

# 1 Effect of the 2022 summer drought across forest types 2 in Europe

3 Mana Gharun<sup>1</sup>, Ankit Shekhar<sup>2,3</sup>, Jingfeng Xiao<sup>4</sup>, Xing Li<sup>5</sup>, Nina Buchmann<sup>2</sup>

4  
5 <sup>1</sup> Institute of Landscape Ecology, University of Münster, Münster, Germany

6 <sup>2</sup> Institute of Agricultural Sciences, ETH Zürich, Zürich, Switzerland

7 <sup>3</sup> Agricultural and Food Engineering Department, Indian Institute of Technology Kharagpur, Kharagpur,  
8 India

9 <sup>4</sup> Earth Systems Research Center, University of New Hampshire, New Hampshire, USA

10 <sup>5</sup> Research Institute of Agriculture and Life Sciences, Seoul National University, Seoul, South Korea

11  
12 Correspondence to: Mana Gharun, mana.gharun@uni-muenster.de

## 13 Abstract

14 Forests in Europe experienced record-breaking dry conditions during the 2022 summer.  
15 The direction in which various forest types respond to climate extremes during their  
16 growing season is contingent upon an array of internal and external factors. These factors  
17 include the extent and severity of the extreme conditions and the tree ecophysiological  
18 characteristics adapted to environmental cues, which exhibit significant regional  
19 variations. In this study we aimed to: 1) quantify the extent and severity of the extreme  
20 soil and atmospheric dryness in 2022 in comparison to two most extreme years in the  
21 past (2003 and 2018), 2) quantify response of different forest types to atmospheric and  
22 soil dryness in terms of canopy browning and photosynthesis, and 3) relate the functional  
23 characteristics of the forests to the emerging responses observed remotely at the canopy  
24 level. For this purpose, we used spatial meteorological datasets between 2000 to 2022  
25 to identify conditions with extreme soil and atmospheric dryness. We used the near-  
26 infrared reflectance of vegetation (NIRv) derived from the MOderate Resolution Imaging  
27 Spectroradiometer (MODIS), and the Global OCO-2 Solar Induced Fluorescence  
28 (GOSIF) as an observational proxy for ecosystem gross productivity, to quantify the  
29 response of forests at the canopy level.

30 In summer 2022, southern regions of Europe experienced exceptionally pronounced  
31 atmospheric and soil dryness. These extreme conditions resulted in a 30% more  
32 widespread decline in GOSIF across forests compared to the drought of 2018, and 60%  
33 more widespread decline compared to the drought of 2003. Although the atmospheric  
34 and soil drought were more extensive and severe (indicated by a larger observed  
35 maximum z-score) in 2018 compared to 2022, the negative impact on forests, as  
36 indicated by declined GOSIF, was significantly larger in 2022. Different forest types were  
37 affected in varying degrees by the extreme conditions in 2022. Deciduous broad-leaved  
38 forests were the most negatively impacted due to the extent and severity of the drought  
39 within their distribution range. In contrast, areas dominated by Evergreen Needle-Leaf  
40 Forests (ENF) in northern Europe experienced a positive soil moisture (SM) anomaly and  
41 minimal negative vapor pressure deficit (VPD) in 2022. These conditions led to enhanced  
42 canopy greening and stronger solar-induced fluorescence (SIF) signals, benefiting from  
43 the warming. The higher degree of canopy damage in 2022, despite less extreme  
44 conditions, highlights the evident vulnerability of European forests to future droughts.

45

46 *Keywords: photosynthesis, soil drought, atmospheric drought, canopy browning, gross*  
47 *primary production*

## 48 **Introduction**

49 The frequency and intensity of drought events have been rising globally, and future global  
50 warming is expected to further increase their occurrence (Seneviratne et al. 2021;  
51 Röthlisberger and Papritz 2023). Particularly over the past two decades, many regions in  
52 Europe have experienced widespread drought conditions, notably during the summers of  
53 2003, 2010, and 2018 (Bastos et al. 2020; Zhou et al. 2023). The extreme conditions  
54 caused widespread ecological disturbances (Müller and Bahn 2022) and reduced the  
55 capacity of forests for carbon uptake, thereby diminishing their potential for mitigating  
56 climate change (van der Woude et al. 2023). Additionally, heatwaves and prolonged  
57 droughts stress vegetation, making it more susceptible to other biotic and abiotic stress  
58 factors. This increased vulnerability leads to higher tree mortality, elevated wildfire risks,

59 and a loss of biodiversity among plants and animals living at the edge of their temperature  
60 tolerance. These conditions also alter phenology and plant development, causing  
61 cascading effects on ecosystem functioning (Seidl et al. 2017).

62 The spatial extent and severity of drought events vary, and their impacts depend on local  
63 ecological characteristics of the forests, species-specific temperature and moisture  
64 thresholds that limit tree functioning, as well as adaptation strategies and acclimation of  
65 trees to more frequent and intense extreme conditions (Gessler et al. 2020). For example,  
66 comparing the 2003 and 2018 extreme years, the year 2018 was characterized by a  
67 climatic dipole, featuring extremely hot and dry weather conditions north of the Alps but  
68 comparably cool and moist conditions across large parts of the Mediterranean. Negative  
69 drought impacts appeared to affect an area 1.5 times larger and to be significantly  
70 stronger in summer 2018 compared to summer 2003 (Buras et al. 2020).

71 In 2022, Europe faced its second hottest and driest year on record, with the summer of  
72 that year being the warmest summer ever recorded. Conditions in summer 2022 led to  
73 record-breaking heatwave and drought events across many regions (Copernicus Climate  
74 Change Service, 2023). Compound drought and heatwave conditions in 2022 caused  
75 widespread crop damage, water shortages, and wildfires across Europe. The hardest-hit  
76 areas were the Iberian Peninsula, France, and Italy, where temperatures exceeded 2.5°C  
77 above normal, and severe droughts persisted from May to August (Tripathy and Mishra  
78 2023). The reduced soil moisture due to precipitation deficits and high temperatures,  
79 contributed to the persistence and severity of drought, creating a positive feedback loop  
80 where dry soils led to even drier conditions (Tripathy and Mishra 2023).

81 Drought and heatwaves have a range of detrimental effects on trees and forests. The  
82 most immediate impact is that elevated air temperatures and increased dryness, whether  
83 in the soil or in the atmosphere, disrupt mesophyll and stomatal conductance, thereby  
84 impairing carbon uptake (Marchin et al. 2022). Plants reduce stomatal conductance under  
85 severe drought to reduce water stress at the expense of reduced rates of photosynthesis  
86 (Oren et al., 1999). Drought also increases the chance of hydraulic failure, which can lead  
87 to tree mortality (Choat et al. 2018). Additionally, rising temperatures reduce the  
88 enzymatic activity in trees, which in turn diminishes the forest's gross primary productivity  
89 (Gourlez de la Motte et al. 2020). Elevated temperatures can also increase respiration

90 rates in both soil and trees, which reduces the forest's net carbon uptake and their ability  
91 to mitigate anthropogenic CO<sub>2</sub> emissions (van der Molen et al. 2011; Anjileli et al. 2021).  
92 Drought also restricts the movement of nutrients in soil water, reducing their availability  
93 to trees and consequently impacting their growth and productivity (Bauke et al. 2022).  
94 Changes in plant water-use and nutrient cycling can trigger feedback loops that magnify  
95 the effects of drought and heat stress. For instance, reduced plant cover can increase  
96 soil temperatures and further accelerate water loss and increase plant water demand  
97 (Haesen et al. 2023). On the other hand, increased atmospheric dryness or reduced soil  
98 moisture levels increase stomatal closure which limits transpiration and leads to higher  
99 leaf temperature that intensifies heat stress on plants (Drake et al. 2018). Reduced  
100 transpiration and photosynthesis elevate surface temperatures and atmospheric CO<sub>2</sub>  
101 concentrations, altering local and regional climate patterns and intensifying the frequency  
102 and severity of extreme events (Humphrey et al. 2018). These effects vary significantly  
103 depending on forest type and species composition. Together with the characteristics of  
104 the extreme events themselves – such as their extent and severity- this variability  
105 complicates our understanding of how drought affects the functionality of different forest  
106 ecosystems (Gharun et al. 2020; Shekhar et al. 2023). These feedback loops highlight  
107 the urgent need to assess how climate extremes impact different forest types, which are  
108 crucial for sequestering significant portions of anthropogenic emissions. Our study aims  
109 to 1) quantify the extent and severity of the extreme conditions in 2022 – focusing on soil  
110 and atmospheric dryness- and compare them to those of two previous extreme years  
111 (2003, 2018), 2) quantify the responses of different forest types to drought in terms of  
112 canopy browning and photosynthesis, and 3) connect the functional characteristics of the  
113 forests with the canopy-level responses observed.

## 114 **Methods**

### 115 *Meteorological dataset*

116 We used Europe-wide gridded datasets covering daily mean air temperature (Tair; °C),  
117 daily mean relative humidity (RH; %) and daily mean soil moisture (SM; m<sup>3</sup>m<sup>-3</sup>) for the  
118 topsoil layer (0-7 cm depth), spanning from 2000-2022. The study area encompasses

119 longitudes from 11°W to 32°E, and latitudes from 35.8°N to 72°N, approximately 4.45  
120 million km<sup>2</sup>. We sourced the Tair and RH datasets from the E-OBS v27.0e dataset which  
121 provides daily data at 0.1°×0.1° spatial resolution (Cornes et al., 2018; Klein et al., 2002).  
122 We calculated daily mean vapor pressure deficit (VPD; kPa) from Tair and RH using  
123 Equation 1 (Dee et al. 2011).

124

$$125 \quad VPD = \left(1 - \frac{RH}{100}\right) \times 0.6107 \times 10^{\frac{7.5 \times Tair}{237.3 + Tair}} \quad (1)$$

126

127 The topsoil SM dataset was extracted from the most recent reanalysis data from  
128 ECMWF's (European Centre for Medium-range Weather Forecasts) new land component  
129 of the fifth generation of European Reanalysis (ERA5-Land) dataset (daily at 0.1°×0.1°  
130 resolution; Munoz-Sabater et al., 2021). ERA5-Land provides soil moisture (SM) data at  
131 an hourly interval with a spatial resolution of 0.1° × 0.1°. For our analysis, we aggregated  
132 the hourly SM data into daily averages. Recent validation studies using in-situ  
133 measurements and satellite data have confirmed the high accuracy of surface SM  
134 simulations from ERA5-Land (Albergel et al., 2012; Lal et al., 2022; Muñoz-Sabater et al.,  
135 2021). Additionally, SM data from ERA5-Land have been utilized to investigate drought  
136 and global SM patterns (see Lal et al., 2023; Shekhar et al., 2024b). We re-sampled the  
137 Tair, VPD, and SM data from daily (0.1° × 0.1°) to 8-day (0.05° × 0.05°) intervals to align  
138 with the temporal and spatial resolution of the vegetation response dataset (see below).

