temperatures and  
272 reduced atmospheric demand.

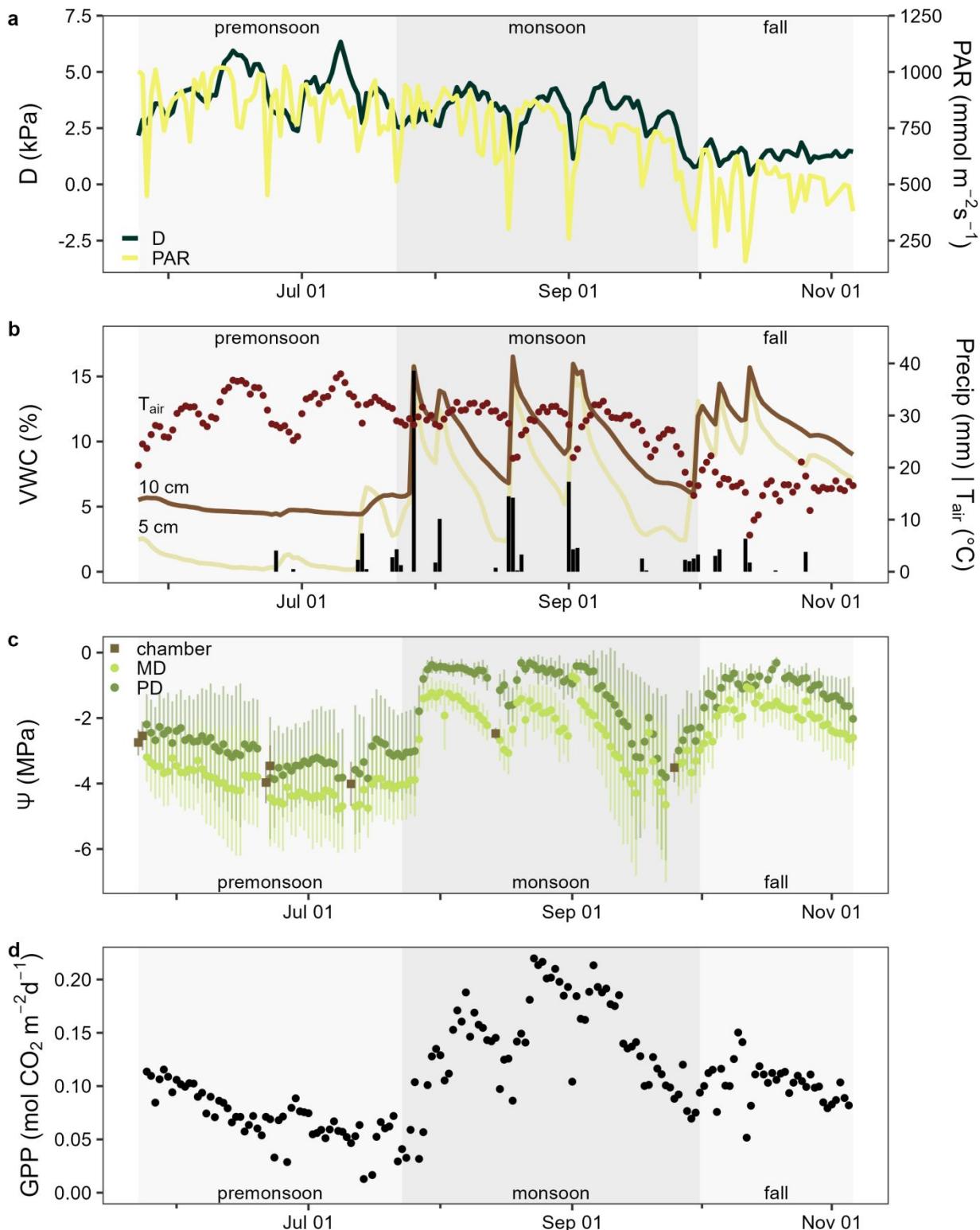


Figure 1: Daily site-level environmental characteristics of a) maximum vapor pressure deficit (D) and photosynthetically active radiation (PAR) b) mean volumetric water content at 5 cm

and 10 cm, total precipitation, and air temperature. Time series of c) chamber and automated daily measurements of mean stem water potential and d) daily total gross primary productivity (GPP). Labeled boxes subdivide the study period into before, during, and after the monsoon season; error bars represent population standard deviation

273 Seasonal GPP dynamics were similarly responsive to precipitation inputs (Fig. 1 b). GPP  
 274 declined during the premonsoon period to near-zero levels, while the onset of the monsoon  
 275 prompted sharp increases. Interestingly, while the first major pulse event was the largest single-  
 276 day total (39 mm on 2021-07-27) that corresponded to almost immediate increases in  $\Psi_{PD}$ , GPP  
 277 rose only modestly. All three peaks in GPP during the monsoon period lagged the moisture  
 278 inputs and lasted more briefly than peaks in  $\Psi_{PD}$  (Fig. 1). Fall GPP averaged  $0.1 \text{ mol CO}_2 \text{ m}^{-2}$   
 279  $\text{d}^{-1}$ , about the same as initial GPP during the premonsoon period.

280 The dynamics of hydraulic regulation can be visually estimated by plotting stem  $\Psi_{MD}$  and  $\Psi_{PD}$   
 281 for each season (Fig. 2). The slope  $\sigma$  appeared similar during the premonsoon and fall periods,  
 282 although  $W$  and therefore stem  $\Psi$  differed substantially between the two seasons. During the  
 283 monsoon period, two slopes were detected via segmented regression, with  $\sigma > 1$  occurring when  
 284 soil moisture was high and  $\Psi_{PD} > -0.6 \text{ MPa}$ ; the same region during the fall season had a  
 285 much shallower slope.

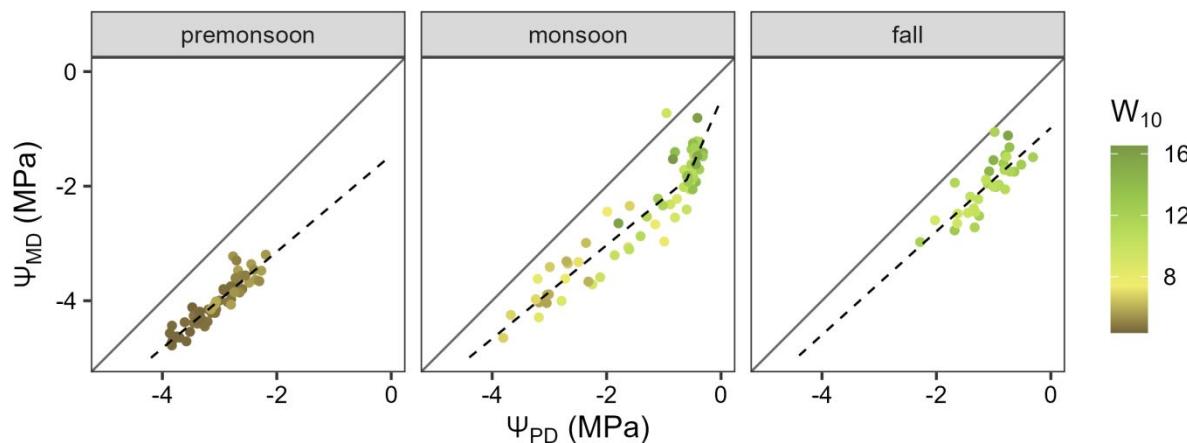


Figure 2: Midday ( $\Psi_{MD}$ ) vs. predawn ( $\Psi_{PD}$ ) stem water potential in each of three seasons. Points are colored by the concurrent daily volumetric water content at 10 cm ( $W_{10}$ ). Solid line

is the 1:1 line. Dashed lines represent linear fits by season, with a segmented regression joined at  $\Psi_{PD} = -0.6$  during the monsoon

286 *Environmental drivers and timescales of  $\sigma$*

287 The hydraulic regulation model (Eqn. 3 - Eqn. 6) fit the data very well (Fig. S1, observed  
 288 vs. predicted  $\Psi_{MD}$   $R^2 = 0.920$ ) with low bias (slope of observed vs. predicted = 0.919).

