

1 **TITLE:** Non-breeding conditions induce carry-over effects on survival of migratory birds
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4 **AUTHORS:** Nathan W. Cooper^{1,7,*}, Scott W. Yanco^{2,3,4}, Clark S. Rushing⁵, T. Scott Sillett¹, and
5 Peter P. Marra⁶

6
7 ¹Migratory Bird Center, Smithsonian's National Zoo and Conservation Biology Institute,
8 Washington, D.C., USA 20008

9
10 ²School for Environment and Sustainability, University of Michigan Ann Arbor, MI,
11 USA 48109

12
13 ³Center for Biodiversity and Global Change, Yale University, New Haven, CT, USA
14 06511

15
16 ⁴Department of Integrative Biology, University of Colorado Denver, Denver, Colorado,
17 USA 80204

18
19 ⁵Warnell School of Forestry and Natural Resources, University of Georgia, Athens, GA,
20 USA 30602

21
22 ⁶The Earth Commons Institute; Department of Biology; McCourt School of Public
23 Policy; Georgetown University, Washington, D.C., USA 20057

24
25 ⁷Lead contact

26
27 *Correspondence: coopern@si.edu

28
29 **SUMMARY**

30 Identifying the processes that limit populations is a foundational objective of ecology and an
31 urgent need for conservation. For migratory animals, researchers must study individuals
32 throughout their annual cycles to determine how environmental conditions limit demographic
33 rates within each period of the annual cycle and also between periods through carry-over effects
34 and seasonal interactions^{1–6}. Our poor understanding of the rates and causes of avian migration
35 mortality⁷ hinders identification of limiting factors and reversal of widespread avian population
36 declines^{8,9}. Here, we implement new methods to estimate apparent survival (hereafter survival)
37 during migration directly from automated telemetry data¹⁰ in Kirtland's Warblers (*Setophaga*

38 *kirtlandii*) and indirectly from mark-recapture data in Black-throated Blue Warblers (*S.*
39 *caerulescens*). Previous experimental and observational studies of our focal species and other
40 migratory songbirds have shown strong effects of Caribbean precipitation and habitat quality on
41 food availability^{11–14}, body condition^{12–19}, migration timing^{11,12,15,16,20–23}, natal dispersal^{24,25},
42 range dynamics²⁶, reproductive success^{20,22,27}, and annual survival^{18–20,23,28–31}. Building on this
43 research, we test the hypotheses that environmental conditions during the non-breeding period
44 affect subsequent survival during spring migration and breeding. We found that reduced
45 precipitation and environmental productivity in the non-breeding period strongly influenced
46 survival in both species, primarily by reducing survival during spring migration. Our results
47 indicate that climate-driven environmental conditions can carry over to affect survival in
48 subsequent periods and thus likely play an important role in year-round population dynamics.
49 These lethal carry-over effects may be widespread and are likely magnified by intensifying
50 climate change.

51

52 **RESULTS**

53 **Weekly survival was lower during migration than in stationary periods**

54 We found that weekly survival probabilities for both species were lower during migration than in
55 the stationary periods. In Kirtland's Warblers (*Setophaga kirtlandii*), we leveraged the species'
56 restricted breeding range, small population size^{32–34}, and a continent-wide automated telemetry
57 network (i.e., Motus Wildlife Tracking Network¹⁰) to directly estimate apparent survival
58 (hereafter survival) during the stationary non-breeding (hereafter non-breeding), spring
59 migration, and breeding periods using a novel extension of a multi-state robust-design Cormack–
60 Jolly–Seber (CJS) model (Figure 1). In Black-throated Blue Warblers (*S. caerulescens*), we used

61 14 years of mark-recapture data from geographically- and demographically-linked populations^{29–}
62 ^{31,35,36} to estimate non-breeding, spring migration, breeding, and fall migration survival using a
63 full-annual-cycle, integrated survival model³⁷. Weekly survival probability for Kirtland’s
64 Warblers during spring migration was 0.899 (95% Credible Interval = 0.846-0.944),
65 considerably lower than in the non-breeding or (0.968 [0.948-0.985]) or breeding (0.963 [0.944-
66 0.979]) periods. Black-throated Blue Warblers also had lower weekly survival during spring
67 (0.975 [0.942-0.996]) and fall migration (0.975 [0.940-0.997]), compared to the breeding (0.994
68 [0.985-0.999]) or non-breeding (0.997 [0.991-1.00]) periods.

