

# ECOGRAPHY

## Research article

### Major axes of variation in tree demography across global forests

Melina de Souza Leite<sup>1,2</sup>, Sean M. McMahon<sup>3,4</sup>, Paulo Inácio Prado<sup>1</sup>,  
Stuart J. Davies<sup>3</sup>, Alexandre Adalardo de Oliveira<sup>1</sup>, Hannes P. De Deurwaerder<sup>1,5</sup>,  
Salomón Aguilar<sup>6</sup>, Kristina J. Anderson-Teixeira<sup>3,7</sup>, Nurfarah Aqilah<sup>8</sup>, Norman A. Bourg<sup>1,7</sup>,  
Warren Y. Brockelman<sup>1,9,10</sup>, Nicolas Castaño<sup>11</sup>, Chia-Hao Chang-Yang<sup>12</sup>,  
Yu-Yun Chen<sup>13</sup>, George Chuyong<sup>14</sup>, Keith Clay<sup>15</sup>, Álvaro Duque<sup>16</sup>, Sisira Ediriweera<sup>17</sup>,  
Corneille E. N. Ewango<sup>18</sup>, Gregory Gilbert<sup>19</sup>, I. A. U. N. Gunatilleke<sup>1,20</sup>, C. V. S. Gunatilleke<sup>20</sup>,  
Robert Howe<sup>21</sup>, Walter Huaraca Huasco<sup>22</sup>, Akira Itoh<sup>1,23,24</sup>, Daniel J. Johnson<sup>1,25</sup>, David Kenfack<sup>1,6</sup>,  
Kamil Král<sup>1,26</sup>, Yao Tze Leong<sup>1,27,28</sup>, James A. Lutz<sup>1,27,28</sup>, Jean-Remy Makana<sup>1,29</sup>, Yadvinder Malhi<sup>1,22</sup>,  
William J. McShea<sup>1,7</sup>, Mohizah Mohamad<sup>30</sup>, Musalmah Nasardin<sup>1,28</sup>, Anuttara Nathalang<sup>1,9</sup>,  
Geoffrey Parker<sup>4</sup>, Renan Parmigiani<sup>1</sup>, Rolando Pérez<sup>6</sup>, Richard P. Phillips<sup>1,31</sup>, Pavel Šamonil<sup>1,26</sup>,  
I-Fang Sun<sup>1,32</sup>, Sylvester Tan<sup>30</sup>, Duncan Thomas<sup>33</sup>, Jill Thompson<sup>1,34</sup>, María Uriarte<sup>35</sup>, Amy Wolf<sup>21</sup>,  
Jess Zimmerman<sup>36</sup>, Daniel Zuleta<sup>1,3</sup>, Marco D. Visser<sup>1,37,\*</sup> and Lisa Hülsmann<sup>1,2,38,39,\*</sup>

EDITOR'S  
CHOICE

<sup>1</sup>Department of Ecology, University of São Paulo, São Paulo, SP, Brazil

<sup>2</sup>Ecosystem Analysis and Simulation Lab, University of Bayreuth, Bayreuth, Germany

<sup>3</sup>Forest Global Earth Observatory, Smithsonian Tropical Research Institute, Washington, DC, USA

<sup>4</sup>Smithsonian Environmental Research Center, Edgewater, MD, USA

<sup>5</sup>Department of Ecology and Evolutionary Biology, Princeton University, Princeton, NJ, USA

<sup>6</sup>Forest Global Earth Observatory, Smithsonian Tropical Research Institute, Panama

<sup>7</sup>Conservation Ecology Center, Smithsonian National Zoo & Conservation Biology Institute, Front Royal, VA, USA

<sup>8</sup>Forest Research Institute Malaysia, Selangor, Kepong, Malaysia

<sup>9</sup>National Biobank of Thailand, National Science and Technology Development Agency, Klongluang, Pathum Thani, Thailand

<sup>10</sup>Institute of Molecular Biosciences, Mahidol University, Salaya, Nakhon Pathom, Thailand

<sup>11</sup>Programa de Ecosistemas y Recursos Naturales, Instituto SINCHI, Bogotá, Colombia

<sup>12</sup>Department of Biological Sciences, National Sun Yat-sen University, Kaohsiung, Taiwan

<sup>13</sup>Department of Natural Resources and Environmental Studies, National Dong Hwa University, Hualien, Taiwan

<sup>14</sup>Department of Plant Science, University of Buea, Buea, Cameroon

<sup>15</sup>Department of Ecology and Evolutionary Biology, Tulane University, New Orleans, LA, USA

<sup>16</sup>Departamento de Ciencias Forestales, Universidad Nacional de Colombia Sede Medellín, Medellín, Colombia

<sup>17</sup>Department of Science and Technology, Faculty of Applied Sciences, Uva Wellassa University, Badulla, Sri Lanka

<sup>18</sup>Faculté de Gestion des Ressources Naturelles Renouvelables & Faculté des Sciences, Université de Kisangani, Kisangani, Democratic Republic of Congo

<sup>19</sup>Environmental Studies Department, University of California, Santa Cruz, CA, USA

<sup>20</sup>Department of Botany, University of Peradeniya, Peradeniya, Sri Lanka

<sup>21</sup>Department of Biology and Cofrin Center for Biodiversity, University of Wisconsin-Green Bay, Green Bay, WI, USA

<sup>22</sup>Environmental Change Institute, School of Geography and the Environment, University of Oxford, Oxford, UK

<sup>23</sup>Osaka City University, Osaka, Japan

<sup>24</sup>Osaka Metropolitan University, Osaka, Japan

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Jens-Christian Svensson

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<sup>25</sup>School of Forest, Fisheries, and Geomatics Sciences, University of Florida, Gainesville, FL, USA

<sup>26</sup>Department of Forest Ecology, Silva Tarouca Research Institute, Brno, Czech Republic

<sup>27</sup>Department of Wildland Resources, Utah State University, Logan, UT, USA

<sup>28</sup>Ecology Center, Utah State University, Logan, UT, USA

<sup>29</sup>Faculte des Sciences, Universite de Kisangani, Kisangani, Democratic Republic of Congo

<sup>30</sup>Forest Department Sarawak, Kuching, Malaysia

<sup>31</sup>Department of Biology, Indiana University, Bloomington, IN, USA

<sup>32</sup>Center for Interdisciplinary Research on Ecology and Sustainability, National Dong Hwa University, Hualien, Taiwan

<sup>33</sup>School of Biological Sciences, Washington State University, Vancouver, WA, USA

<sup>34</sup>UK Centre for Ecology & Hydrology, Edinburgh, UK

<sup>35</sup>Department of Ecology, Evolution, and Environmental Biology, Columbia University, New York, NY, USA

<sup>36</sup>Department Environmental Sciences, University of Puerto Rico, Puerto Rico

<sup>37</sup>Institute of Environmental Sciences, Leiden University, Leiden, the Netherlands

<sup>38</sup>Theoretical Ecology, University Regensburg, Regensburg, Germany

<sup>39</sup>Bayreuth Center of Ecology and Environmental Research (BayCEER), University of Bayreuth, Bayreuth, Germany

**Correspondence:** Melina de Souza Leite ([melina.leite@ib.usp.br](mailto:melina.leite@ib.usp.br)), Marco D. Visser ([m.d.visser@cml.leidenuniv.nl](mailto:m.d.visser@cml.leidenuniv.nl))

The future trajectory of global forests is closely intertwined with tree demography, and a major fundamental goal in ecology is to understand the key mechanisms governing spatio-temporal patterns in tree population dynamics. While previous research has made substantial progress in identifying the mechanisms individually, their relative importance among forests remains unclear mainly due to practical limitations. One approach to overcome these limitations is to group mechanisms according to their shared effects on the variability of tree vital rates and quantify patterns therein. We developed a conceptual and statistical framework (variance partitioning of Bayesian multilevel models) that attributes the variability in tree growth, mortality, and recruitment to variation in species, space, and time, and their interactions – categories we refer to as *organising principles* (OPs). We applied the framework to data from 21 forest plots covering more than 2.9 million trees of approximately 6500 species. We found that differences among species, the *species* OP, proved a major source of variability in tree vital rates, explaining 28–33% of demographic variance alone, and 14–17% in interaction with *space*, totalling 40–43%. Our results support the hypothesis that the range of vital rates is similar across global forests. However, the average variability among species declined with species richness, indicating that diverse forests featured smaller interspecific differences in vital rates. Moreover, decomposing the variance in vital rates into the proposed OPs showed the importance of unexplained variability, which includes individual variation, in tree demography. A focus on how demographic variance is organized in forests can facilitate the construction of more targeted models with clearer expectations of which covariates might drive a vital rate. This study therefore highlights the most promising avenues for future research, both in terms of understanding the relative contributions of groups of mechanisms to forest demography and diversity, and for improving projections of forest ecosystems.

**Keywords:** multilevel models, spatial and temporal variation, species differences, temperate forests, tree demography, tropical forests, variance partitioning, vital rates

## Introduction

Forests are an integral component of the global carbon cycle (Anderson-Teixeira et al. 2021) and are home to a majority of the terrestrial biodiversity (Pillay et al. 2022). Changes in climate and land use threaten forests, but anticipating how these diverse systems might respond is challenged by the broad array of mechanisms that might determine forest structure and function. Further, these mechanisms differ in their influence over space and time, and are difficult to measure at the appropriate scale of their potential influence. A common approach to quantifying forest function is through the analysis of tree demography (Griffith et al. 2016): the growth, survival, and reproduction of individual trees. These vital (i.e. demographic) rates combine to determine key features of forests, such as biomass stocks and fluxes (Needham et al. 2022), structural complexity (Kohyama 1993), and diversity (Lasky et al. 2014). A better understanding of forest demography can advance the

development and testing of ecological theories such as the role of coexistence (Broekman et al. 2019, Hülsmann et al. 2020, 2024) and niche (Kohyama 1993, Lasky et al. 2014) in community ecology. Moreover, demography has been identified as critical for more accurately modelling the terrestrial component in earth system models (Fisher et al. 2018) and projecting the future of the terrestrial carbon sink (Pan et al. 2011). Even small changes, over space and time, in tree vital rates can affect carbon cycles (Needham et al. 2022) and thus the extent to which climate change can be mitigated by forests (Canadell and Raupach 2008).