### 139 *Forest canopy response dataset*

140 In order to assess the forest canopy response to drought stress, we used two satellite-  
141 based proxies:

142 1) The structure-based NIRv (Near-Infrared Reflectance of Vegetation) index derived  
143 from MODIS (Moderate Resolution Imaging Spectroradiometer; 8-day 500m × 500m  
144 MOD09Q1 v6.1 product) which is calculated using surface spectral reflectance at near-  
145 infrared band (R<sub>NIR</sub>) and red band (R<sub>Red</sub>) as shown in Equation 2 (Badgley et al. 2017).  
146 The calculated NIRv at 500m resolution was aggregated to a 0.05°×0.05° resolution  
147 (daily) by averaging.

148

149 
$$NIR_V = R_{NIR} \times \frac{R_{NIR} - R_{Red}}{R_{NIR} + R_{Red}}$$
 (2)

150

151 2) The physiological-based reconstructed global OCO-2 (Observation Carbon  
152 Observatory - 2) solar induced fluorescence (GOSIF) dataset. Solar-induced  
153 fluorescence (SIF) is an energy flux (unit:  $Wm^{-2}\mu m.sr^{-1}$ ) re-emitted as fluorescence by the  
154 chlorophyll *a* molecules in the plants during photosynthesis (Baker, 2008). Recent  
155 extensive research has established a strong link between Solar-Induced Fluorescence  
156 (SIF) and vegetation photosynthesis, validating SIF as an effective proxy for ecosystem  
157 gross primary productivity (GPP) (Li et al. 2018; Magney et al. 2019; Shekhar et al., 2022).  
158 The GOSIF dataset was created by training a Cubist Regression Tree model to gap-fill  
159 SIF retrievals from OCO-2 satellite. This was done using MODIS Enhanced Vegetation  
160 Index (EVI) and meteorological reanalysis data from MERRA-2 (Modern-Era  
161 Retrospective analysis for Research and Applications), which includes photosynthetically  
162 active radiation (PAR), VPD, and air temperature (see Li and Xiao, 2019). We  
163 downloaded GOSIF data set (v2) from the Global Ecology Data Repository  
164 ([http://data.globalecology.unh.edu/data/GOSIF\\_v2/](http://data.globalecology.unh.edu/data/GOSIF_v2/), last accessed on 25 July 2024). The  
165 GOSIF was available from 2000-2022 at 8-day temporal scale with a spatial resolution of  
166  $0.05^\circ \times 0.05^\circ$  (Li and Xiao, 2019).

167 GOSIF signals provide information about physiological response of forest photosynthesis  
168 while NIRv (a recently developed vegetation index) signals provide information about the  
169 health status of the canopy. NIRv is preferred over NDVI and EVI as it can isolate the  
170 vegetation signal, mitigate mixed-pixel issue, and partly address the influences of  
171 background brightness and soil contamination (Zhang et al. 2022). The two vegetation  
172 proxies used in this study are anticipated to offer complementary insights into vegetation  
173 response to drought.

174 *Land cover dataset*

175 In this study, we focused on five different types of forests (and woodlands) across Europe,  
176 namely, evergreen needleleaf forest (ENF), evergreen broadleaf forest (EBF), deciduous

177 broadleaf forest (DBF), mixed forest (MF), and woody savannas (WSA). The spatial  
178 distribution of the five different forest types across Europe is shown in Figure 1. We used  
179 the yearly MODIS land cover product (MCD12C1 version 6.1 at  $0.05^{\circ} \times 0.05^{\circ}$  resolution)  
180 for the years of 2001, 2006, 2011, 2016 and 2021, to extract total areas covered by each  
181 forest type. Area of each grid cell was calculated using trigonometric equations  
182 considering the latitudinal and longitudinal variations arising due to Earth's spherical  
183 shape (Ellipsoid). Only areas that were consistently identified as each forest type over the  
184 five-year period were included in the analysis. This means that only pixels common  
185 across these five years were selected, and with more than 50% of the  $0.05^{\circ} \times 0.05^{\circ}$  pixel  
186 area identified as forests. The forested areas selected for this study encompassed  
187 907,875 km<sup>2</sup>, which represents approximately 24% of Europe's total land area. Out of the  
188 total area about 23% (206'212 km<sup>2</sup>) was dominated by ENFs distributed largely across  
189 Northern Europe (NEU). Approximately 1% (7'000 km<sup>2</sup>) of the area was dominated by  
190 EBFs, located entirely in Mediterranean Europe (MED), and about 10% (92'209 km<sup>2</sup>) was  
191 dominated by DBF which was largely distributed across MED. Approximately 20%  
192 (174'934 km<sup>2</sup>) of the total forested area was dominated by MFs largely dominating Central  
193 Europe (CEU), and about 47% (427'529 km<sup>2</sup>) was dominated by WSA mostly found in  
194 NEU (Figure 1).

#### 195 *Drought detection and statistical data analysis*

196 The focus of our analysis was on the summer months during three extreme years of 2003,  
197 2018, and 2022. For this purpose, we subset VPD, soil moisture (SM), and both  
198 vegetation proxies (NIRv and GOSIF) for the months of June, July, August (JJA) which  
199 consisted of fourteen 8-day periods, for each forested pixel between 2000 and 2022. We  
200 restricted our analysis to the months of June-July-August so our study is 1) comparable  
201 with existing studies focused on the summer drought 2) to capture the peak of the warm  
202 and dry conditions across Europe, that would be most stressful for the vegetation  
203 functioning, from the perspective of heat and water supply.  
204 To account for the impact of the observed greening trend across Europe on vegetation  
205 proxy anomalies during the extreme years (2003, 2018, 2022), we applied a detrending

206 process to the summer mean NIRv and GOSIF data. This detrending was performed  
207 pixel-wise from 2000 to 2022 using a simple linear regression model (Buras et al., 2020).  
208 We then calculated pixel-wise standardized summer anomalies, expressed as z-scores  
209 ( $Var_z$ ), for all variables—VPD, SM, and the detrended NIRv and GOSIF (hereafter  
210 referred to as NIRv and GOSIF)—for each year, including the extreme years, using  
211 Equation 3.

212

213 
$$Var_z \text{ (unitless)} = \frac{Var - Var_{mean}}{Var_{sd}} \quad (3)$$

214

215 where,  $Var_{mean}$  and  $Var_{sd}$  are mean and standard deviation of any variable over the 2000-  
216 2022 period.

217

218 In drought identification studies, classification of ‘normal’ (not to be confused with normal  
219 distribution), ‘drought’ (used synonymously with ‘dry’), or ‘wet’, is largely done using a  
220 standardized index, such as SPI (Standardized Precipitation Index), SPEI (Standardized  
221 Precipitation Evapotranspiration Index), and z-score among others (see Mishra and  
222 Singh, 2011). All studies that use a standardized index for classification, classify “normal”  
223 conditions when the index is between -1 and 1, and “below normal” conditions when the  
224 index is  $< -1$ , and “above normal” conditions when the index  $> 1$  (Jain et al., 2015, Wable  
225 et al., 2019, Dogan et al., 2012, Tsakiris and Vangelis, 2005). In this study, we classified  
226 drought conditions as occurring when soil moisture is below normal ( $SM_z < -1$ ) and VPD  
227 is above normal ( $VPD_z > 1$ ), indicating both soil AND atmospheric dryness. This  
228 threshold-based approach using standardized anomalies aligns with established methods  
229 for drought identification and is pertinent for studying drought impacts on forests. Both  
230 soil moisture and VPD directly affect vegetation functioning, making them effective  
231 proxies for identifying environmental constraints on plant physiological performance.  
232 Furthermore, such classification of ‘normal’ (and thus, ‘above normal’ and ‘below normal’  
233 used in this study) based on z-scores (also called standardized anomalies) can be done  
234 for any meteorological and/or response variables, such as NIRv and GOSIF done in this  
235 study, making the narration of results coherent across different variables.

236 We used the Pearson correlation coefficient ( $r$ ) and partial correlation coefficients (Pr) to  
237 understand the spatial (across space for each year) and temporal (during each year)  
238 correlation of GOSIF and  $\text{NIR}_v$  anomalies with SM and VPD anomalies (Dang et al.,  
239 2022). We calculated the partial correlation coefficient using equations 4-7:

240

241 
$$\text{Pr}(\text{GOSIF}, \text{SM}) = \frac{r(\text{GOSIF}, \text{SM}) - r(\text{GOSIF}, \text{VPD}) \times r(\text{SM}, \text{VPD})}{\sqrt{1 - r(\text{GOSIF}, \text{VPD})^2} - \sqrt{1 - r(\text{SM}, \text{VPD})^2}}$$
 (4)

242

243 
$$\text{Pr}(\text{GOSIF}, \text{VPD}) = \frac{r(\text{GOSIF}, \text{VPD}) - r(\text{GOSIF}, \text{SM}) \times r(\text{SM}, \text{VPD})}{\sqrt{1 - r(\text{GOSIF}, \text{SM})^2} - \sqrt{1 - r(\text{SM}, \text{VPD})^2}}$$
 (5)

244

245 
$$\text{Pr}(\text{NIR}_v, \text{SM}) = \frac{r(\text{NIR}_v, \text{SM}) - r(\text{NIR}_v, \text{VPD}) \times r(\text{SM}, \text{VPD})}{\sqrt{1 - r(\text{NIR}_v, \text{VPD})^2} - \sqrt{1 - r(\text{SM}, \text{VPD})^2}}$$
 (6)

246

247 
$$\text{Pr}(\text{NIR}_v, \text{VPD}) = \frac{r(\text{NIR}_v, \text{VPD}) - r(\text{NIR}_v, \text{SM}) \times r(\text{SM}, \text{VPD})}{\sqrt{1 - r(\text{NIR}_v, \text{SM})^2} - \sqrt{1 - r(\text{SM}, \text{VPD})^2}}$$
 (7)

## 248 **Results**

249 *Severity of the 2022 summer drought compared to 2018 and 2003*

250 Figure 2 shows the extent and magnitude of anomalies (z-score) of VPD and top layer (0-  
251 7 cm) soil moisture content during the summer months in 2003, 2018, and 2022 across  
252 Europe. In summer 2022, particularly southern regions of Europe experienced the most  
253 pronounced increase in atmospheric (z-score  $> 1$ ) and soil dryness (z-score  $< -1$ ) (Figure  
254 2) while in 2018 we observed the most widespread VPD and SM anomalies in northern  
255 Europe (Figure 2).

256 Figure 3 shows the intensity of atmospheric and soil drought via z-score values of VPD  
257 and SM anomalies over the summer months (JJA) in 2003, 2018, and 2022. The total  
258 affected area displayed in Figure 3 is the sum of all pixels within the given z-score bin  
259 during the summer period where z-scores are averaged for each bin for the summer  
260 period. Restricted to forested areas, atmospheric and soil drought was 55% and 58%  
261 more extensive in 2018 compared to 2022, and in both years more extensive than in 2003  
262 (Figure 3). In 2022, 28 Mha of forested areas in Europe experienced an extremely high

263 VPD (z-score > 1), while in 2018, 63 Mha experienced such extreme conditions. In 2022,  
264 21 Mha of forested areas experienced an extremely low soil moisture content (z-score <  
265 -1) while in 2018, 50 Mha of forests in Europe were affected by such extreme conditions.  
266 In 2003 an area of 25 Mha was affected by extremely dry air and a similar area was  
267 affected by extremely dry soil (Figure 3). A comparison of soil drought detected from SM  
268 at 0-100 cm showed a similar result in terms of drought severity and spatial coverage and  
269 thus we used SM at 0-7 cm soil layer for our analysis (see Figure S1).

