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Parametric Study of CO₂ Sequestration in Deep Saline Aquifers Using Data-Driven Models

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Abstract

Large-scale geo-sequestration of anthropogenic carbon dioxide (CO₂) is one of the most promising methods to mitigate the effects of climate change without significant stress on the current energy infrastructure. However, the successful implementation of CO₂ sequestration projects in suitable geological formations, such as deep saline aquifers and depleted hydrocarbon reservoirs, is contingent upon the optimal selection of decision parameters constrained by several key uncertainty parameters. This study performs an in-depth parametric analysis of different CO₂ injection scenarios (water-alternating gas, continuous, intermittent) for aquifers with varying petrophysical properties. The petrophysical properties evaluated in this study include aquifer permeability, porosity, relative permeability, critical gas saturation, and others. Based on the extensive data collected from the literature, we generated a large set of simulated data for different operating conditions and geological settings, which is used to formulate a proxy model using different machine learning methods. The injection is run for 25 years with 275 years of post-injection monitoring. The results demonstrated the effectiveness of the machine learning models in predicting the CO₂ trapping mechanism with a negligible prediction error while ensuring a low computational time. Each model demonstrated acceptable accuracy ($R^2 > 0.93$), with the XGBoost model showing the best accuracy with an R^2 value of 0.999, 0.995, and 0.985 for predicting the dissolved, trapped, and mobile phase CO₂. Finally, a feature importance analysis is conducted to understand the effect of different petrophysical properties on CO₂ trapping mechanisms. The WAG process exhibited a higher CO₂ dissolution than the continuous or intermittent CO₂ injection process. The porosity and permeability are the most influential features for predicting the fate of the injected CO₂. The results from this study show that the data-driven proxy models can be used as a computationally efficient alternative to optimize CO₂ sequestration operations in deep saline aquifers effectively.

Keywords: CO₂ sequestration, Saline aquifers, Machine learning, Reservoir simulation, WAG process

Introduction

The increasing levels of carbon dioxide (CO₂) emission into the atmosphere from human activities such as burning fossil fuels and deforestation have raised serious concerns about their impact on the environment

and climate. This increased emission has led to global warming, causing significant changes in weather patterns, melting polar ice caps, and rising sea levels. To mitigate these impacts, there is a growing need to find ways to capture and sequester anthropogenic CO₂. One such approach is CO₂ sequestration, where CO₂ is captured from the atmosphere and industrial processes and then injected into geological formations such as depleted oil, deep saline aquifers, gas reservoirs, and unmineable coal beds (Klara et al., 2003). The International Energy Agency (IEA) predicts that by 2030, over 220 Mt of CO₂ will be captured annually (IEA, 2022), and the deep saline aquifers have the most significant storage potential for larger-scale migration of the captured CO₂ (Wei et al., 2022; Yu et al., 2021). However, despite offering significant storage potential, implementing carbon capture and storage (CCS) projects in deep saline aquifers is challenging due to the various technical and economic uncertainties. These challenges include issues related to site characterization, leakage, monitoring, and economic feasibility (Bachu, 2008; Lucier & Zoback, 2008). Therefore, properly investigating the site geology and optimizing the relevant parameters are important for successfully implementing CO₂ sequestration projects in deep saline aquifers.

Numerous experimental and numerical studies have evaluated different aspects of CO₂ storage in deep saline aquifers. The experimental studies conducted on CO₂ sequestration in deep saline aquifers are focused on exploring various aspects of the sequestration process, such as investigating different trapping mechanisms (dissolution, mineralization, and residual), determining storage capacity, enhancing the trapping capacity, and evaluating other relevant mechanisms (Bachu, 2002). However, experimental studies on deep saline aquifers for CO₂ sequestration are limited by several factors, including the high-cost experiments, safety concerns associated with the large amount of CO₂, and the limitations of techniques used to monitor and measure the storage amount of CO₂. Additionally, the small scale of the experiments can lead to difficulties in extrapolating results to a larger scale and accurately predicting the behavior of CO₂ storage in the subsurface over long periods (Liu et al., 2019; Oloruntobi & LaForce, 2009).

Physics-based numerical simulation is a widely used approach to study the CO₂ sequestration process in deep saline aquifers. This approach is particularly useful as it offers an economical and non-invasive solution to assess the CO₂ trapping mechanisms in deep saline aquifers, and it can also be used to complement and validate the findings from experimental studies. Kumar et al. (Kumar et al., 2004) conducted a compositional reservoir simulation of a typical CO₂ sequestration project in a deep saline aquifer to comprehend and quantify the sequestered CO₂ in different mechanisms. They altered the geological parameters of the base case model to present that aqueous and mineral trapping is the most efficient mechanism for CO₂ storage. Based on the cases of their study, they concluded that the capacity of CO₂ storage by residual trapping is higher than mineral trapping. Rasheed et al. (Rasheed et al., 2020) concluded that low to medium-level heterogeneous aquifers with good porosity (>20%) are suitable for CO₂ storage. The understanding of the detailed mechanism of subsurface sequestration is limited due to the scarcity of information about the heterogeneity and the geometry of the aquifer selected for the process (Bachu et al., 2007). The outcome of CO₂ sequestration is contingent upon various factors such as the injection rate, pressure, strategy, and the design of the injection well. In their study, Calabrese et al. (Calabrese et al., 2005) aimed to understand the physical and chemical processes that occur during CO₂ sequestration in a depleted gas reservoir located in northern Italy. They concluded that an optimum injection rate must be obtained to maximize the storage capacity. At a higher injection rate, the storage capacity is reduced as the gas channels through high permeability streaks, and with a lower injection rate, denser CO₂ sinks to the bottom of the gas zone and dissolves into the aquifers. The researchers also concluded that factors such as molecular diffusion, dispersion, and geochemistry is less important in assessing CO₂ storage. Water alternating gas (WAG) has been found to be effective in increasing the aquifer's storage capacity and improving the efficiency of the CO₂ sequestration process. Pan et al. (Pan et al., 2016) compared the WAG process with continuous CO₂ injection schemes in their investigation. They concluded that in WAG process, the water helps to create pathways (increased effective porosity by 2.7% and permeability by 8.4%) for the CO₂ to flow into the