Vital rates are influenced by interacting mechanisms across spatial and temporal scales, creating a challenge to the inclusion of demography in forest models (Weng et al. 2015). Many of these mechanisms are difficult or impossible to measure directly, leading to the use of imperfect proxies (Swenson et al. 2020). Besides, data analysis is usually restricted to a few non-interacting proxies, making it even

Table 1. Seven organising principles (OPs) and the mechanisms that are associated with them, i.e. by creating variability of vital rates in the associated dimensions species, space, and time and their interactions. References are example studies for the mechanisms.

Organising principles (OPs)	Related mechanisms and examples
<b>Species</b> Trees of different species have different vital rates	Natural selection in response to biotic and abiotic stressors creates variation in evolutionary strategies that leads to unique geno- and phenotypes among individual trees manifested in different species. Species then display difference in their vital rates, as evidenced as follows: <ul style="list-style-type: none"><li>Species have different growth forms (e.g. shrubs and trees), dispersal abilities, and regeneration strategies (<a href="#">Martínez-Ramos et al. 2021</a>) that are related to different allocation strategies (<a href="#">Rüger et al. 2020</a>), also known as life history strategies, leading to different demographic niches (<a href="#">Condit et al. 2006</a>) and the emergence of interspecific demographic trade-offs, such as growth-mortality, recruitment-mortality (<a href="#">Russo et al. 2008</a>), and stature-recruitment (<a href="#">Rüger et al. 2018</a>)</li><li>All these differences are potentially related to species functional traits (<a href="#">Poorter et al. 2008</a>, <a href="#">Adler et al. 2014</a>)</li></ul>
<b>Space</b> Trees in different locations (quadrats) have different vital rates	Spatial heterogeneity created by variability in soil and topography as well as by differences in stand structure results in spatial differences of resource availability (nutrients, moisture, light) and environmental stressors (e.g. wind). In response, tree vital rates can be consistently higher in some areas than in others ( <a href="#">Arellano 2019</a> ): <ul style="list-style-type: none"><li>Tree mortality may be higher on hilltops given lower water availability in soil and higher wind disturbances (<a href="#">Zuleta et al. 2020</a>)</li><li>Tree growth is faster and mortality higher in nutrient-rich soils (<a href="#">Russo et al. 2005</a>, <a href="#">Lévesque et al. 2016</a>)</li></ul>
<b>Time</b> Trees during different time periods have different vital rates	Environmental conditions are not stable in time but vary with climate and in response to disturbances, jointly affecting all species across a forest (synchronised effects): <ul style="list-style-type: none"><li>Cyclones and other drastic climatic disturbances can kill many trees at once in a forest (<a href="#">Uriarte et al. 2019</a>)</li><li>Severe droughts can decrease growth and/or increase mortality directly (<a href="#">McDowell et al. 2020</a>) or indirectly by increasing the propensity of disease outbreaks (<a href="#">Negrón et al. 2009</a>)</li><li>Irregular masting events and rainfall affect growth and survival of seedlings (<a href="#">Martini et al. 2022</a>)</li></ul>
<b>Species × space</b> Trees of different species in the same location (quadrat) have different vital rates	Due to spatial niche effects, species have different environmental preferences that, in combination with spatial variability, create certain habitats where some species perform better than others. For example: <ul style="list-style-type: none"><li>Species adapted to low light availability have lower mortality in denser areas (<a href="#">Jurinitz et al. 2013</a>)</li><li>Species with more dispersive seeds recruit more in open gaps (<a href="#">Clark et al. 2018</a>)</li><li>Soil fertility affects species in different ways (<a href="#">Russo et al. 2008</a>)</li></ul> Conspecific and/or heterospecific negative density dependence may induce different vital rates in areas with different local population density ( <a href="#">Hülsmann et al. 2020, 2024</a> )
<b>Species × time</b> Trees of different species during the same time period have different vital rates	Species environmental preferences also create temporal niche effects that lead to asynchronous species responses to temporal variability ( <a href="#">Fung et al. 2020</a> ). For example: <ul style="list-style-type: none"><li>Species that are vulnerable to drought have higher mortality than those that are resistant or resilient (<a href="#">Chen et al. 2019</a>)</li><li>Species with more dispersive seeds recruit more in a favourable year (<a href="#">Clark et al. 2018</a>)</li><li>Species with high wood density suffer lower immediate mortality after hurricanes (<a href="#">Uriarte et al. 2019</a>)</li></ul>
<b>Space × time</b> Trees in the same location during different periods have different vital rates	Gap dynamics: large tree falls open temporal gaps in the forest changing the environmental conditions of the surrounding area for a certain period ( <a href="#">Kohyama 1993</a> ): <ul style="list-style-type: none"><li>Fallen trees or trees killed by lightning increase immediate local mortality in the area surrounding it (<a href="#">Gora et al. 2021</a>)</li><li>Open gaps increase light availability, allowing faster growth (<a href="#">Brokaw 1987</a>) of understory trees and recruitment (<a href="#">Wright et al. 2003</a>) but just during specific time periods</li></ul> Climate effects can manifest themselves differently depending on the prevailing basic conditions in a given area. For example: <ul style="list-style-type: none"><li>Drought events increase mortality disproportionately in valleys than on hilltops or ridges (<a href="#">Zuleta et al. 2017</a>)</li><li>Soil nutrients can influence growth response to drought (<a href="#">Lévesque et al. 2016</a>)</li></ul>
<b>Species × space × time + individual</b> Trees of the same species in different locations and during different time periods have different vital rates + Individual trees have different vital rates	Individual variation in vital rates given ontogeny, genetic, and phenotypic variation ( <a href="#">Clark 2010</a> , <a href="#">Clark et al. 2010</a> ), and spatial variation at the microscale ( <a href="#">Schwartz et al. 2020</a> ) <ul style="list-style-type: none"><li>Trees of different sizes and multi-stemmed trees have different mortality (<a href="#">Johnson et al. 2018</a>, <a href="#">Su et al. 2020</a>) and growth rates (<a href="#">Lu et al. 2021</a>)</li><li>Functional traits influence growth depending on the size of the individuals (<a href="#">Gibert et al. 2016</a>)</li><li>Local biotic interactions, as higher-order interactions, change individual vital rates (<a href="#">Li et al. 2020</a>)</li></ul> Phenotypic plasticity changes the observed vital rates of different species due to temporal and spatial environmental conditions. Plasticity may be seen in individual-level functional traits ( <a href="#">Burns and Strauss 2012</a> )

harder to distinguish and compare mechanisms' relative importance. There exists, however, a higher level of information that may guide demographic analyses focused on mechanisms: the patterns in vital rates themselves. The three vital rates and the contextual variables ('dimensions') associated with them offer an opportunity to organise the elements of forest dynamics in ways that help to infer the potential mechanisms that structure forests. For example, through natural selection, species have developed different strategies to acquire and allocate resources. This results in a species dimension that represents the range of phenotypes among species (Díaz et al. 2016) and, thus, also the observed vital rates of individual species (Johnson et al. 2018, Rüger et al. 2018, Needham et al. 2022). Moreover, as resource availability and stressors vary along spatial and temporal dimensions, the environmental conditions of a forest also structure the vital rates of the trees, e.g. soil and topography vary across space (Zuleta et al. 2020) and drought conditions over time (Chen et al. 2019). Finally, all these dimensions (species, space, and time) have interactive effects. Functional traits vary between species and cause differential responses along spatial and temporal dimensions, for example when drought-tolerant species respond differently to a climatic event (Kupers et al. 2019). Gap dynamics change over both space and time, and tree responses change as forest gaps close (Wright et al. 2003). Patterns of how variability in vital rates is partitioned along these key dimensions can thus reveal how important biotic and abiotic drivers are influencing tree demography and, by extension, forest dynamics.

We present a conceptual framework that groups the mechanisms creating variation in vital rates as being related to species, space, and time. Together, these three dimensions and their interactions form seven organising principles (OPs, Table 1). When the mechanisms that drive tree vital rates operate on unique combinations of these dimensions, quantifying the variability in vital rates that each OP describes may provide insights into the strength and relative importance of the mechanisms that might potentially be correlated with that rate (Table 1). The statistical counterpart to this conceptual framework is variance partitioning analysis, a technique that decomposes the variability in the response variable to the groups of interest in the data (Searle et al. 2006) using multilevel models (MLMs) (McMahon and Diez 2007, Visser et al. 2016). In our framework, we decompose forest demographic data across OPs and quantify the relative importance of each OP by estimating and partitioning the variance in each vital rate (Browne et al. 2005). By attributing the total variability in vital rates to the different OPs, a broad assessment of the structure of variation in vital rates can be accomplished (Table 1).

We applied this framework to a set of 21 large (6–52 ha) and globally distributed forest dynamics plots (Davies et al. 2021). We then compared the relative importance of the OPs for each vital rate at each forest with the goal of identifying consistent patterns in which OPs capture variation in vital rates: 1) among vital rates, i.e. investigating if some OPs are more important than others for specific vital rates; 2) across spatial scales (grain size), given the nature of scale

dependency of ecological processes; and 3) among forests globally to understand how patterns may differ depending on forest diversity and structure. In doing so, we provide macroecological patterns of the relative importance of OPs and, thus, the first approximate assessment of their associated mechanisms in generating variation in global forest demography. Our framework aims to facilitate hypothesis-driven research on mechanisms by first describing the higher-level patterns of vital rate variability and giving important insights to the ecological dimensions 'at which the action lies' (Browne et al. 2005).