69

70 **Reduced environmental productivity and precipitation carry over to affect survival during**
71 **migration and breeding**

72 We found that environmental conditions in the Caribbean affected non-breeding survival of
73 Kirtland’s Warblers and carried over to influence survival during migration and breeding in both
74 species (Figure 2, Figure S1). In Kirtland’s Warblers, non-breeding home ranges with higher
75 Enhanced Vegetation Index (EVI) values, a measure of environmental productivity³⁸, were
76 associated with higher survival during the end of the non-breeding period (0.651 [0.370-0.855];
77 Bayesian probability of direction [hereafter Bayesian *pd*] ≥ 0.99), on spring migration (0.620
78 [0.335-0.835]; Bayesian *pd* ≥ 0.99), and in the following breeding period (0.620 [0.341-0.832];
79 Bayesian *pd* ≥ 0.99). In Black-throated Blue Warblers, precipitation during the non-breeding
80 period, as indexed by the Southern Oscillation Index (SOI)^{39–41}, was not significantly associated
81 with non-breeding survival for either sex (males: -0.68 [-2.88-1.24], Bayesian *pd* = 0.76;
82 females: -0.34 [-2.03-1.19], Bayesian *pd* = 0.67). Survival probability on spring migration was
83 positively associated with SOI in both sexes (males: 0.92 [0.02-2.04], Bayesian *p* = 0.98;

84 females: 0.72 [-0.27-1.96], Bayesian $pd = 0.94$). SOI was significantly and negatively associated
85 with male survival in the subsequent breeding period (-1.47 [-3.34-0.34], Bayesian $pd = 0.96$),
86 but only weakly associated with breeding survival of females (-0.88 [-3.17-1.10], Bayesian $pd =$
87 0.80). Thus, following dry, El Niño winters in the Caribbean, Black-throated Blue Warbler
88 survival was lower during spring migration and at least for males, higher over the following
89 breeding season.

90

91 **DISCUSSION**

92 Using new analytical approaches, we demonstrate that non-breeding period environmental
93 conditions have lethal effects that carry over into migration and the subsequent breeding period
94 in two species of migratory songbirds. These effects were evident at local and regional spatial
95 scales and were confirmed in two species with different geographic ranges, using both direct and
96 indirect methods for estimating seasonal survival. By extending a multi-state robust-design CJS
97 model to incorporate the timing of one-way transitions during the annual cycle and account for
98 the spatially heterogeneous layout of the automated-telemetry stations (see STAR Methods), we
99 were able to directly estimate survival of Kirtland's Warblers during an entire migration,
100 whereas previous studies of small birds have provided indirect estimates^{25,28,29,42-44} or direct
101 estimates from one leg of the migratory journey⁴⁵. We found that poor environmental
102 productivity within individual non-breeding home ranges in The Bahamas was associated with
103 lower survival during the non-breeding, spring migration, and breeding periods. In Black-
104 throated Blue Warblers, we extended previously-developed indirect estimation methods to assess
105 the influence of environmental covariates on seasonal survival, the first such application of these
106 methods in non-simulated data³⁷. Spring migration survival of both sexes was lower following

107 dry, El Niño winters in the Caribbean, and males showed higher breeding survival after dry
108 winters. Collectively, these results support previous findings showing that non-breeding climate
109 influences annual survival^{28,30,46–49}, and extend these findings by pinpointing where and when in
110 the annual cycle the effects on mortality manifest. Our results have important implications for
111 forecasting how climate change will influence the population dynamics of migratory birds.
112 Moreover, we argue that they also highlight the need to estimate season-specific vital rates and to
113 determine the factors that affect those vital rates in other imperiled taxa.