270

#### 271 *Forest canopy response to the 2022 drought*

272 The intensity of GOSIF and NIRv anomalies over the summer months (JJA) in 2003,  
273 2018, and 2022 are displayed in Figure 4. The extent shown in Figure 4 is the sum of all  
274 pixels within the given z-score bin during the summer period (z-scores are averaged for  
275 each bin). Compared to 2018, the extremely dry conditions in 2022 led to 30% increase  
276 in forested areas that exhibited declined photosynthesis (17 Mha in 2022 compared to 12  
277 Mha in 2018) (Figure 4). The extent of the canopy browning observed in 2022 was similar  
278 to 2018, which in both years was 120% of the extent of observed canopy browning in  
279 2003 (11 Mha compared to 5 Mha observed in 2003) (Figure 4).

280 Figure 5a shows the GOSIF anomalies (z-score) across all forested areas in Europe. The  
281 intensity and extent of the GOSIF anomalies during the summer months (JJA) in each  
282 year are shown for different forest types in Figure 5b. Across specific forest types, DBFs  
283 showed the largest negative GOSIF anomaly in 2022 but the ENFs showed a positive  
284 GOSIF anomaly in 2022, both in terms of magnitude and in terms of the spatial extent of  
285 negative GOSIF anomalies (Figure 5).

286 Figure 6a shows the anomalies of NIRv (average z-score over the summer months)  
287 across all forested areas in Europe. The intensity and extent of the NIRv anomalies during  
288 the summer months (JJA) in each year are shown for different forest types in Figure 6b.  
289 In terms of canopy browning response (NIRv anomalies), the largest negative NIRv  
290 anomalies in 2022 were observed in southern Europe (Figure 6). Largest negative NIRv  
291 anomalies (indicated by the maximum anomaly) were observed in the DBFs in 2022,  
292 fitting the declined GOSIF signals. The ENFs showed positive NIRv anomalies in 2022,  
293 in terms of magnitude, spatial coverage, and % of total area affected (Figure 6).

294

295 *Relationship between GOSIF and NIRv*

296 In general, the values of NIRv and GOSIF were highly correlated (Figure S2). The  
297 anomalies of NIRv and GOSIF were most correlated across WSAs ( $r^2 = 0.73$  in 2018) and  
298 least correlated across the ENFs (Figure S2). Figure 7 shows the spatial regression  
299 between standardized GOSIF anomalies with (a) VPD and (b) SM and Figure 8 shows  
300 the spatial regression between standardized NIRv anomalies with (a) VPD and (b) SM  
301 over the drought areas in summers 2003, 2018 and 2022. With the increase in VPD (i.e.,  
302 increased atmospheric dryness), GOSIF values declined across all forest types, across  
303 all years, except in 2022 in the WSA, and in 2018 and 2022 in EBFs (Figure 7). With  
304 decrease in soil moisture (i.e., increased soil dryness), GOSIF values also declined  
305 overall ( $r^2 = 0.34$ ), but not as strongly as with the increase in air dryness ( $r^2 = 0.39$ ) (Figure  
306 7). Across different forest types, GOSIF responded most strongly to VPD anomalies in  
307 the MFs (mean  $r^2 = 0.48$ ), and responded most directly to changes in the soil moisture in  
308 the WSA (Figure 7).

309 Between VPD and SM, in general GOSIF anomalies were more correlated with VPD than  
310 with SM anomalies, and the decline in VPD correlated well with the larger GOSIF decline  
311 that we observed in DBFs in 2022 and in ENFs in 2003 (Figure 7). Under typical  
312 conditions (regardless of drought), GOSIF's response to both air dryness and soil  
313 moisture anomalies was more pronounced than the response of NIRv ( $r^2 = 0.39$  with  
314 GOSIF, compared to  $r^2 = 0.29$  for NIRv) (Figure 7, 8).

315 Figure 9 shows the partial correlation coefficient between GOSIF with SM and VPD during  
316 summer months (JJA) for areas identified as affected (Figure 9a) and not affected (Figure  
317 9b) by drought. The SM and VPD values across all forest types correlated well, but across  
318 DBFs the dryness in the atmosphere and the dryness in the soil were most correlated  
319 (Figure 9). Regarding canopy response to VPD, European Needleleaf Forests (ENF)  
320 exhibited the strongest reaction to changes in atmospheric dryness (Figure 9)

321 **Discussion**

322 *Severity of the 2022 summer drought*

323 Although the years 2003, 2018, and 2022 are all categorized as "extreme," the specific  
324 characteristics of the extreme conditions varied significantly among these years. For  
325 example, in 2003, widespread negative anomalies in soil moisture signaled a significant  
326 soil drought, whereas in 2022, widespread positive VPD anomalies indicated a notably  
327 drier atmosphere (Figure 3). It is important to note that ERA-5 Land datasets have been  
328 shown to underestimate the extent of European heatwaves in 2003, 2010, and 2018  
329 (Duveiller et al., 2023), partly due to the use of a static leaf area index in their modeling  
330 framework. Consequently, the SM droughts in the years 2003, 2018, and 2022 may be  
331 more severe than indicated by our study, suggesting that our results might be somewhat  
332 conservative. The extensive summer drought in 2022 primarily impacted southern  
333 Europe, in contrast to the 2003 summer drought, which affected central Europe, and the  
334 2018 drought, which extended to central and northern Europe (Figure 2) (Bastos et al.,  
335 2020). Consequently, the severe dry conditions in 2022 resulted in an average decline in  
336 GOSIF across forests that was 30% more widespread compared to 2018, and 60% more  
337 widespread compared to 2003 (Figure 4). These above-normal dry conditions during the  
338 summer reduced the photosynthetic capacity of plants and, consequently, the  
339 ecosystem's ability to absorb carbon from the atmosphere (Peters et al., 2018; van der  
340 Woude et al., 2023). Although the atmospheric and soil droughts in 2018 were more  
341 extensive and severe compared to 2022 (as indicated by the maximum observed z-  
342 scores), the adverse impact on forests, as reflected by the decline in GOSIF, was greater  
343 in 2022.

344 *Canopy response to soil versus atmospheric dryness*

345 The GOSIF dataset used in this study has been shown to be a reliable proxy for  
346 vegetation gross productivity, as demonstrated by comparisons with ground-based flux  
347 measurements (Shekhar et al. 2022; Pickering et al. 2022). It is important to note that  
348 GOSIF estimates are derived from a machine learning model trained with OCO-2 SIF  
349 observations, MODIS EVI data, and meteorological reanalysis data. As a result, the  
350 meteorological data used in our analyses are not entirely independent of the SIF data.  
351 However, this overlap is unlikely to impact our findings. A recent study that compared

352 GOSIF with original OCO-2 data to assess the impacts of the 2018 U.S. drought found  
353 similar responses to drought between the two datasets (Li et al., 2020).

354 NIRv and SIF signals are well-correlated and effectively capture seasonal patterns in GPP  
355 (Getachew Mengistu et al. 2021). Although the strength of their relationship can vary with  
356 time, location, and forest type (see Figure S2), reductions in SIF signals are directly  
357 associated with decreased photosynthesis. While both SIF and NIRv are reliable  
358 indicators of canopy responses to extreme climate events, SIF is more responsive to  
359 short-term climatic changes (Figure 7).

360 Our analysis showed that across different regions, GOSIF anomalies corresponded more  
361 strongly to increased atmospheric dryness than to increased soil dryness (Figure 7). This  
362 supports the understanding that vapor pressure deficit plays a larger role in controlling  
363 SIF signals for trees over shorter time scales than soil moisture (Pickering et al. 2022).  
364 Over shorter time frames, trees can often mitigate soil moisture deficits through  
365 mechanisms within the rooting zone and by accessing deeper water sources, whereas  
366 there is no such buffer for the impact of atmospheric dryness on tree canopies.

367 Ground-based observations in forest ecosystems, including both ecosystem and tree-  
368 level measurements, have shown that atmospheric dryness can constraint canopy gas  
369 exchange, even when soil moisture is not limiting (Gharun et al. 2014, Fu et al. 2022,  
370 Shekhar et al. 2024a). These findings highlight the importance of considering atmospheric  
371 dryness as a limiting factor for tree photosynthesis during extremely dry conditions and  
372 demonstrate the rapid response of various canopy types to increased levels of  
373 environmental dryness.

#### 374 *Canopy response to drought across different forest types*

375 The spread of drought, measured as the total area across z-scores, exhibited distinct  
376 patterns in different years, leading to varied responses of different forest types to the  
377 climatic anomalies. Impact of drought on forests can significantly differ depending on the  
378 forest type, tree species, species composition, and past exposure to extreme conditions  
379 (Arthur and Dech 2016; Chen et al. 2022). Our analysis showed that conditions in summer  
380 2022 reduced vegetation functioning across DBFs the most, as it was indicated by  
381 declined GOSIF signals (Figure 5). While deciduous broad-leaved forests were most

382 negatively affected by the extreme conditions in 2022, Evergreen Needle-Leaf Forests  
383 (ENF) distributed in northern regions of Europe were not exposed to extremely dry  
384 conditions in 2022 and even showed enhanced canopy greening and GOSIF signals,  
385 through benefiting from the episodic warming (Forzieri et al. 2022). Under similar drought  
386 conditions, the mechanisms to cope with the level of drought stress vary largely among  
387 forest types, and depend on a combination of characteristics that control water loss  
388 through the coordination of stomatal regulation, hydraulic architecture, and root  
389 characteristics (e.g., rooting depth, root distribution, root morphology) (Gharun et al. 2020;  
390 Peters et al. 2023). Stomata of trees exhibit a high sensitivity to VPD fluctuations, causing  
391 a reduction in stomatal conductance as VPD increases, which, in turn, limits the exchange  
392 of CO<sub>2</sub> with the atmosphere during photosynthesis (Bonal and Guehl in 2011; Li et al.  
393 2023). Different tree species show varying degrees of sensitivity in their stomatal  
394 responses to atmospheric dryness (Oren et al., 1999). For example, ring-porous species  
395 tend to maintain robust gas exchange under dry conditions, while diffuse-porous species,  
396 like those in ENFs, exhibit stronger stomatal regulation, reducing stomatal conductance  
397 as water availability decreases (Klein, 2014). This variability places plants on a spectrum  
398 of drought tolerance, reflecting their specific water relations strategies and leading to  
399 different responses among forests in similar climatic regions.