aquifer, and the pressure changes caused by the alternation of the water and CO₂ help to keep the CO₂ in place within the aquifer. Al-Khdheeawi et al. (Al-Khdheeawi et al., 2018a) also concluded that the WAG process improves CO₂ storage efficiency and reduces the vertical CO₂ leakage risk. Several other investigations have explored the different aspects of CO₂ storage in deep saline aquifers, including the effect of uncertain geological and decision parameters of the storage capacity (Khanal & Shahriar, 2022), trapping efficiency, and risk associated with leakage of the storage (Al-Khdheeawi et al., 2018b), application of nanoparticles for quick mixing of CO₂ in brine (Singh et al., 2012).

Despite their versatility, numerical reservoir simulations can be difficult, time-consuming, and computationally intensive. Processing the simulation data for sensitivity analysis and optimization can be time-consuming, and each new simulation requires a significant amount of complex geological data. While a well-validated reservoir simulation model is necessary for mature projects and requires a lot of geological data, more accurate, reliable, and quicker results are often needed for initial screening and evaluation studies.

Machine learning (ML) models can be adapted as a practical tool for CCS projects. ML algorithms are able to learn from historical data, make predictions, and perform optimizations more efficiently and with greater speed than traditional simulation methods. For instance, Song et al. (Song et al., 2020) used an artificial neural network (ANN) to forecast the Capacity for CO₂ storage by considering a synthetic model to generate the training dataset. Kim et al. (Kim et al., 2017) also used ANN to predict the storage efficiency of CO₂ sequestration in deep saline aquifers. They generated 150 different reservoir models by altering properties such as permeability, porosity, residual oil saturation, thickness, and depth, which were then utilized to create a proxy model based on an artificial neural network (ANN). Menad et al. (Menad et al., 2019) utilized advanced machine learning models, such as multilayer perceptron (MLP) and radial basis function neural network (RBFNN), to predict the solubility of CO₂ in brine for monitoring CO₂ sequestration in saline aquifers. They applied different algorithms to optimize the models and concluded that RBFNN optimized with artificial bee colony (ABC) algorithm is the most reliable one for predicting the solubility of CO₂ in brine. Recently Safaei-Farouji et al. (Safaei-Farouji et al., 2022) worked on predicting the carbon trapping efficiency in storage formations by using four different ML algorithms: neuro-fuzzy inference system (ANFIS), extra tree (ET), random forest (RF) and radial basis function (RBF). They extracted 1868 data points of residual and solubility trapping of CO₂ from the published literature for training, testing, and validation. They concluded that RF provides better accuracy in predicting residual and solubility trapping with the R² value of 0.995 and 0.965, respectively. They also concluded that depth and post-injection time are the most influential factors for residual and solubility trapping performance.

In summary, this article investigates the potential effect of different injection scenarios for CO₂ trapping and compared the effectiveness of the WAG process over continuous and intermittent CO₂ injection. This work will use various machine learning algorithms to optimize the ultimate storage capacity of CO₂ in deep saline aquifers. The paper is structured as follows: First, detailed numerical simulations with varying conditions are conducted using a 3D model to evaluate various aspects of CO₂ sequestration in a deep saline aquifer. A finite reservoir is considered for the base case model where CO₂ is injected for 25 years, after which the reservoir will be monitored for a further period of 275 years to evaluate the efficiency of different trapping mechanisms (dissolution, residual, and structural trapping). After that, numerous iterations are conducted by changing operating conditions and reservoir properties. Finally, the generated data is used to train the XGBoost, RF, and ANN models for prediction. The proposed methodology can be utilized to predict CO₂ sequestration efficiency without extensive and time-consuming compositional reservoir simulations.

Materials and Methods

Reservoir Model

The CO₂ sequestration process is simulated in the GEM reservoir simulation package by the computer modeling group (CMG-v2023.40). This model considers a multi-component system of brine and CO₂ through component transport equations and thermodynamic equations between the gas and aqueous phases. In this simulation, the foundation for the mathematical models is established through the application of mass balance equations for every key component or phase present in the system (Celia et al., 2015). The accumulation of CO₂ in a closed system with α phases and i components can be calculated using the governing equation, which takes into account the contributions from convective mass transfer, diffusive/dispersive mass transfer, consumption due to chemical reactions, and the mass of the injected gas. The governing equation is-

$$\sum_{\alpha} \frac{\partial}{\partial t} \rho_{\alpha} \phi s_{\alpha} m_{\alpha i} = \sum_{\alpha} \nabla \cdot (\rho_{\alpha} \mathbf{u}_{\alpha} m_{\alpha i} + \mathbf{j}_{\alpha}) + r^i + \psi_{\alpha i} \quad (1)$$

Where, ρ_{α} is the density of the phase α , ϕ is the porosity, s_{α} is the saturation of the fluid phase α , $m_{\alpha i}$ is the mass fraction of the i component in α phase, \mathbf{u}_{α} is the Darcy flux for phase α , \mathbf{j}_{α} is the nonadvection flux of the i component in α phase and $\psi_{\alpha i}$ represents external sources or sinks of mass of component i in α phase.