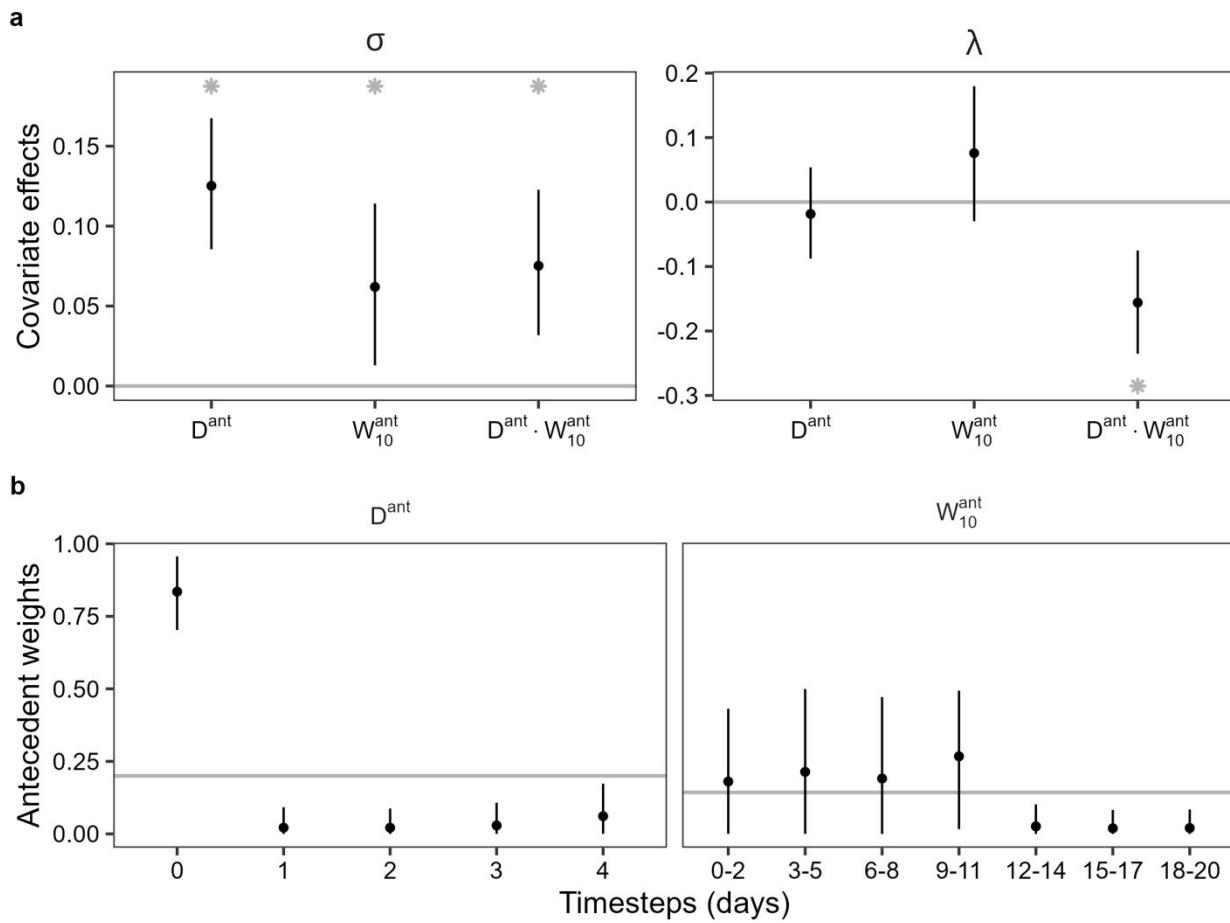


Figure 3: Posterior mean and 95% CI of the a) covariate effects on  $\sigma$  and  $\lambda$  and b) antecedent weights  $\omega$  associated with covariates  $D$  and  $W_{10}$ . Gray horizontal lines indicate the prior means, and asterisks indicate significant covariate effects

289 Temporal variation in hydraulic regulation ( $\sigma$ ) was strongly positively associated with antecedent  
 290 VPD ( $D^{ant}$ ), antecedent soil water content ( $W_{10}^{ant}$ ), and their interaction (Fig. 3 a), indicating that

291 *J. osteosperma* became especially anisohydric under dry atmospheric conditions when soils were  
 292 wet. While the positive effect of D was primarily driven by the atmospheric dryness on the same  
 293 day, soil moisture up to 11 days prior was influential (Fig. 3 b). The pressure drop parameter  $\lambda$   
 294 was negatively associated with the interaction of  $D^{\text{ant}}$  and  $W_{10}^{\text{ant}}$ , although the main effects were  
 295 not significant (Fig. 3 a).

296 *Temporal patterns in  $\sigma$  and GPP*

297 Although general trends in hydraulic regulation can be inferred from grouping  $\Psi_{\text{MD}}$  and  $\Psi_{\text{PD}}$  by  
 298 season (e.g., Fig. 2), the hierarchical Bayesian model permitted combining posterior parameter  
 299 distributions with environmental covariates to produce daily timeseries of predicted  $\sigma$  (e.g.,  
 300 Fig. 4 a), which cannot be determined empirically. During the premonsoon, *J. osteosperma*  
 301 shifted between iso- and anisohydry, with  $\sigma$  values near 1. But during the monsoon season, the  
 302 three main pulse events heralded peaks in  $\sigma$  that signify extreme anisohydry, with  $\sigma$  values well  
 303 above 1, driven by the high VPD and still-wet soils that characterize the post-precipitation  
 304 period. Finally, in the fall, *J. osteosperma* returned to isohydry, and  $\sigma$  fell below 1.

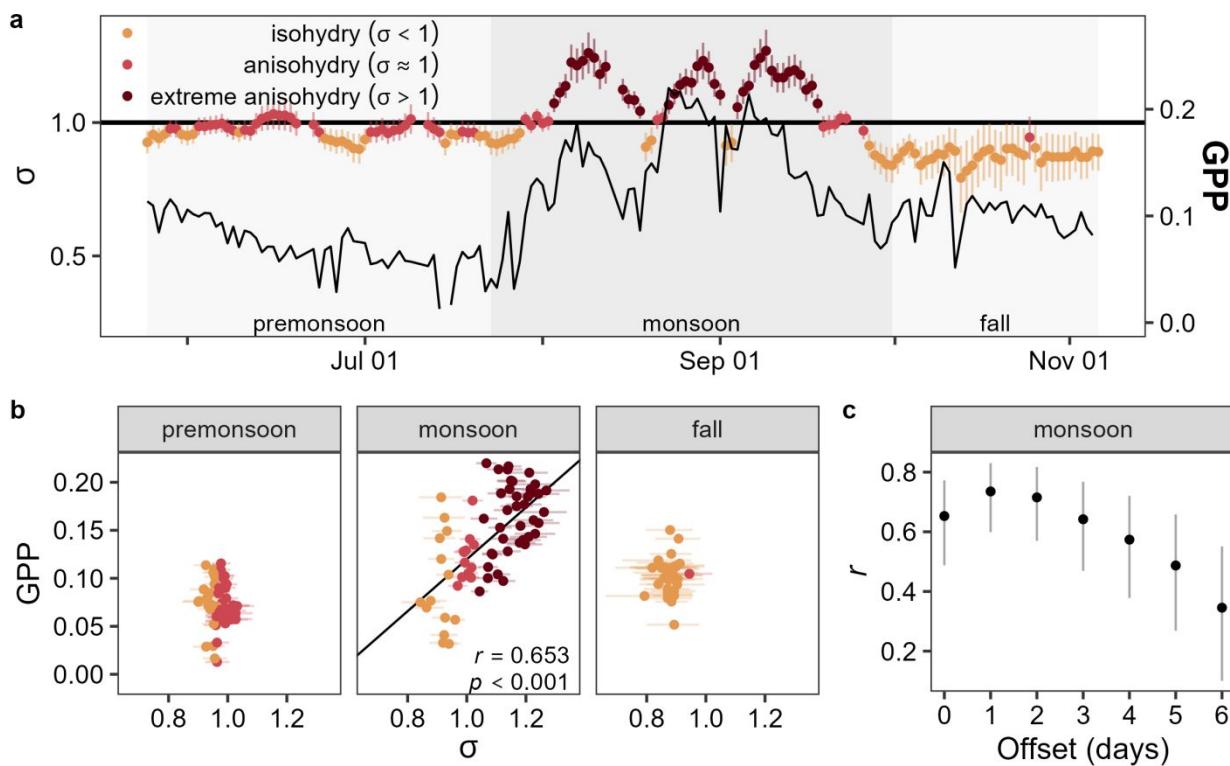


Figure 4: Relationship between predicted  $\sigma$  (posterior mean and 95% CI) and daily GPP shown as a) time series across three seasons, b) bivariate plots for each season, and c) Pearson's correlation coefficient (estimate and 95% confidence interval) during monsoon season across a range of daily offsets where GPP leads  $\sigma$

305 The trends in daily  $\sigma$  corresponded well to observed time-series of GPP (Fig. 4 a), particularly in  
 306 the responsiveness of both  $\sigma$  and GPP to the three main pulse events. Thus,  $\sigma$  and GPP were  
 307 positively correlated during the monsoon period ( $r = 0.653, p < 0.001$ , Fig. 4 b). However, the  
 308 peak in GPP appeared to lead the the peak in  $\sigma$ , as the highest Pearson's correlation between  
 309 GPP and  $\sigma$  was achieved at a 1-2 day offset between the two timeseries (Fig. 4 c).