400 *Vulnerability of forests to more frequent drought*

401 The increased canopy damage observed in 2022, despite less severe conditions  
402 compared to the previous extreme year, suggests a lasting impact on forest canopies that  
403 could lead to a decline in forest resilience in the face of more frequent drought events  
404 (Forzieri et al., 2022). A potential decline in the resilience of forests has significant  
405 implications for vital ecosystem services, including the forest's capacity to mitigate climate  
406 change. Consequently, there is an urgent need to consider these trends when formulating  
407 robust forest-based mitigation strategies. This need is especially critical given future  
408 projections indicating that the frequency and intensity of extreme dryness across Europe  
409 will more than triple by the end of the 21st century (Shekhar et al., 2024b). In this context,  
410 it is increasingly important to investigate the vulnerability of forests to external  
411 perturbations and to develop mitigation strategies tailored to site-specific

412 ecophysiological and environmental factors that influence forest resilience to drought.  
413 Effective management strategies should be based on an understanding of these factors  
414 to mitigate the legacy effects of drought (McDowell et al., 2020; Wang et al., 2023;  
415 Shekhar et al., 2024a).

416

## 417 **Conclusion**

418 The severity of the 2022 summer drought, marked by increased atmospheric dryness,  
419 significantly compromised the photosynthetic capacity of trees, leading to widespread  
420 declines in vegetation functioning, especially in deciduous broad-leaved forests. Our  
421 findings underscore the importance of considering atmospheric dryness as a critical factor  
422 influencing canopy responses during extreme climatic events, alongside soil moisture  
423 deficits. Despite less severe overall conditions compared to previous extreme years, the  
424 greater canopy damage observed in 2022 suggests a growing vulnerability of forests to  
425 drought. This raises concerns about the future climate mitigation capacity of forest  
426 ecosystems, particularly as projections indicate a continued increase in the frequency and  
427 intensity of extreme dryness across Europe.

428

## 429 **Competing interests**

430 The authors have no competing interests to declare.

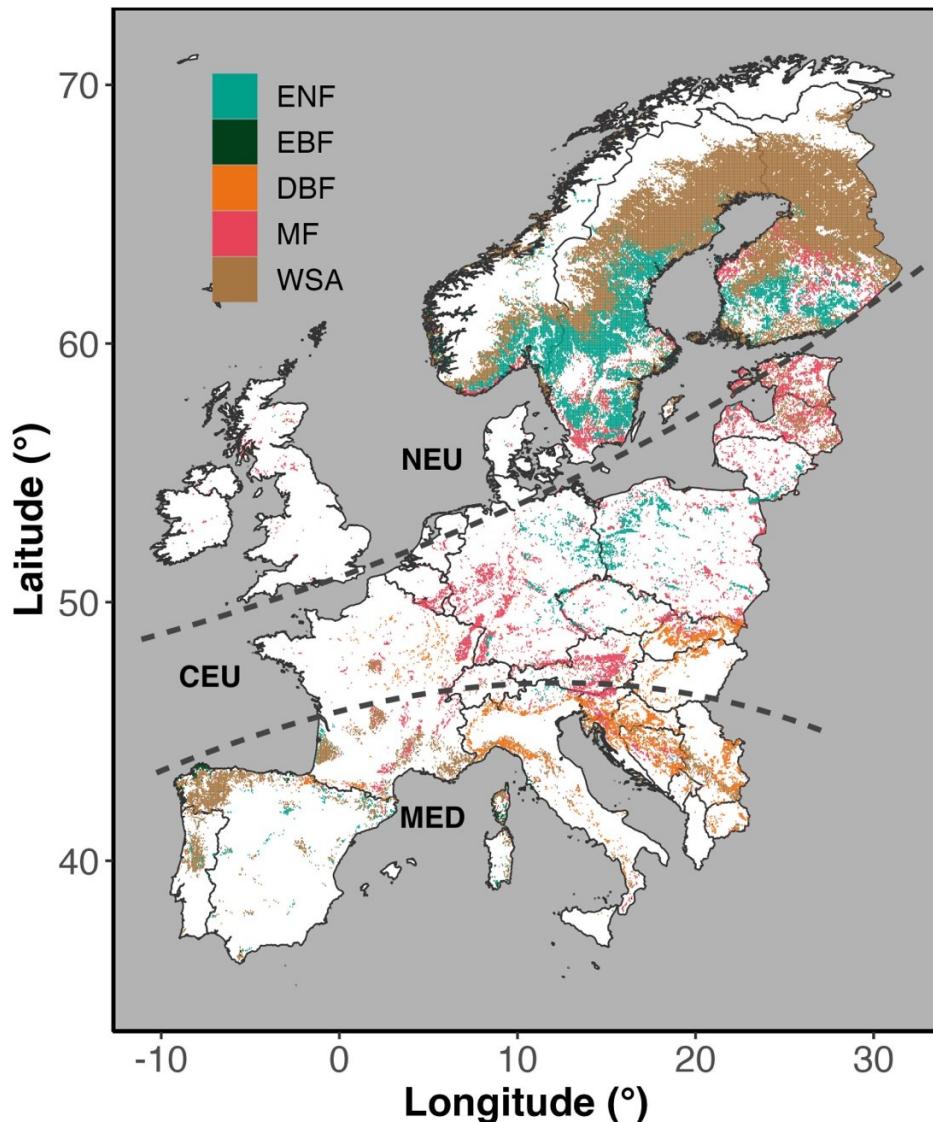
431

## 432 **Financial support**

433 AS acknowledges funding from the SNF funded project EcoDrive (IZCOZ0\_198094). JX  
434 was supported by the National Science Foundation (NSF) (Macrosystem Biology &  
435 NEON-Enabled Science program: DEB-2017870).

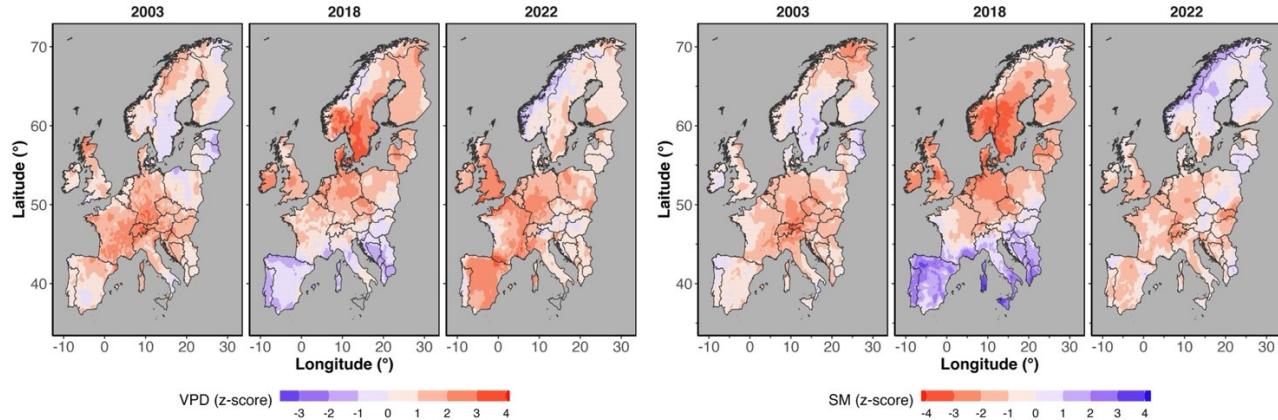
436 **Data availability** The R scripts used for the data analyses and plots are available upon  
437 request from the corresponding author.

438 **Author contributions** MG, AS, NB conceptualized the study. AS, JX, XL: data  
439 processing. MG and AS: data analyses. MG, AS, JX, XL: paper writing, revision and  
440 editing of the paper.



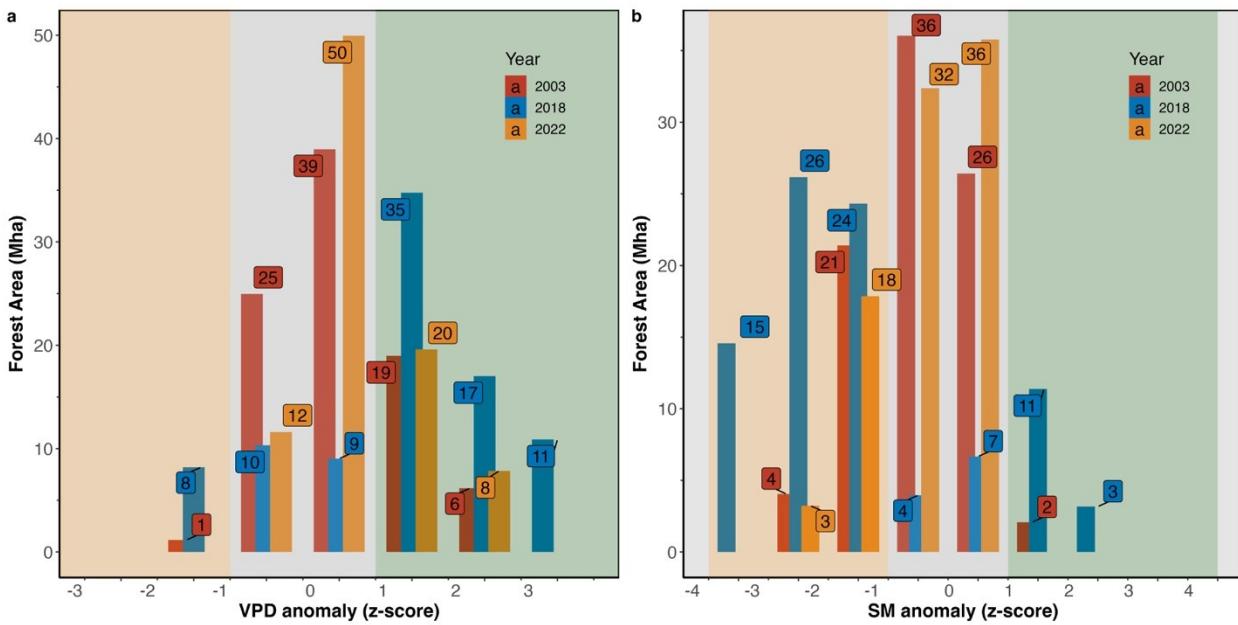
441

442 **Figure 1** Spatial coverage of forests (ENF - evergreen needleleaf forest; EBF - evergreen  
443 broadleaf forest; DBF - deciduous broadleaf forest; MF - mixed forest), and woodlands  
444 (WSA - woody savannas) across Europe. Areas are differentiated into Northern Europe  
445 (NEU), Central Europe (CEU), and Mediterranean Europe (MED) following Markonis et  
446 al. (2021). The map is based on MODIS land cover product MCD12C1 (version 6.1).