## Material and methods

### Tree census data

We used data from 21 forest dynamics plots (Fig. 1A) from the Forest Global Earth Observatory network (ForestGEO, Davies et al. 2021). In each plot, all stems with a diameter  $\geq 1$  cm at 1.3 m above the ground (diameter at breast height, DBH) were mapped, identified, and repeatedly measured using a standardised protocol. Plots used in this study ranged in size between 6 and 52 ha, with an inter-census measurement interval of approximately 5 years (range: 3 to 10 years). The area within each forest plot was subdivided into quadrats of equal sizes (see 'Organizing principles across spatial scales'). All forest plots had at least two censuses. The forest plots covered a wide range of environmental, climatic, and edaphic conditions, with the number of species per plot varying two orders of magnitude from 12 to 1402 (including morpho-species). In total, approximately 2.9 million trees from more than 6500 species were repeatedly censused over periods of 3–40 years in more than 575 ha. For summary information on the plots and further details on how tree census data were processed see the Supporting information.

### Vital rate definition and modelling

We analysed growth, mortality, and recruitment as annual rates by using vital rate information at the level of individual trees and applying the variance partitioning analysis per forest plot and vital rate. Annual individual growth was calculated as DBH increment in millimetres of living trees, divided by the individual's census interval length in years, and modelled using MLMs with a normal distribution.

Variance partitioning of mortality and recruitment is less intuitive than growth, because although every individual has a unique, observable growth rate, individual trees only provide an observable status (e.g. individuals are either alive or dead for mortality rate). However, we can estimate latent mortality and recruitment rates for individuals belonging to the same population, space, and time by calculating per capita vital rates (sensu Kohyama et al. 2018). Further, although the variance of individual binary observations is fixed at 1.68 (the standard deviation of a logistic distribution [see below]), this term has meaning when compared to other sources of measurable variance, such as across populations, years, or spatial aggregations. Therefore, mortality

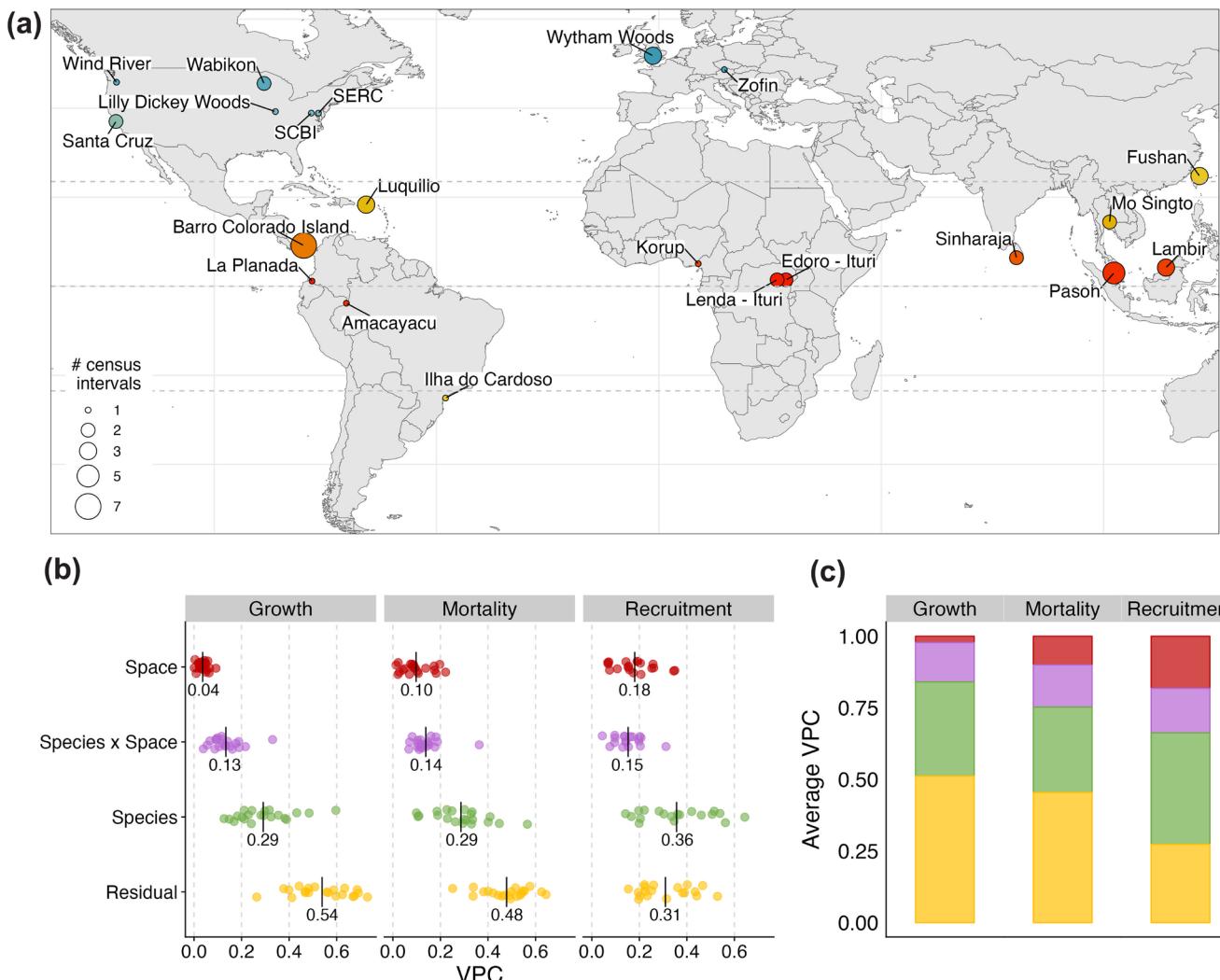


Figure 1. (a) Global distribution of the 21 forest dynamic plots. (b) Variance partition coefficients (VPC) of the *organising principles* (OPs) per vital rate – growth, mortality, and recruitment – with mean values indicated as black vertical lines and numbers. (c) Average VPCs across all plots, where colours correspond to the same OPs as in (b). Models were fitted at the  $5 \times 5$  m grain size. Each forest plot in (a) is coloured by latitude and the size of the circle is related to the number of census intervals.

was estimated from the status of trees – alive or dead – in each consecutive census assuming a binomial distribution (Kohyama et al. 2018). Mortality rates were annualised by using a complementary log-log link function (cloglog), where the log-transformed time in years between individual measurements is included as an offset term (Fortin et al. 2008, Johnson et al. 2018).

Recruitment was defined as the final per capita recruitment rate (Kohyama et al. 2018), which denotes the proportion of trees that are new recruits (i.e. not present in the previous census) and can be interpreted as the probability of an individual tree being new. Recruitment rates were estimated using the same modelling approach as for mortality, i.e. a binomial model with a cloglog link function and time interval length as an offset term. Because there is no time interval associated with individual recruits as they have not been monitored in the previous census, the time interval for recruitment was calculated as the mean time interval of the

survivors in the same quadrat. If there were no survivors in a specific quadrat, we used the mean time interval between the respective censuses from the entire plot.

### Variance partitioning analysis

To quantify the variation in vital rates associated with each OP, we applied variance partitioning to MLMs fitted separately for each vital rate and forest plot. MLMs are particularly useful for variance decomposition as they are able to reflect that ecological datasets contain identifiable hierarchical units, groups, or clusters (McMahon and Diez 2007). MLMs can account for such interdependence by partitioning the total variance into different components of variation due to each cluster (example in Table 1). We included *species*, quadrat (*space*) and census interval (*time*) and their two-way interactions as variance components. With that, we estimated the variance associated with each OP while respecting the

hierarchical structure of the data. Following the convention of MLMs, the general structure of our models (Eq. 1) is:

$$Y_{ijkl} = \beta_0 + s_i + q_j + t_k + sq_{ij} + st_{ik} + qt_{jk} + e_{ijkl} \quad (1)$$

where  $Y_{ijkl}$  is the vital rate for individual observation  $l$ , from species  $i$ , in quadrat  $j$  and time interval  $k$ .  $\beta_0$  is the global intercept.  $s_i$ ,  $q_j$ ,  $t_k$  are effects at the *species*, *space* (quadrat) and *time* level, respectively (commonly termed as random effects in mixed-effects models), while  $sq_{ij}$ ,  $st_{ik}$ ,  $qt_{jk}$  are effects of interactions between OPs: *species*  $\times$  *space*, *species*  $\times$  *time* and *space*  $\times$  *time*. All parameters are represented by a normal distribution with mean zero and their respective variances  $\sigma_s^2$ ,  $\sigma_q^2$ ,  $\sigma_t^2$ ,  $\sigma_{sq}^2$ ,  $\sigma_{st}^2$ ,  $\sigma_{qt}^2$  with no covariance being modelled. The residual variance ( $e_{ijkl}$ ) represents the variance explained by the three-way interaction *species*  $\times$  *space*  $\times$  *time*, and any unexplained variation among observations including non-separable measurement error and individual variation (Table 1). Residual variance in growth models assume a normally distributed error. For mortality and recruitment, modelled with binomial distributions, the residual variance at the linear predictor scale is the expected variance for the binomial distribution ( $\pi^2/6 \sim 1.68$ ) (Nakagawa et al. 2017). We decided not to include the three-way interaction *species*  $\times$  *space*  $\times$  *time* because most of the clusters formed by the combinations of the species, space, and time categories would have only one tree, i.e. not enough observations per cluster especially for the small spatial grains, preventing the model from correctly computing variance among these clusters. The same reasoning applies to the individual variance, where the number of individual trees with only one measurement is high for all forests plots. It means that any variability given to the three-way interaction *species*  $\times$  *space*  $\times$  *time* and the individual variation will be attributed to the residual variance in growth models (normal distribution) and will not be accounted for in the mortality and recruitment models (binomial distribution).

To partition the total variance of the vital rates among the individual OPs, we calculated variance partition coefficients (VPCs) (Browne et al. 2005). The VPC of each OP was calculated as the proportion of its variance to the total variance of the model. It is worth noting that we did not include fixed effects in the models intentionally, in contrast to the usual statistical approach when searching for specific mechanisms. We did this because all mechanisms are considered through OPs, which represent the dimensions at which they generate variability.