310 *Plant water potential relationship to GPP*

311 The initial GPP model (Eqn. 7, Eqn. 8) also fit the data well (Fig. 5 c, observed vs. predicted GP  
 312  $P R^2 = 0.733$ ), although with some degree of bias (slope of observed vs. predicted = 0.732) such  
 313 that some high GPP values were underpredicted. Comparing the GPP and NIRv time series  
 314 (Fig. 5 a), the first major pulse event elicited a strong GPP response prior to any green up  
 315 detected optically by NIRv. Conversely, low, near-zero GPP in mid to late July was not matched  
 316 by extreme lows in scaled PAR or NIRv, resulting in overprediction of low GPP values. GPP  
 317 was positively associated with NIRv and the interaction between NIRv and PAR (Fig. 5 b),  
 318 though PAR alone was not significantly associated with GPP.

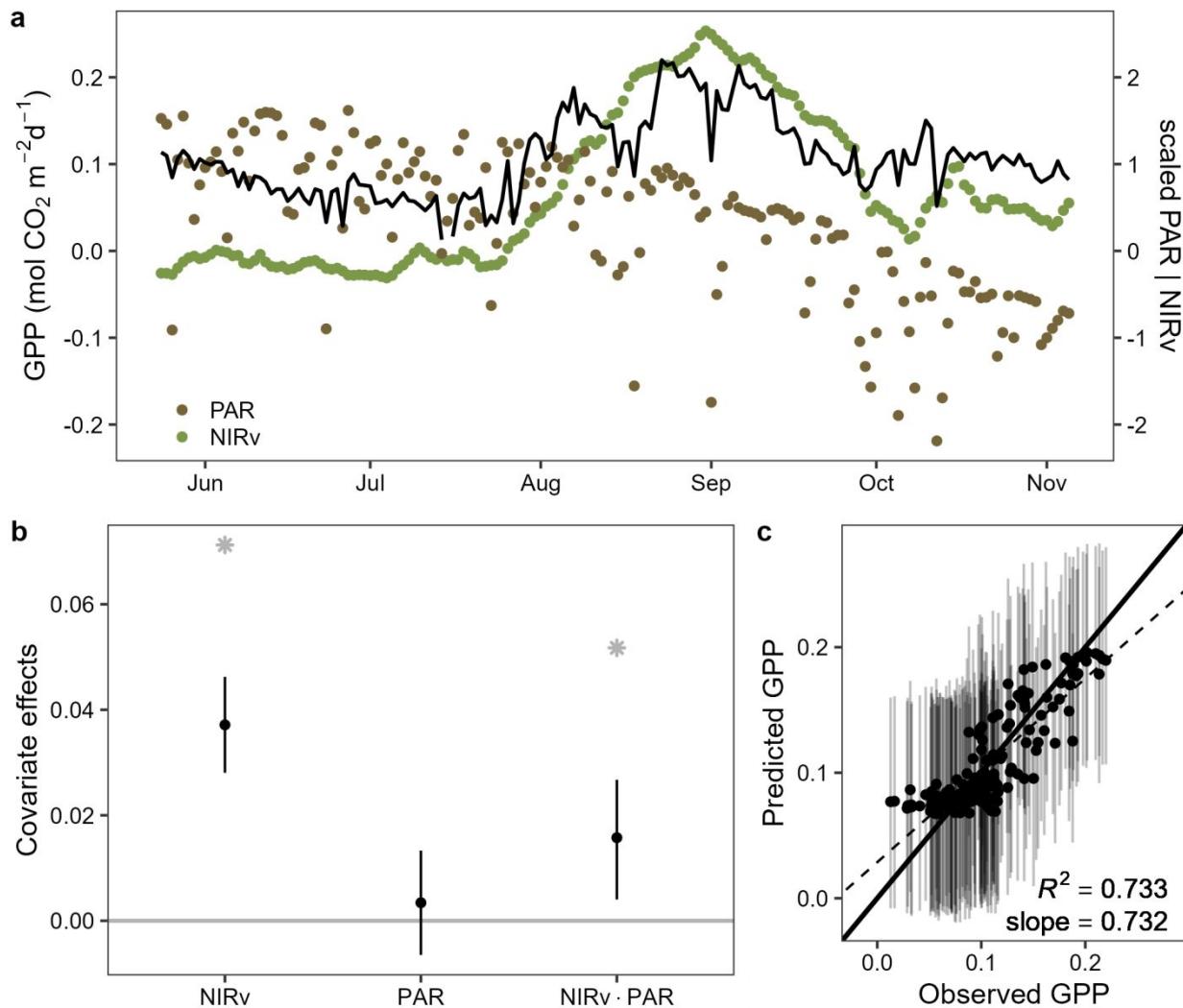


Figure 5: a) Time series of daily total gross primary productivity (GPP) with scaled values of near infrared reflectance of vegetation (NIRv) and photosynthetically active radiation (PAR). Posterior mean and 95% CI of the b) covariate effects and c) predicted versus observed GPP from the first part of the GPP model (Eqn. 7, 8). Gray horizontal lines indicate the prior means, asterisks indicate significant covariate effects, error bars represent the 95% CIs, the solid diagonal is the 1:1 line, and the dashed line represents the line of best fit

319 Residuals from the initial GPP model were interpreted as fluctuations in light use efficiency  
 320 (LUE), and model fit was compare among three functional forms: environmental covariates with  
 321 D and  $W_{10}^{\text{ant}}$  (Eqn. 5, Eqn. 10), soil water content with  $W_{10}^{\text{ant}}$  only (Eqn. 5, Eqn. 11), and predawn  
 322 water potential with  $\overline{\Psi_{PD}}$  only (Eqn. 12). Of the three models, the  $\overline{\Psi_{PD}}$  model had the fewest

323 effective number of parameters (pD), lowest posterior predictive loss ( $D_{\infty}$ ), strongest coefficient  
 324 of determination ( $R^2$ ), and lowest bias (Fig. 6 b,c).