A synthetic single-well CO₂ injection model shown in Fig.1 is used for the base case model in this study. This study employs a limited reservoir size to reduce computational time, with a potential for scaling to accommodate larger reservoirs (Khanal & Weijermars, 2019). The homogeneous model uses uniform porosity and permeability and is 500 m × 500 m × 125 m with a uniform grid system of 31,250 cells divided into 50 × 25 × 25 grids. The total pore and block volumes are calculated as 4.687×10^6 m³ and 3.125×10^7 m³, respectively. 300 m³/day of CO₂ is injected at a depth range of 2365 m – 2375 m through the injector placed at the corner of the model. The perforation of the well was performed on the 23rd to the 25th layer and covered a length of 10 meters. The CO₂ is injected into the aquifer for 25 years, followed by a 275-year post-injection period. The geological properties, including input parameters of the saline aquifer model, are presented in Table 1.

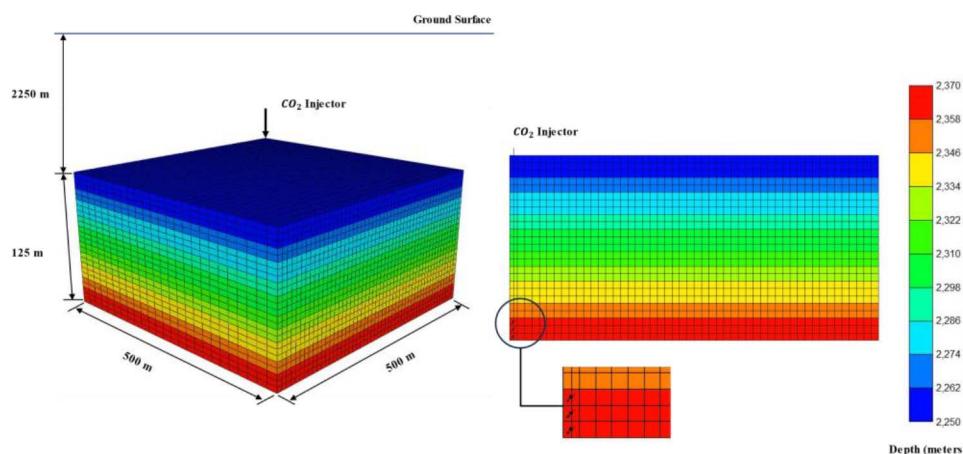


Figure 1—A visual representation of the deep saline aquifer for CO₂ storage is shown through a 3D view on the left and a cross-section view on the right.

Table 1—Saline aquifer system input parameters.

| Aquifer Properties | Values |
|---------------------------------|-------------------|
| Reservoir temperature (°C) | 50 |
| Max. bottom hole pressure (KPa) | 44,500 |
| Grid number | 31,250 (50×25×25) |
| Length (m) | 500 |
| Width (m) | 500 |
| Depth at the top (m) | 225,0 |
| Thickness (m) | 125 |
| Permeability (mD) | 200 |
| Porosity | 0.15 |
| SC _{gas} /PV | 0.2 |

*SC = surface condition, PV = pore volume

Relative permeability is an important parameter and is used to determine the effective permeability of CO₂ and brine. The relative permeability curve shown in Fig. 2 has been sourced from Khanal et al. (Khanal et al., 2024) for the reservoir model used in this study. The thermodynamic equilibrium between the gas and the aqueous phase determines the solubility of CO₂ in the aqueous phase. Peng Robinson's Equation of State (PR EOS) and Henry's law (Ali et al., 2014) are used to calculate the gas phase's fugacity and the components dissolved in the aqueous phase, respectively. The Harvey correlation is used to calculate Henry's law constant at varying temperatures and pressures (Ali et al., 2014).

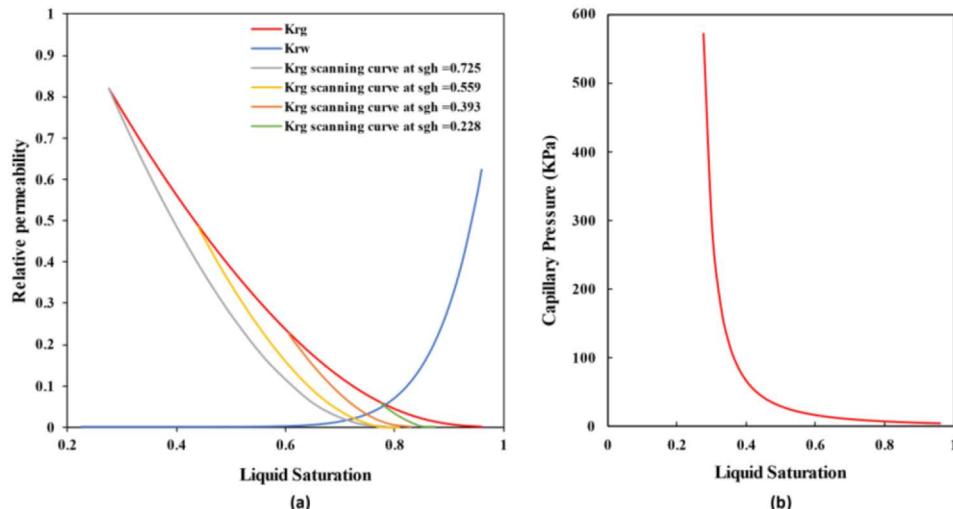


Figure 2—(a) Relative permeability and scanning curve for drainage and imbibition and (b) Capillary pressure curve (Khanal et al., 2024).