114

115 **Seasonal survival on a rapidly changing planet**

116 Our results suggest that the negative effects of non-breeding drought on survival during
117 migration and the breeding period may be an important, but previously undocumented threat to
118 the survival of birds during migration. The Caribbean has been drying since at least the 1950s⁵⁰
119 and climate change is predicted to increase both the frequency and severity of drought across this
120 region^{51–53}. A drying Caribbean has the potential to degrade non-breeding habitat quality and
121 reduce food availability, thereby reducing the ability of individual birds to adequately prepare for
122 migration and to survive through the breeding period.

123 Climate-induced impacts on population vital rates represent a looming and under-
124 quantified threat for species' sustained existence⁵⁴ because the changes to seasonal survival that
125 we documented have the potential to negatively affect population trajectories (Figure 3). Most
126 migratory bird populations are already exhibiting precipitous declines, the drivers of which
127 remain poorly understood⁸. Our findings that environmental conditions in the non-breeding
128 period can lead to changes in vital rates across multiple seasons and that those vital rate changes
129 could endanger population persistence (Figure 3) highlight the urgency for additional research

130 focused on identifying the specific mechanisms that limit populations throughout the annual
131 cycle.

132

133 **What are the mechanisms of lethal carry-over effects?**

134 Our results confirm predictions that environmental conditions during the non-breeding period
135 can have a lethal carry-over effect during spring migration, but do not identify the precise
136 mechanism(s) responsible. Previous work on songbirds has found that dry conditions and poor-
137 quality habitat on the non-breeding grounds result in lower food availability^{11–14} and reduced
138 body condition prior to spring migration^{12–19}. Limited food resources and poor body condition
139 could increase non-breeding period mortality and are known to delay departure on spring
140 migration^{11,12,15,16,55}. Food and body condition in the non-breeding period also presumably limit
141 the ability of birds to meet the physiological demands of spring migration and ultimately, to
142 survive migration. Departing on migration late relative to conspecifics could further enhance risk
143 of mortality if individuals attempt to compensate for late departure by increasing their migration
144 rate²³. Future research will be needed to confirm the precise mechanism(s), but our findings
145 suggest that observed effects on survival during migration and breeding are the result of a chain
146 of sublethal effects that begins on the non-breeding grounds.

147 The carry-over effects of non-breeding environmental conditions on spring migration
148 survival continued to manifest during the breeding period in both species, but in opposite
149 directions. Poor non-breeding conditions were associated with lower breeding survival in
150 Kirtland’s Warblers but with higher breeding survival in male Black-throated Blue Warblers. We
151 attribute this discrepancy to differences in the scale at which we estimated environmental
152 conditions. At the individual level in Kirtland’s Warblers, high environmental productivity on

153 individual non-breeding home ranges likely improved the chances of surviving spring migration
154 (see above), and the following breeding period due to increased body condition, earlier arrival
155 timing, and priority access to high-quality breeding territories^{56,57}. At the population level in
156 Black-throated Blue Warblers, reduced rainfall across the non-breeding range may result in
157 fewer individuals reaching the breeding grounds (see above). This could lead to lower population
158 density during the breeding period, which then increases survival during the breeding season.
159 Previous investigations have not found evidence of density-dependent survival in Black-throated
160 Blue Warblers at Hubbard Brook³¹, but the demographic processes suggested by our integrated
161 models may be operating at a spatial scale larger than the study site. Nevertheless, our results
162 support the hypothesis that non-breeding environmental conditions can have important carry-
163 over effects on survival during both migration and the subsequent breeding season.

164 A possible alternative explanation of our findings for Black-throated Blue Warblers is
165 that reduced survival during migration is not a carry-over effect, but instead a direct effect of
166 ENSO-driven weather during migration and breeding. Although the El Niño-Southern
167 Oscillation (ENSO) has strong and predictable effects on dry and wet season precipitation in the
168 Caribbean, including in Jamaica³⁹⁻⁴¹, it also has teleconnections across the hemisphere. Black-
169 throated Blue Warblers migrate through the southeastern, eastern, and northeastern United States
170 in April and May⁵⁸. Along this migration route, effects of ENSO on temperature and
171 precipitation are only apparent in the southeast between October and March, and are in the
172 opposite direction from the effects observed in Jamaica (i.e., when conditions are dry in Jamaica,
173 they are cool and wet in the southeast)^{59,60}. This regionally variable effect, or lack thereof, of
174 ENSO on weather along the migratory route suggests that a direct effect of ENSO during
175 migration is unlikely. Nonetheless, we cannot entirely rule out that ENSO-related weather