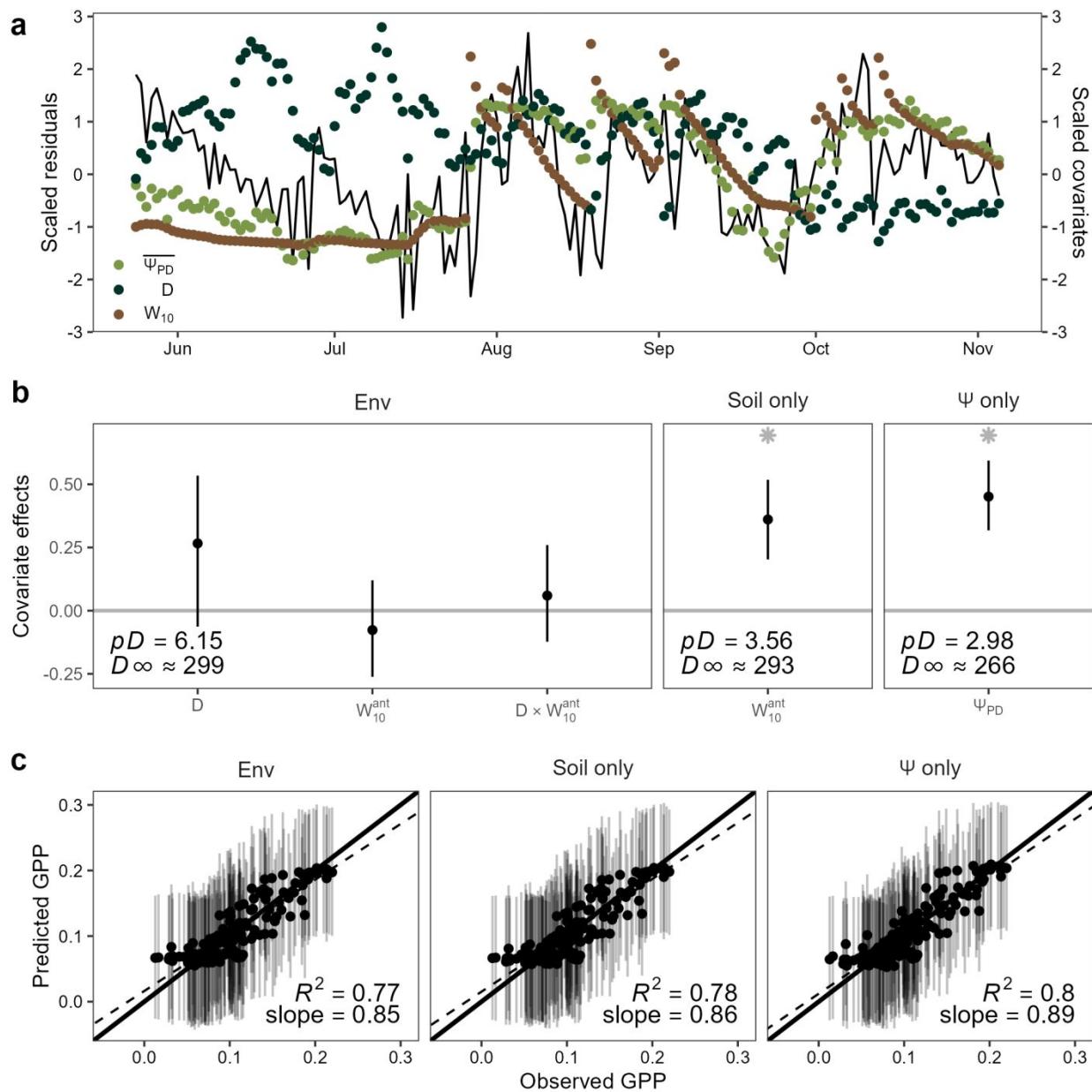


Figure 6: a) Time series of daily residuals from the GPP model with daily maximum vapor pressure deficit (D), volumetric water content at 10 cm ( $W_{10}^{ant}$ ), and site-averaged predawn water potential ( $\bar{\Psi}_{PD}$ ), all standardized to the same scale. Posterior mean and 95% CI of the b) covariate effects and c) predicted versus observed residuals from each residual model: Env (Eqn. 5 & Eqn. 10), Soil only (Eqn. 5 & Eqn. 11), and  $\Psi$  only (Eqn. 12). Gray horizontal lines

indicate the prior means, asterisks indicate significant covariate effects, error bars represent the 95% CIs, the solid diagonal is the 1:1 line, and the dashed line represents the line of best fit

325 Among the LUE models, the  $\overline{\Psi_{PD}}$  fit the GPP residuals modestly well (observed vs. predicted res  
 326 id  $R^2=0.199$ ) while minimizing posterior predictive loss (Fig. 6 b). Combining the initial GPP  
 327 model and the best-performing LUE model using  $\overline{\Psi_{PD}}$  improved the overall  $R^2$  from 0.733 to  
 328 0.800 and substantially reduced bias from 0.732 to 0.89 (Fig. 5 c, Fig. 6 c).

329 The strong performance of  $\overline{\Psi_{PD}}$  for predicting GPP residuals is likely due to their close temporal  
 330 coherence (Fig. 6 a), which outperformed models with  $W_{10}^{\text{ant}}$  alone or in conjunction with D  
 331 (Fig. 6 b). The antecedent weights for  $W_{10}^{\text{ant}}$  (Fig. S2) indicated that GPP residuals lagged soil  
 332 moisture by 3-5 days, but the temporally-weighted soil moisture still did not correlate as strongly  
 333 with GPP residuals as  $\overline{\Psi_{PD}}$  did. Surprisingly, neither D nor  $W_{10}^{\text{ant}}$  was significantly associated with  
 334 the GPP residuals in the environmental covariates model (Fig. 6 b), perhaps because D and PAR  
 335 were highly correlated (Fig. 1) and the initial GPP model already accounted for PAR.

### 336 Discussion

337 In this study, we aimed to improve our understanding of temporal variability in plant hydraulic  
 338 regulation and its relationship to ecosystem carbon uptake. We leveraged contemporaneous,  
 339 high-resolution water potential and carbon flux data to compare temporal trends in daily plant  
 340 water potential, hydraulic behavior, and GPP in a juniper woodland. First, we found that in *J.*  
 341 *osteosperma*, hydraulic regulation varied over the growing season. Increasing anisohydricity was  
 342 observed following precipitation pulses, associated with high soil moisture and high atmospheric  
 343 demand (Fig. 3). Next, we found that GPP and  $\sigma$  were most positively correlated during the  
 344 monsoon season, but with different temporal trajectories following precipitation pulses (Fig. 4).  
 345 Surprisingly, although both  $\Psi_{MD}$  and  $\Psi_{PD}$  responded rapidly to precipitation inputs, plants  
 346 achieved maximum  $\sigma$  1-2 days after peak GPP was reached for a given moisture pulse (Fig. 4 c).  
 347 Together, these results hint at the intriguing possibility that extreme anisohydry can serve to  
 348 maximize soil water extraction and prolong GPP pulses in dryland ecosystems. Finally, predawn  
 349 water potential explained more variability in GPP compared to environmental covariates  
 350 associated with atmospheric and soil moisture conditions (Fig. 6). As a direct metric of water

351 stress, plant water potential closely matched the timing of GPP variability not accounted for by  
352 light availability and vegetation greenness, underscoring water stress as the dominant constraint  
353 on intra-annual GPP dynamics in dryland ecosystems.

354 *Temporally-varying hydraulic behavior*

355 *Juniperus* species are considered more anisohydric than co-occurring pinyon pines due to their  
356 more cavitation-resistant xylem (Linton *et al.*, 1998), higher hydraulic safety margins (Plaut *et*  
357 *al.*, 2012), and lower leaf water potentials (West *et al.*, 2007; Breshears *et al.*, 2009), while their  
358 categorization based on stomatal control is less conclusive (Garcia-Forner *et al.*, 2016a). Due to  
359 less vulnerable xylem in *Juniperus*, low water potentials alone do not suggest less stringent  
360 stomatal control, as they must be interpreted relative to vulnerability curve parameters such as  
361  $\Psi_{50}$ , or the xylem pressure at 50% loss of hydraulic conductance. Here, a 166-day time series of  
362  $\Psi_{PD}$  and  $\Psi_{MD}$  in *J. osteosperma* reveals strong, context-dependent variation in hydraulic  
363 regulation, an intermediate timescale that can potentially bridge the gap between short-term  
364 stomatal response-based definition of iso/anisohydry (Tardieu & Simonneau, 1998) and  
365 definitions that rely on seasonal extremes (Klein, 2014; Martinez-Vilalta *et al.*, 2021). We posit  
366 that response-based metrics (Kannenberg *et al.*, 2022) can be used to quantify plant water use  
367 strategies without distinguishing between active versus passive regulation, yet can enhance  
368 predictive understanding of plant-environment interactions.

369 Dry air in combination with wet soil drove large increases in  $\sigma$  in *J. osteosperma*, attesting to the  
370 importance of VPD as a driver of plant responses (Novick *et al.*, 2016; Grossiord *et al.*, 2020).  
371 As the same drivers were important for hydraulic regulation in the drought-tolerant desert shrub,  
372 *Larrea tridentata* (Guo *et al.*, 2020), transient drops in  $\Psi_{MD}$  may be strategic only during the wet  
373 periods of otherwise water-limited ecosystems, when the reward of carbon uptake exceeds the  
374 risk of embolism. In our study, shifts to extreme anisohydry appeared only as responses to  
375 discrete monsoon precipitation pulses, suggesting that flexible hydraulic behavior enables *J.*  
376 *osteosperma* to take advantage of soil moisture when available. The responsiveness of hydraulic  
377 behavior to soil moisture may explain why, despite similar lateral root densities as pinyon pines  
378 (Schwinnning *et al.*, 2020), junipers tend to be more physiologically responsive to moderate  
379 moisture inputs (Breshears *et al.*, 1997; West *et al.*, 2007; Guo *et al.*, 2018).