Workflow for Developing Machine Learning Models

The data generated from the simulation model is prepared for the testing, training, and validation of the model. Then, feature scaling is applied to transform the input and target data into the same scale. In this work, data normalization has been done into a standard range of -1 to 1 by using the following formula-

$$x' = 2\left(\frac{x - \min(x)}{\max(x) - \min(x)}\right) - 1 \quad (2)$$

Where x represents the original value of the given parameter, x' represents the scaled value of x , $\min(x)$ is the minimum value of x , and $\max(x)$ represents the maximum value of x .

The phase then begins by subdividing the dataset into training, testing, and validation. A K-fold ($K=5$) cross-validation method is applied during the selection of training and testing. Next, the multilayer perceptron (MLP), Random Forest (RF), and Extreme gradient boosting (XGBoost) machine learning models are applied to the target dataset. Then, finally, the performance of each model is analyzed using various statistical indexes described below-

$$\text{Coefficient of determination (R}^2\text{)} : R^2 = 1 - \frac{\sum_{i=1}^n (y_{i,sim} - y_{i,pred})^2}{\sum_{i=1}^n (y_{i,pred} - \bar{y})^2} \quad (3)$$

$$\text{Mean Squared Error (MSE)} : MAE = \frac{1}{n} \sum_{i=1}^n |y_{i,sim} - y_{i,pred}| \quad (4)$$

$$\text{Mean Squared Error (MSE)} : MSE = \frac{1}{n} \sum_{i=1}^n (y_{i,sim} - y_{i,pred})^2 \quad (5)$$

$$\text{Root Mean Squared Error (RMSE)} : RMSE = \sqrt{\frac{1}{n} \sum_{i=1}^n (y_{i,sim} - y_{i,pred})^2} \quad (6)$$

Where, \bar{y} indicates the average value, $y_{i,sim}$ and $y_{i,pred}$ denotes the simulated and predicted values, respectively. [Fig.3](#). Illustrates the general workflow of applying ML models.

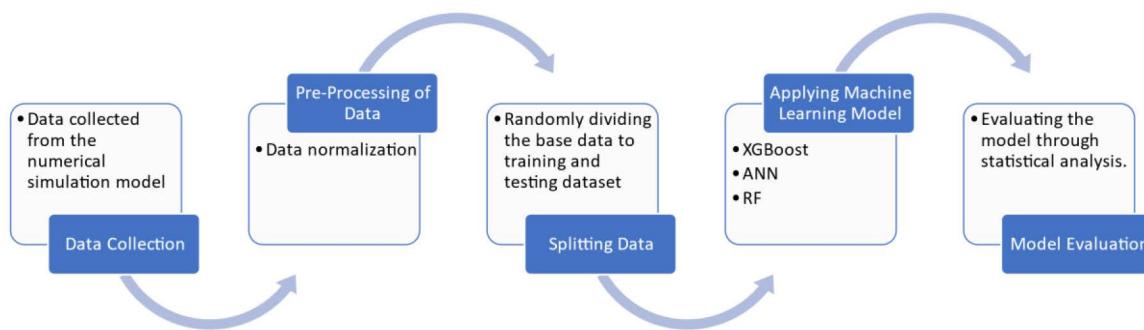


Figure 3—General workflow of applying machine learning algorithms for predicting the CO₂ trapping mechanisms in the deep saline aquifer.

Result and Discussion

Comparison between Continuous, Intermittent, and WAG Injection Processes

This section analyzes continuous and intermittent WAG injection processes to assess the CO₂-trapping efficiency. For all the processes, an equal amount of CO₂ is injected for 25 years and then monitored for 275 years. The comparison between continuous, intermittent, and WAG processes demonstrated that a higher amount of CO₂ dissolved in the WAG process. [Fig.4](#) presents the extent of CO₂ trapped by dissolution, residual, and structural trapping over the simulation period (300 years).

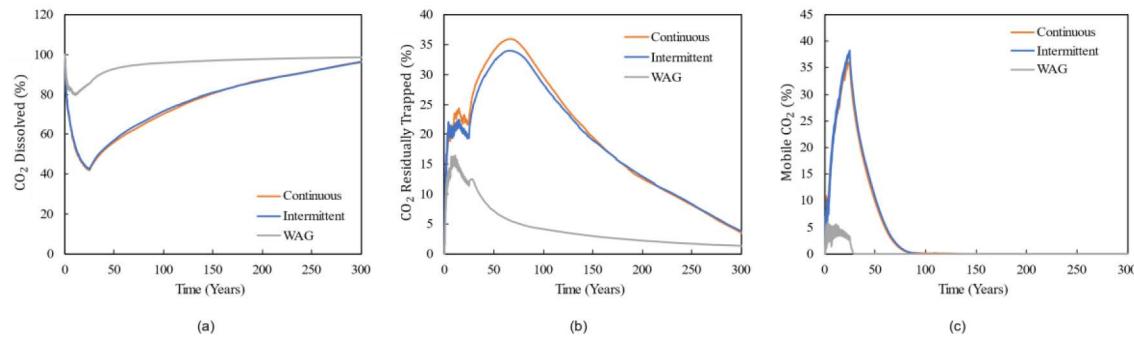


Figure 4—(a) Dissolved, (b) Residually trapped, and (c) Mobile phase CO₂ in Continuous, Intermittent, and WAG injection process after 300 years of period.