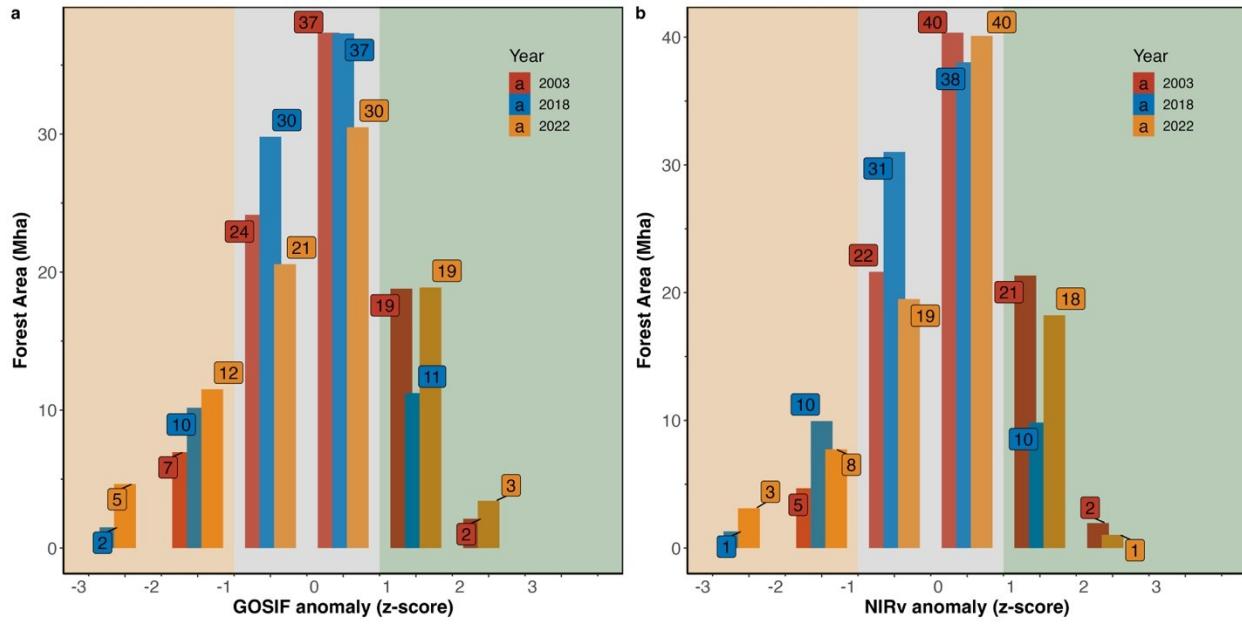


447  
448 **Figure 2** Standardized summer (JJA) anomalies (z-score) of mean vapor pressure deficit  
449 (VPD), and top layer (0-7 cm depth) soil moisture (SM) in 2003, 2018 and 2022, across  
450 the region of Europe.

451  
452  
453



454  
455  
456 **Figure 3** Intensity (z-score) and extent (area affected, Mha) of (a) VPD, and (b) SM  
457 anomalies across forested areas during the summer months (JJA). Z-score, values from  
458 -1 and 1 are considered normal (within 1 standard deviation of the mean). Orange-shaded  
459 area marks below normal and green-shaded area marks above normal conditions.

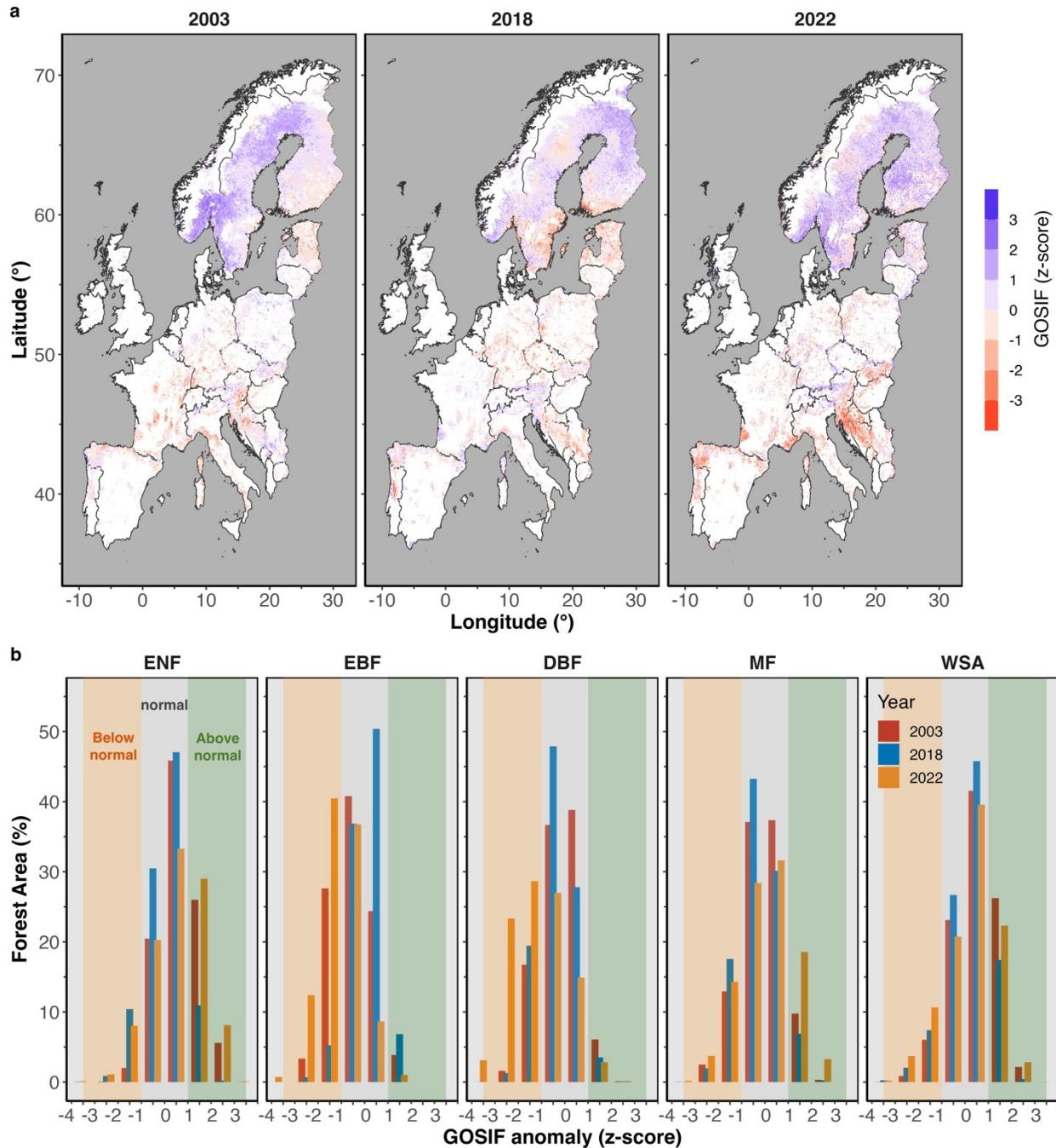


460

461 **Figure 4** Intensity (z-score) and extent (area affected, Mha) for (a) GOSIF, and (b) NIRv  
 462 anomalies across forested areas during the summer months (JJA). Z-score, values from  
 463 -1 and 1 are considered normal (within 1 standard deviation of the mean). Orange-shaded  
 464 area marks below normal and green-shaded area marks above normal conditions.

465

466

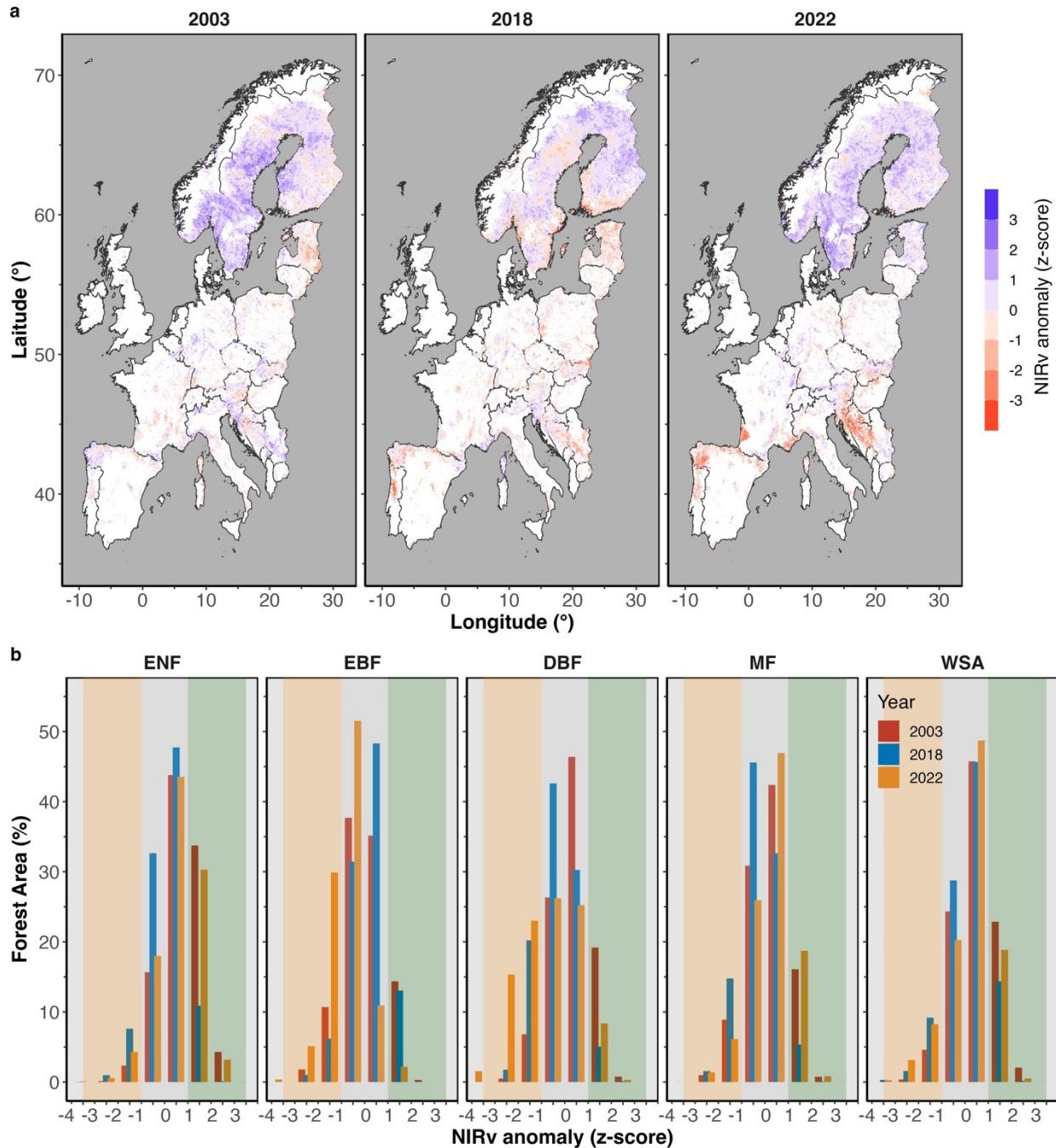


467

468 **Figure 5** (a) GOSIF anomaly (in terms of z-score) across Europe, and (b) area coverage  
 469 (in terms of percentage of total area for each forest type) during the summer months (JJA)  
 470 in 2003, 2018 and 2022. Orange-shaded area marks below normal and green-shaded  
 471 area marks above normal conditions. White areas on the map mark non-forested regions.