380 The positive relationship between  $\sigma$  and GPP during the monsoon season suggests that  
381 temporally-variable hydraulic regulation can maximize carbon uptake during periods of patchy  
382 moisture availability. Most interestingly, the timing of  $\sigma$  and GPP indicates that extreme  
383 anisohydry intensifies after GPP peaks. After a precipitation pulse when soil moisture is high,  
384 GPP may be immediately stimulated, such that relatively high midday water potentials (low  $\sigma$ )  
385 are sufficient to drive water transport along the soil-plant-atmosphere continuum. Because soil  
386 moisture declines rapidly after precipitation pulses, extreme anisohydry ( $\sigma > 1$ ) may serve to  
387 decrease midday water potentials, maintain water transport in drying soils, and possibly confer a  
388 competitive advantage over co-occurring understory species (e.g., Barron-Gafford *et al.*, 2021).  
389 The propensity of *Juniperus* spp. to extract soil water even at low soil water potentials (West *et*  
390 *al.*, 2007) is consistent with extreme anisohydry and prolonged elevation of GPP as soils dry.

391 *Possible mechanisms of temporally-varying hydraulic regulation*

392 The mechanisms underlying temporally-varying hydraulic regulation are not well understood,  
393 but coordination with other temporally-varying physiology and growth responses could play a  
394 role. First, pressure-volume relationships in *Juniperus monosperma* are plastic depending on leaf  
395 hydration (Meinzer *et al.*, 2014), such that as a leaf dehydrates, it experiences more negative  
396 turgor loss point and less elastic cell walls. Conceivably, stomatal regulation of leaf water  
397 potential could also vary with leaf hydration, which may be especially dynamic in evergreen  
398 leaves experiencing pulse-driven precipitation. Accounting for plastic adjustment in turgor loss  
399 point, *J. monosperma* would ultimately lose turgor at -8.2 MPa (Meinzer *et al.*, 2014), and  
400 indeed 99.1% of our individual  $\Psi_{MD}$  observations occurred above this threshold. Temporally-  
401 varying leaf-water relations may indicate that osmotic adjustment, cell wall elastic properties,  
402 and stomatal regulation could vary in concert to maintain turgor across declining leaf hydration.

403 Hydraulic regulation strategies could also be linked to temporal dynamics of foliar ABA during  
404 soil water stress and recovery (Brodribb & McAdam, 2013). In *Callatris rhomboidea*, sustained  
405 water stress led to a decline in ABA such that loss of leaf water potential (and thus guard cell  
406 turgor) drove stomatal closure, with the corollary of low ABA also enabling rapid recovery of  
407 leaf water potential after rewetting (Brodribb & McAdam, 2013). Among *Cupressaceae*,  
408 including *Juniperus* and *Callatris*, the use of leaf desiccation to close stomata during prolonged

409 water stress (Brodribb *et al.*, 2014) could explain why *J. osteosperma* experiences temporally-  
410 varying hydraulic regulation. The hydraulic risk of extreme anisohydry could also be partially  
411 compensated by rapid recovery following rewetting, enabling persistence in seasonally dry  
412 ecosystems.

413 Finally, temporally-varying hydraulic regulation may be associated with the timing of  
414 belowground dynamics that enable increased conductance, such as fine root and mycorrhizal  
415 development (Peek *et al.*, 2006; Lehto & Zwiazek, 2011). In *J. osteosperma*, fine roots grew  
416 when soil water was most available and shifted toward deeper layers as the growing season  
417 progressed (Peek *et al.*, 2006), and root distributions varied depending on cool-season vs. warm-  
418 season precipitation. Rooting dynamics can directly influence plant water potential via  
419 rhizosphere conductance, although this is difficult to quantify empirically (Bristow *et al.*, 1984;  
420 Sperry *et al.*, 2016). Similarly, mycorrhizal symbionts are known to alter root conductivity  
421 (Lehto & Zwiazek, 2011), enhance stomatal conductance (Augé *et al.*, 2015), and increase plant  
422 productivity (Mohan *et al.*, 2014), but the temporal dynamics of plant-mycorrhizae relationships  
423 under field conditions are poorly understood and merit further investigation (Gehring *et al.*,  
424 2017).

#### 425 *Implications for hydraulic modeling*

426 Plant hydraulic schemes are becoming increasingly represented in vegetation and land surface  
427 models (Kennedy *et al.*, 2019b; Eller *et al.*, 2020; Sabot *et al.*, 2020). The link we observed  
428 between hydraulic strategy and GPP reinforces the value of these approaches for improved  
429 predictions of GPP, especially in dryland ecosystems where patchy resource availability leads to  
430 widespread underpredictions of both the magnitude and variability of carbon fluxes (Biederman  
431 *et al.*, 2017; MacBean *et al.*, 2021b; Barnes *et al.*, 2021). Temporal heterogeneity in plant  
432 hydraulic strategy and spatial heterogeneity in topoedaphic characteristics may also interact, as  
433 evidenced by high variance in stem  $\Psi$  among seven co-located trees (Fig. 1 c), and contribute to  
434 model underperformance in dryland ecosystems. However, if transient anisohydry does indeed  
435 represent a life history strategy to maximize carbon uptake during pulses of moisture availability,  
436 then models will need to allow for vegetation hydraulic strategies to vary over time in order to  
437 correctly estimate dryland GPP.

438 One avenue of model development operationalizes the trade-off between carbon gain and  
439 hydraulic costs (Sperry *et al.*, 2016; Wolf *et al.*, 2016; Mencuccini *et al.*, 2019), a subset of  
440 stomatal optimization models that accounts for the cavitation risk of low plant  $\Psi$  (Wang *et al.*,  
441 2020). Temporally-variable water use strategies may arise as an emergent property of such  
442 models (Kannenberg *et al.*, 2022), but likely only where the hydraulic costs of anisohydry and  
443 the forfeited carbon gain of isohydry are simultaneously represented. Alternatively, improving  
444 the temporal fidelity of optimization models could involve explicit implementation at multiple  
445 timescales (daily, weekly) to represent plant physiological acclimation to a changing  
446 environment (Joshi *et al.*, 2022). It remains an open question how best to account for transient  
447 hydraulic strategies in modeling frameworks, and further research regarding when, where, and  
448 how such strategies arise is necessary to evaluate their role in improving estimation of dryland  
449 carbon fluxes.

450 *Importance of plant water potential at large scales*

451 In our study of a single growing season, we found that predawn water potential matches the  
452 temporal pattern of LUE even more strongly than antecedent soil moisture, which comports with  
453 the critical role of water potential to plant physiology. Importantly, predawn water potential  
454 improved GPP model fit even though measurement scales varied greatly, with NIRv derived  
455 from a 500 m pixel, GPP from a flux tower, and stand water potential averaged from 7 trees  
456 within the tower footprint, suggesting that the theoretical foundation connecting plant hydraulics  
457 to ecosystem productivity is robust to significant scale mismatch. Inclusion of predawn water  
458 potential rather than VPD and soil moisture improved not only model fit of GPP, but also the  
459 significantly reduced model bias (Fig. 6), primarily by accounting for the transition between dry  
460 season and first monsoon pulse, wherein high predawn water potentials signal physiological  
461 readiness for photosynthesis even though vegetation greenness is still lagging. Overprediction of  
462 low premonsoon GPP and underprediction of high monsoon GPP were strongly ameliorated by  
463 concomitant shifts in predawn water potential.

464 However, interpreting model residuals as indicative of variations in LUE must be approached  
465 with caution, as these residuals also encompass data uncertainty and unaccounted factors. This  
466 consideration is particularly important when extrapolating our findings to broader contexts or

467 different temporal scales. Despite these considerations, the substantial improvement of GPP  
468 predictions with the inclusion of predawn water potential underscores its promise as a valuable  
469 indicator for capturing intra-annual variability of dryland GPP and warrants additional  
470 investigation.

471 While continuous timeseries of plant water potential remain rare, new technology and collective  
472 efforts are poised to increase accessibility to this key metric. At large scales, promising pathways  
473 are being explored to develop remote sensing-based proxies of plant water potential using  
474 thermal (Farella *et al.*, 2022) and microwave (Konings *et al.*, 2021) observations. Current  
475 initiatives to collect and aggregate soil and plant water potential in conjunction with flux tower  
476 measurements, including the Ameriflux ‘Year of Water’ and the PSInet Research Coordination  
477 Network database, are anticipated to improve water potential data availability and spur synthesis  
478 beyond single-site studies. We believe that an expanded network of water potential  
479 measurements co-located at existing flux tower sites is essential to calibrate and evaluate both  
480 model and remote sensing approaches for estimating productivity.