After the injection period (25 years), ~85% of the injected CO₂ is dissolved in the WAG process, whereas in the intermittent and continuous injection process, ~42% of the injected CO₂ is dissolved (Fig. 5). In the WAG process, the larger CO₂-water contact surface area increases the lateral spreading of CO₂ (Doughty, 2010); hence, more CO₂ comes into contact with the fresh brine and get dissolved. During the initial period, it is observed that more CO₂ tends to be trapped residually in the intermittent injection process. However, no significant differences in dissolution trapping are observed for the intermittent and continuous injection process. The reason behind the initial differences in residual trapping for intermittent and continuous processes is due to the time lag between imbibition and maximum CO₂ gas saturation achieved before imbibition starts. As no significant imbibition is expected to happen before the injection stops (Doughty, 2007), the intermittent injection process experiences more imbibition in the initial period. After several imbibition and drainage cycles, the maximum saturation of CO₂ is reduced in the intermittent injection process. Therefore, less residual trapping is observed after ~9 years in intermittent injection compared to the continuous injection process (Fig. 5).

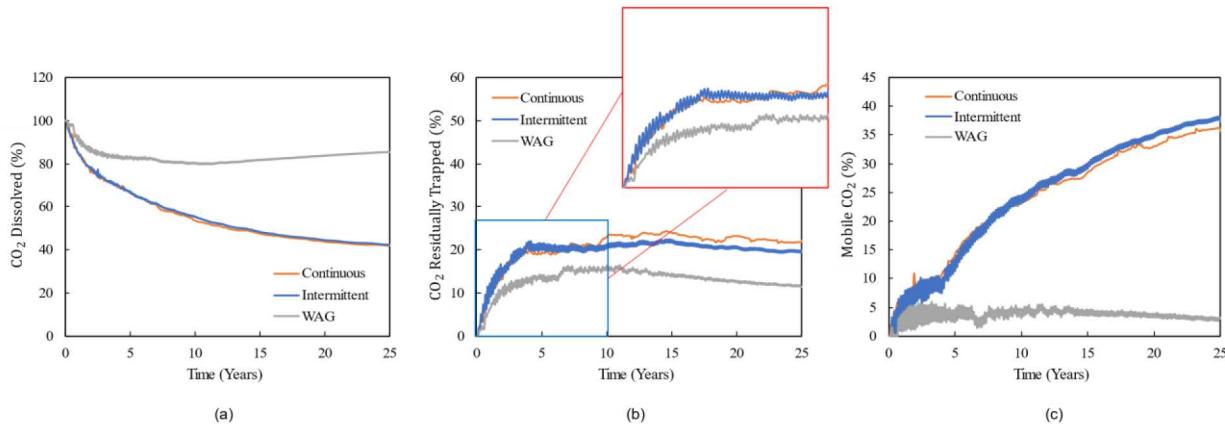


Figure 5—(a) Dissolved, (b) Residually trapped, and (c) Mobile CO₂ in Continuous, Intermittent, and WAG injection process after 25 year of period. The red box in (b) is the zoomed section of the blue box.

Effects of Wag Ratio

The WAG ratio significantly affects the trapping efficiency of CO₂ in the reservoir. This study uses WAG ratios of 1:1, 1:2, 1:3, and 2:1 to examine the effect on CO₂ trapping efficiency (Fig. 6a-b). For all the cases, an equal amount of CO₂ is injected over a period of 25 years. It is observed that increasing the CO₂ injection cycle time increases the residual trapping and mobile phase of CO₂. On the other hand, increasing the water injection cycle time leads to more dissolution trapping, as more water would be available to dissolve the injected gas. The findings are important as they suggest that CO₂ injection for a longer period of time

in the WAG process could increase the mobile phase CO_2 , which may increase the risk for CO_2 leakage through caprock, whereas increasing the water injection cycle time will lead to a lower mobile phase CO_2 and increase the dissolution trapping.