472

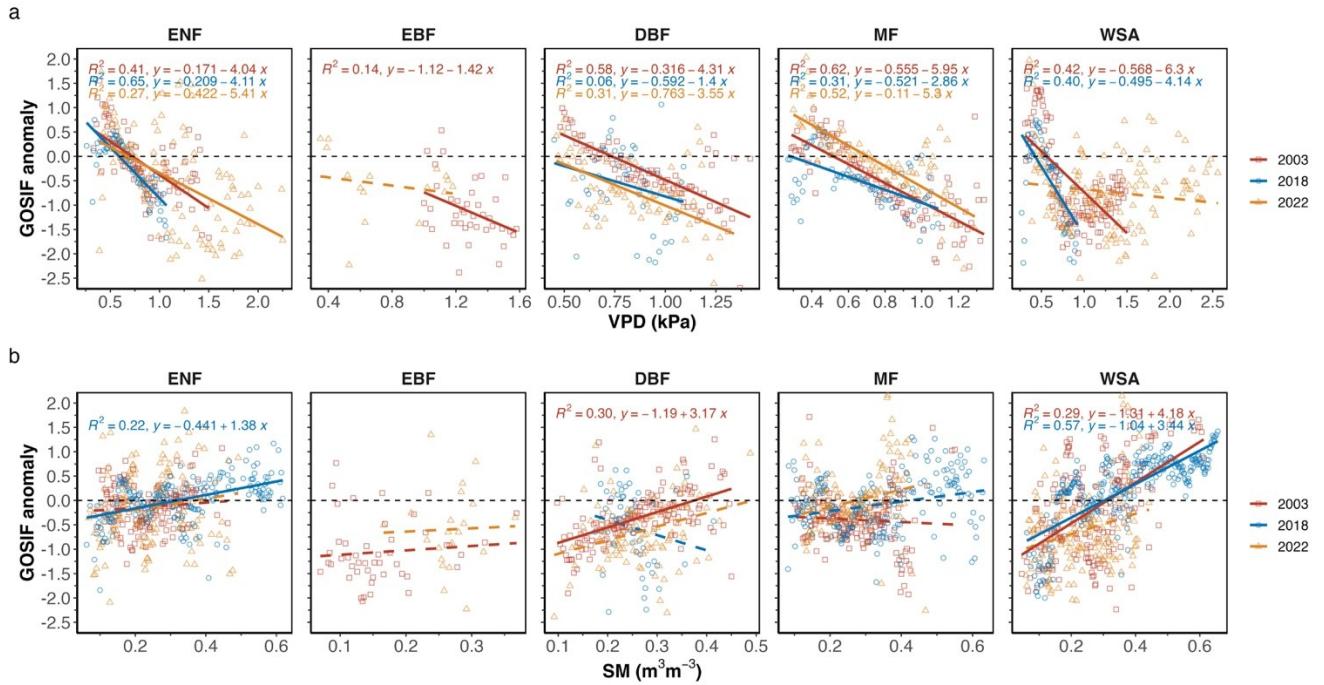
473



474

475

476 **Figure 6** (a) NIRv anomaly (in terms of z-score) across Europe, and (b) area coverage  
 477 (in terms of percentage of total area for each forest type) during the summer months (JJA)  
 478 in 2003, 2018 and 2022. In panel (b) Orange-shaded area marks below normal and green-  
 479 shaded area marks above normal conditions. White areas on the map mark non-forested  
 480 regions.



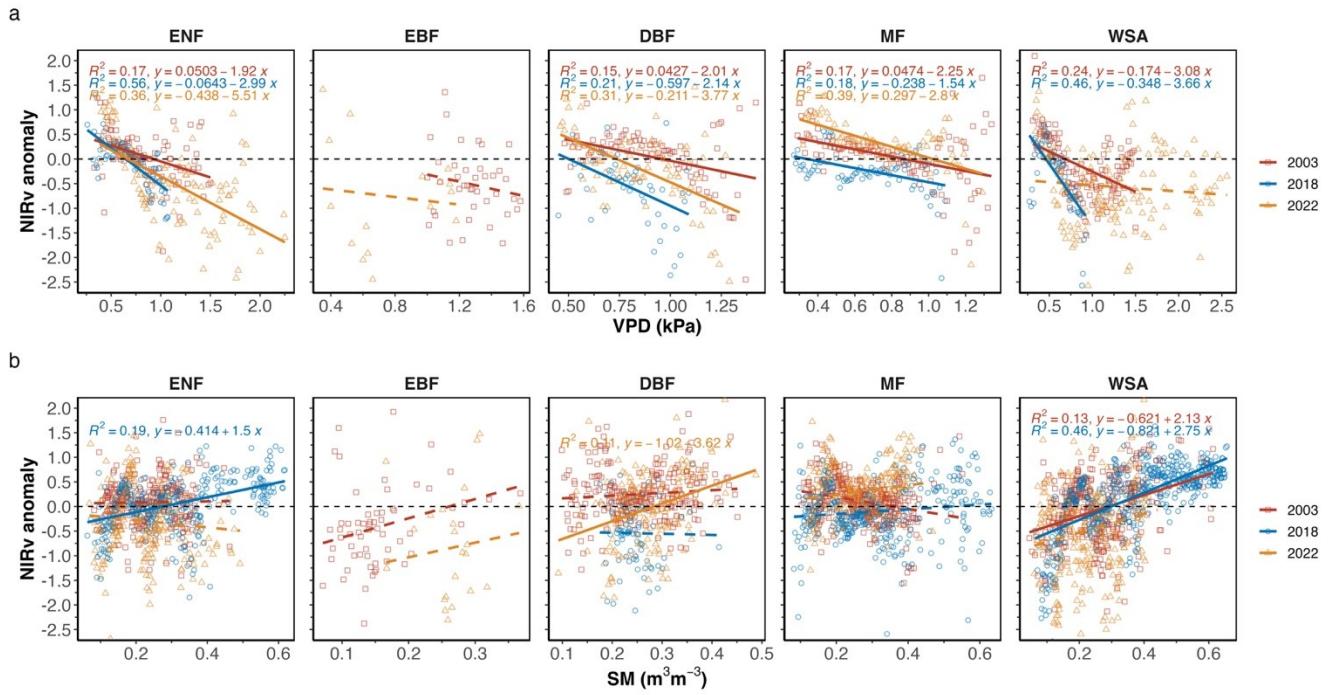
481

482 **Figure 7** Spatial regression between standardized GOSIF anomalies with (a) VPD and  
483 (b) SM over the drought areas during the summer months (JJA) 2003, 2018 and 2022.  
484 Dashed lines mark a non-significant relationship ( $p > 0.05$ ).

485

486

487



488

489

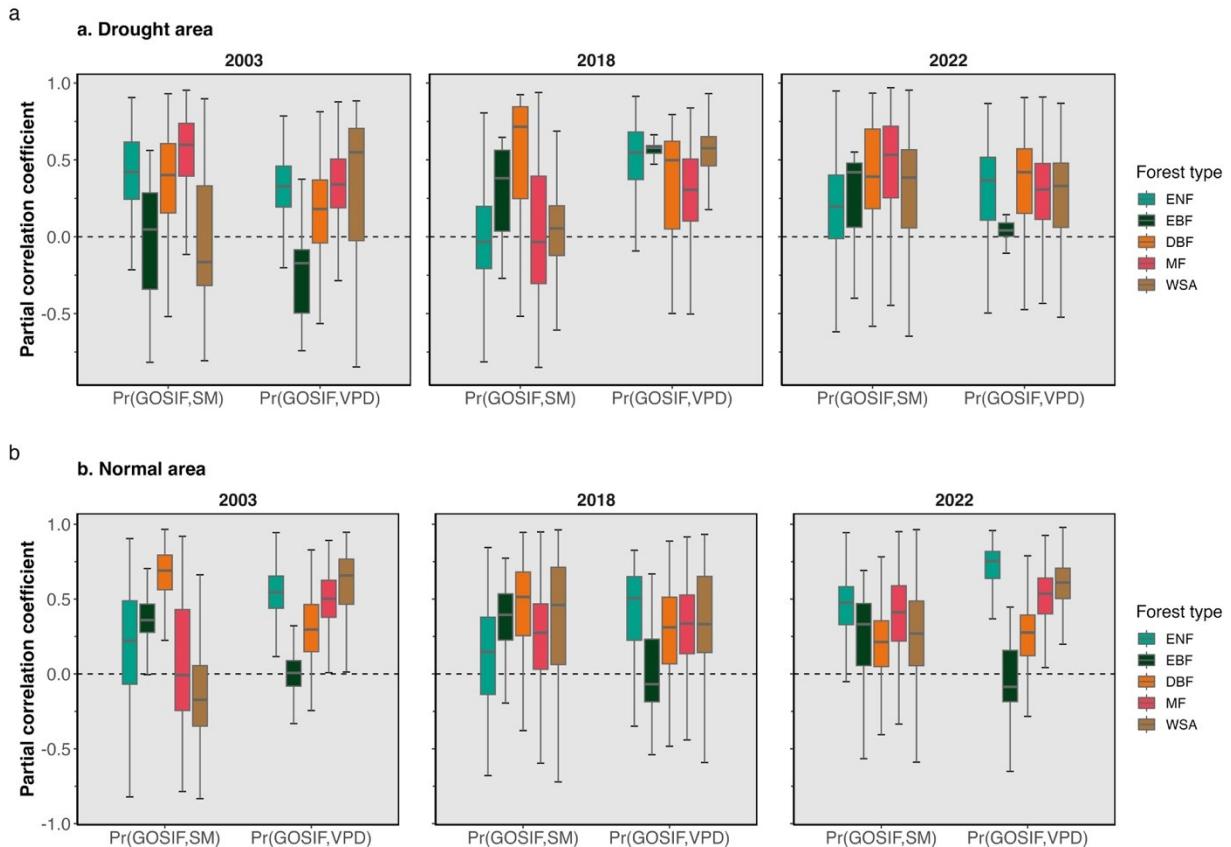
490 **Figure 8.** Spatial (over all pixels) regression between standardized NIRv anomalies with  
491 (a) VPD and (b) SM over the drought areas and normal areas in 2003, 2018, and 2022  
492 during the summer months (JJA).

493

494

495

496



497  
498  
499  
500  
501  
502  
503

504 **References**

505 Albergel, C., De Rosnay, P., Balsamo, G., Isaksen, L., and Munoz-Sabater, J.: Soil  
506 moisture analyses at ECMWF: evaluation using global ground-based in situ  
507 observations, *J. Hydrometeorol.*, 13, 1442–1460, <https://doi.org/10.1175/JHM-D11-0107.1>, 2012.

509 Anjileli, H., Huning, L. S., and Moftakhari, H.: Extreme heat events heighten soil  
510 respiration, *Sci. Rep.*, 11, 6632, <https://doi.org/10.1038/s41598-021-85764-8>, 2021.

511 Arthur, C. M., and Dech, J. P.: Species composition determines resistance to drought in  
512 dry forests of the Great Lakes - St. Lawrence forest region of central Ontario, *J.  
513 Veg. Sci.*, 27, 914–925, 2016.

514 Badgley, G., Field, C. B., and Berry, J. A.: Canopy near-infrared reflectance and  
515 terrestrial photosynthesis, *Sci. Adv.*, 3, e1602244,  
516 <https://doi.org/10.1126/sciadv.1602244>, 2017.