All data analyses were performed using R ([www.r-project.org](http://www.r-project.org), ver. 4.3.1), using the R package ‘brms’ (Bürkner 2017) to build Bayesian MLMs. For all estimated parameters, we used ‘brms’ default weakly informative prior distributions. For each model, we ran three Monte Carlo Markov chains with 3000 iterations, discarding the first 1000 iterations and thinning with an interval of five, resulting in 1200 posterior samples. We checked the convergence of the chains using the Gelman–Rubin criterion and by visually inspecting trace plots of estimated coefficients.

## Analysis framework

### Organising principles among vital rates

To assess the relative importance of the OPs among vital rates, we compared the VPC results for each vital rate among the 21 forest plots. However, because 16 forests had too few census intervals (i.e. fewer than three, see below), we fit a reduced version of the variance partitioning analysis (Eq. 1) without the temporal OPs (dropping the variances  $\sigma_t^2$ ,  $\sigma_{st}^2$ ,  $\sigma_{qt}^2$ ). We ran separate analysis for each time interval of the same forest plot and averaged variances for forests with more than one census interval.

### Temporal organising principles

Currently, a bottleneck of our analysis is the scarcity of data for the temporal dimension of vital rates variability. For variance partitioning analysis, the estimation of the variance of a grouping variable (i.e. *time* OP) with a small number of levels may be biased towards zero (Oberpriller et al. 2022). In our data, only five forest plots in the (sub)tropics (Supporting information) presented a reasonable number of census intervals (i.e. at least four censuses spanning between 20 and 40 years) to be considered suitable for the VPC analysis including temporal OPs (Eq. 1). We fit these MLMs to ten random subsets of 5 ha each sampled from the full forest plots, where each subset was composed of five non-overlapping quadrats of 1 ha. This procedure was necessary to restrict computational time resulting from the large number of observations, especially on the large plots that are species-rich and of high tree density (i.e. Barro Colorado Island 50 ha, Lambir 50 ha, Pasoh 50 ha, Fig. 1a, Supporting information). Variance estimates of the OPs for each forest plot were averaged across estimates of the ten subsets.

### Organising principles across spatial scales

To assess how the relative importance of OP varies with spatial scale, i.e. how the choice of a specific grain size impacts VPCs, we divided each forest plot into non-overlapping quadrats with increasing size: 5  $\times$  5 m (0.0025 ha), 10  $\times$  10 m (0.01 ha), 20  $\times$  20 m (0.04 ha), 50  $\times$  50 m (0.25 ha), and 100  $\times$  100 m (1 ha). Depending on the size of the plot, we trimmed the data to fit within a rectangular region with edges that were even multiples of 100 m, discarding the data outside this area. This guaranteed that each plot could be evenly divided into quadrats of 1 ha and that the same area was analysed at all spatial scales. We ran variance partitioning analyses without and with temporal OPs, and averaged VPCs over all forest plots for each grain size and vital rate.

### Organising principles across a global species richness gradient

Globally, species richness is one of the most distinguishing characteristics of forests and strongly correlates, for instance, with latitude (Keil and Chase 2019), precipitation (Adler and Levine 2007), and biome history (Wiens and Donoghue 2004). The plots used in this analysis spanned two orders of magnitude in the number of species (12–1402, including morphospecies), offering a unique opportunity to explore if and how sources of variability in vital rates are associated

with species diversity. We therefore assessed how log-transformed rarefied species richness (c.f. Supporting information) is associated with the VPCs of *species*, *space*, *species*  $\times$  *space* and residual OP using Dirichlet regression from the R package 'DirichletReg' (Maier 2021), which is appropriate for response variables that are multiple categories of proportional data (Douma and Weedon 2019).

### Robustness analyses

We performed four extra analyses to make sure our VPCs estimates from the forest plots were robust to 1) different forest plot sizes (6–50 ha) for the models without temporal OPs, by subsampling and comparing VPCs of the same forest (Lambir) with the entire plot data; 2) to the approach of computing average VPCs for the model with temporal OPs from subsampled plots (10 samples of 5 ha each); 3) to changes in the modelling procedure, by including or excluding temporal OPs from the VPC analysis; and 4) to the presence of rare species on VPCs by excluding or including rare species. All these analyses are presented in the Supporting information.

VPCs estimates from all forest plots were robust to changes in plot size and subsampling data. VPC estimates also remained reliable after removing temporal OPs. Specifically, our main results were also robust to the presence of rare species, though excluding or regrouping rare species does result in small decreases in the species VPC, balanced by an increase in the residual and species  $\times$  space VPC (Supporting information).

## Results

### Organising principles among vital rates

When comparing the relative importance of the OPs for all 21 forests, we found that, despite large differences among the

plots with respect to climate, environment, species richness, etc., the relative importance of the OPs was relatively similar (Fig. 1). Generally, *species* was the most important VPC for explaining variance in all three vital rates, after the *residual* VPC. At the smallest spatial grain (quadrats at 5  $\times$  5 m), average *species* and *species*  $\times$  *space* VPCs varied little among vital rates, ranging from 29 to 36%, and 13 to 15%, respectively. The average *space* VPC was smaller for growth (4%), intermediate for mortality (10%), and larger for recruitment (19%). Residual VPCs were on average about half of the total variance for growth and mortality (55 and 47%, respectively) but smaller for recruitment (31%).

### Temporal organising principles

When analysing demographic data from the five forest plots with more than four consecutive censuses (grain size 5  $\times$  5 m), we found that *species* remained the most important VPC to explain variance in tree vital rates, except for growth, where the *species*  $\times$  *space* VPC was larger for four of the five plots (Fig. 2). Temporal OPs (*time*, *species*  $\times$  *time*, and *space*  $\times$  *time*) were especially important for mortality and recruitment, where VPCs of *space*  $\times$  *time* (on average 10 and 15%, respectively) were larger than VPCs of *species*  $\times$  *space* (on average 6 and 10%, respectively).

### Organising principles across spatial scales

When comparing average VPCs across five spatial grain sizes, we found that the relative importance of residual variation increased with grain size for all vital rates and more accentuated for growth (Fig. 3). For instance, for the models including temporal OPs (Fig. 3b), residual variation increased from 46% at the smallest grain (quadrats at 5  $\times$  5 m) to 71% at the largest grain (100  $\times$  100 m). In turn, the spatial OPs – *space*, *species*  $\times$  *space* and *space*  $\times$  *time* – consistently decreased in

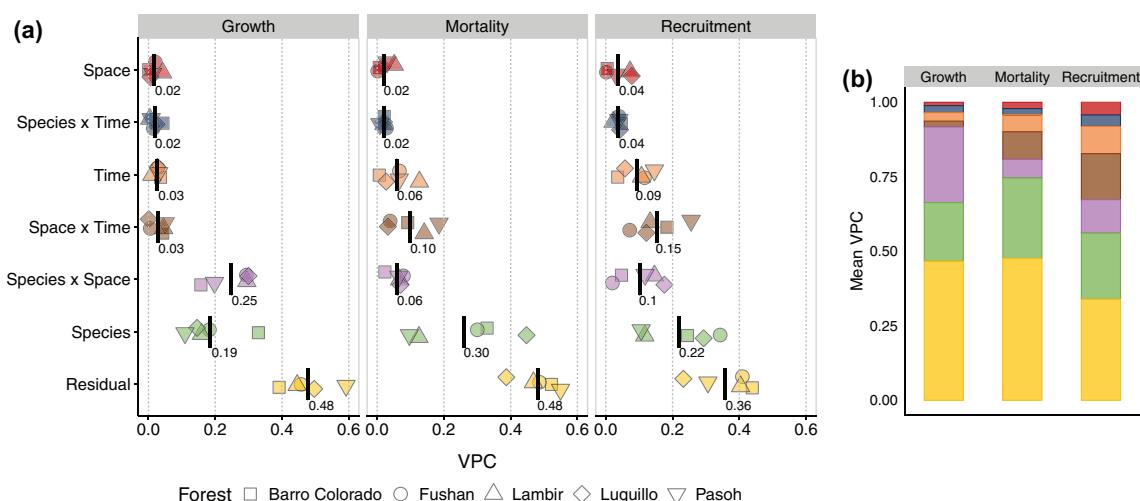


Figure 2. (a) Variance partition coefficient (VPC) of the organising principles (OPs) per vital rate – growth, mortality, and recruitment – for the five forest plots with at least four censuses (different point shapes). Average VPCs across plots are presented as black lines and numbers. (b) Average VPCs across the five plots, where colours correspond to the same OPs as in (a). Models were fitted at the 5  $\times$  5 m grain size. See Fig. 1a for forest plot locations.