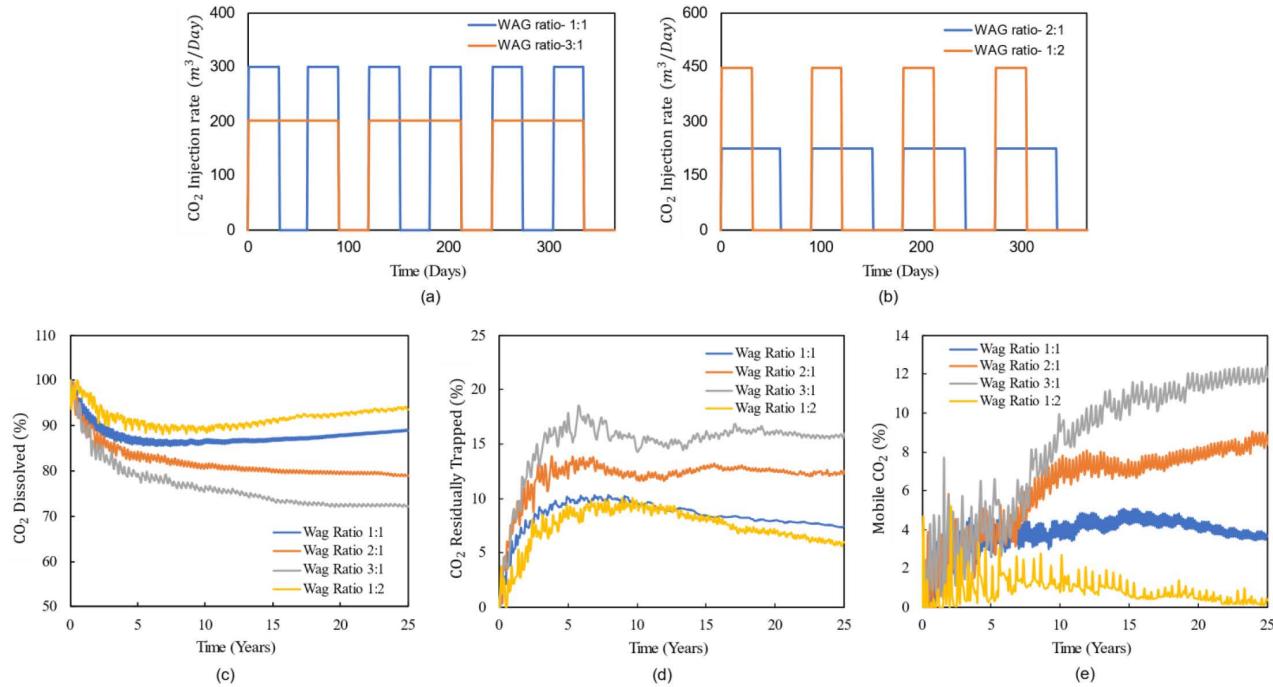


Figure 6—(a) and (b) Presents the WAG ratio for one year used in this study, and (c), (d), and (e) are the dissolved, residually trapped, and mobile phase CO_2 in different WAG ratios after 25 years of period.

Results of Proxy Models

Data Generation for Proxy Model. The WAG process with a ratio of 1:1 is selected as the base case model for developing the ML model. We generated numerous iterations of the base case by changing different uncertain properties of the reservoir. To represent different geological storage, nine parameters were varied. The range of values for the selected parameters is based on previous literature (Krevor et al., 2012; Medina et al., 2011; Michael et al., 2010; Vo Thanh & Lee, 2022; Zhao et al., 2010). Table 2 Represents the statistical overview of the input parameters considered in this study. In total, 90 geological realizations are considered during the WAG process to investigate the uncertainty of reservoir heterogeneities. For each case after five years of interval, the amount of CO_2 trapped by dissolution, residual, and structural is considered as one data point. In total, 5,551 data points are generated from the simulations for training, testing, and validating the ML model. In order to avoid bias in selecting the training and testing dataset, this study employs a five-fold cross-validation method.

Table 2—Statistical overview of the input parameters.

| Parameter | Range | Reference |
|---------------------------------|-------------|---|
| Porosity | 0.10 - 0.30 | (Medina et al., 2011; Michael et al., 2010) |
| Permeability (mD) | 20 - 1000 | (Medina et al., 2011; Michael et al., 2010) |
| Permeability ratio, k_v/k_h | 0.01 - 1 | (Zhao et al., 2010) |
| Maximum residual gas saturation | 0.1-0.5 | (Vo Thanh & Lee, 2022) |
| Krg at Connate liquid | 0.5-0.9 | (Krevor et al., 2012) |
| Exponent for calculating Krw | 5-17 | (Krevor et al., 2012) |

| Parameter | Range | Reference |
|-------------------------------------|----------|-----------------------|
| Exponent for calculating Krg | 1-4 | (Krevor et al., 2012) |
| Endpoint Saturation: Critical gas | 0.02-0.2 | (Krevor et al., 2012) |
| Endpoint Saturation: Critical water | 0.2-0.4 | (Krevor et al., 2012) |

Results of Machine Learning Models. In this study, three different ML models (RF, MLP, and XGBoost) are applied to predict the fate of the CO₂ inside the reservoir. The capacity to capture nonlinear relationships and the ability to handle high-dimensional and heterogeneous data made these models well-recognized in CCS projects (Khanal & Shahriar, 2022; You et al., 2020). However, the hyperparameters of the ML models need to be tuned to improve the prediction accuracy. This study uses the RandomSearchCV optimization technique to identify the optimal set of hyperparameters. As it can effectively explore the hyperparameter space by randomly sampling values from redefined distributions, especially when dealing with large search space of hyperparameters. Moreover, it is faster than other hyperparameter tuning techniques. Table 3 summarizes the hyperparameters and their optimal values for RF, MLP, and XGBoost models.

Table 3—Hyperparameters and their optimal values for the ML models used in this study.

| Models | Prediction Target | Parameters |
|---------|---------------------------|--|
| XGBoost | CO ₂ Dissolved | {‘colsample_bytree’: 0.84, ‘learning_rate’: 0.189, ‘max_depth’: 5, ‘n_estimators’: 177, ‘subsample’: 0.96} |
| | CO ₂ Trapped | {‘colsample_bytree’: 0.83, ‘learning_rate’: 0.108, ‘max_depth’: 8, ‘n_estimators’: 117, ‘subsample’: 0.82} |
| | CO ₂ Mobile | {‘colsample_bytree’: 0.91, ‘learning_rate’: 0.112, ‘max_depth’: 8, ‘n_estimators’: 168, ‘subsample’: 0.85} |
| MLP | CO ₂ Dissolved | {‘solver’: ‘lbfgs’, ‘hidden_layer_sizes’: (64, 32), ‘alpha’: 0.1, ‘activation’: ‘relu’} |
| | CO ₂ Trapped | |
| | CO ₂ Mobile | |
| RF | CO ₂ Dissolved | {‘bootstrap’: True, ‘max_depth’: 21, ‘min_samples_leaf’: 1, ‘min_samples_split’: 2, ‘n_estimators’: 1281} |
| | CO ₂ Trapped | |
| | CO ₂ Mobile | |