176 patterns along the migratory route, or other unmeasured factors, directly contribute or interact
177 with non-breeding effects to influence survival during migration. However, our conclusion that
178 ENSO-driven changes in Jamaican precipitation carries over to affect migration survival is also
179 bolstered by multiple observational and experimental studies that have documented the effects of
180 non-breeding precipitation and habitat quality on food availability, body condition, migration
181 timing, natal dispersal, reproductive success, and range dynamics (see references above).

182

183 **Implications**

184 Estimating seasonal survival and documenting how climate-driven environmental conditions
185 affect survival throughout the annual cycle are key to advancing our understanding of migratory
186 bird population dynamics and the conservation of these imperiled species. As we demonstrate
187 here, the Motus Wildlife Tracking System¹⁰ provides new opportunities for directly estimating
188 migration survival. However, continued expansion of the Motus network and analytical advances
189 will be needed to realize this potential for the vast majority of other species, which have larger
190 ranges and more complex migration patterns than Kirtland's Warblers^{32,34}. In the meantime, the
191 indirect approach we developed and implemented for estimating the effects of environmental
192 conditions on seasonal survival in Black-throated Blue Warblers can be replicated for any
193 species with the necessary data from demographically-linked populations. Regardless, more
194 research and innovation in tracking technology will be needed to understand the immediate
195 causes of mortality (e.g., starvation, collisions, predation).

196 Documenting survival throughout the annual cycle is a critical next step, but seasonal
197 survival estimates must be combined with other demographic data (e.g., reproduction, dispersal,
198 population size) to develop full-annual-cycle population models that can pinpoint where and

199 when in the annual cycle populations are limited and/or regulated^{25,43,61}. Such models will allow
200 researchers to explore how processes throughout the annual cycle combine and interact in
201 potentially nuanced ways to determine the population dynamics of migratory species. Only by
202 further adoption of these approaches will we be able to identify the multiple and likely interacting
203 causes of widespread migratory species declines and begin to slow and reverse them through
204 conservation and management. Similar efforts will be needed across migratory taxa, many of
205 which are also experiencing alarming declines^{5,62}.

206

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223

224 **AUTHOR CONTRIBUTIONS**

225 All authors designed the study, NWC, PPM, and TSS performed the research, SWY and CSR
226 provided new analytical methods and analyzed the data, NWC, SWY, and CSR wrote the
227 manuscript, and all authors substantially edited the manuscript.

228

229 **DECLARATION OF INTERESTS**

230 The authors declare no competing interests.

231

232 **Figure 1. Seasonal survival estimation and predicted effects of environmental conditions on**
233 **spring migration survival**

234 Conceptual figure illustrating: (A) non-breeding and breeding locations for the two species and
235 differences in survival estimation methodology, (B) the periods where survival was either
236 directly (solid) or indirectly (dashed) estimated throughout the annual cycle for each species, and
237 (C) the predicted effects of non-breeding environmental conditions on spring migration survival.

238 Artwork by Sally L. Bornbusch.

239

240 **Figure 2. Reduced environmental productivity and precipitation carry over to affect**
241 **survival during migration and breeding**

242 Conditional effect (lines) and 95% credible intervals (shaded areas) of environmental conditions
243 on apparent weekly survival. Conditional effect of March enhanced vegetation index (EVI) on

244 weekly survival of Kirtland's Warblers (left) during late non-breeding (blue), migration (green),
245 and breeding (red) periods. Tick marks along x-axis depict the distribution of observations across
246 the domain of EVI. Conditional effects of non-breeding Southern Oscillation Index (SOI) on
247 female (top right) and male (bottom right) Black-throated Blue Warblers during non-breeding
248 (blue), spring migration (green), and breeding (red) periods. As with Kirtland's Warbler,
249 tick marks along the x-axis depict the distribution of observations across the domain of SOI. See
250 also Figure S1.