#### 481 *Conclusions*

482 Though classically considered anisohydric, *J. osteosperma* exhibited multiple hydraulic  
483 regulation strategies within a growing season. Extreme anisohydry was only evident after  
484 monsoon precipitation pulses, while soils were rapidly drying yet carbon uptake was high. This  
485 suggests that temporally flexible hydraulic regulation allows *J. osteosperma* to avoid extreme  
486  $\Psi_{MD}$  and xylem cavitation during seasonal drought and prolong high carbon uptake following  
487 episodic precipitation events. Furthermore, plant water potential significantly improved GPP  
488 model fit and reduced bias despite significant scale mismatch, heralding the immense potential of  
489 using plant water stress to increase the temporal fidelity of ecosystem carbon predictions.

#### 490 **Data availability**

491 Data and code are organized as a research compendium in a public GitHub repository  
492 (<https://github.com/jessicaguo/juniper-ecohydraulics>) and archived on Zenodo  
493 (<https://doi.org/10.5281/zenodo.10951221>).

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511 **Author contributions**

512 JSG and SAK conceived of and designed study with input from WRLA. Field data were  
513 collected by SAK, processed by SAK and JSG, and analyzed by JSG. Remote sensing data were  
514 processed by MLB and analyzed by JSG with input from WKS. JSG prepared the manuscript  
515 with input from all co-authors. All authors reviewed the results and approved the final version of  
516 the manuscript.

517

518 **References**

519 **Ahlström A, Raupach MR, Schurgers G, Smith B, Arneth A, Jung M, Reichstein M,**  
520 **Canadell JG, Friedlingstein P, Jain AK, et al. 2015.** The dominant role of semi-arid  
521 ecosystems in the trend and variability of the land CO<sub>2</sub> sink. *Science* **348**: 895–899.

522 **Augé RM, Toler HD, Saxton AM. 2015.** Arbuscular mycorrhizal symbiosis alters stomatal  
523 conductance of host plants more under drought than under amply watered conditions: a meta-  
524 analysis. *Mycorrhiza* **25**: 13–24.

525 **Badgley G, Anderegg LDL, Berry JA, Field CB. 2019.** Terrestrial gross primary production:  
526 Using NIRV to scale from site to globe. *Global Change Biology* **25**: 3731–3740.

527 **Badgley G, Field CB, Berry JA. 2017.** Canopy near-infrared reflectance and terrestrial  
528 photosynthesis. *Science Advances* **3**: e1602244.

529 **Barnes ML, Farella MM, Scott RL, Moore DJP, Ponce-Campos GE, Biederman JA,**  
530 **MacBean N, Litvak ME, Breshears DD. 2021.** Improved dryland carbon flux predictions with  
531 explicit consideration of water-carbon coupling. *Communications Earth & Environment* **2**: 1–9.

532 **Barron-Gafford GA, Knowles JF, Sanchez-Cañete EP, Minor RL, Lee E, Sutter L, Tran**  
533 **N, Murphy P, Hamerlynck EP, Kumar P, et al. 2021.** Hydraulic redistribution buffers climate  
534 variability and regulates grass-tree interactions in a semiarid riparian savanna. *Ecohydrology* **14**:  
535 e2271.

536 **Berry JA, Beerling DJ, Franks PJ. 2010.** Stomata: Key players in the earth system, past and  
537 present. *Current Opinion in Plant Biology* **13**: 232–239.

538 **Biederman JA, Scott RL, Bell TW, Bowling DR, Dore S, Garatuza-Payan J, Kolb TE,**  
539 **Krishnan P, Kroccheck DJ, Litvak ME, et al. 2017.** CO<sub>2</sub> exchange and evapotranspiration  
540 across dryland ecosystems of southwestern north america. *Global Change Biology* **23**: 4204–  
541 4221.

542 **Biederman JA, Scott RL, Goulden ML, Vargas R, Litvak ME, Kolb TE, Yepez EA, Oechel**  
543 **WC, Blanken PD, Bell TW, et al.** 2016. Terrestrial carbon balance in a drier world: The effects  
544 of water availability in southwestern north america. *Global change biology* **22**: 1867–1879.

545 **Breshears DD, Myers OB, Johnson SR, Meyer CW, Martens SN.** 1997. Differential use of  
546 spatially heterogeneous soil moisture by two semiarid woody species: *Pinus edulis* and *juniperus*  
547 *monosperma*. *The Journal of Ecology* **85**: 289.

548 **Breshears DD, Myers OB, Meyer CW, Barnes FJ, Zou CB, Allen CD, McDowell NG,**  
549 **Pockman WT.** 2009. Tree die-off in response to global change-type drought: Mortality insights  
550 from a decade of plant water potential measurements. *Frontiers in Ecology and the Environment*  
551 **7**: 185–189.

552 **Bristow KL, Campbell GS, Calissendorff C.** 1984. The Effects of Texture on the Resistance to  
553 Water Movement within the Rhizosphere. *Soil Science Society of America Journal* **48**: 266–270.

554 **Brodribb TJ, McAdam SAM.** 2013. Abscisic acid mediates a divergence in the drought  
555 response of two conifers. *Plant Physiology* **162**: 1370–1377.

556 **Brodribb TJ, McAdam SAM, Jordan GJ, Martins SCV.** 2014. Conifer species adapt to low-  
557 rainfall climates by following one of two divergent pathways. *Proceedings of the National  
558 Academy of Sciences* **111**: 14489–14493.

559 **Eller CB, Rowland L, Mencuccini M, Rosas T, Williams K, Harper A, Medlyn BE, Wagner**  
560 **Y, Klein T, Teodoro GS, et al.** 2020. Stomatal optimization based on xylem hydraulics (SOX)  
561 improves land surface model simulation of vegetation responses to climate. *New Phytologist*  
562 **226**: 1622–1637.

563 **Farella MM, Fisher JB, Jiao W, Key KB, Barnes ML.** 2022. Thermal remote sensing for plant  
564 ecology from leaf to globe. *Journal of Ecology* **110**: 1996–2014.

565 **Feldman AF, Short Gianotti DJ, Konings AG, McColl KA, Akbar R, Salvucci GD,**  
566 **Entekhabi D.** 2018. Moisture pulse-reserve in the soil-plant continuum observed across biomes.  
567 *Nature Plants* **4**: 1026–1033.

568 **Feng X, Ackerly DD, Dawson TE, Manzoni S, McLaughlin B, Skelton RP, Vico G, Weitz**  
569 **AP, Thompson SE. 2019.** Beyond isohydricity: The role of environmental variability in  
570 determining plant drought responses. *Plant, Cell & Environment* **42**: 1104–1111.

571 **Garbulsky MF, Peñuelas J, Papale D, Filella I. 2008.** Remote estimation of carbon dioxide  
572 uptake by a mediterranean forest. *Global Change Biology* **14**: 2860–2867.

573 **Garcia-Forner N, Adams HD, Sevanto S, Collins AD, Dickman LT, Hudson PJ, Zeppel MJ,**  
574 **Jenkins MW, Powers H, Martínez-Vilalta J, et al. 2016a.** Responses of two semiarid conifer  
575 tree species to reduced precipitation and warming reveal new perspectives for stomatal  
576 regulation. *Plant, Cell & Environment* **39**: 38–49.

577 **Garcia-Forner N, Sala A, Biel C, SavÁ© R, Martínez-Vilalta J. 2016b.** Individual traits as  
578 determinants of time to death under extreme drought in *pinus sylvestris* l. *Tree Physiology* **36**:  
579 1196–1209.

580 **Gehring CA, Swaty RL, Deckert RJ. 2017.** Chapter 16 - Mycorrhizas, Drought, and Host-Plant  
581 Mortality. In: Johnson NC, Gehring C, Jansa J, eds. Elsevier, 279–298.

582 **Gelfand AE, Ghosh SK. 1998.** Model choice: A minimum posterior predictive loss approach.  
583 *Biometrika* **85**: 1–11.

584 **Gelman A, Carlin JB, Stern HS, Rubin DB. 2014.** *Bayesian data analysis*. Boca Raton, FL,  
585 USA: Taylor & Francis.

586 **Gelman A, Rubin DB. 1992.** Inference from iterative simulation using multiple sequences.  
587 *Statistical Science* **7**: 457–472.

588 **Grantz K, Rajagopalan B, Clark M, Zagona E. 2007.** Seasonal shifts in the north american  
589 monsoon. *Journal of Climate* **20**: 1923–1935.