Figure 7 shows the performance of ML models for predicting the dissolved, trapped, and mobile phase CO₂. A higher correlation factor (R²) and lower evaluation matrices generally correspond to higher predictive model accuracy. For a better understanding of the prediction accuracy, the statistical accuracy of each model is detailed in Table 4. For all the models, the R² values are satisfactory. However, the XGBoost model shows the best accuracy with R² values of 0.999, 0.995, and 0.985 for predicting the dissolved, trapped, and mobile phase CO₂, respectively. For all the cases, the R² value for mobile phase CO₂ is lower compared to the dissolved and trapped CO₂. As time goes on, the mobile phase CO₂ approaches zero due to dissolution and residual trapping. Since all the data for dissolved, trapped, and mobile phase CO₂ are recorded after every five-year interval, time is strongly correlated with the amount of CO₂ trapped by these mechanisms. But as the mobile phase CO₂ diminishes to zero sooner than the dissolved and trapped phase CO₂, less data is available to train the machine learning models, which leads to a lower R² value.

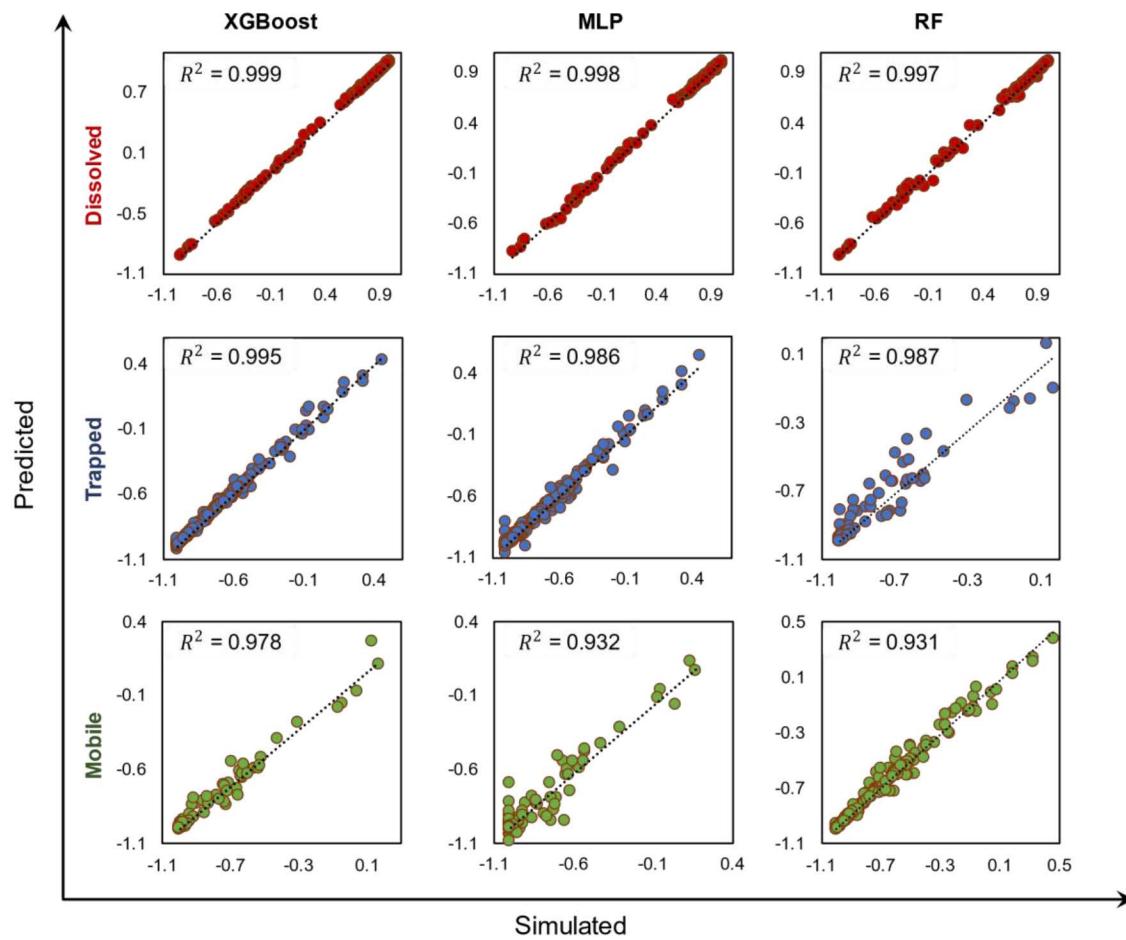


Figure 7—Performance of predictive models obtained from XGBoost, RF, and MLP models.

Table 4—Statistical accuracy of the proposed machine learning models in predicting dissolved, trapped, and mobile phase CO_2 .

| Models | Prediction Target | Evaluation Metrics | | | |
|---------|-------------------------|--------------------|---------|---------|---------|
| | | R^2 | MAE | MSE | RMSE |
| XGBoost | CO_2 Dissolved | 0.99962 | 0.00306 | 0.00004 | 0.00620 |
| | CO_2 Trapped | 0.99514 | 0.00667 | 0.00025 | 0.01569 |
| | CO_2 Mobile | 0.97876 | 0.00456 | 0.00033 | 0.01823 |
| MLP | CO_2 Dissolved | 0.99862 | 0.00633 | 0.00014 | 0.01174 |
| | CO_2 Trapped | 0.98672 | 0.01409 | 0.00067 | 0.02594 |
| | CO_2 Mobile | 0.93292 | 0.00977 | 0.00105 | 0.03240 |
| RF | CO_2 Dissolved | 0.99799 | 0.00440 | 0.00019 | 0.01414 |
| | CO_2 Trapped | 0.98773 | 0.00937 | 0.00062 | 0.02494 |
| | CO_2 Mobile | 0.93123 | 0.00839 | 0.00108 | 0.03281 |