251

252 **Figure 3. Potential effects of non-breeding grounds drying on population growth**

253 Indicative population growth rate (i.e., lambda) as a function of non-breeding Enhanced
254 Vegetation Index (EVI). To characterize the potential for variance in non-breeding
255 environmental conditions to drive population growth trends via the survival dynamics described
256 in this paper, we developed an ad hoc population projection model based on our conditional
257 survival estimates and estimates from the literature. Briefly, this projection matrix estimates
258 population growth as a single-age class, four-season "cyclic" model. We derived estimated per
259 capita fecundity from Bocetti et al.⁶³ and a static estimate of fall migration survival from
260 Rockwell et al.²⁸. The conditional estimates for the other three seasons came from our fitted
261 survival models and considered EVI conditions ranging from the 10-90% quantiles of EVI
262 observed in our dataset. At or above median EVI, population growth is relatively stable and near
263 1 but drops quickly for below-normal EVI conditions. This suggests that drying conditions may
264 impact population abundance via reduced survival. These findings should be treated as indicative
265 of a possible effect and future work to develop a more complete population abundance model is
266 underway.

267

268 **STAR METHODS**

269 **RESOURCE AVAILABILITY**

270 **Lead Contact**

271 Further information and requests for resources should be directed to and will be fulfilled by the
272 lead contact, Nathan W. Cooper (coopern@si.edu).

273

274 **Materials Availability**

275 This study did not generate unique reagents.

276

277 **Data and Code Availability**

278 All data has been deposited at Figshare and is publicly available as of the date of publication.
279 The DOI is listed in the key resources table. All original code has been deposited at GitHub and
280 is publicly available as of the date of publication. The DOI is listed in the key resources table.
281 Any additional information required to reanalyze the data reported in this paper is available from
282 the lead contact upon request.

283

284 **EXPERIMENTAL MODEL AND STUDY PARTICIPANT DETAILS**

285 We used mist nets and conspecific song playback to capture 136 adult male Kirtland's Warblers
286 on their non-breeding grounds on Cat Island, The Bahamas (24.33° N, 75.45° W). We used mist
287 nest to capture 314 (89 ♀, 205 ♂) adult Black-throated Blue Warblers at Hubbard Brook
288 Experimental Forest in New Hampshire, USA (43.94°N, -71.70°W) and 116 (45 ♀, 71 ♂) adults
289 at Copse Mountain, Jamaica, West Indies (18.29°N, -77.96°W). All handling of birds was

290 approved by Animal Care and Use Committee Permits from the Smithsonian Institution and
291 Dartmouth.

292

293 **METHOD DETAILS**

294 **Field methods**

295 From 2017-2019, we captured and color-banded male Kirtland's Warblers (2017 $n = 58$, 2018 n
296 $= 40$, 2019 $n = 38$) across Cat Island, The Bahamas (24.67°N , -75.74°W). After capture, we
297 attached a 0.35 g coded radio tag (NTQBW-2, Lotek Wireless) using a leg-loop harness³³. In
298 2017, we used handheld telemetry to estimate departure date on spring migration by searching at
299 least once every three days from April 1 through May 5¹⁵. In 2018 and 2019, we used an array of
300 four automated-telemetry stations (hereafter stations) to determine departure date on spring
301 migration^{64,65}. We acquired detections during migration from the Motus Wildlife Tracking
302 System¹⁰, using the package *motus*⁶⁶ in Program R⁶⁷. To resight individuals on the breeding
303 grounds in the Lower Peninsula of Michigan (44.62°N , -84.27°W), where 95-97% of the
304 Kirtland's Warbler population breeds³², we used handheld telemetry and detections from 12
305 stations located in breeding habitat³³. Once located, we attempted to re-sight each individual and
306 determine their fate once per week through June.