590 **Grossiord C, Buckley TN, Cernusak LA, Novick KA, Poulter B, Siegwolf RTW, Sperry JS,**  
591 **McDowell NG. 2020.** Plant responses to rising vapor pressure deficit. *New Phytologist* **226**:  
592 1550–1566.

593 **Guo JS, Hultine KR, Koch GW, Kropf H, Ogle K. 2020.** Temporal shifts in iso/anisohydry  
594 revealed from daily observations of plant water potential in a dominant desert shrub. *New*  
595 *Phytologist* **225**: 713–726.

596 **Guo JS, Hungate BA, Kolb TE, Koch GW. 2018.** Water source niche overlap increases with  
597 site moisture availability in woody perennials. *Plant Ecology* **219**: 719–735.

598 **Haberstroh S, Lobo-do-Vale R, Caldeira MC, Dubbert M, Cuntz M, Werner C. 2022.** Plant  
599 invasion modifies isohydricity in mediterranean tree species. *Functional Ecology* **36**: 2384–2398.

600 **Hochberg U, Rockwell FE, Holbrook NM, Cochard H. 2018.** Iso/anisohydry: A  
601 plantenvironment interaction rather than a simple hydraulic trait. *Trends in Plant Science* **23**:  
602 112–120.

603 **Jones HG. 1998.** Stomatal control of photosynthesis and transpiration. *Journal of Experimental*  
604 *Botany* **49**: 387–398.

605 **Joshi J, Stocker BD, Hofhansl F, Zhou S, Dieckmann U, Prentice IC. 2022.** Towards a  
606 unified theory of plant photosynthesis and hydraulics. *Nature Plants*: 1–13.

607 **Kannenberg SA, Barnes ML, Bowling DR, Driscoll AW, Guo JS, Anderegg WRL. 2023.**  
608 Quantifying the drivers of ecosystem fluxes and water potential across the soil-plant-atmosphere  
609 continuum in an arid woodland. *Agricultural and Forest Meteorology* **329**: 109269.

610 **Kannenberg SA, Driscoll AW, Malesky D, Anderegg WRL. 2021.** Rapid and surprising  
611 dieback of Utah juniper in the southwestern USA due to acute drought stress. *Forest Ecology*  
612 *and Management* **480**: 118639.

613 **Kannenberg SA, Guo JS, Novick KA, Anderegg WRL, Feng X, Kennedy D, Konings AG,**  
614 **Martínez-Vilalta J, Matheny AM. 2022.** Opportunities, challenges and pitfalls in  
615 characterizing plant water-use strategies. *Functional Ecology* **36**: 24–37.

616 **Kennedy D, Swenson S, Oleson KW, Lawrence DM, Fisher R, Lola da Costa AC, Gentile**  
617 **P. 2019a.** Implementing Plant Hydraulics in the Community Land Model, Version 5. *Journal of*  
618 *Advances in Modeling Earth Systems* **11**: 485–513.

619 **Kennedy D, Swenson S, Oleson KW, Lawrence DM, Fisher R, Lola da Costa AC, Gentine**  
620 **P. 2019b.** Implementing Plant Hydraulics in the Community Land Model, Version 5. *Journal of*  
621 *Advances in Modeling Earth Systems* **11**: 485–513.

622 **King DA, Turner DP, Ritts WD. 2011.** Parameterization of a diagnostic carbon cycle model for  
623 continental scale application. *Remote Sensing of Environment* **115**: 1653–1664.

624 **Klein T. 2014.** The variability of stomatal sensitivity to leaf water potential across tree species  
625 indicates a continuum between isohydric and anisohydric behaviours (S Niu, Ed.). *Functional*  
626 *Ecology* **28**: 1313–1320.

627 **Konings AG, Saatchi SS, Frankenberg C, Keller M, Leshyk V, Anderegg WRL, Humphrey**  
628 **V, Matheny AM, Trugman A, Sack L, et al. 2021.** Detecting forest response to droughts with  
629 global observations of vegetation water content. *Global Change Biology* **27**: 6005–6024.

630 **Lehto T, Zwiazek JJ. 2011.** Ectomycorrhizas and water relations of trees: a review. *Mycorrhiza*  
631 **21**: 71–90.

632 **Linton MJ, Sperry JS, Williams DG. 1998.** Limits to water transport in *juniperus osteosperma*  
633 and *pinus edulis*: Implications for drought tolerance and regulation of transpiration. *Functional*  
634 *Ecology* **12**: 906–911.

635 **Loik ME, Breshears DD, Lauenroth WK, Belnap J. 2004.** A multi-scale perspective of water  
636 pulses in dryland ecosystems: Climatology and ecohydrology of the western USA. *Oecologia*  
637 **141**: 269–281.

638 **MacBean N, Scott RL, Biederma JA, Peylin P, Kolb T, Litvak ME, Krishnan P, Meyers**  
639 **TP, Arora VK, Bastrikov V, et al. 2021a.** Dynamic global vegetation models underestimate net  
640 **CO<sub>2</sub> flux mean and inter-annual variability in dryland ecosystems.** *Environmental*  
641 *Research Letters* **16**: 094023.

642 **MacBean N, Scott RL, Biederma JA, Peylin P, Kolb T, Litvak ME, Krishnan P, Meyers**  
643 **TP, Arora VK, Bastrikov V, et al. 2021b.** Dynamic global vegetation models underestimate net  
644 **CO<sub>2</sub> flux mean and inter-annual variability in dryland ecosystems.** *Environmental*  
645 *Research Letters* **16**: 094023.

646 **Martínez-Vilalta J, Poyatos R, Aguadé D, Retana J, Mencuccini M.** 2014. A new look at  
647 water transport regulation in plants. *New Phytologist* **204**: 105–115.

648 **Martínez-Vilalta J, Santiago LS, Poyatos R, Badiella L, de Cáceres M, Aranda I, Delzon  
649 S, Vilagrosa A, Mencuccini M.** 2021. Towards a statistically robust determination of minimum  
650 water potential and hydraulic risk in plants. *New Phytologist* **232**: 404–417.

651 **McDowell N, Pockman WT, Allen CD, Breshears DD, Cobb N, Kolb T, Plaut J, Sperry J,  
652 West A, Williams DG, et al.** 2008. Mechanisms of plant survival and mortality during drought:  
653 Why do some plants survive while others succumb to drought? *New Phytologist* **178**: 719–739.

654 **Meinzer FC, Woodruff DR, Marias DE, McCulloh KA, Sevanto S.** 2014. Dynamics of leaf  
655 water relations components in co-occurring iso- and anisohydric conifer species. *Plant, Cell &  
656 Environment* **37**: 2577–2586.

657 **Mencuccini M, Manzoni S, Christoffersen B.** 2019. Modelling water fluxes in plants: from  
658 tissues to biosphere. *New Phytologist* **222**: 1207–1222.

659 **Mohan JE, Cowden CC, Baas P, Dawadi A, Frankson PT, Helmick K, Hughes E, Khan S,  
660 Lang A, Machmuller M, et al.** 2014. Mycorrhizal fungi mediation of terrestrial ecosystem  
661 responses to global change: mini-review. *Fungal Ecology* **10**: 3–19.

662 **Monteith JL.** 1972. Solar radiation and productivity in tropical ecosystems. *Journal of Applied  
663 Ecology* **9**: 747–766.

664 **Monteith JL, Moss CJ, Cooke GW, Pirie NW, Bell GDH.** 1977. Climate and the efficiency of  
665 crop production in britain. *Philosophical Transactions of the Royal Society of London. B,  
666 Biological Sciences* **281**: 277–294.

667 **Moritz S, Bartz-Beielstein T.** 2017. imputeTS: Time series missing value imputation in r. *R J.*  
668 **9**: 207.

669 **Nolan RH, Tarin T, Santini NS, McAdam SAM, Ruman R, Eamus D.** 2017. Differences in  
670 osmotic adjustment, foliar abscisic acid dynamics, and stomatal regulation between an isohydric  
671 and anisohydric woody angiosperm during drought. *Plant, Cell & Environment* **40**: 3122–3134.

672 **Novick KA, Ficklin DL, Baldocchi D, Davis KJ, Ghezzehei TA, Konings AG, MacBean N,**  
673 **Raoult N, Scott RL, Shi Y, et al.** 2022. Confronting the water potential information gap. *Nature*  
674 *Geoscience* **15**: 158–164.