517 Bastos, A., Fu, Z., Ciais, P., Friedlingstein, P., Sitch, S., Pongratz, J., Weber, U.,  
518 Reichstein, M., Anthoni, P., Arneth, A., Haverd, V., Jain, A., Joetzjer, E., Knauer, J.,  
519 Lienert, S., Loughran, T., McGuire, P. C., Obermeier, W., Padrón, R. S., Shi, H.,  
520 Tian, H., Viovy, N., and Zaehle, S.: Impacts of extreme summers on European  
521 ecosystems: a comparative analysis of 2003, 2010 and 2018, *Phil. Trans. R. Soc.  
522 B*, 375, 20190507, <http://dx.doi.org/10.1098/rstb.2019.0507>, 2020.

523 Bauke, S. L., Amelung, W., Bol, R., Brandt, L., Brüggemann, N., Kandeler, E., Meyer,  
524 N., Or, D., Schnepf, A., Schloter, M., Schulz, S., Siebers, N., von Sperber, C., and  
525 Vereecken, H.: Soil water status shapes nutrient cycling in agroecosystems from  
526 micrometer to landscape scales, *J. Plant Nutr. Soil Sci.*, 185, 773–792,  
527 <https://doi.org/10.1002/jpln.202200357>, 2022.

528 Bonal, D., and Guehl, J.-M.: Contrasting patterns of leaf water potential and gas  
529 exchange responses to drought in seedlings of tropical rainforest species, *Funct.  
530 Ecol.*, 15, 490–496, 2011.

531 Buras, A., Rammig, A., and Zang, C. S.: Quantifying impacts of the 2018 drought on  
532 European ecosystems in comparison to 2003, *Biogeosciences*, 17, 1655–1672,  
533 <https://doi.org/10.5194/bg-17-1655-2020>, 2020.

534 Chen, Y., Vogel, A., and Wagg, C.: Drought-exposure history increases  
535 complementarity between plant species in response to a subsequent drought, *Nat.  
536 Commun.*, 13, 3217, <https://doi.org/10.1038/s41467-022-30954-9>, 2022.

537 Choat, B., Brodribb, T. J., Brodersen, C. R., Duursma, R. A., López, R., and Medlyn, B.  
538 E.: Triggers of tree mortality under drought, *Nature*, 558, 531–539,  
539 <https://doi.org/10.1038/s41586-018-0240-x>, 2018.

540 Cornes, R. C., van der Schrier, G., van den Besselaar, E. J. M., and Jones, P. D.: An  
541 Ensemble Version of the E-OBS Temperature and Precipitation Data Sets, *J.  
542 Geophys. Res. Atmos.*, 123, 9391–9409, <https://doi.org/10.1029/2017JD028200>,  
543 2018.

544 Dang, C., Shao, Z., Huang, X., Qian, J., Cheng, G., Ding, Q., and Fan, Y.: Assessment  
545 of the importance of increasing temperature and decreasing soil moisture on global  
546 ecosystem productivity using solar-induced chlorophyll fluorescence, *Glob. Change  
547 Biol.*, 28, 2066–2080, <https://doi.org/10.1111/gcb.16043>, 2022.

548 Dee, D. P., Uppala, S. M., Simmons, A. J., Berrisford, P., Poli, P., Kobayashi, S.,  
549 Andrae, U., Balmaseda, M. A., Balsamo, G., Bauer, P., Bechtold, P., Beljaars, A. C.  
550 M., van de Berg, L., Bidlot, J., Bormann, N., Delsol, C., Dragani, R., Fuentes, M.,  
551 Geer, A. J., Haimberger, L., Healy, S. B., Hersbach, H., Hólm, E. V., Isaksen, L.,  
552 Källberg, P., Köhler, M., Matricardi, M., McNally, A. P., Monge-Sanz, B. M.,  
553 Morcrette, J. J., Park, B. K., Peubey, C., de Rosnay, P., Tavolato, C., Thépaut, J.  
554 N., and Vitart, F.: The ERA-Interim reanalysis: configuration and performance of the  
555 data assimilation system, *Q. J. R. Meteorol. Soc.*, 137, 553–597,  
556 <https://doi.org/10.1002/qj.828>, 2011.

557 Dogan, S., Berkay, A., and Singh, V. P.: Comparison of multi-monthly rainfall-based  
558 drought severity indices, with application to semi-arid Konya closed basin, Turkey,  
559 *J. Hydrol.*, 470–471, 255–268, 2012.

560 Drake, J. E., Tjoelker, M. G., Vårhammar, A., Medlyn, B. E., Reich, P. B., Leigh, A.,  
561 Pfautsch, S., Blackman, C. J., López, R., Aspinwall, M. J., Crous, K. Y., Duursma,  
562 R. A., Kumarathunge, D., De Kauwe, M. G., Jiang, M., Nicotra, A. B., Tissue, D. T.,  
563 Choat, B., and Atkin, O. K.: Trees tolerate an extreme heatwave via sustained  
564 transpirational cooling and increased leaf thermal tolerance, *Glob. Change Biol.*, 24,  
565 2390–2402, <https://doi.org/10.1111/gcb.14037>, 2018.

566 Duveiller, G., Pickering, M., Muñoz-Sabater, J., Caporaso, L., Boussetta, S., Balsamo,  
567 G., and Cescatti, A.: Getting the leaves right matters for estimating temperature  
568 extremes, *Geosci. Model Dev.*, 16, 7357–7373, <https://doi.org/10.5194/gmd-16-7357-2023>, 2023.

570 Forzieri, G., Dakos, V., and McDowell, N. G.: Emerging signals of declining forest  
571 resilience under climate change, *Nature*, 608, 534–539,  
572 <https://doi.org/10.1038/s41586-022-04959-9>, 2022.

573 Fu, Z., Ciais, P., and Prentice, I. C.: Atmospheric dryness reduces photosynthesis along  
574 a large range of soil water deficits, *Nat. Commun.*, 13, 989,  
575 <https://doi.org/10.1038/s41467-022-28652-7>, 2022.

576 Gessler, A., Bottero, A., Marshall, J., and Arend, M.: The way back: recovery of trees  
577 from drought and its implication for acclimation, *New Phytol.*, 228, 1704–1709,  
578 <https://doi.org/10.1111/nph.16703>, 2020.

579 Mengistu, A. G., Tsidu, G. M., Koren, G., Kooreman, M. L., Boersma, K. F., Tagesson,  
580 T., Ardö, J., Nouvellon, Y., and Peters, W.: Sun-induced fluorescence and near-  
581 infrared reflectance of vegetation track the seasonal dynamics of gross primary  
582 production over Africa, *Biogeosciences*, 18, 2843–2857, <https://doi.org/10.5194/bg-18-2843-2021>, 2021.

584 Gharun, M., Vervoort, R. W., Turnbull, T. L., and Adams, M. A.: A test of how coupling  
585 of vegetation to the atmosphere and climate spatial variation affects water yield

586 modelling in mountainous catchments, *J. Hydrol.*, 514, 202–213,  
587 <https://doi.org/10.1016/j.jhydrol.2014.04.037>, 2014.

588 Gharun, M., Hörtnagl, L., Paul-Limoges, E., Ghiasi, S., Feigenwinter, I., Burri, S.,  
589 Marquardt, K., Etzold, S., Zweifel, R., Eugster, W., and Buchmann, N.: Physiological  
590 response of Swiss ecosystems to 2018 drought across plant types and elevation,  
591 *Philos. Trans. R. Soc. B*, 375, 20190521, <http://doi.org/10.1098/rstb.2019.0521>,  
592 2020.

593 Gourlez de la Motte, L., Beauclaire, Q., Heinesch, B., Cuntz, M., Foltýnová, L., Šigut, L.,  
594 Kowalska, N., Manca, G., Ballarin, I. G., Vincke, C., Roland, M., Ibrom, A.,  
595 Lousteau, D., Siebiche, L., Neiryink, J., and Longdoz, B.: Non-stomatal processes  
596 reduce gross primary productivity in temperate forest ecosystems during severe  
597 edaphic drought, *Philos. Trans. R. Soc. B*, 375, 20190527,  
598 <https://doi.org/10.1098/rstb.2019.0527>, 2020.

599 Haesen, S., Lembrechts, J. J., De Frenne, P., Lenoir, J., Aalto, J., Ashcroft, M. B.,  
600 Kopecký, M., Luoto, M., Maclean, I., Nijs, I., Niittynen, P., van den Hoogen, J.,  
601 Arriga, N., Brúna, J., Buchmann, N., Čiliak, M., Collalti, A., De Lombaerde, E.,  
602 Descombes, P., and Van Meerbeek, K.: ForestClim—Bioclimatic variables for  
603 microclimate temperatures of European forests, *Glob. Change Biol.*, 29, 2886–  
604 2892, <https://doi.org/10.1111/gcb.16678>, 2023.

605 Humphrey, V., Zscheischler, J., Ciais, P., Gudmundsson, L., Sitch, S., and Seneviratne,  
606 S. I.: Sensitivity of atmospheric CO<sub>2</sub> growth rate to observed changes in terrestrial  
607 water storage, *Nature*, 560, 628, <https://doi.org/10.1038/s41586-018-0424-4>, 2018.

608 Jain, V. K., Pandey, R. P., Jain, M. K., and Byun, H. R.: Comparison of drought indices  
609 for appraisal of drought characteristics in the Ken River Basin, *Weather and Climate  
610 Extremes*, 8, 1–11, <https://doi.org/10.1016/j.wace.2015.05.002>, 2015.

611 Klein Tank, A. M. G., Wijngaard, J. B., Können, G. P., Böhm, R., Demarée, G.,  
612 Gocheva, A., Mileta, M., Pashiardis, S., Hejkrlik, L., Kern-Hansen, C., Heino, R.,  
613 Bessemoulin, P., Müller-Westermeier, G., Tzanakou, M., Szalai, S., Pálsdóttir, T.,  
614 Fitzgerald, D., Rubin, S., Capaldo, M., and Petrovic, P.: Daily dataset of 20th-  
615 century surface air temperature and precipitation series for the European Climate  
616 Assessment, *Int. J. Climatol.*, 22, 1441–1453, <https://doi.org/10.1002/joc.773>, 2002.

617 Klein, T.: The variability of stomatal sensitivity to leaf water potential across tree species  
618 indicates a continuum between isohydric and anisohydric behaviours, *Funct. Ecol.*,  
619 28, 1313–1320, <https://doi.org/10.1111/1365-2435.12289>, 2014.

620 Lal, P., Shekhar, A., Gharun, M., and Das, N. N.: Spatiotemporal evolution of global  
621 long-term patterns of soil moisture, *Sci. Total Environ.*, 867, 161470,  
622 <https://doi.org/10.1016/j.scitotenv.2023.161470>, 2023.

623 Lal, P., Singh, G., Das, N. N., Colliander, A., and Entekhabi, D.: Assessment of ERA5-  
624 land volumetric soil water layer product using in situ and SMAP soil moisture  
625 observations, *Geosci. Rem. Sens. Lett. IEEE*, 19, 1–5,  
626 <https://doi.org/10.1109/LGRS.2022.3223985>, 2022.

627 Li, X., Frankenberg, C., Berry, J. A., Joiner, J., Guanter, L., Tol, C., and Badgley, G.:  
628 Solar-induced chlorophyll fluorescence is strongly correlated with terrestrial  
629 photosynthesis for a wide variety of biomes: first global analysis based on OCO-2  
630 and flux tower observations, *Glob. Change Biol.*, 24, 3990–4008, 2018.