307 Black-throated Blue Warblers were studied at Hubbard Brook Experimental Forest in
308 New Hampshire, USA (43.94°N , -71.70°W) from 1986 – 2000 and at Copse Mountain, Jamaica,
309 West Indies (18.29°N , -77.96°W) from 1986 – 1998. Birds were captured in mist nets and color-
310 banded. The Hubbard Brook study plot (64 ha) was searched for color-banded individuals every
311 1-7 days during breeding (mid-May to mid-August). Sampling on the Jamaica study plot (7 ha)
312 occurred during 3–5-day periods in mid-October–early November, and during 2–5-day periods in

313 mid-March. Following Sillett and Holmes²⁹, we built encounter histories from two periods per
314 site: mid-May to early June and late July to mid-August at Hubbard Brook, and October and
315 March in Jamaica. Although no banded individuals were encountered during both the breeding
316 and non-breeding period, the Hubbard Brook and Jamaica populations have been shown to be
317 geographically linked via stable isotopes³⁵ and light-level geolocators³⁶, and demographically
318 linked via correlation of reproductive rates in Hubbard Brook and juvenile recruitment rates in
319 Jamaica^{29–31}.

320

321 QUANTIFICATION AND STATISTICAL ANALYSIS

322 Kirtland's Warbler Seasonal Survival

323 To characterize non-breeding environmental conditions experienced by Kirtland's Warblers, we
324 estimated an individual-specific Enhanced Vegetation Index (EVI) value as a proxy for relative
325 environmental productivity³⁸. We used EVI values measured during March because of previous
326 research indicating the relevance of March rainfall for annual differences in arrival timing and
327 reproductive success on the breeding grounds²⁰ and apparent annual survival²⁸. We gathered EVI
328 values from within a 250-meter radius of the centroid of individual non-breeding locations
329 (calculated using the *sf* package⁶⁸) in Program R⁶⁷ and took the mean of these values across both
330 the spatial and temporal buffer to generate an individual-level covariate for survival modeling
331 (see below). All environmental annotations were done using the *rstoot* package⁶⁹ in Program R⁶⁷,
332 which draws from the NASA MODIS daily EVI dataset
333 (<https://modis.gsfc.nasa.gov/data/dataproducts/mod09.php>).

334 To estimate seasonal survival for Kirtland's Warblers, we constructed a variation of a
335 three-state Cormack-Jolly-Seber (CJS) survival model^{70–72}. We defined states based on

336 individual migratory progression such that detections on Cat Island (known non-breeding
337 grounds) were assigned to the non-breeding state, detections on the breeding grounds were
338 assigned to the breeding state, and all other observations were considered migratory. Because our
339 period of study extended from the end of the non-breeding period through breeding, we restricted
340 transitions between states to be “one way” – individuals could only transition from non-breeding
341 to migration and migration to breeding (and from any living state to the “dead” state). Kirtland’s
342 Warblers are obligate migrants; therefore, we modeled the transitions between states using the
343 Weibull cumulative distribution function^{73,74}. This function increases the probability of transition
344 from one state to the next as time proceeds, asymptotically approaching one. Thus, individuals
345 were not permitted to remain in a lower order state indefinitely but instead must eventually either
346 proceed with migration or die. In this way, the model also estimated population-level migration
347 phenology.

348 The clustered distribution of Motus stations resulted in detection patterns that
349 superficially resembled “trap happy” and “trap shy” individuals; given that a bird was detected
350 on one day, the probability of detection was substantially higher than for individuals that had not
351 been detected on the previous day because detected birds were more likely to be in a high-
352 density portion of the array (and vice versa). To account for this potential source of bias, we used
353 the “robust design”⁷⁵, wherein primary occasions were defined as weeks and secondary
354 occasions were defined as days.

355 We fit a model that estimated seasonal survival as a function of non-breeding EVI
356 conditions for all three seasons. We also fit a null model that considered survival as constant
357 within seasons. We used Deviance Information Criterion (DIC) to evaluate relative support
358 between the fitted models⁷⁶. We used uninformative priors for most parameters in the model, but

359 parameters estimating constant seasonal survival probabilities were approximately based on
360 Rockwell et al. ²⁸ (note that exact moment-matching from those estimates resulted in impossible
361 parameter combinations so we used more permissive prior distributions). We also constrained
362 prior distributions associated with transition processes to exclude biologically implausible
363 transitions (e.g., transitions from non-breeding to migration deep into the breeding season).

364 We fit models in a Bayesian frameworks using JAGS⁷⁷, which was called from within R
365 using the *jagsUI* package⁷⁸. We drew four chains of 500,000 MCMC samples with a burn-in
366 period of 50,000, thinned to every 250th sample to reduce serial autocorrelation within the chains.
367 We considered models converged at R-hat <1.05.