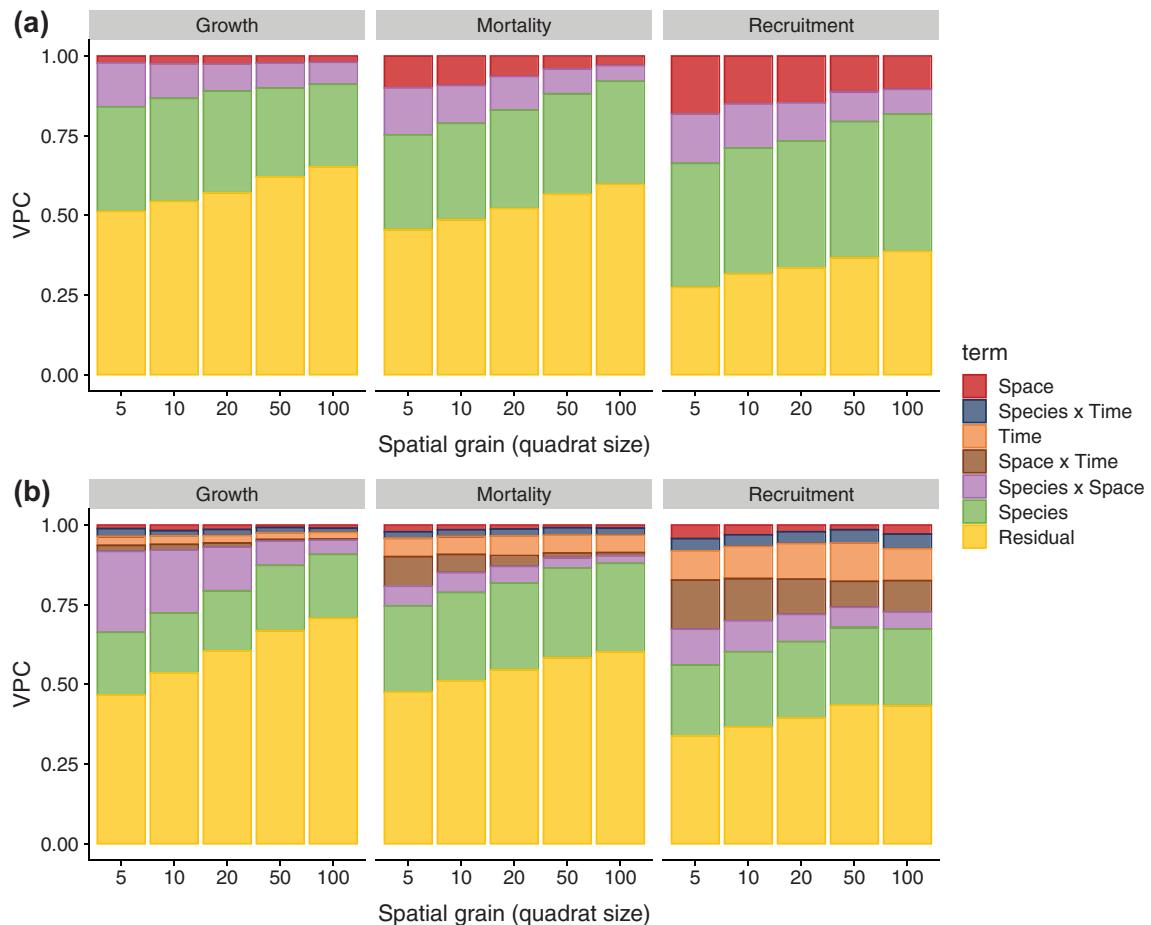


Figure 3. Average variance partition coefficients (VPCs) of each organising principle (OP) across five spatial grain sizes (quadrats from  $5 \times 5$  to  $100 \times 100$  m) for the vital rates growth, mortality, and recruitment: (a) reduced models without temporal OPs for all 21 forests plots, and (b) full models with temporal OPs for the five (sub)tropical forest plots with enough censuses (Barro Colorado Island, Fushan, Lambir, Luquillo, and Pasoh).

relative importance with increasing spatial grain for all vital rates. The OPs *species* and *species × time* remained almost equally important across spatial grains.

#### Organising principles across a global species richness gradient

While the *species* OP was the most important VPC for the vital rates throughout the forests, we also found that the importance of the *species* VPC decreased with species richness for recruitment and growth, but not for mortality (Fig. 4). The decrease in the *species* VPC for growth and recruitment was led by a decrease in the species standard deviation (Supporting information). This result was robust to the presence of rare species (Supporting information). The other OPs showed no significant changes with species richness (Fig. 4).

## Discussion

Innumerable mechanisms operate and interact in forests and leave fingerprints of their integrated effects in tree vital rates, i.e. growth, survival, and recruitment, which together drive

forest dynamics. Here, we used a conceptual and statistical framework to identify organising principles (OPs, Table 1) and quantify the associated variability among vital rates for more than 2.9 million trees of approximately 6500 species in 21 forests across the globe. This, in turn, allows a first assessment of the relative importance of the groups of mechanisms underlying each OP, offering a first step in narrowing down which of the mechanisms are critical for structuring global forests. In the following sections, we summarise our most striking findings, discuss some potentially important mechanisms, and provide recommendations for an agenda to study tree vital rates.

#### Species is a major source of variability in tree vital rates

We found that *species* was the most important OP for all tree vital rates, explaining on average between 29 and 36% of the demographic variance across all forest sites (Fig. 1). *Species* in interaction with *space* added another 13–15% variance explained, meaning that a total of 42–51% of demographic variation can be partitioned towards species differences and species-specific responses to spatial heterogeneity (Table 1).

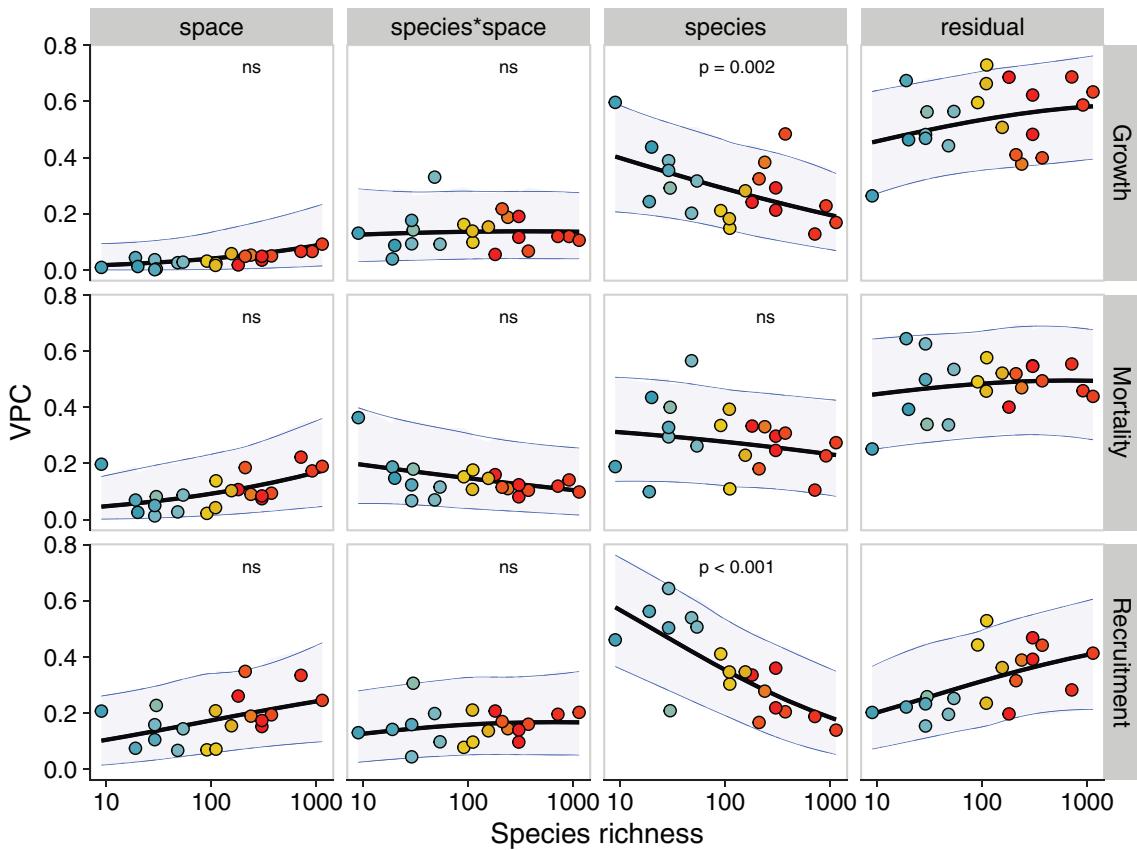


Figure 4. Variance partition coefficients (VPCs) for the organising principles (OPs) *species*, *space*, *species x space*, and *residual* against species richness of 21 forest plots. OPs were estimated with the reduced model (Eq. 1) without temporal OPs. Black lines are fitted relationships obtained from Dirichlet regressions of VPCs against species richness; shaded blue areas are the 95% prediction intervals. p-values are shown only for the significant values after Bonferroni correction ( $\alpha = 0.016$ ). Residual VPCs are reference categories and thus were not tested for significance. Each forest plot (dots) is coloured by absolute latitude as in Fig. 1a. Species richness on the x-axis is at the logarithmic scale with base 10.

In contrast, *space* and *time* OPs explained relatively little variability in vital rates (Fig. 1–2). Our results, therefore, suggest that – at least at the temporal and spatial scales covered by our datasets – spatio-temporally varying factors alone contribute less to demographic variance than evolutionary history and adaptations to the environment. Grouping individuals into species thus creates a globally important cluster of demographic variation that appears consistently most important across a wide range of forests.

Our results on the importance of species variability to demographic rates support numerous ongoing research agendas. Efforts to include more realistic representation of species strategies in global vegetation models appear to be a promising route (Fisher et al. 2018, Anderegg et al. 2022), regardless of whether forest dynamics are studied in local tree neighbourhoods or larger spatial units (Fig. 3). We expect that accounting for species differences can explain up to  $\sim 36\%$  of demographic variation, while additionally accounting for small-scale species–environment associations (Messier et al. 2010, Lasky et al. 2014) might further improve this to almost half of the variation explained. More critically, however, our work shows that there are clear limits to the improvement that more realistic representations of

species can bring. Studies including species strategies typically rely on functional traits (Rubio and Swenson 2022) or demographic trade-offs (Rüger et al. 2020, Russo et al. 2021), i.e. simplifications that explain only about half of the among-species variation (Visser et al. 2016). Nevertheless, the global importance of species in clustering demographic variance and its consistency across spatial scales indicates that endeavours seeking to better map species differences may have been undervalued compared to those focussing on spatial and temporal effects.

#### Temporal variability acts mostly on recruitment and mortality and in interaction with space

In contrast to variability among species, temporal OPs played a minor role for variability in tree vital rates, as time interval alone was responsible for only 3–7% of total variability for plots with sufficient data. Although these data probably have the most comprehensive temporal coverage of large forest areas currently available, our findings might reflect the relatively short time frame (20–40 years), the low temporal resolution (approximately five years), and the fact that we could only analyse data from five tropical and

subtropical forests. Nevertheless, variability between census intervals was detected in recruitment and to a lesser degree in mortality but was relatively unimportant for growth (Fig. 2). A possible explanation is that growth rates fluctuate within shorter periods than our 5-year census interval can capture (Dobbertin 2005), while recruitment and mortality may exhibit several bad or good years in a row (Schwartz et al. 2020).