675 **Novick KA, Ficklin DL, Stoy PC, Williams CA, Bohrer G, Oishi A, Papuga SA, Blanken**  
676 **PD, Noormets A, Sulman BN, et al.** 2016. The increasing importance of atmospheric demand  
677 for ecosystem water and carbon fluxes. *Nature Climate Change* **6**: 1023.

678 **Novick KA, Konings AG, Gentine P.** 2019. Beyond soil water potential: An expanded view on  
679 isohydricity including land–atmosphere interactions and phenology. *Plant, Cell & Environment*  
680 **42**: 1802–1815.

681 **Noy-Meir I.** 1973. Desert ecosystems: Environment and producers. *Annual Review of Ecology*  
682 *and Systematics* **4**: 25–51.

683 **Ogle K, Barber JJ, Barron-Gafford GA, Bentley LP, Young JM, Huxman TE, Loik ME,**  
684 **Tissue DT.** 2015. Quantifying ecological memory in plant and ecosystem processes. *Ecology*  
685 *Letters* **18**: 221–235.

686 **Ogle K, Reynolds JF.** 2004. Plant responses to precipitation in desert ecosystems: Integrating  
687 functional types, pulses, thresholds, and delays. *Oecologia* **141**: 282–294.

688 **Peek MS, Leffler AJ, Hipps L, Ivans S, Ryel RJ, Caldwell MM.** 2006. Root turnover and  
689 relocation in the soil profile in response to seasonal soil water variation in a natural stand of utah  
690 juniper (*juniperus osteosperma*). *Tree Physiology* **26**: 1469–1476.

691 **Pierrat Z, Nehemy MF, Roy A, Magney T, Parazoo NC, Laroque C, Pappas C, Sonnentag**  
692 **O, Grossmann K, Bowling DR, et al.** 2021. Tower-based remote sensing reveals mechanisms  
693 behind a two-phased spring transition in a mixed-species boreal forest. *Journal of Geophysical*  
694 *Research: Biogeosciences* **126**: e2020JG006191.

695 **Plaut JA, Yepez EA, Hill J, Pangle R, Sperry JS, Pockman WT, McDowell NG.** 2012.  
696 Hydraulic limits preceding mortality in a piñonjuniper woodland under experimental drought.  
697 *Plant, Cell & Environment* **35**: 1601–1617.

698 **Plummer M.** 2003. Proceedings of the 3rd international workshop on distributed statistical  
699 computing. In: Vienna, Austria.

700 **Plummer M.** 2022. *Rjags: Bayesian graphical models using MCMC*.

701 **Poulter B, Frank D, Ciais P, Myneni RB, Andela N, Bi J, Broquet G, Canadell JG,**  
702 **Chevallier F, Liu YY.** 2014. Contribution of semi-arid ecosystems to interannual variability of  
703 the global carbon cycle. *Nature* **509**: 600–603.

704 **R Core Team.** 2021. *R: A language and environment for statistical computing*. Vienna, Austria:  
705 R Foundation for Statistical Computing.

706 **Running SW, Nemani RR, Heinsch FA, Zhao M, Reeves M, Hashimoto H.** 2004. A  
707 continuous satellite-derived measure of global terrestrial primary production. *Bioscience* **54**:  
708 547–560.

709 **Sabot MEB, De Kauwe MG, Pitman AJ, Medlyn BE, Ellsworth DS, Martin-StPaul NK,**  
710 **Wu J, Choat B, Limousin J-M, Mitchell PJ, et al.** 2022. One stomatal model to rule them all?  
711 Toward improved representation of carbon and water exchange in global models. *Journal of*  
712 *Advances in Modeling Earth Systems* **14**: e2021MS002761.

713 **Sabot MEB, De Kauwe MG, Pitman AJ, Medlyn BE, Verhoef A, Ukkola AM, Abramowitz**  
714 **G.** 2020. Plant profit maximization improves predictions of european forest responses to  
715 drought. *New Phytologist* **226**: 1638–1655.

716 **Schwinning S, Litvak ME, Pockman WT, Pangle RE, Fox AM, Huang C-W, McIntire CD.**  
717 **2020.** A 3-dimensional model of *Pinus edulis* and *Juniperus monosperma* root distributions in  
718 New Mexico: implications for soil water dynamics. *Plant and Soil* **450**: 337–355.

719 **Smith WK, Dannenberg MP, Yan D, Herrmann S, Barnes ML, Barron-Gafford GA,**  
720 **Biederman JA, Ferrenberg S, Fox AM, Hudson A, et al.** 2019. Remote sensing of dryland  
721 ecosystem structure and function: Progress, challenges, and opportunities. *Remote Sensing of*  
722 *Environment* **233**: 111401.

723 **Sperry JS, Wang Y, Wolfe BT, Mackay DS, Anderegg WRL, McDowell NG, Pockman WT.**  
724 **2016.** Pragmatic hydraulic theory predicts stomatal responses to climatic water deficits. *New*  
725 *Phytologist* **212**: 577–589.

726 **Tardieu F, Simonneau T. 1998.** Variability among species of stomatal control under fluctuating  
727 soil water status and evaporative demand: Modelling isohydric and anisohydric behaviours.  
728 *Journal of Experimental Botany* **49**: 419–432.

729 **Vines S, Gilks W, Wild P. 1996.** Fitting bayesian multiple random effects models. *Statistics and*  
730 *Computing* **6**: 337–346.

731 **Wang X, Biederman JA, Knowles JF, Scott RL, Turner AJ, Dannenberg MP, Köhler P,**  
732 **Frankenberg C, Litvak ME, Flerchinger GN, et al. 2022.** Satellite solar-induced chlorophyll  
733 fluorescence and near-infrared reflectance capture complementary aspects of dryland vegetation  
734 productivity dynamics. *Remote Sensing of Environment* **270**: 112858.

735 **Wang Y, Sperry JS, Anderegg WRL, Venturas MD, Trugman AT. 2020.** A theoretical and  
736 empirical assessment of stomatal optimization modeling. *New Phytologist* **227**: 311–325.

737 **West AG, Hultine KR, Jackson TL, Ehleringer JR. 2007.** Differential summer water use by  
738 *pinus edulis* and *juniperus osteosperma* reflects contrasting hydraulic characteristics. *Tree*  
739 *Physiology* **27**: 1711–1720.

740 **Williams AP, Cook BI, Smerdon JE. 2022.** Rapid intensification of the emerging southwestern  
741 North American megadrought in 20202021. *Nature Climate Change* **12**: 232–234.

742 **Wolf A, Anderegg WRL, Pacala SW. 2016.** Optimal stomatal behavior with competition for  
743 water and risk of hydraulic impairment. *Proceedings of the National Academy of Sciences* **113**:  
744 E7222–E7230.

745 **Wu G, Guan K, Li Y, Novick KA, Feng X, McDowell NG, Konings AG, Thompson SE,**  
746 **Kimball JS, De Kauwe MG. 2021.** Interannual variability of ecosystem iso/anisohydry is  
747 regulated by environmental dryness. *New Phytologist* **229**: 2562–2575.

748 **Yuan W, Liu S, Zhou G, Zhou G, Tieszen LL, Baldocchi D, Bernhofer C, Gholz H,**  
749 **Goldstein AH, Goulden ML, et al. 2007.** Deriving a light use efficiency model from eddy  
750 covariance flux data for predicting daily gross primary production across biomes. *Agricultural*  
751 *and Forest Meteorology* **143**: 189–207.

752 **Zeng Y, Hao D, Huete A, Dechant B, Berry J, Chen JM, Joiner J, Frankenberg C, Bond-**  
753 **Lamberty B, Ryu Y, et al. 2022.** Optical vegetation indices for monitoring terrestrial  
754 ecosystems globally. *Nature Reviews Earth & Environment* **3**: 477–493.

755 **Zhang Y, Song C, Sun G, Band LE, Noormets A, Zhang Q. 2015.** Understanding moisture  
756 stress on light use efficiency across terrestrial ecosystems based on global flux and remote-  
757 sensing data. *Journal of Geophysical Research: Biogeosciences* **120**: 2053–2066.

758 **Zhao M, Running SW. 2010.** Drought-induced reduction in global terrestrial net primary  
759 production from 2000 through 2009. *Science* **329**: 940–943.

760 **Supporting Information**

761 **Fig. S1** Model fit of  $\Psi_{MD}$

762 **Fig. S2** Antecedent weights of soil moisture for two GPP residual models