Feature Importance Analysis. We investigated the effect of input parameters on the fate of CO_2 inside the reservoir by calculating the feature importance plot for the machine learning models evaluated in this study (XGBoost and RF). In the feature analysis of RF and XGBoost, there must be a strong relationship of time with dissolved, mobile, and trapped CO_2 , as with time, the fate of the CO_2 changes within the reservoir. However, we excluded time in the feature analysis to underscore the importance of other parameters in relation to the different phases of CO_2 . Fig.8 shows that for both models, porosity is the most important

parameter for predicting the dissolved CO_2 , which means dissolution trapping of CO_2 is greatly related to porosity. If the reservoir porosity is low, the pressure within the reservoir will increase with CO_2 injection. As the pressure-dependent Henry's constant controls the gas dissolution, the dissolution of CO_2 will also increase (Khanal et al., 2024). For the trapped CO_2 , the most important feature is the vertical permeability, with a score of 0.60 in RF and 0.33 in the XGBoost model. This indicates that compared to the horizontal permeability, the vertical permeability of the reservoir significantly influences the trapping of CO_2 . In the case of predicting the mobile phase CO_2 , vertical permeability is again the most crucial feature, with an importance factor of 0.62, and the second most important feature is horizontal permeability in the RF model. This is because, with high permeability, more brine can flow towards the perforation zone and will come in contact with the fresh brine; hence, more CO_2 will dissolve. We can also relate this with the feature importance plot, which shows that the second most important feature for predicting dissolved CO_2 is the vertical permeability of the reservoir.

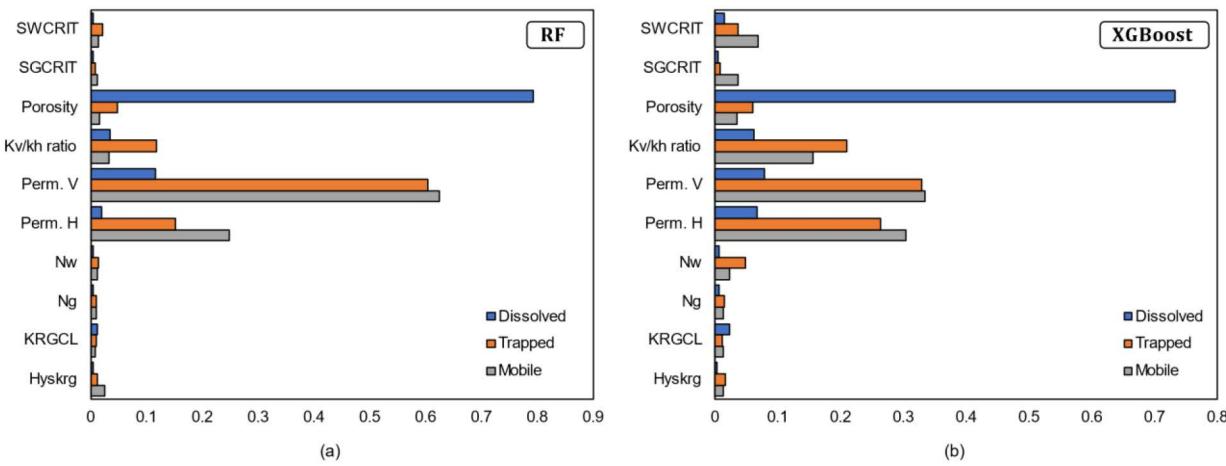


Figure 8—Feature importance plot for dissolved, residual, and mobile phase CO_2 for (a) RF and (b) XGBoost model. SWCRIT: Critical water saturation, SGCRIT: Critical gas saturation, Perm. V: Vertical permeability, Perm. H: Horizontal permeability, KRGCL: Krg at Connate Liquid, Hyskrg: Hysteresis residual gas saturation, Nw: Exponent for calculating Krw and Ng: Exponent for calculating Krg .

Conclusions

This study investigated different CO_2 injection schemes in underground CO_2 storage projects. A proxy model for a deep saline aquifer is developed using the machine learning model for predicting the CO_2 trapping scenarios. Moreover, a feature importance analysis has been conducted to understand the effect of different parameters on dissolved, trapped, and mobile phase CO_2 . The following conclusion can be drawn from the findings of this work:

1. The WAG process is more suitable for safe CO_2 trapping in deep saline aquifers (Fig.5). In the WAG process, more CO_2 comes in contact with the brine solution; hence, the dissolution rate is higher compared to the intermittent and continuous injection process. However, further investigation on the effect of injection rate, injection position, and economic analysis is required before concluding which injection process is optimal for CCS projects.
2. Increasing the water injection cycle time in the WAG process is more suitable for CO_2 storage through dissolution trapping (Fig.6).
3. A proxy-based reservoir model can be developed using the ML models to accurately predict the CO_2 trapping mechanisms, which offers saving in both time and resources.
4. Three different ML models (XGBoost, RF, and MLP) are used to predict the dissolved, trapped, and mobile CO_2 scenarios. The R^2 , MAE, MSE, and RMSE values are satisfactory for all the models.

However, the XGBoost model shows better prediction accuracy than other models, with an R^2 value of 0.999, 0.995, and 0.985 for predicting the dissolved, trapped, and mobile phase CO_2 , respectively (Table 4).

5. Porosity is the most important feature for predicting dissolved CO_2 , while vertical permeability is the most crucial parameter for predicting trapped and mobile phase CO_2 (Fig. 8).

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