Temporal effects were most important in interaction with space which, for instance, could indicate gap dynamics that jointly affect vital rates of most trees (Kohyama 1993). This interpretation is consistent with the result that the *space × time* OP was more important for mortality and recruitment than for growth – as mortality in gaps is known to be ‘spatially contagious’, with falling trees killing multiple neighbours (Araujo et al. 2021), and the resulting gaps generally favour recruitment for many species (Brokaw 1987). Additionally, some of the variability in the *space × time* OP could be the result of climatic events acting differently depending on local conditions, such as droughts that harm trees more in valleys than on ridges (Zuleta et al. 2017).

Our results on temporal OPs support a research agenda that should analyse the importance of climatic and/or temporal effects on vital rates in interaction with spatial effects. Moreover, we advocate for datasets with higher temporal resolution and longer time series, which would allow the capture of larger but infrequent disturbances (Šamonil et al. 2013), thereby revealing more of the demographic importance of environmental fluctuations and temporal niches (Fung et al. 2020).

### Small spatial grain variability is important for tree vital rates

Spatial OPs were important for vital rate variability mostly in interaction with species for growth, and time for mortality and recruitment (Fig. 1–2), indicating the importance of spatial niches and patch dynamics (previous section). Alone, *space* was the least important OP and only created considerable variability in models without time (Fig. 1). However, it may be possible that some spatial variability could still be present in the residual variance, since we used a simple, discrete spatial structure without accounting for spatial autocorrelation or more sophisticated spatial analysis.

Spatially acting mechanisms were best detected by dividing the plots into quadrats of  $5 \times 5$  m (Fig. 3), indicating that trees interact and respond to local conditions at scales of a few metres through local mechanisms such as gap dynamics, competition, crown damage, and micro-topography (Schwartz et al. 2020). Further decreasing the spatial grain would then move below the scale of tree crowns, and begin to merely assign quadrats to single trees, here reflected by residual variance. With increasing grain size, less variability is explained by spatial mechanisms (Cáceres et al. 2012). Consequently, vital rates become less predictable at larger spatial grain. Nevertheless, even at the largest quadrat size of  $100 \times 100$  m, spatial OPs still explained a reasonable part of the variability, with the consequence that tree species also

seem to distinctly respond to environmental heterogeneity over larger areas (de Knecht et al. 2010), probably because of topography, water resources, and soil nutrients (Russo et al. 2005, 2008, Zuleta et al. 2020).

### Large proportion of unexplained variability in tree vital rates

Residual variance was consistently the dominant component of the vital rate VPCs across sites and in the temporal and spatial analyses. In our variance partitioning analyses, residual variance represents the variance in the response that cannot be attributed to any of the grouping factors (here, the OPs). On one hand, this result encourages more detailed models that might include covariates that ‘explain’ differences among individual trees. For instance, both growth and mortality are known to differ across ontogeny, and thus tree size (e.g. DBH) should be able to explain some of the residual variance (Hülsmann et al. 2018). Moreover, functional traits at the individual level (Su et al. 2020) and structures that explicitly deal with spatial (Wiegand et al. 2017) and temporal autocorrelation may explain additional differences in individual vital rates. On the other hand, there are intrinsic limits to what can be explained by even the most detailed models, as the residual variance also includes inherent noise. This noise is the result of misattribution of species, mapping error, or measurement error (Dettlo et al. 2019) and chaotic behaviour known to exist in many biological systems (Benincà et al. 2015).

### Globally, variability among species declines with species richness

Across plots, increasing species richness was associated with decreasing relative importance of the *species* OP in growth and recruitment (Fig. 4). This trend was robust to one of the most probable sources of bias, i.e. differences in species rarity between forest plots. Although species richness can strongly correlate with other environmental drivers (e.g. latitude, rainfall, biogeography), we consider that the decreasing relative importance of the species OP with species richness reflects a true macroecological pattern that could be further explored. Moreover, the decrease in the species VPC was determined by a decrease in the respective variance estimates, and not by an increase of variances related to the other OPs (Supporting information). Similarly, Condit et al. (2006) found across ten tropical forests (seven in common with this study) that the range of species-specific mortality and growth rates decreased with higher species richness.

These results underpin that – in contrast to expectations of niche theory – the most diverse forests feature the lowest interspecific variation in vital rates. Following the rationale of niche theory, diverse forests should have more demographic niches than low-diversity forests, as more niches allow more species to have equivalent fitness, thus favouring species coexistence (Chesson 2000). The lack of evidence for wider demographic ranges in species-rich forests (this study, Condit et al. 2006, Clark 2010) suggests that

demographic niches play a minor role for large-scale diversity patterns, hinting towards more neutral dynamics (Hubbell 2006). However, coexistence is inherently high dimensional, and comparing mean species values across low dimensions (a few vital rates) only partly represents the full niche space (Clark 2010).

## Conclusions

As the mechanisms that influence vital rates can be grouped by the dimensions at which they operate and interact, patterns of how variance is partitioned along key dimensions can reveal how important various biotic and abiotic mechanisms are in influencing tree demography and hence forest dynamics. Here, we have shown that variance partitioning of vital rates among key ecological dimensions, i.e. species, space, and time, has the potential to provide a first step in identifying the structuring processes of global forest dynamics. We found that species differences were a major source of variability in tree vital rates, while temporal variability acted mostly on recruitment and in interaction with spatial variability. Small grain sizes captured most of the spatial variability, but there were still larger proportions of unexplained variability in vital rates, probably due to individual variation. Most intriguing, we found that, globally, variability among species declined with species richness. In summary, species in highly diverse forests present redundant vital rates that do not add to the diversity of demographic types, highlighting the challenges of studying and predicting changes in hyper-diverse systems.

The proposed framework highlights the most promising avenues for future research both in terms of understanding the relative contributions of groups of mechanisms to forest demography and diversity, and for predicting forest ecosystems. We hope future studies may benefit from using this approach as a conceptual and modelling approach to narrow down which of the mechanisms are critical for structuring global forests.

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## Author contributions

**Melina de Souza Leite**: Conceptualization (lead); Data curation (equal); Formal analysis (lead); Funding acquisition (equal); Investigation (lead); Methodology (equal); Project administration (lead); Software (lead); Validation (lead); Visualization (lead); Writing – original draft (equal); Writing – review and editing (lead). **Sean M. McMahon**: Conceptualization (equal); Formal analysis (supporting); Funding acquisition (equal); Investigation (equal); Methodology (equal); Supervision (equal); Writing – original draft (equal); Writing – review and editing (equal). **Paulo Inácio Prado**: Conceptualization (equal); Formal analysis (supporting); Funding acquisition (equal); Investigation (equal); Methodology (equal); Supervision (equal); Writing – review and editing (equal). **Stuart J. Davies**: Funding acquisition (supporting); Resources (supporting); Writing – review and editing (supporting). **Alexandre A. de Oliveira**: Conceptualization (supporting); Funding acquisition (supporting); Writing – review and editing (supporting). **Hannes P. De Deurwaerder**: Resources (supporting); Writing – review and editing (supporting). **Salomón Aguilar**: Resources (supporting); Writing – review and editing (supporting). **Kristina J. Anderson-Teixeira**: Resources (supporting); Writing – review and editing (supporting). **Nurfarah Aqilah**: Resources (supporting); Writing – review and editing (supporting). **Norman A. Bourg**: Resources (supporting); Writing – review and editing (supporting). **Warren Y. Brockelman**: Resources (supporting); Writing – review and editing (supporting). **Nicolas Castaño**: Resources (supporting); Writing – review and editing (supporting). **Chia-Hao Chang-Yang**: Resources (supporting); Writing – review and editing (supporting). **Yu-Yun Chen**: Resources (supporting); Writing – review and editing (supporting). **George Chuyong**: Resources (supporting); Writing – review and editing (supporting). **Keith Clay**: Resources (supporting); Writing – review and editing (supporting). **Álvaro Duque**: Resources (supporting); Writing – review and editing (supporting). **Sisira Ediriweera**: Resources (supporting); Writing – review and editing (supporting). **Corneille E. N. Ewango**: Resources (supporting); Writing – review and editing (supporting). **Gregory Gilbert**: Resources (supporting); Writing – review and editing (supporting). **I. A. U. N. Gunatilleke**: Resources (supporting); Writing – review and editing (supporting). **C. V. S. Gunatilleke**: Resources (supporting); Writing – review and editing (supporting). **Robert Howe**: Resources (supporting); Writing – review and editing (supporting). **Walter Huaraca Huasco**: Resources (supporting); Writing – review and editing (supporting). **Akira Itoh**: Resources (supporting); Writing – review and editing (supporting). **Daniel J. Johnson**: Resources (supporting); Writing – review and

editing (supporting). **David Kenfack**: Resources (supporting); Writing – review and editing (supporting). **Kamil Král**: Resources (supporting); Writing – review and editing (supporting). **Yao Tze Leong**: Resources (supporting); Writing – review and editing (supporting). **James A. Lutz**: Resources (supporting); Writing – review and editing (supporting). **Jean-Remy Makana**: Resources (supporting); Writing – review and editing (supporting). **Yadvinder Malhi**: Resources (supporting); Writing – review and editing (supporting). **William J. McShea**: Resources (supporting); Writing – review and editing (supporting). **Mohizah Mohamad**: Resources (supporting); Writing – review and editing (supporting). **Musalmah Nasardin**: Resources (supporting); Writing – review and editing (supporting). **Anuttara Nathalang**: Resources (supporting); Writing – review and editing (supporting). **Geoffrey Parker**: Resources (supporting); Writing – review and editing (supporting). **Renan Parmigiani**: Resources (supporting); Writing – review and editing (supporting). **Rolando Pérez**: Resources (supporting); Writing – review and editing (supporting). **Richard P. Phillips**: Resources (supporting); Writing – review and editing (supporting). **Pavel Samonil**: Resources (supporting); Writing – review and editing (supporting). **I-Fang Sun**: Resources (supporting); Writing – review and editing (supporting). **Sylvester Tan**: Resources (supporting); Writing – review and editing (supporting). **Duncan Thomas**: Resources (supporting); Writing – review and editing (supporting). **Jill Thompson**: Resources (supporting); Writing – review and editing (supporting). **Maria Uriarte**: Resources (supporting); Writing – review and editing (supporting). **Amy Wolf**: Resources (supporting); Writing – review and editing (supporting). **Jess Zimmerman**: Resources (supporting); Writing – review and editing (supporting). **Daniel Zuleta**: Resources (supporting); Writing – review and editing (supporting). **Marco D. Visser**: Conceptualization (equal); Formal analysis (supporting); Investigation (equal); Methodology (equal); Supervision (equal); Writing – original draft (equal); Writing – review and editing (equal). **Lisa Hülsmann**: Conceptualization (equal); Data curation (equal); Formal analysis (supporting); Funding acquisition (equal); Investigation (equal); Methodology (equal); Software (equal); Supervision (equal); Validation (equal); Writing – original draft (equal); Writing – review and editing (equal).