631 Li, X., and Xiao, J.: A global, 0.05-degree product of solar-induced chlorophyll  
632 fluorescence derived from OCO-2, MODIS, and reanalysis data, *Remote Sens.*, 11,  
633 517, <https://doi.org/10.3390/rs11050517>, 2019.

634 Li, X., Xiao, J., Kimball, J. S., Reichle, R. H., and Frankenberg, C.: Synergistic use of  
635 SMAP and OCO-2 data in assessing the responses of ecosystem productivity to the  
636 2018 U.S. drought, *Remote Sens. Environ.*, 251, 112062, 2020.

637 Li, F., Xiao, J., Chen, J., Ballantyne, A., Jin, K., Li, B., Abraha, M., and John, R.: Global  
638 water use efficiency saturation due to increased vapor pressure deficit, *Science*,  
639 381, 672–677, <https://doi.org/10.1126/science.adf5041>, 2023.

640 Magney, T. S., Frankenberg, C., Sun, Y., Joiner, J., Porcar-Castell, A., and Baker, I.:  
641 Mechanistic evidence for tracking the seasonality of photosynthesis with solar-  
642 induced fluorescence, *Proc. Natl Acad. Sci. USA*, 116, 11640–11645, 2019.

643 Marchin, R. M., Backes, D., Ossola, A., Leishman, M. R., Tjoelker, M. G., and Ellsworth,  
644 D. S.: Extreme heat increases stomatal conductance and drought-induced mortality  
645 risk in vulnerable plant species, *Glob. Change Biol.*, 28, 1133–1146,  
646 <https://doi.org/10.1111/gcb.15976>, 2022.

647 Markonis, Y., Kumar, R., Hanel, M., Rakovec, O., Máca, P., and AghaKouchak, A.: The  
648 rise of compound warm-season droughts in Europe, *Sci. Adv.*, 7, eabb9668,  
649 <https://doi.org/10.1126/sciadv.abb9668>, 2021.

650 McDowell, N. G., Allen, C. D., Anderson-Teixeira, K., Aukema, B. H., Bond-Lamberty,  
651 B., Chini, L., Clark, J. S., Crowther, T., Frey, S., and HilleRisLambers, J.: Pervasive  
652 shifts in forest dynamics in a changing world, *Science*, 368, eaaz9463, 2020.

653 Mishra, A. K., and Singh, V. P.: Drought modeling – A review, *J. Hydrol.*, 403, 157–175,  
654 <https://doi.org/10.1016/j.jhydrol.2011.03.049>, 2011.

655 Müller, L. M., and Bahn, M.: Drought legacies and ecosystem responses to subsequent  
656 drought, *Glob. Change Biol.*, 28, 5086–5103, <https://doi.org/10.1111/gcb.16270>,  
657 2022.

658 Muñoz-Sabater, J., Dutra, E., Agustí-Panareda, A., Albergel, C., Arduini, G., Balsamo,  
659 G., Boussetta, S., Choulga, M., Harrigan, S., Hersbach, H., Martens, B., Miralles, D.  
660 G., Piles, M., Rodríguez-Fernández, N. J., Zsoter, E., Buontempo, C., and Thépaut,  
661 J. N.: ERA5-Land: A state-of-the-art global reanalysis dataset for land applications,  
662 *Earth Syst. Sci. Data*, 13, 4349–4383, <https://doi.org/10.5194/essd-13-4349-2021>,  
663 2021.

664 Peters, W., van der Velde, I. R., van Schaik, E., Miller, J. B., and Ciais, P.: Increased  
665 water-use efficiency and reduced CO<sub>2</sub> uptake by plants during droughts at a  
666 continental scale, *Nat. Geosci.*, 11, 744–748, <https://doi.org/10.1038/s41561-018-0212-7>, 2018.

668 Peters, R. L., Steppe, K., Pappas, C., Zweifel, R., Babst, F., Dietrich, L., von Arx, G.,  
669 Poyatos, R., Fonti, M., Fonti, P., Grossiord, C., Gharun, M., Buchmann, N., Steger,  
670 D. N., and Kahmen, A.: Daytime stomatal regulation in mature temperate trees  
671 prioritizes stem rehydration at night, *New Phytol.*, 239, 533–546,  
672 <https://doi.org/10.1111/nph.18964>, 2023.

673 Pickering, M., Cescatti, A., and Duveiller, G.: Sun-induced fluorescence as a proxy for  
674 primary productivity across vegetation types and climates, *Biogeosciences*, 19,  
675 4833–4864, <https://doi.org/10.5194/bg-19-4833-2022>, 2022.

676 Oren, R., Sperry, J. S., Katul, G. G., Pataki, D. E., Ewers, B. E., Phillips, N., and  
677 Schäfer, K. V. R.: Survey and synthesis of intra- and interspecific variation in  
678 stomatal sensitivity to vapour pressure deficit, *Plant Cell Environ.*, 22, 1515–1526,  
679 <https://doi.org/10.1046/j.1365-3040.1999.00513.x>, 1999.

680 Röthlisberger, M., and Papritz, L.: Quantifying the physical processes leading to  
681 atmospheric hot extremes at a global scale, *Nature Geosci.*, 16, 210–216,  
682 <https://doi.org/10.1038/s41561-023-01126-1>, 2023.

683 Seidl, R., Thom, D., Kautz, M., Martin-Benito, D., Peltoniemi, M., Vacchiano, G., Wild,  
684 J., Ascoli, D., Petr, M., Honkaniemi, J., Lexer, M. J., Trotsiuk, V., Mairoti, P.,  
685 Svoboda, M., Fabrika, M., Nagel, T. A., and Reyer, C. P. O.: Forest disturbances  
686 under climate change, *Nature Clim. Change*, 7, 395–402,  
687 <https://doi.org/10.1038/nclimate3303>, 2017.

688 Seneviratne, S. I., Zhang, X., Adnan, M., Badi, W., Dereczynski, C., Di Luca, A., Ghosh,  
689 S., Iskandar, I., Kossin, J., Lewis, S., Otto, F., Pinto, I., Satoh, M., Vicente-Serrano,  
690 S. M., Wehner, M., and Zhou, B.: Weather and Climate Extreme Events in a  
691 Changing Climate, in: *Climate Change 2021: The Physical Science Basis.*  
692 Contribution of Working Group I to the Sixth Assessment Report of the  
693 Intergovernmental Panel on Climate Change, edited by: Masson-Delmotte, V., Zhai,  
694 P., Pirani, A., Connors, S. L., Péan, C., Berger, S., Caud, N., Chen, Y., Goldfarb, L.,  
695 Gomis, M. I., Huang, M., Leitzell, K., Lonnoy, E., Matthews, J. B. R., Maycock, T. K.,  
696 Waterfield, T., Yelekçi, O., Yu, R., and Zhou, B., Cambridge University Press,

697 Cambridge, United Kingdom and New York, NY, USA, 1513–1766,  
698 <https://doi.org/10.1017/9781009157896.013>, 2021.

699 Shekhar, A., Hörtnagl, L., Buchmann, N., and Gharun, M.: Long-term changes in forest  
700 response to extreme atmospheric dryness, *Global Change Biol.*, 29, 5379–5396,  
701 <https://doi.org/10.1111/gcb.16846>, 2023.

702 Shekhar, A., Hörtnagl, L., Paul-Limoges, E., Etzold, S., Zweifel, R., Buchmann, N., and  
703 Gharun, M.: Contrasting impact of extreme soil and atmospheric dryness on the  
704 functioning of trees and forests, *Sci. Total Environ.*, 916, 169931,  
705 <https://doi.org/10.1016/j.scitotenv.2024.169931>, 2024a.

706 Shekhar, A., Humphrey, V., Buchmann, N., and Gharun, M.: More than three-fold  
707 increase of extreme dryness across Europe by end of 21st century, under review in  
708 *Weather Climate Extremes*, <https://doi.org/10.21203/rs.3.rs-3143908/v2>, 2024b.

709 Shekhar, A., Buchmann, N., and Gharun, M.: How well do recently reconstructed solar-  
710 induced fluorescence datasets model gross primary productivity?, *Remote Sens.*  
711 *Environ.*, 283, 113282, 2022.

712 Tripathy, K. P., and Mishra, A. K.: How unusual is the 2022 European compound  
713 drought and heatwave event?, *Geophys. Res. Lett.*, 50, e2023GL105453,  
714 <https://doi.org/10.1029/2023GL105453>, 2023.

715 Tsakiris, G., and Vangelis, H.: Establishing a drought index incorporating  
716 evapotranspiration, *Eur. Water*, 9/10, 3–11, 2005.

717 van der Woude, A. M., Peters, W., Joetzjer, E., Lafont, S., Koren, G., Ciais, P.,  
718 Ramonet, M., Xu, Y., Bastos, A., Botía, S., Sitch, S., de Kok, R., Kneuer, T.,  
719 Kubistin, D., Jacotot, A., Loubet, B., Herig-Coimbra, P.-H., Loustau, D., and Luijckx,  
720 I. T.: Temperature extremes of 2022 reduced carbon uptake by forests in Europe,  
721 *Nat. Commun.*, 14, 6218, <https://doi.org/10.1038/s41467-023-41851-0>, 2023.

722 van der Molen, M. K., Dolman, A. J., Ciais, P., Eglin, T., Gobron, N., Law, B. E., Meir,  
723 P., Peters, W., Phillips, O. L., Reichstein, M., Chen, T., Dekker, S. C., Doubková,  
724 M., Friedl, M. A., Jung, M., van den Hurk, B. J. J. M., de Jeu, R. A. M., Kruijt, B.,  
725 Ohta, T., Rebel, K. T., Plummer, S., Seneviratne, S. I., Sitch, S., Teuling, A. J., van  
726 der Werf, G. R., and Wang, G.: Drought and ecosystem carbon cycling, *Agric. For.*  
727 *Meteorol.*, 151, 765–773, <https://doi.org/10.1016/j.agrformet.2011.01.018>, 2011.

728 Zhang, J., Xiao, J., Tong, X., Zhang, J., Meng, P., Li, J., Liu, P., and Yu, P.: NIRv and  
729 SIF better estimate phenology than NDVI and EVI: Effects of spring and autumn  
730 phenology on ecosystem production of planted forests, *Agric. For. Meteorol.*, 315,  
731 108819, 2022.

732 Zhou, S., Yu, B., and Zhang, Y.: Global concurrent climate extremes exacerbated by  
733 anthropogenic climate change, *Sci. Adv.*, 9(10), eab01638,  
734 <https://doi.org/10.1126/sciadv.abo1638>, 2023.

735 Wable, P. S., Jha, M. K., and Shekhar, A.: Comparison of drought indices in a semi-arid  
736 river basin of India, *Water Resour. Manag.*, 33, 75–102,  
737 <https://doi.org/10.1007/s11269-018-2089-z>, 2019.

738 Wang, B., Chen, T., Xu, G., Wu, G., and Liu, G.: Management can mitigate drought  
739 legacy effects on the growth of a moisture-sensitive conifer tree species, *For. Ecol.*  
740 *Manage.*, 544, 121196, <https://doi.org/10.1016/j.foreco.2023.121196>, 2023.

741