368

369 **Black-throated Blue Warbler Seasonal Survival**

370 We used an integrated survival model³⁷ to estimate seasonal survival probabilities of male and
371 female Black-throated Blue Warblers. For each data set (breeding and non-breeding), the
372 integrated model fits a Cormack-Jolly-Seber (CJS) model to estimate apparent survival within
373 and between stationary periods while accounting for imperfect detection⁷⁹. In New Hampshire,
374 the survival of individuals between re-sighting occasions was determined by either the 3-month
375 breeding period survival probability (May – August; $\phi_{s,t}^{\text{Sum}}$) or the 9-month between-breeding
376 period survival (August – May; $\phi_{s,t}^{BB}$), where s denotes sex-specific survival probabilities in year
377 t . In Jamaica, the survival of individuals between re-sighting occasions was determined by the 6-
378 month non-breeding survival probability (October – April; $\phi_{s,t}^{\text{Win}}$) or the 6-month between-non-
379 breeding survival (April – Oct; $\phi_{s,t}^{BW}$). We accounted for imperfect detection by modeling the
380 observed status of individuals as a function of the sex- and occasion-specific detection
381 probabilities ($p_{s,t}^j$), where j denotes occasion (May, August, October, or March). In their original

382 analysis of these data, Sillett and Holmes²⁹ found no support for sex-specific detection
383 probabilities in October and March, so we did not model sex variation on that parameter.
384 Furthermore, because observer effort was consistent during the May and October sampling
385 periods, we followed Sillett and Holmes²⁹ and did not include annual variation in those detection
386 probabilities.

387 The two survival models can be integrated by parameterizing the between-period survival
388 probabilities in terms of the underlying seasonal survival rates:

389
$$\phi_{s,t}^{BW} = \phi_{s,t}^{Spr} \phi_t^{Sum} \phi_{s,t}^{Aut}$$

390
$$\phi_{s,t}^{BB} = \phi_{s,t}^{Aut} \phi_{t+1}^{Win} \phi_{s,t+1}^{Spr}$$

391 which allows estimation of the latent 2-month spring and fall survival probabilities and the
392 inclusion of covariates on seasonal survival probabilities³⁷. We chose covariates based on
393 previous research linking environmental conditions to annual survival in Black-throated Blue
394 Warblers. In Jamaica, annual survival of Black-throated Blue Warblers from 1986-1999 was
395 strongly associated with the phase of the El Niño Southern Oscillation (ENSO;^{29,30}). During El
396 Niño years in Jamaica, reduced rainfall^{39,40} likely led to decreased availability of arthropod prey
397 in the winter dry season and, hence, to lower survival. Wetter La Niña years likely resulted in
398 increased food availability and higher survival. We hypothesized that ENSO mediates warbler
399 survival via direct effects on non-breeding survival or by influencing body condition at
400 departure, which should influence subsequent survival during spring migration or over the
401 following breeding season. We included mean monthly values of the standardized Southern
402 Oscillation Index (SOI) as a predictor of non-breeding, spring, and breeding survival
403 probabilities for both sexes and predicted that survival would be positively associated with SOI
404 (positive SOI values indicate La Niña conditions; negative values signify El Niño conditions).

405 Posterior distributions for parameters were estimated using Markov chain Monte Carlo
406 implemented using the *nimble* package⁸⁰ in R⁶⁷. Encounter histories were summarized using the
407 multi-state array (m-array) format for computational efficiency. Model fitting revealed that the
408 default random walk samplers used by nimble resulted in high correlation between the posterior
409 samples for the latent fall and spring mean survival probabilities and thus poor mixing. Nimble's
410 automated factor slice sampler resulted in lower correlation and improved mixing. We used the
411 default samplers for all other parameters. We ran three chains for 75,000 iterations each after
412 discarding the first 25,000 iterations as burn-in and saving every 5th sample to reduce
413 autocorrelation in the posterior samples, resulting in 30,000 posterior samples of each parameter.
414 Convergence was confirmed when R-hat was <1.05 and through visual inspections of trace plots.

415

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