#### Transparent peer review

The peer review history for this article is available at <https://publons.com/publon/10.1111/ecog.07187>.

#### Data availability statement

The forest data that support the findings of this study are available from the ForestGEO network. For some of the sites, the data is publicly available at <https://forestgeo.si.edu/explore-data>. **Restrictions apply**, however, to the availability of the data from other sites, which were used under license for the current study, and so are not publicly available. Data are, however, available from the authors upon reasonable request

and with the permission of the principal investigators of the ForestGEO sites. We provide an example of data preparation and analysis workflow from a forest plot with public available data together with the code for all results and analyses on the Zenodo repository: <https://zenodo.org/doi/10.5281/zenodo.10046359> (Leite et al. 2024).

#### Supporting information

The Supporting information associated with this article is available with the online version.

#### References

Adler, P. B. and Levine, J. M. 2007. Contrasting relationships between precipitation and species richness in space and time. – *Oikos* 116: 221–232.

Adler, P. B., Salguero-Gómez, R., Compagnoni, A., Hsu, J. S., Ray-Mukherjee, J., Mbeau-Ache, C. and Franco, M. 2014. Functional traits explain variation in plant life history strategies. – *Proc. Natl Acad. Sci. USA* 111: 740–745.

Anderegg, L. D. L., Griffith, D. M., Cavender-Bares, J., Riley, W. J., Berry, J. A., Dawson, T. E. and Still, C. J. 2022. Representing plant diversity in land models: an evolutionary approach to make ‘Functional Types’ more functional. – *Global Change Biol.* 28: 2541–2554.

Anderson-Teixeira, K. J., Herrmann, V., Morgan, R. B., Bond-Lamberty, B., Cook-Patton, S. C., Ferson, A. E., Muller-Landau, H. C. and Wang, M. M. H. 2021. Carbon cycling in mature and regrowth forests globally. – *Environ. Res. Lett.* 16: 053009.

Araujo, R. F., Grubinger, S., Celes, C. H. S., Negrón-Juárez, R. I., Garcia, M., Dandois, J. P. and Muller-Landau, H. C. 2021. Strong temporal variation in treefall and branchfall rates in a tropical forest is related to extreme rainfall: results from 5 years of monthly drone data for a 50 ha plot. – *Biogeosciences* 18: 6517–6531.

Arellano, G. 2019. Calculation of narrower confidence intervals for tree mortality rates when we know nothing but the location of the death/survival events. – *Ecol. Evol.* 9: 9644–9653.

Benincà, E., Ballantine, B., Ellner, S. P. and Huisman, J. 2015. Species fluctuations sustained by a cyclic succession at the edge of chaos. – *Proc. Natl Acad. Sci. USA* 112: 6389–6394.

Broekman, M. J. E., Muller-Landau, H. C., Visser, M. D., Jongejans, E., Wright, S. J. and de Kroon, H. 2019. Signs of stabilisation and stable coexistence. – *Ecol. Lett.* 22: 1957–1975.

Brokaw, N. V. L. 1987. Gap-phase regeneration of three pioneer tree species in a tropical forest. – *J. Ecol.* 75: 9.

Browne, W. J., Subramanian, S. V., Jones, K. and Goldstein, H. 2005. Variance partitioning in multilevel logistic models that exhibit overdispersion. – *J. R. Stat. Soc. A* 168: 599–613.

Bürkner, P.-C. 2017. brms: an R package for Bayesian multilevel models using stan. – *J. Stat. Softw.* 80: 1–28.

Burns, J. H. and Strauss, S. Y. 2012. Effects of competition on phylogenetic signal and phenotypic plasticity in plant functional traits. – *Ecology* 93: S126–S137.

Cáceres, M. et al. 2012. The variation of tree beta diversity across a global network of forest plots. – *Global Ecol. Biogeogr.* 21: 1191–1202.

Canadell, J. G. and Raupach, M. R. 2008. Managing forests for climate change mitigation. – *Science* 320: 1456–1457.

Chen, Y., Uriarte, M., Wright, S. J. and Yu, S. 2019. Effects of neighborhood trait composition on tree survival differ between drought and postdrought periods. – *Ecology* 100: e02766.

Chesson, P. 2000. Mechanisms of maintenance of species diversity. – *Annu. Rev. Ecol. Syst.* 31: 343–366.

Clark, J. S. 2010. Individuals and the variation needed for high species diversity in forest trees. – *Science* 327: 1129–1132.

Clark, A. T., Detto, M., Muller-Landau, H. C., Schnitzer, S. A., Wright, S. J., Condit, R. and Hubbell, S. P. 2018. Functional traits of tropical trees and lianas explain spatial structure across multiple scales. – *J. Ecol.* 106: 795–806.

Clark, J. S., Bell, D., Chu, C., Courbaud, B., Dietze, M., Hersh, M., HilleRisLambers, J., Ibáñez, I., LaDeau, S., McMahon, S., Metcalf, J., Mohan, J., Moran, E., Pangle, L., Pearson, S., Salk, C., Shen, Z., Valle, D. and Wyckoff, P. 2010. High-dimensional coexistence based on individual variation: a synthesis of evidence. – *Ecol. Monogr.* 80: 569–608.

Condit, R. et al. 2006. The importance of demographic niches to tree diversity. – *Science* 313: 98–101.

Davies, S. J. et al. 2021. ForestGEO: understanding forest diversity and dynamics through a global observatory network. – *Biol. Conserv.* 253: 108907.

Detto, M., Visser, M. D., Wright, S. J. and Pacala, S. W. 2019. Bias in the detection of negative density dependence in plant communities. – *Ecol. Lett.* 22: 1923–1939.

Díaz, S. et al. 2016. The global spectrum of plant form and function. – *Nature* 529: 167.

Dobbertin, M. 2005. Tree growth as indicator of tree vitality and of tree reaction to environmental stress: a review. – *Eur. J. For. Res.* 124: 319–333.

Douma, J. C. and Weedon, J. T. 2019. Analysing continuous proportions in ecology and evolution: a practical introduction to beta and Dirichlet regression. – *Methods Ecol. Evol.* 10: 1412–1430.

Fisher, R. A. et al. 2018. Vegetation demographics in Earth system models: a review of progress and priorities. – *Global Change Biol.* 24: 35–54.

Fortin, M., Bédard, S., DeBlois, J. and Meunier, S. 2008. Predicting individual tree mortality in northern hardwood stands under uneven-aged management in southern Québec, Canada. – *Ann. For. Sci.* 65: 205–205.

Fung, T. et al. 2020. Temporal population variability in local forest communities has mixed effects on tree species richness across a latitudinal gradient. – *Ecol. Lett.* 23: 160–171.

Gibert, A., Gray, E. F., Westoby, M., Wright, I. J. and Falster, D. S. 2016. On the link between functional traits and growth rate: meta-analysis shows effects change with plant size, as predicted. – *J. Ecol.* 104: 1488–1503.

Gora, E. M., Bitzer, P. M., Burchfield, J. C., Gutierrez, C. and Yanoviak, S. P. 2021. The contributions of lightning to biomass turnover, gap formation and plant mortality in a tropical forest. – *Ecology* 102: e03541.

Griffith, A. B., Salguero-Gómez, R., Merow, C. and McMahon, S. 2016. Demography beyond the population. – *J. Ecol.* 104: 271–280.

Hubbell, S. P. 2006. Neutral theory and the evolution of ecological equivalence. – *Ecology* 87: 1387–1398.

Hülsmann, L., Bugmann, H., Cailleret, M. and Brang, P. 2018. How to kill a tree: empirical mortality models for 18 species and their performance in a dynamic forest model. – *Ecol. Appl.* 28: 522–540.

Hülsmann, L., Chisholm, R. A. and Hartig, F. 2020. Is variation in conspecific negative density dependence driving tree diversity patterns at large scales? – *Trends Ecol. Evol.* 36: 151–163.

Hülsmann, L. et al. 2024. Latitudinal patterns in stabilizing density dependence of forest communities. – *Nature* 2024: <https://doi.org/10.1038/s41586-024-07118-4>.

Johnson, D. J. et al. 2018. Climate sensitive size-dependent survival in tropical trees. – *Nat. Ecol. Evol.* 2: 1436–1442.

Jurinitz, C. F., de Oliveira, A. A. and Bruna, E. M. 2013. Abiotic and biotic influences on early-stage survival in two shade-tolerant tree species in Brazil's Atlantic forest. – *Biotropica* 45: 728–736.

Keil, P. and Chase, J. M. 2019. Global patterns and drivers of tree diversity integrated across a continuum of spatial grains. – *Nat. Ecol. Evol.* 3: 390–399.

de Knecht, H. J., van Langevelde, F., Coughenour, M. B., Skidmore, A. K., de Boer, W. F., Heitkönig, I. M. A., Knox, N. M., Slotow, R., van der Waal, C. and Prins, H. H. T. 2010. Spatial autocorrelation and the scaling of species–environment relationships. – *Ecology* 91: 2455–2465.

Kohyama, T. 1993. Size-structured tree populations in gap-dynamic forest – the forest architecture hypothesis for the stable coexistence of species. – *J. Ecol.* 81: 131.

Kohyama, T. S., Kohyama, T. I. and Sheil, D. 2018. Definition and estimation of vital rates from repeated censuses: choices, comparisons and bias corrections focusing on trees. – *Methods Ecol. Evol.* 9: 809–821.

Kupers, S. J., Wirth, C., Engelbrecht, B. M. J., Hernández, A., Condit, R., Wright, S. J. and Rüger, N. 2019. Performance of tropical forest seedlings under shade and drought: an inter-specific trade-off in demographic responses. – *Sci. Rep.* 9: 18784.

Lasky, J. R., Uriarte, M., Boukili, V. K. and Chazdon, R. L. 2014. Trait-mediated assembly processes predict successional changes in community diversity of tropical forests. – *Proc. Natl Acad. Sci. USA* 111: 5616–5621.

Leite, M. de Souza et al. 2024. Data and code from: Major axes of variation in tree demography across global forests. – Zenodo Repository, <https://zenodo.org/doi/10.5281/zenodo.10046359>.

Lévesque, M., Walther, L. and Weber, P. 2016. Soil nutrients influence growth response of temperate tree species to drought. – *J. Ecol.* 104: 377–387.

Li, Y., Mayfield, M. M., Wang, B., Xiao, J., Kral, K., Janik, D., Holik, J. and Chu, C. 2020. Beyond direct neighbourhood effects: higher-order interactions improve modelling and predicting tree survival and growth. – *Natl. Sci. Rev.* 8: nwaa244.

Lu, R., Qiao, Y., Wang, J., Zhu, C., Cui, E., Xu, X., He, Y., Zhao, Z., Du, Y., Yan, L., Shen, G., Yang, Q., Wang, X. and Xia, J. 2021. The U-shaped pattern of size-dependent mortality and its correlated factors in a subtropical monsoon evergreen forest. – *J. Ecol.* 109: 2421–2433.

Maier, M. J. 2021. DirichletReg: dirichlet regression. – R package ver. 0.7-1. <https://github.com/maiermarco/DirichletReg>.

Martínez-Ramos, M., Gallego-Mahecha, M. del M., Valverde, T., Vega, E. and Bongers, F. 2021. Demographic differentiation among pioneer tree species during secondary succession of a Neotropical rainforest. – *J. Ecol.* 109: 3572–3586.

Martini, F., Chang-Yang, C.-H. and Sun, I.-F. 2022. Variation in biotic interactions mediates the effects of masting and rainfall fluctuations on seedling demography in a subtropical rainforest. – *J. Ecol.* 110: 762–771.

McDowell, N. G. et al. 2020. Pervasive shifts in forest dynamics in a changing world. – *Science* 368: eaaz9463.

McMahon, S. M. and Diez, J. M. 2007. Scales of association: hierarchical linear models and the measurement of ecological systems. – *Ecol. Lett.* 10: 437–452.

Messier, J., McGill, B. J. and Lechowicz, M. J. 2010. How do traits vary across ecological scales? A case for trait-based ecology. – *Ecol. Lett.* 13: 838–848.

Nakagawa, S., Johnson, P. C. D. and Schielzeth, H. 2017. The coefficient of determination R<sup>2</sup> and intra-class correlation coefficient from generalized linear mixed-effects models revisited and expanded. – *J. R. Soc. Interface* 14: 20170213.

Needham, J. F. et al. 2022. Demographic composition, not demographic diversity, predicts biomass and turnover across temperate and tropical forests. – *Global Change Biol.* 28: 2895–2909.

Negrón, J. F., McMillin, J. D., Anhold, J. A. and Coulson, D. 2009. Bark beetle-caused mortality in a drought-affected ponderosa pine landscape in Arizona, USA. – *For. Ecol. Manage.* 257: 1353–1362.

Oberpriller, J., de Souza Leite, M. and Pichler, M. 2022. Fixed or random? On the reliability of mixed-effects models for a small number of levels in grouping variables. – *Ecol. Evol.* 12: e9062.

Pan, Y., Birdsey, R. A., Fang, J., Houghton, R., Kauppi, P. E., Kurz, W. A., Phillips, O. L., Shvidenko, A., Lewis, S. L., Canadell, J. G., Ciais, P., Jackson, R. B., Pacala, S. W., McGuire, A. D., Piao, S., Rautiainen, A., Sitch, S. and Hayes, D. 2011. A large and persistent carbon sink in the world's forests. – *Science* 333: 988–993.

Pillay, R., Venter, M., Aragon-Osejo, J., González-del-Pliego, P., Hansen, A. J., Watson, J. E. and Venter, O. 2022. Tropical forests are home to over half of the world's vertebrate species. – *Front. Ecol. Environ.* 20: 10–15.

Poorter, L., Wright, S. J., Paz, H., Ackerly, D. D., Condit, R., Ibarra-Manríquez, G., Harms, K. E., Licona, J. C., Martínez-Ramos, M., Mazer, S. J., Muller-Landau, H. C., Peña-Claros, M., Webb, C. O. and Wright, I. J. 2008. Are functional traits good predictors of demographic rates? Evidence from five neotropical forests. – *Ecology* 89: 1908–1920.

Rubio, V. E. and Swenson, N. G. 2022. Functional groups, determinism and the dynamics of a tropical forest. – *J. Ecol.* 110: 185–196.

Rüger, N., Comita, L. S., Condit, R., Purves, D., Rosenbaum, B., Visser, M. D., Wright, S. J. and Wirth, C. 2018. Beyond the fast–slow continuum: demographic dimensions structuring a tropical tree community. – *Ecol. Lett.* 21: 1075–1084.

Rüger, N., Condit, R., Dent, D. H., DeWalt, S. J., Hubbell, S. P., Lichstein, J. W., Lopez, O. R., Wirth, C. and Farrior, C. E. 2020. Demographic trade-offs predict tropical forest dynamics. – *Science* 368: 165–168.

Russo, S. E., Davies, S. J., King, D. A. and Tan, S. 2005. Soil-related performance variation and distributions of tree species in a Bornean rain forest. – *J. Ecol.* 93: 879–889.

Russo, S. E., Brown, P., Tan, S. and Davies, S. J. 2008. Interspecific demographic trade-offs and soil-related habitat associations of tree species along resource gradients. – *J. Ecol.* 96: 192–203.

Russo, S. E. et al. 2021. The interspecific growth–mortality trade-off is not a general framework for tropical forest community structure. – *Nat. Ecol. Evol.* 5: 174–183.

Šamonil, P., Doleželová, P., Vašíčková, I., Adam, D., Valtera, M., Král, K., Janík, D. and Šebková, B. 2013. Individual-based approach to the detection of disturbance history through spatial scales in a natural beech-dominated forest. – *J. Veg. Sci.* 24: 1167–1184.

Schwartz, N. B., Feng, X., Muscarella, R., Swenson, N. G., Umaña, M. N., Zimmerman, J. K. and Uriarte, M. 2020. Topography and traits modulate tree performance and drought response in a tropical forest. – *Front. For. Global Change* 3: 136.

Searle, S. R., Casella, G. and McCulloch, C. E. 2006. Variance components. – Wiley.

Su, S.-H., Guan, B. T., Chang-Yang, C.-H., Sun, I.-F., Wang, H.-H. and Hsieh, C.-F. 2020. Multi-stemming and size enhance survival of dominant tree species in a frequently typhoon-disrupted forest. – *J. Veg. Sci.* 31: 429–439.

Swenson, N. G., Worthy, S. J., Eubanks, D., Iida, Y., Monks, L., Petprakob, K., Rubio, V. E., Staiger, K. and Zambrano, J. 2020. A reframing of trait–demographic rate analyses for ecology and evolutionary biology. – *Int. J. Plant Sci.* 181: 33–43.

Uriarte, M., Thompson, J. and Zimmerman, J. K. 2019. Hurricane María tripled stem breaks and doubled tree mortality relative to other major storms. – *Nat. Commun.* 10: 1362.

Visser, M. D., Bruijning, M., Wright, S. J., Muller-Landau, H. C., Jongejans, E., Comita, L. S. and de Kroon, H. 2016. Functional traits as predictors of vital rates across the life cycle of tropical trees. – *Funct. Ecol.* 30: 168–180.

Weng, E. S., Malyshev, S., Lichstein, J. W., Farrior, C. E., Dybinski, R., Zhang, T., Sheviakova, E. and Pacala, S. W. 2015. Scaling from individual trees to forests in an Earth System Modeling Framework using a mathematically tractable model of height-structured competition. – *Biogeosciences* 12: 2655–2694.

Wiegand, T., Uriarte, M., Kraft, N. J. B., Shen, G., Wang, X. and He, F. 2017. Spatially explicit metrics of species diversity, functional diversity, and phylogenetic diversity: insights into plant community assembly processes. – *Annu. Rev. Ecol. Evol. Syst.* 48: 329–351.

Wiens, J. J. and Donoghue, M. J. 2004. Historical biogeography, ecology and species richness. – *Trends Ecol. Evol.* 19: 639–644.

Wright, S. J., Muller-Landau, H. C., Condit, R. and Hubbell, S. P. 2003. Gap-dependent recruitment, realized vital rates, and size distributions of tropical trees. – *Ecology* 84: 3174–3185.

Zuleta, D., Duque, A., Cardenas, D., Muller-Landau, H. C. and Davies, S. J. 2017. Drought-induced mortality patterns and rapid biomass recovery in a terra firme forest in the Colombian Amazon. – *Ecology* 98: 2538–2546.

Zuleta, D., Russo, S. E., Barona, A., Barreto-Silva, J. S., Cardenas, D., Castaño, N., Davies, S. J., Detto, M., Sua, S., Turner, B. L. and Duque, A. 2020. Importance of topography for tree species habitat distributions in a terra firme forest in the Colombian Amazon. – *Plant Soil* 450: 133–149.